

A case of mass occurrence of *Prorocentrum minimum* in the Kiel Fjord

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ABSTRACT: A case of mass development of the potentially toxic dinoflagellate *Prorocentrum minimum* was recorded in the inner reaches of the Kiel Fjord in July/August 1983. Although this species has been sporadically recorded at a number of locations in Norwegian, Danish and Swedish coastal waters, this is the first record of a massive bloom in this part of the Baltic Sea. The bloom seems to have been favoured by relatively high temperatures, high insolation and calm weather conditions prevailing at the time. The dinoflagellate did not prove toxic either to mussel beds in adjoining areas of the Kiel Bight, affected by the bloom, or to other test organisms in the laboratory.

INTRODUCTION

During the months of July and August 1983 a conspicuous bloom was observed in the harbor area of the Kiel Fjord. Microscopic examination of the water revealed the presence of large numbers of a *Prorocentrum* species later identified, according to Dodge (1975, 1982), as *P. minimum* (Pavillard) Schiller.

The dinoflagellate has a heart-shaped form, is 18 to 20 µm long and 15 to 16 µm wide, with a slight depression at its wider end. The surface of the plates is covered with minute spines (Fig. 1a, b, c). It moves about swiftly in the water with an oscillating movement and alternate turning about its long axis.

In view of the close affinity of this species to *Prorocentrum triangulatum* and *P. mariae-lebouriae*, considered as morphotypes of the same species (Dodge 1975, Tangen 1980), and its characteristic shape, the taxonomic position of this bloom-forming dinoflagellate was re-examined and established for reference purposes as *P. minimum* (Pavillard) Schiller var. *triangulatum* (Hulbert ex Adachi) (Elbrächter, pers. comm.). However, for practical purposes the name *P. minimum* will be used throughout this paper.

There are to date no records of mass occurrence of *Prorocentrum minimum* in the Baltic Sea. Pankow (1976) refers only to *P. micans* and 3 species of *Exuviella* including *E. baltica* (all of which are now

included in the genus *Prorocentrum*) as the Baltic representatives of the genus. However, *P. minimum* has now been reported, albeit in relatively low concentrations (about 30 cells ml⁻¹), near Bornholm, Denmark (Baltic Marine Environment Protection Commission 1982).

The same species was, however, described as causing brown water due to its mass development together with other dinoflagellates in Oslo Fjord during August to September 1979 (Tangen 1980) and as potentially toxic to humans through shellfish poisoning (Okaichi & Imatomi 1979, Tangen 1983, Langeland et al. 1984). It was also reported that this organism appeared for the first time (at about the same time of the year) on the west coast of Sweden (Granéli & Granéli 1982) and also in the Kattegat (Edler et al. 1982, Pingree et al. 1982, Nielsen & Ærtebjerg-Nielsen 1983, Pedersen 1983, Granéli et al. 1984).

In view of the unusual occurrence of this bloom in the Kiel Fjord, and its potential danger due to the accumulation of algal toxins by edible shellfish, it was decided to monitor the occurrence of *Prorocentrum minimum* on a day-to-day basis during the weeks following its initial detection.

METHODS

Unconcentrated water samples preserved in Lugol's iodine were used for determining the numerical abundance of *Prorocentrum minimum* in the surface

