

Succession of bacterial grazing defense mechanisms against protistan predators in an experimental microbial community

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ABSTRACT: We studied the effects of 2 bacterivorous protistan predators on the phenotypic and taxonomic successions of an experimental bacterial community growing in continuous culture with the cryptophyte *Cryptomonas phaseolus*. Two predators were inoculated to the cultivation system, the mixotrophic flagellate *Ochromonas* sp., an interception feeder, and the filter-feeding heterotrophic ciliate *Cyclidium glaucoma*. The system was sampled every second day over a period of 3 mo. The competition between algae and bacteria for limiting phosphorus ($20 \mu\text{g P l}^{-1} \text{ d}^{-1}$) was reflected in pronounced fluctuations of the ratio of algal versus bacterial biomass. High grazing pressure exerted on bacteria by the flagellate led to a strong decrease of bacterial abundance and biomass, with only few cells left freely dispersed. A succession of various grazing-resistant morphologies was observed, such as microcolonies, large aggregates and filaments. A population affiliated with *Aquabacterium* sp. formed grazing-resistant co-colonies with a *Caulobacter* spp., indicating that these 2 bacteria might find a common refuge from protistan grazing within aggregates or biofilms. During the course of the experiment these large colonies were succeeded by grazing-protected morphotypes related to the *Flexibacter sancti-Flavobacterium ferrugineum* lineage of the Sphingobacterales. In contrast, the ciliate induced only a slight reduction of bacterial abundance, and bacteria even recovered to higher than original numbers after a few weeks. However, pronounced successions of microbial populations were also observed in the presence of the ciliate.

KEY WORDS: Bacteria–algae interaction · Bacterial assemblage composition · Bacterial succession · Continuous cultivation system · *Cyclidium glaucoma* · Grazing defense mechanism · *Ochromonas* sp. · Predator–prey interaction

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INTRODUCTION

Grazing by bacterivorous protists can affect major phenotypic and physiological characteristics and the species composition of a bacterial assemblage. Changes in bacterial morphology due to protistan grazing are well documented (for review see Hahn & Höfle 2001, Jürgens & Matz 2002). In contrast, fewer studies have investigated the relationship between genotypic and phenotypic assemblage changes, and different bacteria seem to follow contrasting strategies to respond to a particular predator. Some bacterial phylotypes show pronounced phenotypic plasticity, as

reflected by an increase or decrease in size, the formation of filaments or the establishment of microcolonies or aggregates (Hahn & Höfle 2001, Hahn et al. 2004).

Fluorescence *in situ* hybridization (FISH) is a powerful tool to simultaneously study the phenotypic characteristics and phylogenetic identity of cells in mixed bacterial assemblages. Using FISH, grazing-induced changes in bacterial assemblage composition have been repeatedly observed both in the laboratory (Pernthaler et al. 1997, 2001, Šimek et al. 1997, Posch et al. 2001) and in field studies (Jürgens et al. 1999, Šimek et al. 1999, 2001a,b, 2003, Langenheder & Jürgens 2001, Gasol et al. 2002). Such shifts could be caused directly

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by size-selective feeding, but also indirectly, for instance as a result of a change in the abundance of bacterial competitors or in the nutrient release by predators (see Fig. 5 in Jürgens & Matz 2002).

In most of the aforementioned studies, changes in bacterial assemblages caused by protistan grazing have been induced experimentally and they were studied over very short time periods (hours and days). In contrast, our knowledge about successions of distinct bacterial genotypic populations or of their grazing defense mechanisms in response to continuous protistan grazing pressure is still scarce (Masín et al. 2003). There are first indications that grazing-induced shifts of bacterial assemblages are not only linked to the differential vulnerability of individual bacterial populations, but at the same time to the dominance of different predator species (Posch et al. 1999, Pernthaler et al. 2001). Likely, grazing by bacterivorous protists does not provoke a simple binary (yes/no) answer. Instead, the specific characteristics of the grazing regime (e.g. different nutrition modes, different size selectivity, varying generation times of predators, assemblage composition of predators) will be of relevance for the development of any diverse bacterial assemblage. Individual bacterial genotypes might thus be characterized by different levels of vulnerability to grazing per se, but at the same time the grazing-induced responses might be strongly influenced by the characteristics of protistan feeding.

We studied the long-term interactions between a mixed experimental bacterial assemblage and 2 protistan predators with different feeding behavior. Exudates of the autotrophic alga *Cryptomonas phaseolus* were used as a natural carbon source by the accompanying bacteria. The sampling interval was adapted to match the calculated doubling time of organisms as determined by the dilution rate (theoretically 40 bacterial generations).

Our hypotheses were (1) that grazing by 2 protistan species with different feeding modes will cause different responses of the bacterial community, e.g. in abundance, size, biovolume and bacterial assemblage composition; (2) that there will be successions of different bacterial genotypic populations; and (3) that the bacterial assemblage will show a succession of grazing protection mechanisms when exposed to predation over an extended period of time.

MATERIALS AND METHODS

Continuous cultivation system. The experiment was performed in a 2-stage continuous cultivation system. Details of the setup have been described earlier in Posch et al. (1999). Briefly, the first stage (2.4 l, further

described as I-Stage) was inoculated with the cryptophyte *Cryptomonas phaseolus* and its accompanying mixed bacterial assemblage (see Pernthaler et al. 2001 for a phylogenetic analysis of dominant community members). The algae were grown on WC-medium (Guillard & Lorenzen 1972) with a decreased phosphorus content of $50 \mu\text{g P l}^{-1}$ at 18°C in a light:dark rhythm of 16:8 h. Initially, the system was tested with distilled water and autoclaved before inoculation of the *C. phaseolus* and bacteria stockcultures. It was maintained as batch culture for 3 d, and afterwards as a chemostat for 79 d with a dilution rate of 0.39 d^{-1} (corresponding to a P-loading of $19.5 \mu\text{g P l}^{-1} \text{ d}^{-1}$). Three replicate vessels (780 ml each) were used as second stages. Two predator species were inoculated in the second stage vessels at Day 7, the mixotrophic flagellate *Ochromonas* sp. strain DS, an interception feeder (further described as II-Och), and the heterotrophic scuticociliate *Cyclidium glaucoma*, a filter feeder (further described as II-Cyc). The third vessel served as control without any predators (further described as II-Con). The dilution rate of the second stage vessels was 0.35 d^{-1} . Protists were precleaned from bacteria by washing and filtering the stockcultures with sterilized WC-medium onto 3 μm filters and inoculating them in *C. phaseolus* stockculture to remove their accompanying bacteria. This process was repeated 3 times within a month before the experiment was started.

Samples were taken every second day over a period of 79 d in order to follow the theoretical generation time of organisms, given by the dilution rate of the second stage. Sampling was carried out with a special construction to avoid contamination by other bacteria, which is particularly important for long-term experiments. Only a small volume of each vessel was removed every second day (75 ml, $<10\%$ of the total vessel volume), so that there was minimal disturbance of the chemostat system by sampling. Counting and sizing of organisms was done within 1 d after sampling.

Determination of bacterial and protistan abundance and biomass. Subsamples of 40 ml were fixed with alkaline Lugol's solution (0.5 % final concentration), followed by formaldehyde (3 % final concentration), and decolorized with several drops of sodium thiosulfate (Sherr & Sherr 1993). Samples were stained with DAPI ($4',6'$ -diamidino-2'-phenylindole, $6.7 \mu\text{g ml}^{-1}$ final concentration) and filtered onto black polycarbonate filters (Osmonics, 0.22 μm pore size, 25 mm diameter). Microscopic evaluation was carried out with a Zeiss Axiophot and a Zeiss Axioplan epifluorescence microscope at a magnification of $1600\times$ and $320\times$ for counting at least 1000 bacteria and 400 protists per sample, respectively. Ten to 20 images of single bacteria (corresponding to 500–700 cells), 50

images of microcolonies in II-Och, and 10 images of *Cryptomonas phaseolus*, *Ochromonas* sp. and *Cyclidium glaucoma* were recorded with CCD cameras (Vosskühler VDS CCD-1300 and Cohu) mounted on the microscopes, and processed with the image analysis program LUCIA_D (<http://www.lim.cz>). At least 500 to 700 bacteria, 300 algae and 50 predators were measured. Area and perimeter of each cell were measured, and cell length and width and the mean cell volume (MCV) were calculated applying geometric approximation (Massana et al. 1997). Details on image processing and the use of macros can be found in Posch et al. (1997). Bacterial carbon content was calculated using the volume to carbon conversion factor published by Loferer-Krößbacher et al. (1998). Bacterial carbon was estimated to amount for 50% of bacterial dry weight. For the determination of protistan carbon, we applied the formula published by Menden-Deuer & Lessard (2000). Total biomass was acquired by multiplying mean cellular carbon content with total abundance. Bacterial phosphorus content was calculated applying a conversion factor of 2 fg P cell⁻¹, while algal phosphorus content was assumed to be 900 fg P cell⁻¹. These conversion factors have been determined for the same microbial community in earlier experiments (Mindl et al. 2005, this issue). For *Ochromonas* sp., a conversion factor of 303 fg P cell⁻¹ was used (Rothhaupt 1996a), whereas no P conversion factor was available for *C. glaucoma*. Therefore, we calculated the ciliate's phosphorus content indirectly via subtraction of bacterial and algal P from the original phosphorus content of 50 µg l⁻¹ of the medium. Running averages were calculated by averaging 3 consecutive sampling dates (6 d).

