

Consistency of seasonal changes in an estuarine fish assemblage

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ABSTRACT: Data on the abundance of all fish species collected at weekly intervals from the intake screens of the Oldbury-upon-Severn Power Station in the inner Severn Estuary, U. K., between 1972 and 1977, have been analysed using classification and multi-dimensional scaling (MDS) ordination techniques. The results show that in each year the structure of the fish community in the shallows of the estuary underwent similar cyclical changes. These changes were largely attributable to a sequential immigration and emigration of different species, particularly estuarine-dependent marine species, and were not driven directly by variations in water temperature, salinity or freshwater discharge from the river. However, comparisons between the data for years with the driest and wettest winters show that the pattern of change in faunal composition was modified under extreme environmental conditions.

INTRODUCTION

Many marine species of teleosts use estuaries as nursery areas, thereby benefiting from such features as the presence of a greater productivity, a reduced incidence of piscivorous predators and a lower salinity than in their natal environment (Cronin & Mansueti 1971, Beal 1980, Whitfield 1983, Claridge & Potter 1984). Although the juveniles of some of these species also frequently exploit protected inshore marine environments, this important group of marine fish has often been referred to as estuarine-dependent (Clark et al. 1969, McHugh 1976, Beal 1980, Fortier & Leggett 1982, Claridge et al. 1986). Estuaries are also utilized by catadromous and anadromous species (including lampreys) as a migratory route between feeding and spawning areas, and their upper reaches are sometimes colonized by the more salt-tolerant freshwater fish (McHugh 1967, Abou-Seedo & Potter 1979, Day et al. 1981, Dando 1984). In addition, a few teleosts have become adapted to spending the whole of their life cycle in estuarine environments. The abundance of individual species and the composition, abundance and diversity of the total fish fauna undergo seasonal changes in estuaries in many parts of the world (Dahlberg & Odum 1970, McErlean et al. 1973, Haedrich & Haedrich 1974, Gallaway & Strawn 1975, Livingston 1976, Hoff & Ibara 1977, van den Broek 1979, Quinn

1980, MacDonald et al. 1984). While these changes have frequently been related to salinity and/or temperature, Blaber & Blaber (1980) consider that these variables probably do not affect the distribution of the juveniles of estuarine-dependent species.

Detailed information on the biology of the common fish species and the abundance, composition and diversity of the fish assemblage has been obtained for the Severn Estuary over a 5 yr period using samples collected at regular intervals from the intake screens of local power stations (e.g. Claridge & Gardner 1977, 1978, Titmus et al. 1978, Claridge & Potter 1983, 1984, 1985, Claridge et al. 1985, 1986, Potter & Claridge 1985). While these studies showed that consistent seasonal changes took place each year in the numbers of the common species and the fish community as a whole, the changes which these both underwent in the inner Severn Estuary during the winter were more pronounced in the year when salinities fell to an unusually low level (Claridge & Potter 1984, Claridge et al. 1986).

The current study has re-examined the extensive data set for the fish fauna of the inner Severn Estuary in an attempt to determine whether seasonal changes in species abundance, composition and diversity are related to changes in environmental variables, or whether they are primarily the result of 'endogenous' seasonal migrations of larvae, postlarvae and juve-

niles. Attempts to answer this question have utilized the results of classification and ordination of monthly data for the whole 5 yr study period and of weekly data for the years with the driest and wettest winters.