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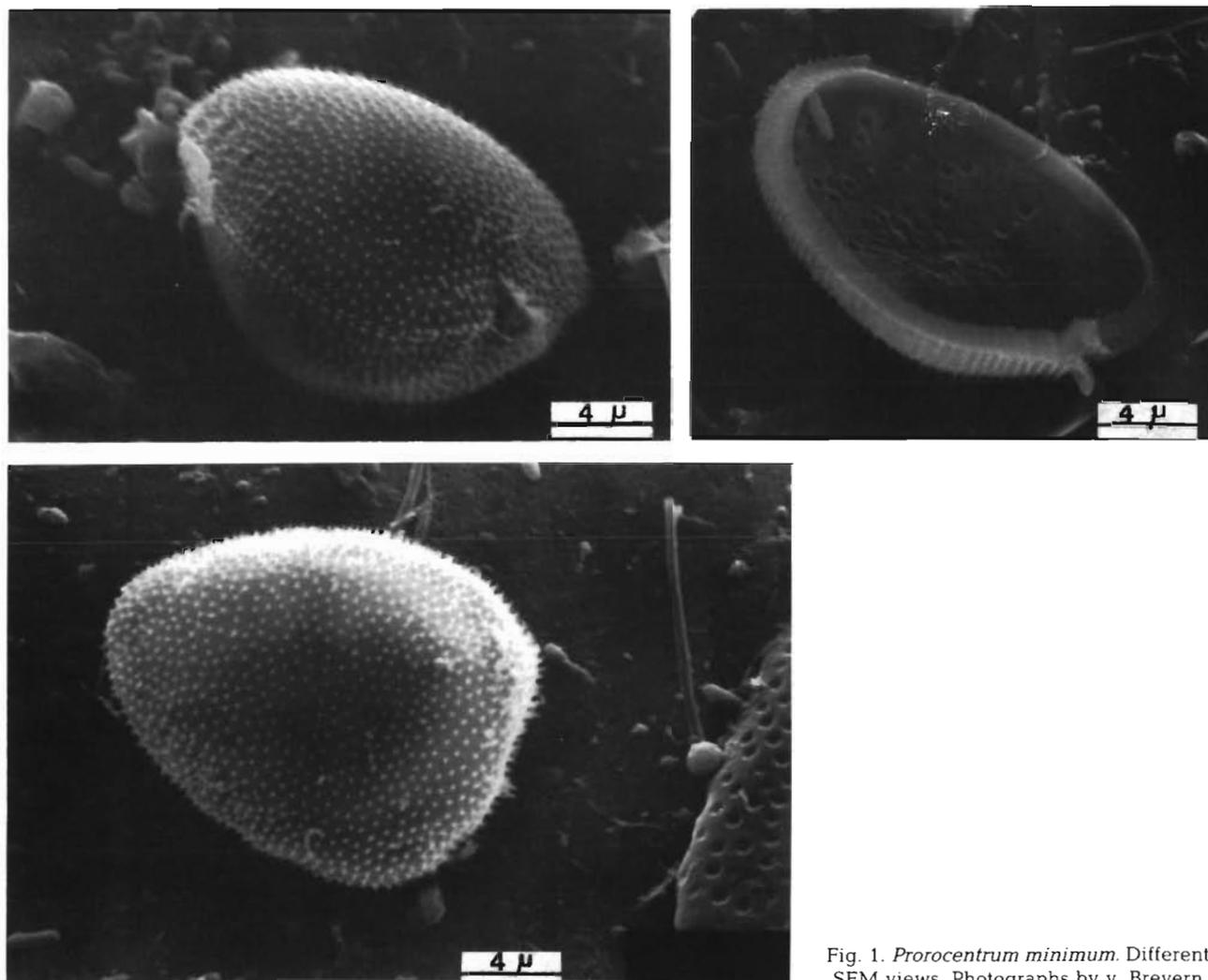


Fig. 1. *Prorocentrum minimum*. Different SEM views. Photographs by v. Brevern

layer by the Utermöhl technique (1958). On one occasion, towards the end of the monitoring period, the vertical distribution of *P. minimum* near the pier of the Institut für Meereskunde, Kiel (Stn 2 in Fig. 2) was determined in a similar fashion to a depth of 5 m. In addition, plankton samples were collected here with a small 40 μm mesh net from the water surface for determination of the most common accompanying larger species.

Sampling operations and observations on some environmental parameters were carried out between noon and 1500 h. Salinity and temperature were measured by a Switchgear instrument; wind velocity and direction as well as light intensity data were obtained from the Marine Meteorology Department of the Institut für Meereskunde, Kiel. On 2 longitudinal transects of the Kiel Fjord (Fig. 4a, b) carried out on 25 July and 16 August, counts were made of *Prorocentrum minimum*, and essential nutrient levels (ammonia, nit-

rite, nitrate, phosphate) and total phosphorus were determined according to Grasshoff (1976).

During the above transects, 2 samples were taken at each of the 8 stations: one as close as possible to the sea floor (approximately 0.7 to 1.2 m above the bottom) and the second combined from 2 to 3 levels in the mixed layer close to the surface. Chlorophyll *a* determination followed the UNESCO (1966) recommendations with modification by Derenbach (1969).