Counting and sizing was done in triplicates at 2 sampling occasions, with 3 parallel sampling bottles and 3 parallel preparations for all bacterial and protistan parameters. Statistical comparisons of parameters of cell morphology and biomass were performed by 1-way ANOVA.

Determination of bacterial assemblage composition. Five to 7 ml of the fixed subsample (see above) were filtered onto white polycarbonate filters (Millipore, Type GTTP, 47 mm diameter, 0.2 µm pore size), rinsed with distilled water, and stored at -20°C until further processing. Sampling dates were selected for hybridization if significant changes were found either in bacterial or protistan abundance of II-Och or II-Cyc (altogether 7 timepoints).

Fluorescence *in situ* hybridization (FISH) was either carried out with fluorescently monolabeled oligonucleotide probes or with horseradish peroxidase labeled probes and catalyzed reporter deposition (CARD) (Pernthaler et al. 2002). No differences between detection rates by the 2 methods were found

in the experimental assemblage (Sekar et al. 2003). CARD-FISH was performed with the probe HGC-69a, targeted to the *Actinobacteria* (Sekar et al. 2003), with probes ALF-968 and BET-42a, specific for the α- and β-subclasses of Proteobacteria, respectively, and with probe EUB I-III, which detects most bacteria including *Verrucomicrobia* and *Planctomycetes* (Daims et al. 1999, Glöckner et al. 1999). Cell attachment and permeabilization on membrane filters and CARD-FISH was carried out as previously described with fluorescein-labeled tyramides for signal amplification (Pernthaler et al. 2002, Sekar et al. 2003). In addition, 3 specific probes labeled with the fluorescent dye Cy3 were applied for 3 microbial groups previously found in the experimental assemblage (Pernthaler et al. 2001). Probe CAU663 is targeted to bacteria affiliated with the *Caulobacter* spp.-*Brevundimonas* spp. lineage (α-Proteobacteria) (probe sequence: 5'-GGA GTT CCA CAT ACC TCT-3') (Neef 1997). Probe BET3-446 is targeted to an '*Aquabacterium* sp.' (α-Proteobacteria), and CF1-853, specific for bacteria from the *Flavobacterium ferrugineum*-*Flexibacter sancti* clade of the *Sphingobacteriales* (Pernthaler et al. 2001). FISH of filter sections was carried out as previously described (Alfreider et al. 1996, Glöckner et al. 1996), with the difference that cells had been pretreated as required for the CARD-FISH procedure (Sekar et al. 2003). The formamide concentrations in the hybridization buffers were 30% for probe CAU663, 40% for probe BET3-446, and 35% for probe CF1-853.

The filter sections were counter-stained with DAPI (1 µg ml⁻¹) and inspected with a Zeiss Axiophot epifluorescence microscope at a magnification of 1250×. At least 1000 DAPI-stained cells were counted per sample (500 cells for β-Proteobacteria, which were more frequent), corresponding to 15 to 500 hybridized cells. In the vessel II-Och, where besides single cells, microcolonies and aggregates also appeared, only single cells were analyzed by FISH. In addition, the food vacuoles of *Cyclidium glaucoma* and *Ochromonas* sp. were inspected for hybridized bacteria. For the mixotrophic flagellate this was only feasible using CARD-FISH because the chlorophyll autofluorescence masked the signals of the Cy3 monolabeled probes. To verify the precision of abundance determinations, 3 replicate hybridizations were made from different sampling dates with probes BET3-446 and BET-42a. Ten images of bacteria were recorded with a CCD camera from each filter section. Composite pictures of cells from these images were arranged with the image analysis program LUCIA_D to visualize the succession of morphotypes of different bacterial clades with and without predators and the cell lengths of hybridized cells was measured.

RESULTS

Bacterial and protistan abundance (Fig. 1)

In I-Stage and II-Con, similar patterns occurred in the course of the experiment. Initially the abundance of *Cryptomonas phaseolus* decreased slightly, remained stable until Day 35, and dropped to the lowest value at Day 51. Thereafter, algae recovered, with fluctuating abundances until the end of the experiment. Bacterial abundance showed the opposite trend, a slight initial increase to more or less stable cell numbers in the beginning, followed by a strong rise to a

maximum at Day 51. Subsequently, bacterial abundance dropped, remained stable for some days and increased again at the end of the experiment. The ratio of bacterial versus algal biomass changed rapidly between Days 40 and 60, from 0.2 to 0.8, and fluctuated between <0.2 and 0.4 thereafter (Fig. 2).

In II-Och, algae followed the same patterns as observed in I-Stage and II-Con. *Ochromonas* sp. grew exponentially until Day 19 and its numbers remained more or less stable thereafter (Fig. 1). Bacterial abundances decreased by more than 1 order of magnitude shortly after inoculation of the predator and remained at this low level (about 8% of the initial abundance) until the end of the experiment.

Algal numbers in II-Cyc showed the same trend as in the other variants, but formed higher abundances than in all other vessels and a short, but pronounced initial peak (Fig. 1). *Cyclidium glaucoma* increased exponentially at the beginning, but after a few days, the population densities started to drastically oscillate. These oscillations were most pronounced during the transient algae–bacteria imbalance situation between Days 40 to 60 (Fig. 2), when algal numbers were minimal. Bacterial abundance decreased after the introduction of the ciliate, but recovered slowly to higher than initial cell numbers at Day 51. The peak of bacterial abundances in the I-Stage between Days 40 to 60 was also reflected in II-Cyc.

Bacterial and protistan biomass and phosphorus (Tables 1 & 2)

Bacteria in I-Stage had the smallest mean cell volume in comparison with all other variants. Therefore, bacte-

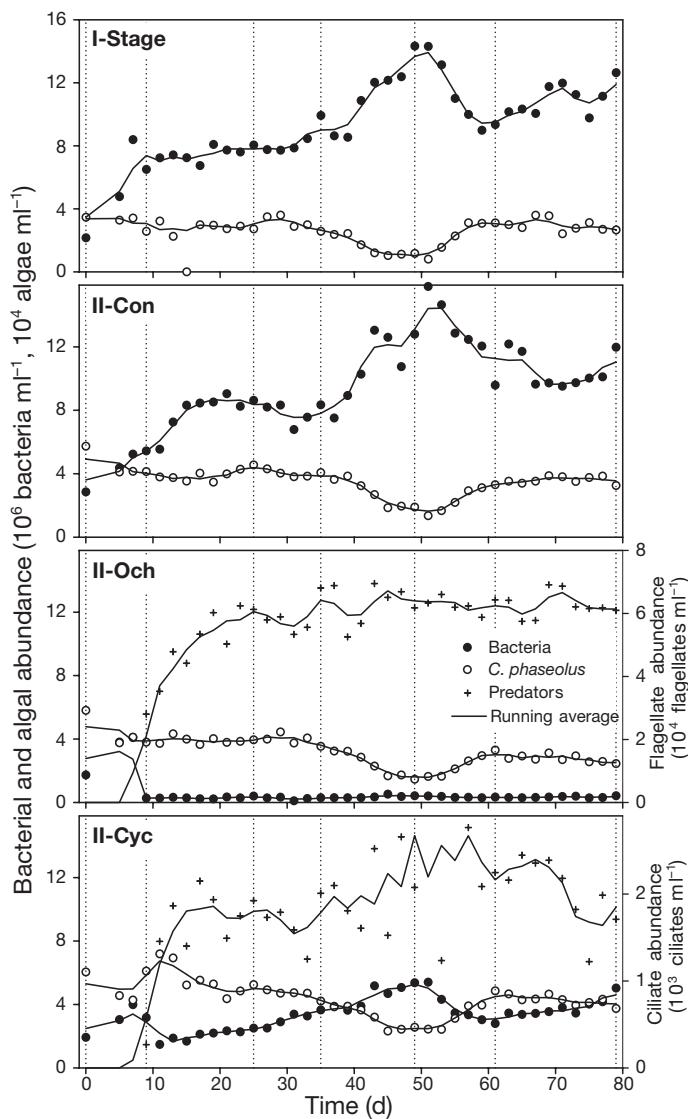


Fig. 1. Bacterial, algal and protistan abundances in the different vessels of the continuous cultivation system. Dotted lines indicate sampling dates selected for fluorescence *in situ* hybridization (FISH) analysis. Note the different units for bacterial and algal abundance on the left y-axis. Abbreviations as in Fig. 1 legend