MATERIALS AND METHODS

Fish entrained over 24 h periods on the cooling water intake screens of the Oldbury-upon-Severn Power Station were collected in most weeks between July 1972 and June 1977. This method yielded more species than that collated by Lloyd (1941) from different sources, as well as a large total number of individuals whose standard lengths ranged from 17 mm to well in excess of 1000 mm (Claridge et al. 1986). This indicates that there is no marked bias in this sampling method. The numbers of each fish species in each sample were adjusted when necessary to correspond to a daily intake of 2.2×10^9 l, the volume of water which typically passed through the screen each day during the autumn and winter, when fish abundance was greatest (see also Claridge & Potter 1983, 1984, Claridge et al. 1986). Note that since 2 morphologically very similar gobies *Pomatoschistus minutus* and *Pomatoschistus lozanoi*, were separated only during a 13 mo period, they have subsequently been considered as collectively comprising the *Pomatoschistus minutus* complex (Claridge et al. 1985). Fish species were separated into the following categories: catadromous (C), i.e. migrating from fresh water into the sea to breed; anadromous (A), i.e. migrating from the sea into fresh water to breed; estuarine (E), i.e. typically occurring and breeding in estuaries; freshwater (F), i.e. typically occurring and breeding in fresh water; marine estuarine dependent (MED), i.e. marine species which enter the estuary in large numbers; marine straggler (MS), i.e. marine species abundant in marine environments but only infrequently found in estuaries. The data given in this paper for time of spawning and peak abundance and the standard length at the time of peak abundance were compiled from information provided in our previous papers on the Severn Estuary listed in the references.

Salinity was recorded at the time of sampling, while water temperature represents the mean of the intake values recorded by power station authorities during the week before sampling. The mean freshwater discharge rate of the River Severn in the week before the collection of each sample was obtained from values recorded at Gloucester by the Severn-Trent Water Authority.

Comprehensive details of the classification and multi-dimensional scaling (MDS) ordination techniques employed in this paper and the rationale for their

use are given in Field et al. (1982). The numbers of each species (including rare species) were transformed using a double square root transformation ($\sqrt{\sqrt{X}}$) so that the small number of numerically dominant species do not totally dominate the measure of similarity between 2 times. Mean monthly values for species abundance were used for examining seasonal trends over the whole 5 yr period (Fig. 1 to 3), while comparisons between the July to June periods incorporating the driest (1975/76) and wettest winters (1976/77) used the numbers for the individual weekly samples (Fig. 5 to 6).

A similarity matrix, comparing each sample with every other sample, was constructed for both the 5 and 2 yr abundance data using the Bray-Curtis measure of similarity (Bray & Curtis 1957). This matrix was both classified using group average sorting (Lance & Williams 1967) to produce a dendrogram, and employed for the MDS ordination technique (MDS-CAL 5-MS Program; Kruskal 1977). Salinity and temperature were superimposed on the corresponding monthly and weekly points in the MDS arrays (Fig. 2b, c & 5c, d). The data set for the 5 yr period was also subjected to an inverse analysis (which clusters species rather than samples) to ascertain whether certain species tended to group together (Fig. 3).

It should be noted that since scaling and orientation of MDS plots are purely arbitrary, the axes in Fig. 2 and 5 are unlabelled. What is significant is the relative distance of samples from each other, as this reflects the relative dissimilarities in species composition and abundance. The only information used by the MDS algorithm is the rank-ordered similarity matrix, i.e. the information that Sample 1 is more similar to Sample 2 than it is to Sample 3, for all such comparisons.

Diversity indices were calculated for the once weekly data for the dry (1975/76) and wet years (1976/77) using data transformed to \log_e . Species richness (D) was obtained using the formula of Margalef (1968), while the Shannon-Wiener (H') and Evenness (J) indices were determined according to Pielou (1966).

RESULTS

Data for 1972–1977

The mean monthly temperatures, salinities and freshwater discharge rates for the 1972–1977 period are given in Claridge & Potter (1984). Since there was a high correlation coefficient for the linear relation between the mean monthly values for salinity and freshwater discharge for this period ($r = -0.954$), any trends shown by salinity after superimposition on the MDS arrays will closely reflect the converse situation

for freshwater discharge. However, it is important to recognize that, while high freshwater discharge may cause downstream displacement of some species in the river, its contribution to water movement in the inner estuary at Oldbury is far less than that due to tidal action (Claridge & Potter 1984).

The ranking by abundance of all 78 species and specifically of the 15 most numerous species collected from Oldbury over the 5 yr study period emphasizes the vast contribution made by the estuarine-dependent marine species to the fish community of the inner Severn Estuary (Tables 1 & 2). Thus, if as seems possible, *Pomatoschistus lozanoi* as well as *P. minutus*, which together constitute the *P. minutus* complex, are eventually shown to belong to this category (Claridge et al. 1985), the estuarine-dependent species would have accounted for 91.5 % of the total fish numbers of 146 828 recorded between July 1972 and June 1977. Ten estuarine-dependent marine teleosts (including the *P. minutus* complex) ranked amongst the top 15 species, together with the estuarine common goby, the anadromous twaite shad and river lamprey, the catadromous European eel and the three-spined stickleback, the last of which is regarded by Wheeler (1969) as typically a freshwater species in the southern part of its distribution in the United Kingdom (Table 2).