Toxicity tests of *Prorocentrum minimum* were carried out on crabs and mice with methanol extracts as recommended by Okaichi & Imatomi (1979).

RESULTS AND DISCUSSION

Prorocentrum minimum appeared for the first time between 11 July (no counts of the organism) and 27 July, 1983 (first quantitative estimation). The weather

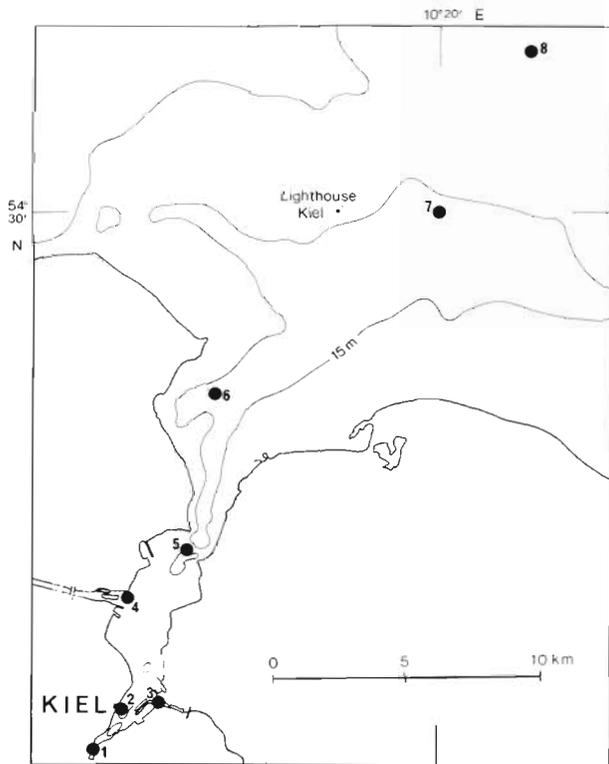


Fig. 2. Location of the sampling stations on 25 July and 16 August 1983. Stn 2 and 3 were at the Institute Pier and in the Schwentine River Estuary, respectively

conditions during these initial 17 d were very calm, and insolation reached its maximum intensity for the whole investigation period. During the same period, 6 days exhibited winds from 000° to 090° , considered as being onshore with regard to the Kiel Fjord. On these days mean wind speed was as low as 3.7 m s^{-1} and insolation reached a mean value of $2.5 \pm 0.8 \times 10^4 \text{ KJ m}^{-2}$. On the other hand, only 3 days exhibited offshore winds (180° to 270°) with an average speed of 4.6 m s^{-1} and an insolation of $1.9 \pm 0.6 \times 10^4 \text{ KJ m}^{-2}$. Further information is given in Table 1. Due to these conditions, water exchange between Kiel Fjord and Kiel Bight was largely inhibited, while phytoplankton growth was enhanced by the high light energy supply.

The results of the surface cell counts during the time of monitoring showed 2 distinct peaks, one during the last days of July and the second toward the end of August (Fig 3, Table 2). An indication of the vertical distribution of *Prorocentrum minimum* in the Kiel Fjord during the bloom period is given in Table 3. It can be seen that the bloom occurred in the upper 2 m followed by a sharp drop further down. Similar results have been obtained in regard to *P. micans* in a sea area west of the Island of Sylt (Wandschneider 1979). Maximum cell concentration, recorded on 20 August 1983, was $31.2 \times 10^6 \text{ l}^{-1}$. An average carbon content of 290

pg cell^{-1} for *P. minimum* can be estimated using the recommendations of the BMB (Edler 1979).

Salinity and temperature determinations (Table 1, Fig. 3 & 4a, b) were found to be in the same range as those of Tangen's records from the inner and outer reaches of Oslo Fjord for the same time of the year (Tangen 1980). Prevailing winds, mainly in a northerly direction, never exceeded 7.5 m s^{-1} .