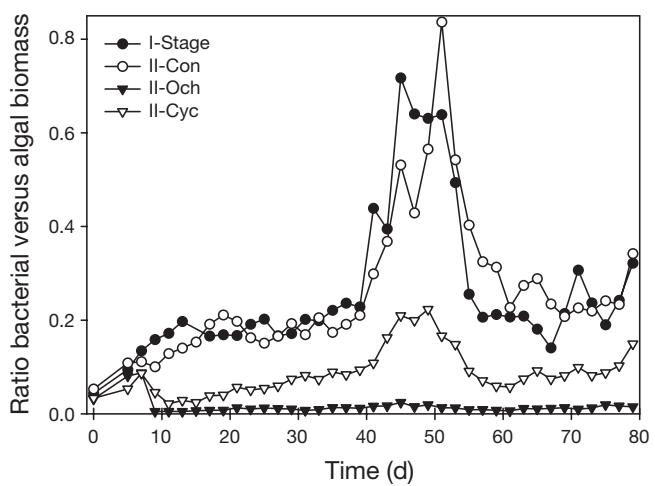


Fig. 2. Bacterial versus algal biomass ratio in the 4 vessels of the continuous cultivation system. I-Stage: *Cryptomonas phaseolus* and accompanying bacteria; II-Con: control; II-Och: *Ochromonas* sp.; II-Cyc: *Cyclidium glaucoma*

Table 1. Community parameters of bacteria, algae and predators. Values are averages calculated after the introduction of the predators ($t = 68$ d, $n = 35$). Asterisks indicate a significant difference (1-way ANOVA, $p < 0.001$) of the vessels II-Och and II-Cyc compared to vessel II-Con (control). Bac: bacteria; Alg: algae; Pre: predators (*Ochromonas* sp. in II-Och, *Cyclidium glaucoma* in II-Cyc); Σ : sum of all parameters; ALF-968, CAU663, BET-42a, BET3-446, CF1-853, HGC-69a: fluorescence *in situ* hybridization (FISH) probes

Community parameter	I-Stage	II-Con	II-Och			II-Cyc	
			Total	Single cells	Microcolonies		
Abundance (cells ml ⁻¹)	10^6 Bac	9.90	10.00	0.44*	0.32*	0.12*	3.44*
	10^4 Alg	2.55	3.39	3.05*	—	—	4.25
	10^4 Pre	—	—	5.97	—	—	0.20
Biovolume (mm ³ l ⁻¹)	Bac	1.153	1.882	0.081*	0.043*	0.038*	0.777*
	Alg	9.53	14.15	13.45	—	—	17.51*
	Pre	—	—	5.42	—	—	2.27
	Σ	10.69	16.03	18.96	—	—	20.56
Biomass (µg C l ⁻¹)	Bac	344	518	22*	12*	10*	208*
	Alg	1432	2111	2000	—	—	2615*
	Pre	—	—	886	—	—	318
	Σ	1776	2629	2908	—	—	3141
Phosphorus ^a (µg P l ⁻¹)	Bac	19.8	20.1	0.9*	0.6*	0.2*	6.9*
	Alg	23.0	30.5	27.4	—	—	38.3*
	Pre	—	—	18.1	—	—	4.8
	Σ	42.8	50.6	46.4	—	—	50.0
Bacterial assemblage composition (% of DAPI)	ALF-968	5.6	7.6	13.2	—	—	27.3*
	CAU663	2.5	3.0	8.6*	—	—	4.4
	BET-42a	86.5	81.9	43.9	—	—	44.8*
	BET3-446	41.4	37.2	20.8*	—	—	19.2*
	CF1-853	2.9	4.6	19.1*	—	—	14.5
	HGC-69a	0.5	0.7	1.6	—	—	0.8

^aPhosphorus content of the predators was evaluated as mentioned in the text

Table 2. Single-cell characteristics of bacteria, algae and predators. Values are averages calculated after the introduction of the predators ($t = 68$ d, $n = 35$). Asterisks indicate a significant difference in single-cell characteristics (1-way ANOVA, $p < 0.001$) of the vessels II-Och and II-Cyc compared to vessel II-Con (control). Bac: bacteria; Alg: algae; Pre: predators (*Ochromonas* sp. in II-Och, *Cyclidium glaucoma* in II-Cyc); MCL: mean cell length; MCW: mean cell width; MCV: mean cell volume; CCC: cellular carbon content

Cellular parameter	I-Stage	II-Con	II-Och			II-Cyc	
			Total	Single cells	Microcolonies		
MCL (µm)	Bac	1.47	1.83	—	1.43*	3.49*	1.46*
	Alg	13.42	13.92	13.92	—	—	13.68
	Pre	—	—	5.49	—	—	17.52
MCW (µm)	Bac	0.34	0.39	—	0.37*	0.35*	0.47*
	Alg	7.29	7.53	7.76	—	—	7.51
	Pre	—	—	5.49	—	—	10.92
MCV (µm ³)	Bac	0.118	0.190	—	0.135*	0.322*	0.229*
	Alg	381	422	448	—	—	413
	Pre	—	—	90	—	—	1126
CCC (fg C cell ⁻¹)	Bac	34.9	51.6	—	38.0*	81.6*	60.3
	10^3 Alg	57.1	62.9	66.6	—	—	61.6
	10^3 Pre	—	—	2.7	—	—	158.0

rial biomass was about 1 third lower than in II-Con, although bacterial abundances in these 2 vessels were nearly equal. In II-Och, we distinguished between freely dispersed bacterial cells and microcolonies (Table 2). Single cells quickly became significantly shorter than in

II-Con, whereas cells in microcolonies were always significantly longer. Although the abundance of single cells was twice as high as of cells in microcolonies, total bacterial biomass of these 2 cell types was approximately equal. Bacterial widths were significantly larger in II-Cyc

than in the control vessel (Table 2); the same holds true for bacterial cell volumes. Bacteria in this treatment had the highest cellular carbon content and formed relatively high biomasses. Algal phosphorus content of the second stage was highest in the vessel containing the ciliate (II-Cyc) and lowest in the vessel II-Och, where the flagellate might have incorporated phosphorus due to its mixotrophic life strategy.

Changes of the bacterial assemblage due to grazing

In II-Och, only a few freely dispersed cells were present. The remaining bacteria (up to 56 % of total numbers) formed microcolonies with mean numbers of 5 to 7 cells per colony, large aggregates (containing more than 25 cells per aggregate) and long, single filaments. While microcolony forming cells could be counted easily due to their loose cell contact, this was not possible for densely packed aggregates. Therefore, only single cells were evaluated by FISH. Microcolonies also occurred in stage II-Cyc, but at a later timepoint, in lower densities and with other cell morphologies. The proportion of single cells was much higher in this vessel (II-Cyc).

FISH was performed for selected sampling dates that reflected drastic population changes of either bacteria or predators (see Fig. 1). The first 2 timepoints (Day 0 and Day 9) describe the situation before and shortly after inoculation of the 2 predators. The 3 subsequent timepoints (Days 25, 35 and 49) mirror the slight (II-Och) or pronounced (II-Cyc) recovery of bacterial abundances and the eventual peak of bacterial numbers in I-Stage and II-Con. During this period, predators reached very high abundances. The sixth datapoint describes the situation of bacterial abundance decrease while predators remained stable (Day 61). The last timepoint describes the community composition at the very end of the experiment, when bacterial numbers recovered in I-Stage, II-Con, and II-Cyc (Day 79).

Approximately 95 % of DAPI-stained cells could be visualized at several timepoints in hybridizations with probe EUB I-III (data not shown). Slightly lower detection rates (87 %) were found in II-Och at the last sampling date. Nearly 100 % of total bacterial abundance could be accounted for by the different probes (Fig. 3). At our level of analysis, the composition of the bacterial assemblages remained remarkably stable in I-Stage and II-Con even during large shifts in algal to bacterial biomass ratios. β -Proteobacteria formed the dominating group in these vessels (56 to 93 % of DAPI-stained cells). About 50 % of β -Proteobacteria could be assigned to '*Aquabacterium* sp.' by probe BET3-446.