The representatives of the estuarine-dependent marine species collected from Oldbury consisted very largely of 0+ fish which entered the shallows at vari-

able sizes and times (Tables 1 & 2). The time taken for the juveniles of these species to pass from the spawning grounds in the Bristol Channel to the shallows of the inner estuary was variable, ranging from 14 to 18 wk in the bass to 38 to 42 wk in the sea snail (Claridge et al. 1986). While the twaite shad was also represented by 0+ fish, these were the products of spawning in the upper part of the inner estuary of the lower reaches of the river (Claridge & Gardner 1978). The river lamprey was represented by sexually maturing adults which were migrating from the sea to their spawning grounds in the shallower regions of the river and its tributaries (Abou-Seedo & Potter 1979). Both the yellow and silver forms of the European eel were caught at Oldbury, the latter presumably embarking on their migration to marine breeding areas. The common goby was the only one of the 15 most abundant species believed to be truly estuarine (Claridge et al. 1985). Although a number of freshwater species were obtained from Oldbury, by far the most abundant of these was the three-spined stickleback.

The view that the results from Oldbury are representative of the inner estuary is supported by the observation that the species composition of the samples from this power station are similar to those collected from Berkeley a further 5 km upstream, even though the intake at the latter is located in deeper water (Claridge et al. 1986).

The classification dendrogram for the monthly data

Table 1. A list of the 78 species (including 1 species complex and 1 hybrid) and their life cycle category collected from Oldbury between Jul 1972 and Jun 1977. Total numbers have been corrected to correspond to four 24 h samples each month at an intake of 2.2×10^9 l. Scientific nomenclature follows Wheeler (1978, pers. comm.). A: anadromous; C = catadromous; E: estuarine; F: freshwater; MS: marine straggler; MED: marine estuarine dependent (see 'Materials and Methods' for fuller description of these life cycle categories)

Species	Common name	Number	Life cycle category
1. <i>Pomatoschistus minutus</i> complex	Sand goby	42859	MED
2. <i>Merlangius merlangus</i>	Whiting	41471	MED
3. <i>Platichthys flesus</i>	Flounder	14480	MED
4. <i>Dicentrarchus labrax</i>	Bass	10779	MED
5. <i>Liparis liparis</i>	Sea snail	9900	MED
6. <i>Trisopterus minutus</i>	Poor cod	4228	MED
7. <i>Liza ramada</i>	Thin-lipped grey mullet	3895	MED
8. <i>Alosa fallax</i>	Twaite shad	3879	A
9. <i>Anguilla anguilla</i>	Eel	3687	C
10. <i>Clupea harengus</i>	Herring	2869	MED
11. <i>Sprattus sprattus</i>	Sprat	1801	MED
12. <i>Gasterosteus aculeatus</i>	Three-spined stickleback	1268	F+E?
13. <i>Lampetra fluviatilis</i>	River lamprey	956	A
14. <i>Trisopterus luscus</i>	Bib	911	MED
15. <i>Pomatoschistus microps</i>	Common goby	744	E
16. <i>Ciliata septentrionalis</i>	Northern rockling	381	MED
17. <i>Solea solea</i>	Sole	374	MED
18. <i>Salmo salar</i>	Salmon	267	A
19. <i>Pollachius pollachius</i>	Pollack	257	MED
20. <i>Trisopterus esmarkii</i>	Norway pout	242	MS

Table 1 (continued)