The longitudinal transects of the Kiel Fjord showed high cell abundance of *Prorocentrum minimum* only in the inner part ($16.591 \times 10^6 \text{ l}^{-1}$ on 25 July, 1983 at Stn 1; Table 2), where higher nutrient levels were observed (Fig. 4a). On 11 August the oxygen-saturation at the surface was 113% and the pH was 9.1 (Nellen, pers. comm.), indicating the high production of the bloom. Chlorophyll data from the sampling site (Stn 2) and the contribution of cells $<20 \mu\text{m}$ recorded on 25 July, 1983 during the initial stages of the bloom showed unusually high levels: $42.2 \mu\text{g l}^{-1}$ total chlorophyll *a* content of which $31.3 \mu\text{g l}^{-1}$ was contained in the organisms $<20 \mu\text{m}$. This also appeared at Stn 1 and 3 to 5, where the same proportion of chlorophyll *a* was seen in the $<20 \mu\text{m}$ fraction during the same transect.

According to Smetacek & Hendrikson (1979), a phytoplankton carbon:chl *a* ratio of $100 \pm 20:1$ could be assumed for dinoflagellate summer populations. Although there are some uncertainties in this assessment, the chl *a* concentration of $42.2 \mu\text{g l}^{-1}$ on 25 July indicates a density of some $15 \times 10^6 \text{ cells l}^{-1}$. Obviously not all of them were *Prorocentrum minimum*. Thus it can be assumed that the observation period missed the real starting point of the bloom by at least 2 or 3 d. This is supported by the fact that the chl *a* concentration on 25 July at Stn 1 (Hörn) was more than $97 \mu\text{g l}^{-1}$, the highest value recorded in the Kiel Fjord area.

The phytoplankton composition throughout the bloom remained fairly constant with a number of accompanying species consistently present. Among these, *Prorocentrum micans*, *Ceratium tripos*, *Cerataulina pelagica* and *Dinophysis acuminata* were always present. On 26 August *Chaetoceros socialis* was particularly conspicuous. In most of the plankton samples examined from Stn 2 (Pier of Institut für Meereskunde), a number of freshwater plankton organisms were recorded. These included a summer form of *Ceratium hirundinella* with a third antapical horn and numerous rotifers, which evidently reached the Fjord waters via runoff from the Schwentine River (Stn 3 in Fig. 2).

The massive red tide caused by *Prorocentrum minimum* in the Kiel Fjord in late summer 1983 constitutes a new record for this part of the Baltic Sea. Since its first occurrence as a cause of brown water in Oslo

Table 1. Density of *Prorocentrum minimum*, salinity and temperature (at 1200 h), wind speed and direction (mean of light period) and light intensity (integrated for the light period) at Stn 2 (pier of the Institute)

Date	Cell density 10 ⁶ l ⁻¹	S ‰	T °C	Onshore m s ⁻¹	Wind Parallel m s ⁻¹	Offshore m s ⁻¹	Light 10 ⁴ KJ m ⁻²
23 Jul	–	–	20.6	2.7			2.62
24 Jul	–	–	21.5	4.4			2.53
25 Jul	–	–	22.0	5.6			2.51
26 Jul	–	–	22.3		2.7		2.19
27 Jul	2.5	–	19.1		4.3		1.93
28 Jul	16.2	14.5	23.4		4.6		1.66
29 Jul	3.6	14.5	22.8			5.6	0.68
30 Jul	18.3	14.8	20.8		6.3		2.07
31 Jul	5.4	14.9	21.1	2.3			2.34
1 Aug	2.8	14.8	22.6			8.1	1.37
2 Aug	1.1	15.7	20.4			6.1	1.80
3 Aug	3.7	14.8	19.7	4.9			1.96
4 Aug	1.5	14.7	19.0		3.0		0.49
5 Aug	1.2	14.5	19.5	5.9			1.80
6 Aug	3.6	14.0	19.9	3.4			1.96
7 Aug	3.3	14.3	20.6	3.2			2.09
8 Aug	3.2	14.1	21.4	3.3			2.07
9 Aug	0.2	14.2	21.5	3.7			2.28
10 Aug	2.1	14.2	21.4	3.2			2.33
11 Aug	3.6	14.4	21.2		3.3		2.32
12 Aug	2.2	14.4	21.0			5.2	0.90
13 Aug	–	–	–	5.8			2.07
14 Aug	–	–	–		3.9		1.57
15 Aug	0.8	15.1	19.8			4.3	2.23
16 Aug	0.4	15.3	18.0			5.8	1.80
17 Aug	0.9	15.1	19.6		2.2		1.15
18 Aug	0.4	15.1	19.0	4.6			1.17
19 Aug	2.3	15.1	19.8	4.8			1.60
20 Aug	31.2	14.5	20.4		4.0		1.69
21 Aug	4.0	15.1	20.4	5.5			1.85
22 Aug	0.4	15.1	21.5	2.2			1.34
23 Aug	2.8	15.5	20.5			2.6	1.68
24 Aug	3.4	15.1	21.9	3.3			1.59
25 Aug	9.2	15.0	21.4	3.2			1.15
26 Aug	3.2	15.3	20.4	3.0			1.97
				n = 19	9	7	
				\bar{x} = 3.95	3.81	5.49	
				SD = 1.20	1.22	1.69	