In II-Och, β -Proteobacteria decreased strongly, and constituted less than 30 % of bacteria in this treatment

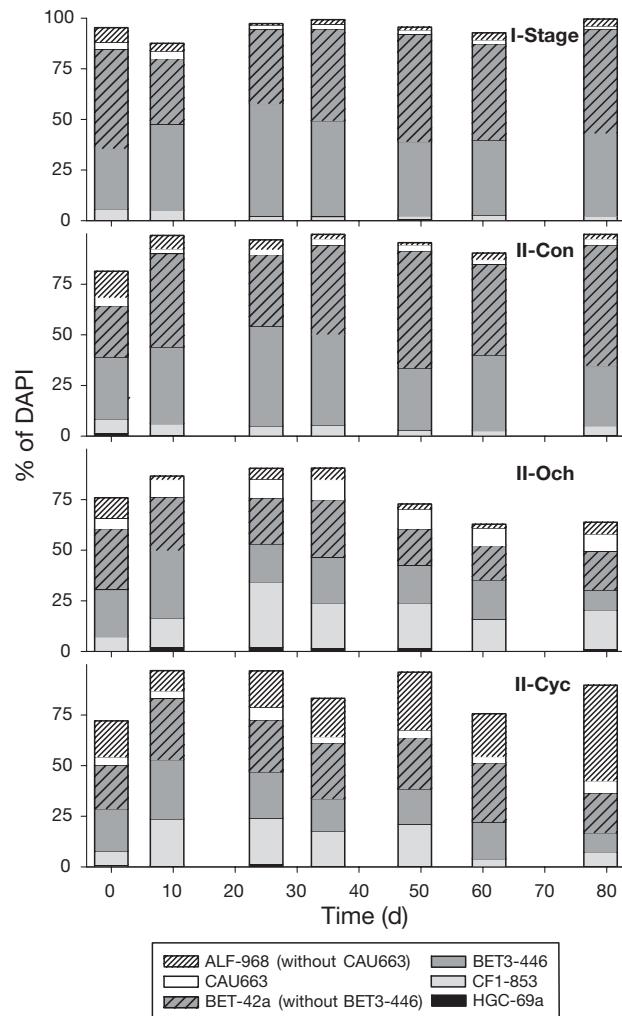


Fig. 3. Relative contribution of different bacterial groups to the assemblages, as determined by fluorescence *in situ* hybridization (FISH). Probes: ALF-968 (α -Proteobacteria without *Caulobacter* spp.), CAU663 (*Caulobacter* spp.), BET-42a (β -Proteobacteria without '*Aquabacterium* sp.'), BET3-446 ('*Aquabacterium* sp.'), CF1-853 (members of the *Flexibacter sancti-Flavobacterium ferrugineum* lineage), HGC-69a (*Actinobacteria*). Abbreviations as in Fig. 1 legend

at the end of the experiment (Figs. 3 & 4). Bacteria of the *Flexibacter sancti-Flavobacterium ferrugineum* lineage of *Sphingobacteriales* (Probe CF1-853) increased from 7 % to a maximum of 32 % of DAPI-stained cells at Day 25 (Fig. 3). Both single cells and microcolonies with loose cell contact were detected with this probe, and were counted separately (Fig. 5). The abundance of single cells decreased slightly thereafter, whereas the abundance of cells in microcolonies remained high. *Caulobacter*-like bacteria hybridized by Probe CAU663 increased in proportion in II-Och (maximum, 10 % at Day 35), whereas other α -Proteobacteria decreased slightly.

In II-Cyc we observed a strong relative decrease in the numbers of all β -Proteobacteria and of 'Aqua-bacterium sp.' (Fig. 3). Bacteria of the *Flexibacter sancti-Flavobacterium ferrugineum* lineage strongly increased until Day 49 (up to 21%) (Fig. 5), but then

dropped by almost 1 order of magnitude at Day 61. At the end of the experiment, a slight recovery was observed. Bacteria hybridized with probe CAU663 were rare and remained rather stable, but other α -Proteobacteria strongly increased to 53% at the end of

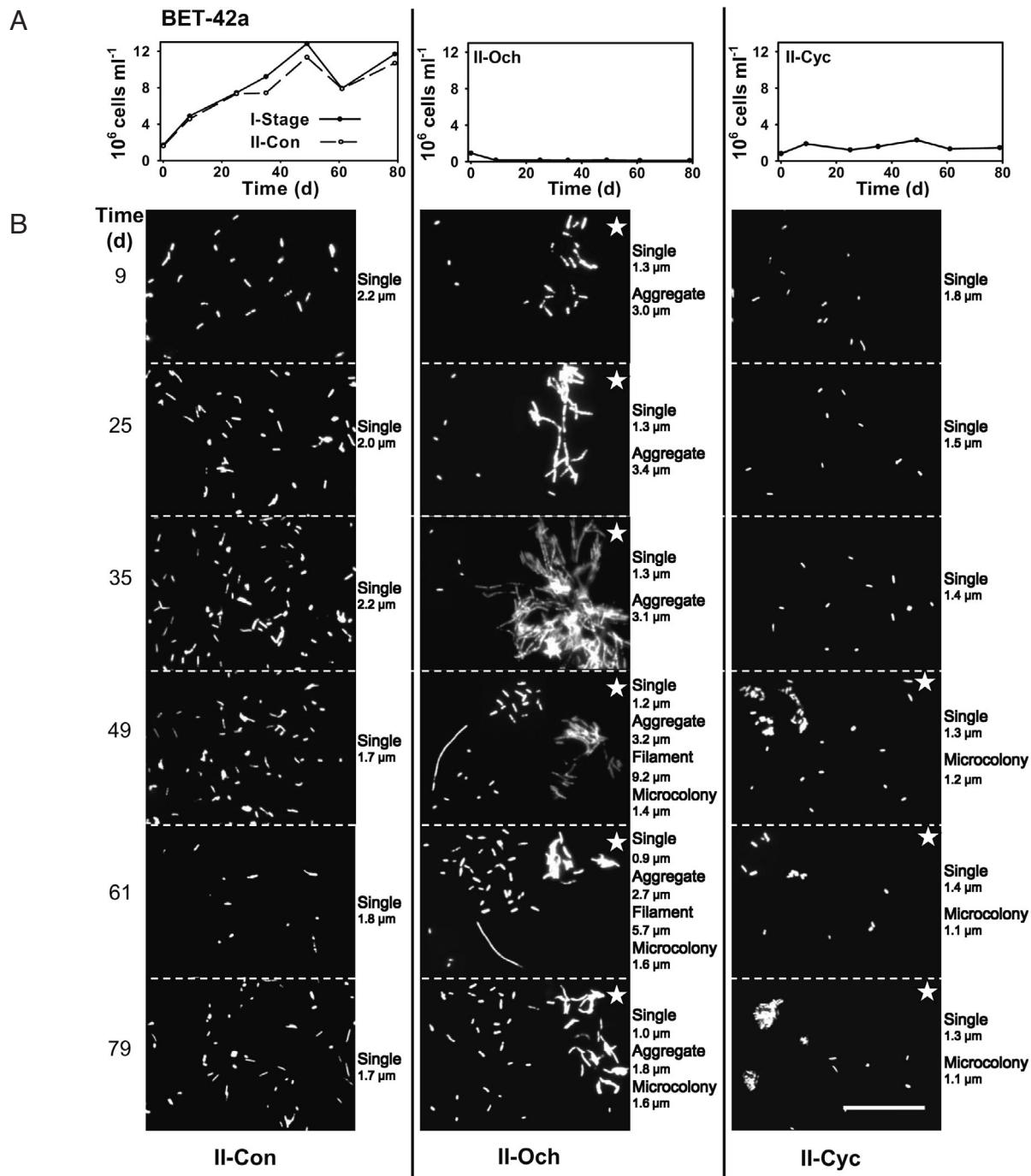


Fig. 4. Succession in (A) abundance and (B) morphotype of bacterial populations detected by probe BET-42a in the vessels II-Con, II-Och and II-Cyc. On the right side of each picture in (B) is a short description of the cell types and the mean cell lengths. White stars in II-Och and II-Cyc indicate that pictures do not represent real abundances of bacteria, but are composed of different photographs in order to show diverse morphotypes. The scale bar at the bottom is 10 μ m and applies to all pictures.

Abbreviations as in Fig. 1 legend

the experiment (Fig. 6). In this treatment, a succession of bacterial populations was observed that contrasted with the development of these populations in I-Stage and II-Con: a decline of β -Proteobacteria and a ra-

pid increase of bacteria from the *Flexibacter sancti-Flavobacterium ferrugineum* lineage was followed by a strong decrease of the latter group and a bloom of α -Proteobacteria.