	Species	Common name	Number	Life cycle category
21.	<i>Mullus surmuletus</i>	Red mullet	214	MS
22.	<i>Crystallogobius linearis</i>	Crystal goby	201	MS
23.	<i>Atherina boyeri</i>	Sand smelt	156	E+M
24.	<i>Ciliata mustela</i>	Five-bearded rockling	147	MED
25.	<i>Perca fluviatilis</i>	Perch	132	F
26.	<i>Pleuronectes platessa</i>	Plaice	80	MS
27.	<i>Gadus morhua</i>	Cod	71	MS
28.	<i>Rutilus rutilus</i>	Roach	63	F
29.	<i>Salmo trutta</i>	Trout	47	A
30.	<i>Merluccius merluccius</i>	Hake	44	MS
31.	<i>Conger conger</i>	Conger eel	38	MS
32.	<i>Buena jeffreysii</i>	Jeffreys' goby	28	MS
33.	<i>Limanda limanda</i>	Dab	28	MS
34.	<i>Abramis brama</i>	Bronze bream	25	F
35.	<i>Petromyzon marinus</i>	Sea lamprey	24	A
36.	<i>Ammodytes tobianus</i>	Sand eel	19	MS
37.	<i>Scomber scombrus</i>	Mackerel	19	MS
38.	<i>Gaidropsarus vulgaris</i>	Three-bearded rockling	18	MS
39.	<i>Zeugopterus punctatus</i>	Topknot	16	MS
40.	<i>Gobius paganellus</i>	Rock goby	15	MS
41.	<i>Scophthalmus rhombus</i>	Brill	12	MS
42.	<i>Syngnathus rostellatus</i>	Nillson's pipefish	11	MED
43.	<i>Engraulis encrasicolus</i>	Anchovy	10	MS
44.	<i>Cyprinus carpio</i>	Carp	10	F
45.	<i>Syngnathus acus</i>	Great pipefish	10	MS
46.	<i>Trachurus trachurus</i>	Horse mackerel	10	MS
47.	<i>Crenimugil labrosus</i>	Thick-lipped grey mullet	9	MS
48.	<i>Cyclopterus lumpus</i>	Lumpsucker	9	MS
49.	<i>Callionymus lyra</i>	Dragonet	8	MS
50.	<i>Trigla lucerna</i>	Tub gurnard	8	MS
51.	<i>Abramis bjoerkna</i>	Silver bream	7	F
52.	<i>Raniceps raninus</i>	Tadpole fish	7	MS
53.	<i>Aphia minuta</i>	Transparent goby	7	MS
54.	<i>Sardina pilchardus</i>	Pilchard	6	MS
55.	<i>Rutilus erythrophthalmus</i>	Rudd	6	F
56.	<i>Crenilabrus melops</i>	Corkwing	6	MS
57.	<i>Eutrigla gurnadus</i>	Grey gurnard	6	MS
58.	<i>Lophius piscatorius</i>	Angler fish	6	MS
59.	<i>Carassius carassius</i>	Crucian carp	5	F
60.	<i>Leuciscus leuciscus</i>	Dace	5	F
61.	<i>Micromesistius poutassou</i>	Blue whiting	5	MS
62.	<i>Agonus cataphractus</i>	Armed bullhead	4	MS
63.	<i>Pungitius pungitius</i>	Ten-spined stickleback	4	F
64.	<i>Scophthalmus maximus</i>	Turbot	4	MS
65.	<i>Rutilus rutilus</i> × <i>Abramis brama</i>	Roach and bronze bream	3	F
66.	<i>Labrus bergylta</i>	Ballan wrasse	2	MS
67.	<i>Centrolabrus exoletus</i>	Rock cook	2	MS
68.	<i>Aspitrigla cuculus</i>	Red gurnard	2	MS
69.	<i>Spinachia spinachia</i>	Fifteen-spined stickleback	2	MS
70.	<i>Scyliorhinus caniculus</i>	Lesser spotted dogfish	1	MS
71.	<i>Gobio gobio</i>	Gudgeon	1	F
72.	<i>Carassius auratus</i>	Goldfish	1	F
73.	<i>Leuciscus cephalus</i>	Chub	1	F
74.	<i>Entelurus aequoreus</i>	Snake pipefish	1	MS
75.	<i>Molva molva</i>	Ling	1	MS
76.	<i>Gobius niger</i>	Black goby	1	E
77.	<i>Trigloporus lastoviza</i>	Streaked gurnard	1	MS
78.	<i>Arnoglossus laterna</i>	Scaldfish	1	MS