Fjord in 1979 (Tangen 1980), *P. minimum* has expanded its area of distribution in subsequent years through the Skagerrak and Kattegat into Swedish and Danish coastal waters, as reported above by several investigators.

A combination of physical factors appears to have been responsible for the development of the bloom at that particular time. The relatively high temperatures of the surface water in the Fjord (20 to 22 °C), coupled with high light intensity and moderate winds for most of the duration of the bloom, may all or in part have been contributing factors to the development of the bloom. These factors have been suggested as the main causes of red tides in various marine environments (Lee 1980). The last factor (wind) as a cell concentrat-

ing mechanism in surface waters of the inner reaches of the Kiel Fjord may have been the most important of the three, due to its speed and direction and the special configuration of this water body.

While the initiation of the bloom seems to be supported by favourable meteorological conditions, its maintenance is due to the permanent availability of nutrients. Both phosphate-P and nitrogen compounds are available in the mixed layer of the Kiel Fjord at considerably higher levels than in Kiel Bight. Nitrate and ammonia supplied nitrogen during the initial phase of the bloom, while ammonia was available towards the end in unusually high amounts, indicating a high remineralization rate. Thus the onset of the bloom could be considered as being balanced between

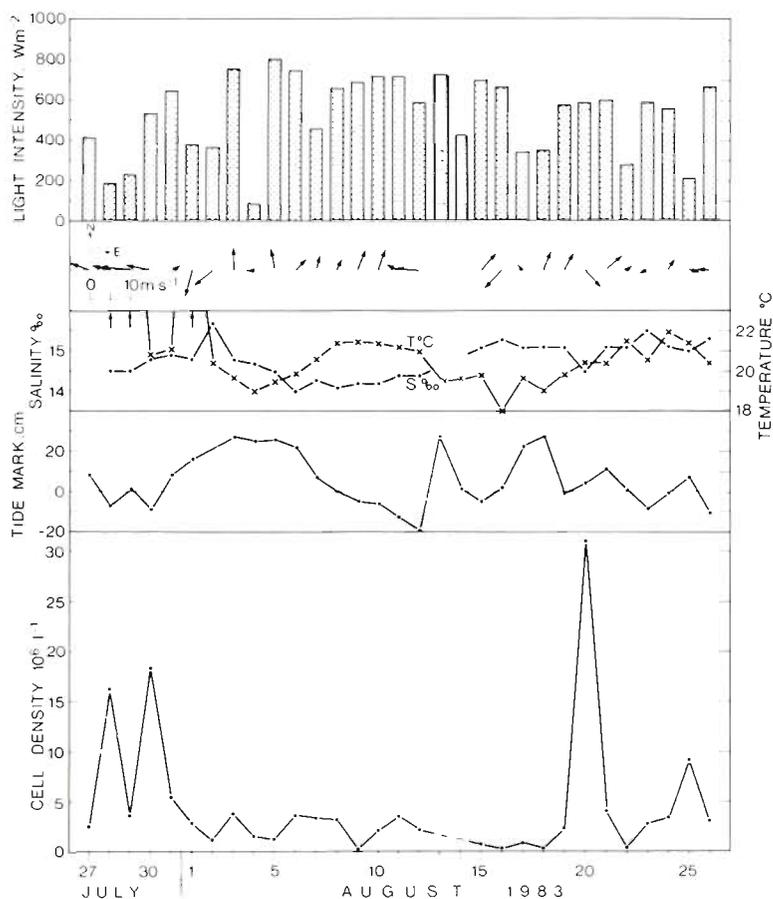


Fig. 3. Hydrological and meteorological data, and cell counts of *Prorocentrum minimum*. Light intensity, wind speed and direction, salinity, temperature (3 arrows indicate data off top of scale, see Table 1), tide mark (zero calibrated to sea level), abundance of *P. minimum* (dotted lines indicate no data available)