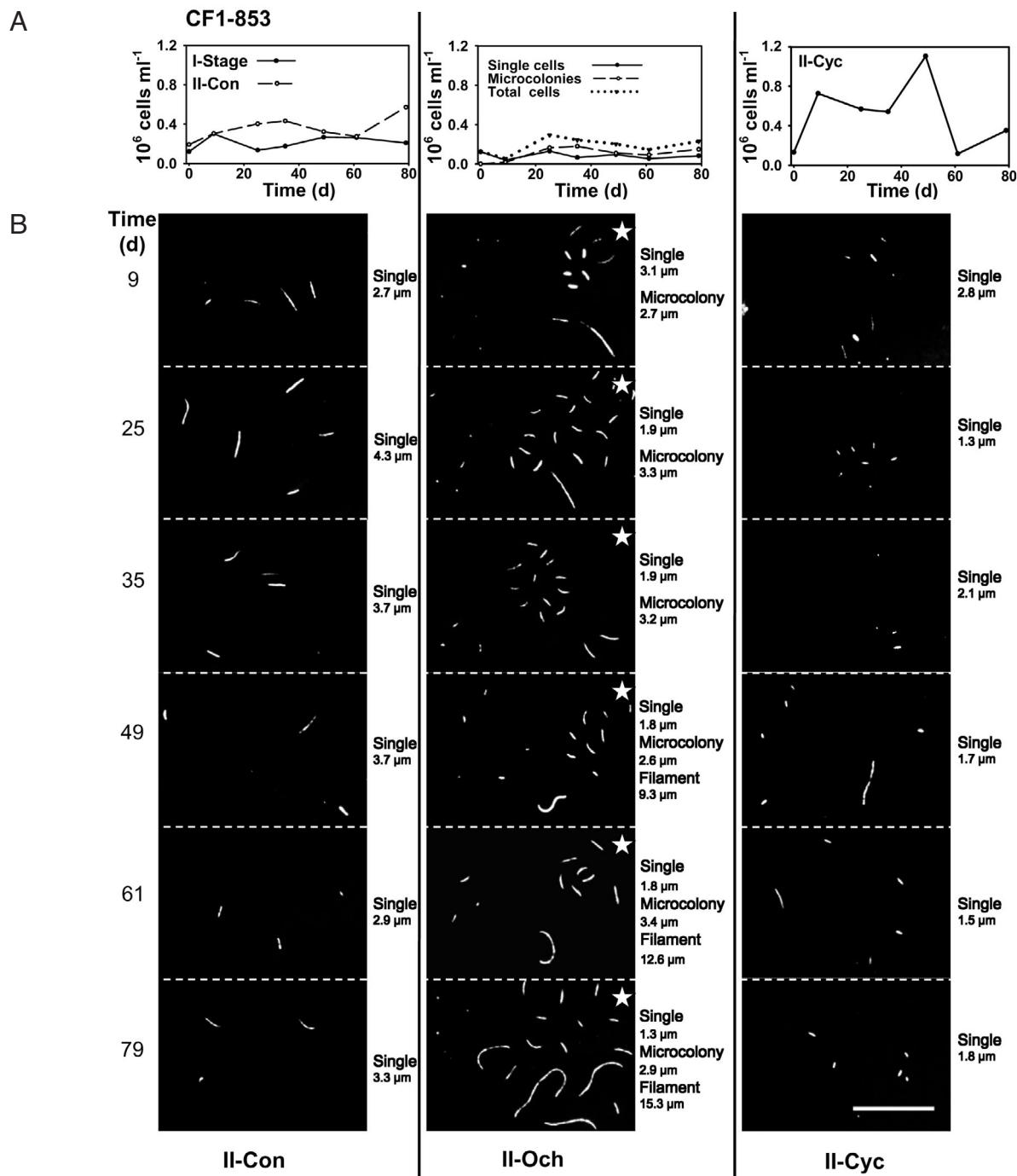


Fig. 5. Succession in (A) abundance and (B) morphology of the bacterial populations detected by probe CF1-853 in the vessels II-Con, II-Och and II-Cyc. On the right side of each picture in (B) is a short description of the cell types and the mean cell lengths. White stars in II-Och indicate that pictures do not represent real abundances of bacteria but are composed of different photographs in order to show diverse morphotypes. The scale bar at the bottom is 10 μm and applies to all pictures.

Abbreviations as in Fig. 1 legend

Morphological shift of the different bacterial groups

β -Proteobacteria (Fig. 4) showed a high phenotypic plasticity, and 5 different morphotypes could be

observed. In I-Stage and II-Con, these bacteria formed medium-sized single rods, whereas in II-Och, all single cells were smaller. Immediately after inoculation of *Ochromonas* sp., large aggregates occurred, which

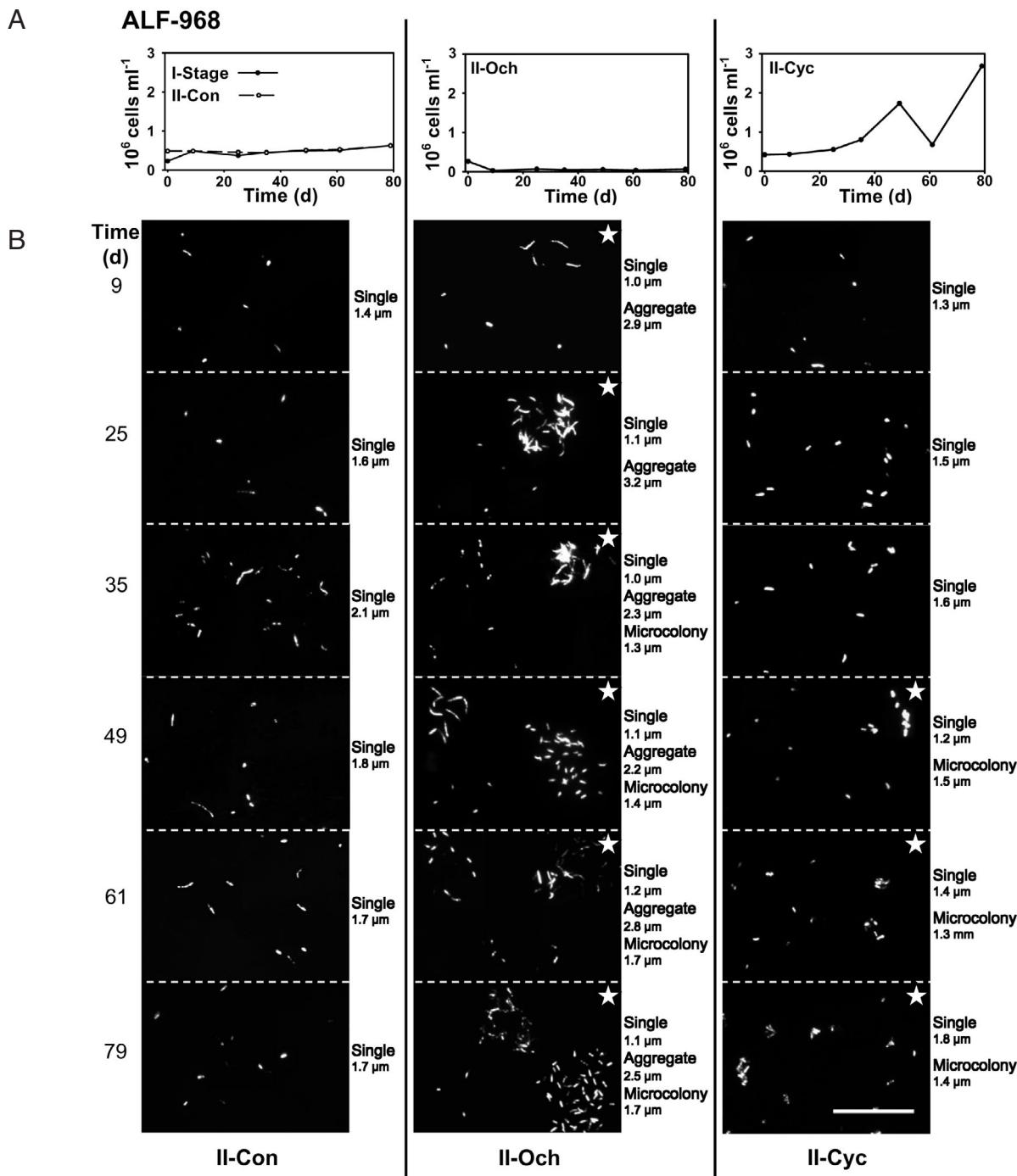


Fig. 6. Succession in (A) abundance and (B) morphology of the bacterial populations detected by probe ALF-968 in the vessels II-Con, II-Och and II-Cyc. On the right side of each picture in (B) is a short description of the cell types and the mean cell lengths. White stars in II-Och and II-Cyc indicate that pictures do not represent real abundances of bacteria but are composed of different photographs in order to show diverse morphotypes. The scale bar at the bottom is $10 \mu\text{m}$ and applies to all pictures.

Abbreviations as in Fig. 1 legend

contained primarily bacteria related to *Aquabacterium* sp. (probe BET3-446), with some cells affiliating with *Caulobacter* spp. (probe CAU663). These aggregates were not very abundant, but became very large (more than 25 cells per aggregate). At Day 49, 2 more morphotypes of β -Proteobacteria were discernable: long rare filaments and small microcolonies with loose cell contact, which increased in numbers until the end of the experiment. These 2 morphotypes could not be detected with probe BET3-446. All β -Proteobacteria in II-Cyc formed rod-shaped single cells until Day 49. Afterwards, microcolonies with tight cell contact appeared, which became very abundant. Cells in these colonies could not be detected with probe BET3-446.

Bacteria related to the *Flexibacter sancti*–*Flavobacterium ferrugineum* lineage (Fig. 5) formed single rods in I-Stage and II-Con. In II-Och, there was a morphological separation between small single cells and microcolonies containing 5 to 7 larger cells with loose cell contact. These colonies became dominant during the experiment (up to 53% in comparison to abundances of single cells). At Day 49, a third cell type hybridizing with CF1-853 appeared, consisting of long filaments, which slightly increased in abundance until the end of the experiment. In the ciliate vessel (II-Cyc), only 1 cell type was hybridized by probe CF1-853, i.e. small rod-shaped single cells.

α -Proteobacteria (Fig. 6) fell into a large range of size classes in I-Stage and II-Con, from small coccoid forms to rod-shaped bacteria. In II-Och, shortly after inoculation of *Ochromonas* sp., *Caulobacter* spp. was present in aggregates dominated by '*Aquabacterium* sp.' (see above). The remaining α -Proteobacteria formed very small single cells until Day 35. At this timepoint small colonies of α -Proteobacteria with loose cell contact appeared, which became more abundant towards the end of the experiment. In the vessel containing *Cyclidium glaucoma*, α -Proteobacteria formed rod-shaped single cells. At Day 49, small colonies appeared also in II-Cyc, but these colonies had direct cell contact, and were not hybridized by probe CAU663.

DISCUSSION

Effects of predators with differing feeding strategies

Ochromonas sp. reached very high abundances shortly after inoculation and minimized bacterial abundance dramatically (Fig. 1). In contrast to former studies on the same experimental community no bloom of *Actinobacteria* was observed in II-Och (Pernthaler et al. 2001, T. Posch et al. unpubl. data). However, we still found a clear relationship between grazing pressure and bacterial size structure. Shortly after intro-

duction of the predator, the bacterial assemblage responded by a bidirectional shift towards small single cells and very large and complex morphologies, such as aggregates, microcolonies and filaments. This grazing-protective strategy has also been found in previous studies (Güde et al. 1989, Pernthaler et al. 1996, Hahn & Höfle 1999, Jürgens & Matz 2002). *Ochromonas* sp. preferentially feeds on bacteria larger than $0.2 \mu\text{m}^3$ ($\sim 0.7 \mu\text{m}$ in diameter, Andersson et al. 1986). Although there was a shift towards smaller cells, 78% of all free-living bacterial cells found in II-Och were within the edible size range ($> 0.2 \mu\text{m}^3$), which probably explains why these cells did not recover in abundance. On the one hand, morphological changes could be a consequence of size selectivity by the flagellate. On the other hand, the observed phenomena could also be caused by the extremely high total grazing rates irrespective of size selectivity. Due to its mixotrophic life strategy, *Ochromonas* sp. can maintain high densities, thus exerting high grazing pressure on bacteria even when only little particulate food is available. When grown under light, the flagellate increasingly uses its ability to grow photosynthetically, if bacterial densities fall below a concentration of $< 10^6 \text{ cells ml}^{-1}$ (Rothhaupt 1996b, 1997), and it has to take up phosphate. As a consequence, bacteria in II-Och had to compete for phosphorus not only with the algae (Mindl et al. 2005), but also with a predominantly phototrophic bacterivore. Another effect of the phosphorus-incorporation of the flagellate is that algae had the lowest phosphorus content in this vessel.

In contrast, exclusively heterotrophic protists such as *Cyclidium glaucoma* release surplus phosphorus and other nutrients, and therefore stimulate P-limited algal and bacterial growth (Eccleston-Parry & Leadbeater 1995, Rothhaupt 1997). This beneficial effect was found in II-Cyc, where algae reached higher densities than in all other vessels. Interestingly, bacteria grazed by the filter feeder *C. glaucoma* reacted with an increase in cell width rather than in cell length (Table 2). This phenomenon was also observed in earlier chemostat experiments (Posch et al. 2001). Considering the complex structures of the ciliate feeding apparatus, it is conceivable that such increased cell width might provide at least partial protection against grazing by *C. glaucoma*.

Effects on assemblage composition (Table 3)

The ability to form grazing-resistant morphologies seems to be widespread among various bacterial phylotypes (Table 3). Field studies as well as laboratory experiments typically show that freshwater β -Proteobacteria are negatively affected by protistan grazing if

Table 3. Examples for the effects of protistan grazing on the bacterial assemblage composition observed in (a) field and (b) chemostat studies applying fluorescence *in situ* hybridization (FISH) (only examples for freshwater systems). Abbreviations for oligonucleotide probes: *Bacteria*: EUB = EUB338; *Actinobacteria*: Ac1 = Ac1-847, HGC = HGC-69a; *Archaea*: ARCH = ARCH915; α -Proteobacteria: ALF = ALF-968, CAU = CAU663, PS = *Pseudomonas* sp.; β -Proteobacteria: BET = BET-42a, BONE = BONE23a; BET3 = BET3-446, R-BT = R-BT65; γ -Proteobacteria: GAM = GAM-42a; *Cytophaga–Flavobacterium–Bacteroides*: CF = CF-319a, CF1 = CF1-853, R-FL = R-FL615. FRACT = size fractionation, LONG = longitudinal transect, TRANS = transplant experiment

(a) Lake	Applied probes	EUB (%)	Natural assemblages	Effects of grazing	Time/ Notes	Source
Small fishless pond	EUB HGC ALF1b BET BONE GAM CF	60–85	BET > CF	Decrease in BET & CF Increase in ALF & GAM Mainly ALF formed filaments	0–140 h	(1)
Rímov reservoir	EUB ALF BET GAM CF	45–85	BET ~ CF > ALF > GAM	Increase in ALF & CF ALF formed filaments	0–96 h	(2)
		45–60	BET > CF > ALF > GAM	Decrease in BET increase in ALF & CF CF formed filaments	0–96 h	(2)
Three small fishless ponds	EUB ARCH ALF1b			Aggregates dominated by ALF, freely dispersed cells dominated by BET	Data from Daphnia-free variants	(3)
Rixdorf	BET GAM CF	70–95	BET ~ CF	Increase in BET & ALF Decrease in CF	0–94 h	
Freudenholm		60–80	BET > CF > ALF	Increase in BET & ARCH Decrease in CF	0–116 h	
Hasselburg		50–80	BET > CF > ALF ~ ARCH	Increase in ALF	0–94 h	
Sau Reservoir	EUB ALF BET GAM CF	60–90	CF > BET > ALF ~ GAM	Decrease in BET & CF Biomass dominance of BET	April, July 1997 LONG	(4)
Rímov reservoir	EUB ALF BET R-BT GAM CF R-FL	75–85	BET > CF > GAM > ALF	Increase in BET & CF R-FL formed filaments Aggregates by BET & CF	0–96 h FRACT	(5)
Rímov reservoir	EUB ALF BET R-BT GAM CF R-FL	50–85	BET ~ CF > R-BT > GAM	Decrease in BET & R-BT	0–96 h FRACT TRANS	(6)

Table continued on next page

they morphologically resemble cocci or small single rods (Jürgens et al. 1999, Posch et al. 2001, Šimek et al. 2001b, Gasol et al. 2002). Via food vacuole inspection, we found that 85 % of all ciliates and 23 % of all flagellates had ingested β -Proteobacteria. This hints at an unresolved paradox within aquatic microbial communities, since this phylogenetic group (together with *Acti-*

nobacteria) forms a prominent fraction in many freshwater systems (Methé & Zehr 1999, Glöckner et al. 2000, Burkert et al. 2003). However, at present only few groups of the β -Proteobacteria are known to be strongly affected by grazing, e.g. '*Aquabacterium* sp.' (Pernthaler et al. 2001) and members of the freshwater beta I clade (Šimek et al. 2001b).

Table 3 (continued)

(b) Chemostat predators	Applied probes	EUB (%)	Natural assemblages	Effects of grazing	Time/ Notes	Source
<i>Bodo saltans</i>	EUB ALF BET GAM CF	80–97	BET > ALF ~ GAM > CF	Decrease in BET Increase in ALF	12 d	(7)
<i>Ochromonas</i> sp.	EUB PS BET GAM	81	GAM > BET > PS	Increase in flocs (PS)	261 d	(8)
<i>Cyclidium glaucoma</i> , <i>Bodo saltans</i> , <i>Ochromonas</i> sp.	Ac1 CAU BET3	–	BET3 ~ CF > CAU > Ac1	Dominance of Ac1 Increase in ALF & CF1 Decrease in BET3	12–16 d	(9)
<i>Cyclidium glaucoma</i>	ALF BET CF	–	ALF > BET > CF	Decrease in BET & CF	16d	(10)

(1) Jürgens et al. (1999), (2) Šimek et al. (1999), (3) Langenheder & Jürgens (2001), (4) Šimek et al. (2001a), (5) Šimek et al. (2001b), (6) Šimek et al. (2003), (7) Šimek et al. (1997), (8) Hahn & Höfle (1999), (9) Pernthaler et al. (2001), (10) Posch et al. (2001)

Some pelagic freshwater β -Proteobacteria might be able to rapidly increase their growth rates upon environmental change (Jürgens et al. 1999, Šimek et al. 2001b, Burkert et al. 2003). Other species of β -Proteobacteria potentially exhibit high phenotypic plasticity, i.e. they can form various morphotypes such as microcolonies, filaments and aggregates (Jürgens et al. 1999, Šimek et al. 2001b, Hahn et al. 2004). In our system, such large and complex structures were obviously resistant to protistan grazing, as they were not found in protists food vacuoles and maintained relatively high abundances. β -Proteobacteria related to the drinking-water biofilm bacterium *Aquabacterium* sp. (Kalmbach et al. 1999) were able to form morphotypes that were resistant even to strong grazing (Fig. 4). They constituted $46 \pm 8\%$ of all β -Proteobacteria in I-Stage, II-Con and II-Cyc, and this proportion remained stable throughout the experiment with slightly lower values for II-Och ($42 \pm 7\%$). These bacteria formed large branched colonies consisting of at least 25 cells in the vessel containing the flagellate. They are phylogenetically closely related to stream biofilm bacteria (Brümmer et al. 2003), and to a freshwater isolate (MWH55) that was found to develop microcolonies when exposed to grazing by *Ochromonas* sp. strain DS (Hahn et al. 2004). In our experiment, this extreme response to grazing pressure was only observed in II-Och but not in II-Cyc, where grazing rates were much lower. At present, we can only speculate if this phenomenon is linked to the contrasting size selectivity of the 2 preda-

tors, or if it is due to different remineralization rates of the mixotrophic and the heterotrophic protist.