Table 2. *Prorocentrum minimum*. Surface density (10^6 cells l^{-1}) at Stn 1 to 8 during the transects

Stations	July 25	August 16
1	16.591	0.450
2	9.609	1.202
3	1.162	2.504
4	5.698	0.069
5	0.566	0.138
6	0.465	0.133
7	0.813	0.013
8	0.026	-

Table 3. *Prorocentrum minimum*. Depth distribution (10^6 cells l^{-1}) at Stn 2 at noon on August 26, 1983

Depth, m	Distribution
0	3.25
1	3.06
2	2.04
3	0.49
4	0.43
5	0.77

new and regenerated production, while regeneration dominated later, although nitrate values increased little in August. Phosphate-P and nitrate-N were introduced into Kiel Fjord from 2 different sources: while phosphate is released in considerable amounts from anoxic sediments, nitrate is delivered by river run-off and originates from agricultural fertilizers. Steidinger (1975) pointed to the support of a bloom by continuous availability of nutrients as the second stage in the life cycle of red tides. Although her statement referred to toxic blooms, it is valid also for the present record.

There is to date little information from other areas regarding the impact of environmental factors directly related to this particular species of dinoflagellate. The general euryhaline and eurythermal characteristics of this species (Birnhak & Farrow 1965, Campbell 1973, Tangen 1980) would both provide an explanation for its eastward expansion into the Baltic Sea. However, pertinent information is available in regard to other species of the genus *Prorocentrum*. Thus, Braarud & Heimdal (1970) recorded the presence of *P. micans* as the predominant species at Drobak, at the entrance to the highly polluted waters of Oslo Fjord, with a concentration of 20×10^6 cells l^{-1} . Fudge (1977) found a positive correlation between the cell density of *P.*