Members of the α -Proteobacteria usually do not form a dominant fraction of freshwater bacterioplankton (Methé & Zahr 1999, Glöckner et al. 2000, Klammer et al. 2002). Interestingly, these bacteria were favored in many investigations that feature an experimental increase in grazing pressure (Table 3, Jürgens et al. 1999, Šimek et al. 1999). Bacteria hybridized by probe CAU663 contributed $48 \pm 18\%$ of all α -Proteobacteria in I-Stage and II-Con, with higher values for II-Och ($68 \pm 18\%$) and lower values for II-Cyc ($18 \pm 7\%$). The typical long stalks of single *Caulobacter* spp. cells (Poindexter et al. 2000) might have caused difficulties especially for the filter-feeding ciliate which only rarely ingested bacteria detected by probe CAU663. Members of α -Proteobacteria seemed to be generally less vulnerable to grazing by *Cyclidium glaucoma*, although they were also found in the ciliate's food vacuoles. However, a variety of morphotypes related to α -Proteobacteria dominated the bacterial assemblage in II-Cyc at the end of the experiment. Again, it seems to be a paradox that members of α -Proteobacteria can coexist with strong grazing pressure but are rare in freshwaters. This suggests that factors other than protistan grazing also substantially affect the abundance of particular bacterial genotypes in freshwater systems.

Apart from morphological changes of single taxonomic groups we could also observe bacterial coexis-

tence in aggregates as a result of strong grazing pressure. The large branched colonies formed by members of the BET3-446 cluster were always associated with single cells affiliated with *Caulobacter* spp. A similar phenomenon has been described from samples of a meso-eutrophic reservoir, where large aggregates of β -Proteobacteria were interspersed with members of the *Sphingobacteriales* (Šimek et al. 2001b). Thus, protistan grazing does not only directly affect individual phylogenetic clusters, but may also modify the interactions between different taxonomic groups.

Bacteria of the *Flexibacter sancti-Flavobacterium ferrugineum* lineage of *Sphingobacteriales* also exhibited a wide spectrum of morphotypes. Interestingly, microcolonies and long filaments only occurred in the presence of *Ochromonas* sp., but not with *Cyclidium glaucoma* (Fig. 5). This again illustrates that microbial predation-defence mechanisms may be triggered only by particular predator species. Since the probe is targeted to a lineage harbouring a number of distinct phylotypes (Pernthaler et al. 2001), the presence of both filaments and microcolonies may either reflect the phenotypic plasticity of a single population, or of co-occurring bacterial populations affiliated with this phylogenetic clade.

Successions in mixed microbial assemblages

It is likely that microbial assemblages respond differently to gradual and to rapid changes of grazing-induced mortality. In most experimental studies, 'unnaturally' strong shifts of grazing pressure are induced, either by inoculating high numbers of protistan predators or by decoupling the microbial food web by size fractionations (Table 3). Such experimental set-ups will lead to rather extreme responses, which probably occur in nature only during particular time periods. In contrast, it is not well understood how a mixed bacterial assemblage is affected by varying levels of protistan grazing pressure, e.g. throughout the season. Freshwater pelagic bacteria are typically confronted with changing intensities of grazing pressure, as well as with shifts in the species composition of the dominant bacterivorous protists (Sanders et al. 1989, Sommaruga & Psenner 1995, Cleven & Weisse 2001). Therefore, it is likely that seasonal community patterns of bacterivorous protists are reflected in successions of bacterial phylotypes. In addition, these successions will also be determined by bottom-up factors, e.g. the availability of limiting substrates and nutrients.

Unfortunately, it is hardly feasible to experimentally mimic protistan species successions. However, it is possible to examine if there might be successions of mixed bacterial assemblages if exposed to different

grazing modes over an extended period of time. In our experiment, we attempted to model a microbial community which features some of the complexity of natural planktonic systems, where interactions between bacteria, algae, and bacterivorous protists occur simultaneously. We hypothesized that there might be no stable long-term response of the bacterial assemblage in such a complex system even if the physico-chemical parameters were tightly controlled. This was verified both by the more pronounced genotypic shifts in the grazing exposed assemblages (Fig. 3), and the successions of different grazing resistant morphotypes (Figs. 4 to 6).

Another very basic interaction that may introduce instability both in our system and in the freshwater pelagic zone is the competition for limiting phosphorus. The repeatedly observed oscillation of bacterial versus algal biomass ratios in the cultivation system (Fig. 2, Mindl et al. 2005) is most likely caused by this phenomenon. According to the so-called phytoplankton-bacteria paradox (Bratbak & Thingstad 1985), the ratio of bacterial versus algal biomass increases with decreasing availability of phosphorus. In previous chemostat experiments with media containing $>40 \mu\text{g P l}^{-1} \text{d}^{-1}$ (Mindl et al. 2005), the ratio of bacterial versus algal biomass ranged around 0.05 to 0.1. These ratios remained stable over periods of 2 mo. In contrast, the use of media with $20 \mu\text{g P l}^{-1} \text{d}^{-1}$ led to a bacteria:algae ratio of 0.2 to 0.3, followed by considerable oscillations (Fig. 2 and Mindl et al. 2005). A similar situation may occur in natural oligo- to mesotrophic freshwater systems. Since bacterial carbon supply in the first stage of our system exclusively depends on exudates of *Cryptomonas phaseolus*, these imbalance situations only lasted for several days.

Implications for future studies

Presently it is still poorly understood which bacterial phylotypes in freshwaters are preferentially eliminated by bacterivores, and how the development of different grazing protection mechanisms by specific bacterial populations is affected by the availability of nutrients. Future studies are required to explore how (1) the intensity of grazing, (2) the community composition and (3) the size selectivity of protistan grazers will influence the phenotypes and the seasonal dynamics of different bacterial populations in pelagic systems.

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LITERATURE CITED

- Alfreider A, Pernthaler J, Amann R, Sattler B, Glöckner FO, Wille A, Psenner R (1996) Community analysis of the bacterial assemblages in the winter cover and pelagic layers of a high mountain lake by *in situ* hybridization. Appl Environ Microbiol 62:2138–2144
- Andersson A, Larsson U, Hagström Å (1986) Size-selective grazing by a microflagellate on pelagic bacteria. Mar Ecol 33:51–57
- Bratbak G, Thingstad TF (1985) Phytoplankton–bacteria interactions: an apparent paradox? Analysis of a model system with both competition and commensalism. Mar Ecol Prog Ser 25:23–30
- Brümmer IHM, Felske A, Wagner-Döbler I (2003) Diversity and seasonal variability of *beta*-Proteobacteria in biofilms of polluted rivers: analysis by temperature gradient gel electrophoresis and cloning. Appl Environ Microbiol 69: 4463–4473
- Burkert U, Warnecke F, Babenzien D, Zwirnmann E, Pernthaler J (2003) Members of a readily enriched β -Proteobacterial clade are common in surface waters of a humic lake. Appl Environ Microbiol 69:6550–6559
- Cleven EJ, Weisse T (2001) Seasonal succession and taxon-specific bacterial grazing rates of heterotrophic nanoflagellates in Lake Constance. Aquat Microb Ecol 23:147–161
- Daims H, Bruhl A, Amann R, Schleifer KH, Wagner M (1999) The domain-specific probe EUB338 is insufficient for the detection of all *Bacteria*: development and evaluation of a more comprehensive probe set. Syst Appl Microbiol 22: 434–444
- Eccleston-Parry JD, Leadbeater BSC (1995) Regeneration of phosphorus and nitrogen by four species of heterotrophic nanoflagellates feeding on three nutritional states of a single bacterial strain. Appl Environ Microbiol 61:1033–1038
- Gasol JM, Comerma M, Garcia JC, Armengol J, Casamayor EO, Kojacká P, Šimek K (2002) A transplant experiment to identify the factors controlling bacterial abundance, activity, production, and community composition in a eutrophic canyon-shaped reservoir. Limnol Oceanogr 47:62–77
- Glöckner FO, Amann R, Alfreider A, Pernthaler J, Psenner R, Trebesius K, Schleifer KH (1996) An *in situ* hybridization protocol for detection and identification of planktonic bacteria. Syst Appl Microbiol 19:403–406
- Glöckner FO, Fuchs BM, Amann R (1999) Bacterioplankton compositions of lakes and oceans: a first comparison based on fluorescence *in situ* hybridization. Appl Environ Microbiol 65:3721–3726
- Glöckner FO, Zaichikov E, Belkova N, Denissova L, Pernthaler J, Pernthaler A, Amann R (2000) Comparative 16S rRNA analysis of lake bacterioplankton reveals globally distributed phylogenetic clusters including an abundant group of *Actinobacteria*. Appl Environ Microbiol 66: 5053–5065
- Güde H (1989) The role of grazing on bacteria in plankton succession. In: Sommer U (ed) Plankton ecology: succession in plankton communities. Springer Verlag, Berlin, p 337–364
- Guillard RRL, Lorenzen CJ (1972) Yellow-green algae with chlorophyllide c. J Phycol 8:10–14
- Hahn M, Höfle MG (1999) Flagellate predation on a bacterial model community: interplay of size-selective grazing, specific bacterial cell size, and bacterial community composition. Appl Environ Microbiol 65:4863–4872
- Hahn M, Höfle MG (2001) Grazing of protozoa and its effect on populations of aquatic bacteria. FEMS Microbiol Ecol 35:113–121
- Hahn MW, Lunsdorf H, Janke L (2004) Exopolymer production and microcolony formation by planktonic freshwater bacteria: defence against protistan grazing. Aquat Microb Ecol 35:297–308
- Jürgens K, Matz C (2002) Predation as a shaping force for the phenotypic and genotypic composition of planktonic bacteria. Anton Leeuw Int J Gen Mol Microbiol 81:413–434
- Jürgens K, Pernthaler J, Schalla S, Amann R (1999) Morphological and compositional changes in a planktonic bacterial community in response to enhanced protozoan grazing. Appl Environ Microbiol 65:1241–1250
- Kalmbach S, Manz W, Wecke J, Szewzyk U (1999) *Aquabacterium* gen. nov., with description of *Aquabacterium citratiphilum* sp. nov., *Aquabacterium parvum* sp. nov. and *Aquabacterium commune* sp. nov., three *in situ* dominant bacterial species from the Berlin drinking water system. Int J Syst Bacteriol 49:769–777
- Klammer S, Posch T, Sonntag B, Griebler C, Mindl B, Psenner R (2002) Dynamics of bacterial abundance, biomass, activity, and community composition in the oligotrophic Traunsee and the Traun river (Austria). Water Air Soil Poll Focus 2:137–163
- Langenheder S, Jürgens K (2001) Regulation of bacterial biomass and community structure by metazoan and protozoan predation. Limnol Oceanogr 46:121–134
- Loferer-Kröbbacher M, Klima J, Psenner R (1998) Determination of bacterial cell dry mass by transmission electron microscopy and densiometric image analysis. Appl Environ Microbiol 64:668–694
- Mašín M, Jezbera J, Nedoma J, Straškrabová V, Hejzlar J, Šimek K (2003) Changes in bacterial community composition and microbial activities along the longitudinal axis of two canyon-shaped reservoirs with different nutrient loading. Hydrobiologia 504:99–113
- Massana R, Gasol JM, Bjørseth PK, Blackburn N and 5 others (1997) Measurement of bacterial size via image analysis of epifluorescence preparations: description of an inexpensive system and solutions to some of the most common problems. Sci Mar 61:397–407
- Menden-Deuer S, Lessard EJ (2000) Carbon to volume relationships for dinoflagellates, diatoms, and other protist plankton. Limnol Oceanogr 45:569–579
- Methé BA, Zehr JP (1999) Diversity of bacterial communities in Adirondack lakes: do species assemblages reflect lake water chemistry? Hydrobiologia 401:77–96
- Mindl B, Sonntag B, Pernthaler J, Vrba J, Psenner R, Posch T (2005) Effects of phosphorus loading on interactions of algae and bacteria: reinvestigation of the 'phytoplankton–bacteria paradox' in a continuous cultivation system. Aquat Microb Ecol 38:203–213
- Neef A (1997) Anwendung der *in situ* Einzelzell-Identifizierung von Bakterien zur Populationsanalyse in komplexen mikrobiellen Biozönosen. PhD thesis, Technical University, Munich
- Pernthaler J, Sattler B, Šimek K, Schwarzenbacher A, Psenner R (1996) Top-down effects on the size-biomass distribution of a freshwater bacterioplankton community. Aquat Microb Ecol 10:255–263
- Pernthaler J, Posch T, Šimek K, Vrba J, Amann R, Psenner R (1997) Contrasting bacterial strategies to coexist with a flagellate predator in an experimental microbial assemblage. Appl Environ Microbiol 63:596–601
- Pernthaler J, Posch T, Šimek K, Vrba J and 5 others (2001) Predator-specific enrichment of *Actinobacteria* from a cosmopolitan freshwater clade in mixed continuous culture. Appl Environ Microbiol 67:2145–2155
- Pernthaler A, Pernthaler J, Amann R (2002) Fluorescence *in*

- situ hybridization and catalyzed reporter deposition for the identification of marine bacteria. *Appl Environ Microbiol* 68:3094–3101
- Poindexter JS, Pujara KP, Staley JT (2000) *In situ* reproductive rate of freshwater *Caulobacter* ssp. *Appl Environ Microbiol* 66:4105–4111
- Posch T, Pernthaler J, Alfreider A, Psenner R (1997) Cell-specific respiratory activity of aquatic bacteria studied with the tetrazolium reduction method, cyto-clear slides, and image analysis. *Appl Environ Microbiol* 63:867–873
- Posch T, Šimek K, Vrba J, Pernthaler J, Nedoma J, Sattler B, Sonntag B, Psenner R (1999) Predator-induced changes of bacterial size-structure and productivity studied on an experimental microbial community. *Aquat Microb Ecol* 18: 235–246
- Posch T, Jezbera J, Vrba J, Šimek K, Pernthaler J, Andreatta S, Sonntag B (2001) Size selective feeding in *Cyclidium glaucoma* (Ciliophora, Scuticociliatida) and its effects on bacterial community structure: a study from a continuous cultivation system. *Microb Ecol* 42:217–227
- Rothhaupt KO (1996a) Utilization of substitutable carbon and phosphorus sources by the mixotrophic chrysophyte *Ochromonas* sp. *Ecology* 77:706–715
- Rothhaupt KO (1996b) Laboratory experiments with a mixotrophic chrysophyte and obligately phagotrophic and phototrophic competitors. *Ecology* 77:716–724
- Rothhaupt KO (1997) Nutrient turnover by freshwater bacterivorous flagellates: differences between a heterotrophic and a mixotrophic chrysophyte. *Aquat Microb Ecol* 12:65–70
- Sanders RW, Porter KG, Bennett SJ, DeBiase AE (1989) Seasonal patterns of bacterivory by flagellates, ciliates, rotifers, and cladocerans in a freshwater planktonic community. *Limnol Oceanogr* 34:673–687
- Sekar R, Pernthaler A, Pernthaler J, Warnecke F, Posch T, Amann R (2003) An improved protocol for quantification of freshwater *Actinobacteria* by fluorescence *in situ* hybridization. *Appl Environ Microbiol* 69:2928–2935
- Sherr EB, Sherr BF (1993) Preservation and storage of samples for enumeration of heterotrophic protist. In: Sherr EB, Sherr BF, Kemp P (eds) *Handbook of methods in aquatic microbial ecology*. Lewis Publishers, Boca Raton, FL, p 207–212
- Šimek K, Vrba J, Pernthaler J, Posch T, Hartman P, Nedoma J, Psenner R (1997) Morphological and compositional shifts in an experimental bacterial community influenced by protists with contrasting feeding modes. *Appl Environ Microbiol* 63:587–595
- Šimek K, Kojacká P, Nedoma J, Hartman P, Vrba J, Dolan JR (1999) Shifts in bacterial community composition associated with different microzooplankton size fractions in a eutrophic reservoir. *Limnol Oceanogr* 44:1634–1644
- Šimek K, Armengol J, Comerma M, Garcia JC, Kojacká P, Nedoma J, Hejzlar J (2001a) Changes in the epilimnetic bacterial community composition, production, and protist-induced mortality along the longitudinal axis of a highly eutrophic reservoir. *Microb Ecol* 42:359–371
- Šimek K, Pernthaler J, Weinbauer MG, Horňák K, Dolan JR, Nedoma J, Mašín M, Amann R (2001b) Changes in bacterial community composition and dynamics and viral mortality rates associated with enhanced flagellate grazing in a mesoeutrophic reservoir. *Appl Environ Microbiol* 67: 2723–2733
- Šimek K, Horňák K, Mašín M, Christaki U, Nedoma J, Weinbauer MG, Dolan JR (2003) Comparing the effects of resource enrichment and grazing on a bacterioplankton community of a meso-eutrophic reservoir. *Aquat Microb Ecol* 31:123–135
- Sommaruga R, Psenner R (1995) Trophic interactions within the microbial food web in Piburger See (Austria). *Arch Hydrobiol* 132:257–278

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