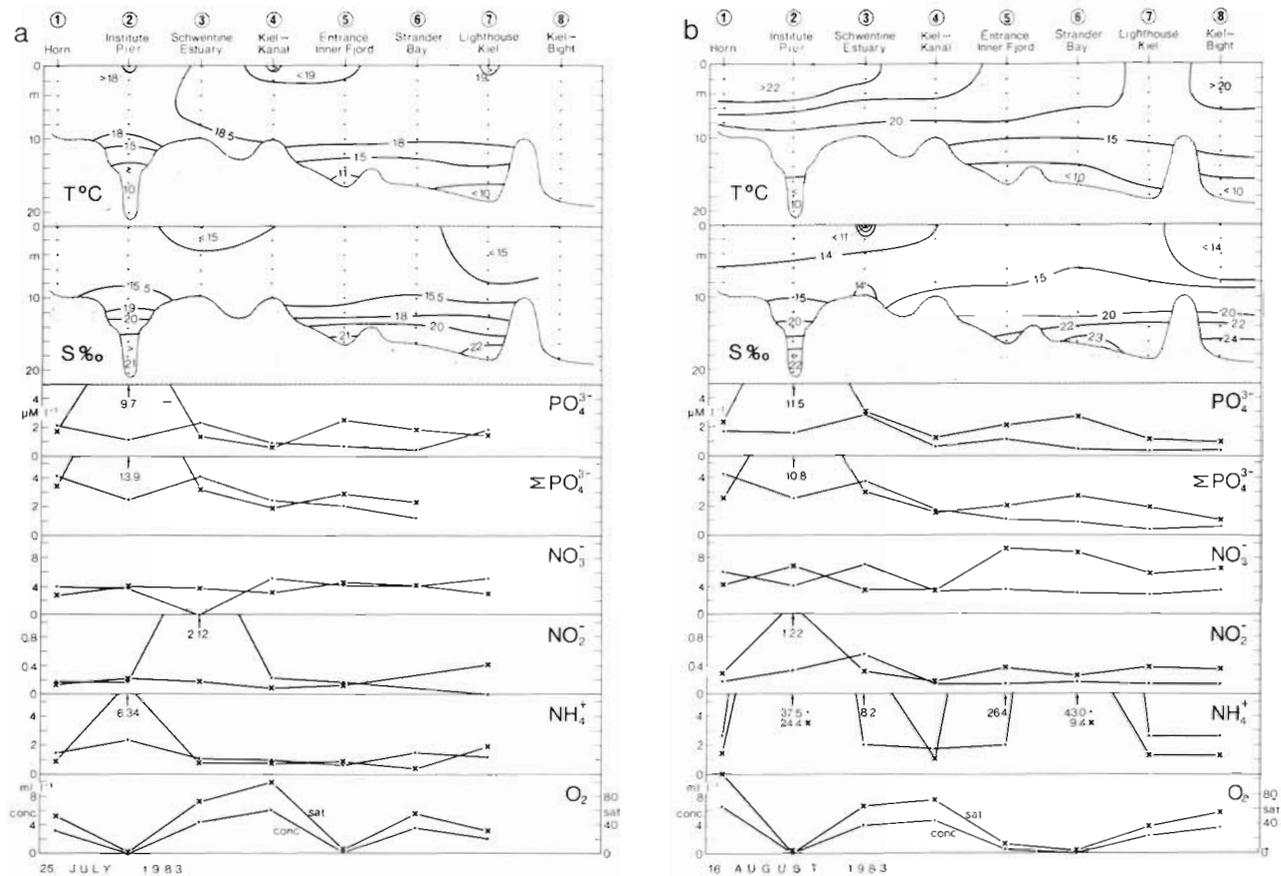


Fig. 4. Environmental parameters in transects on (a) 25 July and (b) 16 August 1983. Temperature and salinity versus depth. Nutrients and total phosphate in $\mu\text{M l}^{-1}$; (x) sampling depth near the bottom, (.) representative sample of the mixed layer. Oxygen samples taken close to the bottom, given as concentration and as % of full saturation

triestinum and the surface temperature in Marsamxetto Harbour, Malta, during May 1970. At that time, the surface temperature during the period of red water ranged from 20.2° to 22.6°C, with the highest values recorded in the surface waters and gradually decreasing with depth. A close relation was also found by the same investigator between cell numbers and insolation.

Regarding the origin of the *Prorocentrum minimum* bloom in the Kiel Fjord, the following suggestions may be made: a southward subsurface process of advection might originate from the Skagerrak, similar to the transport of the closely related morphotype *P. mariae lebouriae* from the southern Chesapeake Bay to the northern bay, where it upwells and forms red tides (Tyler & Seliger 1981). Such a possibility could be envisaged in view of the fact that subsurface chlorophyll maxima (up to 30 $\mu\text{g chl a l}^{-1}$) consisting largely of *P. minimum* have been reported from the Skagerrak area (Pingree et al. 1982) and from the surface waters at the main sampling station in the Kiel Fjord (Stn 2) during the initial stages of the bloom (42.2 $\mu\text{g chl a l}^{-1}$).

A second suggestion that might be considered is that the seed population for *Prorocentrum minimum* red tide in the Kiel Fjord might originate from the resting cysts of this species which became resuspended due to environmental factors, thus creating a motile population capable of active asexual divisions. This seems to be the case of many neritic and estuarine bloom species producing resting cysts (Steidinger & Haddad 1981).

Both these suggestions, which are not necessarily unrelated, would have to be substantiated by long-term monitoring of the waters at discrete depths for analysis of phytoplankton species in the light of available data on the hydrography of the region and the dynamics of phytoplankton bloom patches in the Skagerrak, Kattegat and Bornholm Basin (Astheimer 1983).

There was no evidence of toxicity of the bloom on the basis of laboratory tests carried out on crabs, mice and fish, although not on mussels, which are known to accumulate the toxin venerupin produced by *Prorocentrum minimum*. Due to the fact that the presence of *P. minimum* was also recorded on a qualitative basis in

the adjoining Schlei and Flensburg Fjord waters, where mussel beds are located, the Health Authorities in these areas were alerted to be on the look-out for any ill effects to humans due to this potentially toxic dinoflagellate, through consumption of the mussels. This danger, however, did not materialize.

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