Table 2. Aspects of the biology of the 15 most abundant species collected from Oldbury between Jul 1972 and Jun 1977. A: anadromous; C: catadromous; E: estuarine; F: freshwater; MS: marine straggler; MED: marine estuarine dependent. Data on the 2 species in the *Pomatoschistus minutus* complex is based only on the period between Jun 1974 and Jul 1975 (see Claridge et al. 1985)

Species	Time of peak spawning	Time of peak abundance	Modal standard length class at peak abundance (mm)	Life cycle category
Sand goby complex				
<i>Pomatoschistus minutus</i>	Mar–May	Jan	50–52	MED
<i>Pomatoschistus lozanoi</i>	Apr–Jun	Jan	38–40	MED?
Whiting	Apr	Sep–mid Nov	85–89	MED
Flounder	Mar & Apr	Lat Jun–Aug	40–44	MED
Bass	May	Sep	45–49	MED
Sea snail	Jan–Mar	Dec–mid Jan	65–69	MED
Poor cod	Mar–May	Sep–mid Oct	65–69	MED
Thin-lipped grey mullet	May & Jun	Late Sep–early Dec	45–49	MED
Twaite shad	May & Jun	Aug & Sep	35–39	A
European eel		Nov	270–350	C
Herring	Mar & Apr	Aug & Sep	60–64	MED
Sprat	Mar & Apr	Late Jul–early Sep	40–44	MED
Three-spined stickleback	Apr–Jun	Dec–Mar	42–46	FW
River lamprey	Mar & Apr	Oct–Jan	250–320	A
Bib	Mar–May	Sep–mid Oct	70–74	MED
Common goby	Jun & Jul	Dec–Mar	25–39	E

for the whole 5 yr period shows that at a similarity level of just over 50 % of the samples separated into 3 main groups (Fig. 1). These groups correspond to the following 3 periods: July to November, November to March and April to July. No pronounced distinction could be seen, however, between the samples for these 'time' groups on the MDS ordination array (Fig. 2a). Since the stress factor for the MDS array was low (0.129), the ordination results are likely to be a reasonably accurate representation of the relation between samples. Thus, the allocation of samples to groups in the cluster analysis should not be treated too definitively. It is the experience of K. R. Clarke (pers. comm.) that group average linking will have a tendency to produce clusters on the dendrogram even when the samples form a genuine continuum.

The samples for the individual months on the MDS arrays for each of the 5 yr showed a gradual but very pronounced clockwise progression from July to November at the bottom through sequential months to November to March on the top left to April to June on the top right (Fig. 2a). While the November to March samples tended to be characterized by low salinities and temperatures, as would be expected for this time of the year, their positions in the array were more closely correlated with the annual cycle than with either of these 2 parameters (cf. Fig. 2a, b, c). Moreover, although salinity and temperature were similar for the July to November and the April to July periods, they were clearly separated on the array (cf. Fig. 2a, b, c).

Although 100 % levels of similarity were present between snake pipefish and gudgeon and between

scaldfish and goldfish in the classification dendrogram (Fig. 3), this reflects the capture of each species on only one occasion and in the same month as the other member of its pair (Table 1). In general, the species caught at Oldbury showed no marked tendency to separate into well-defined groups, although 4 abundant species of gadoid (whiting, bib, pollack, five-bearded rockling) did group together at the 55 % level. The same was true of the freshwater species in the case of roach and bronze bream at the 68 % level and of dace and carp at the 73 % level but none of these last 4 species was ever particularly abundant (Table 1). The fact that 9 of the 11 common teleosts grouped together at the 50 % level reflects the overall dominance of the samples by these species.

Comparisons between 1975/76 and 1976/77

The temperatures between July 1975 and June 1976 showed similar cyclical trends and similar maximum and minimum values to those recorded in the same months in 1976/77 (Fig. 4a). Thus, the highest respective temperatures in these 2 yr were found in August 1975 (21.7 °C) and July 1976 (22.5 °C), with the lowest temperatures being recorded in February 1976 (3.8 °C) and January 1977 (3.9 °C) respectively. While salinity also showed a cyclical pattern in 1975/76 and 1976/77, the minimum values differed markedly (Fig. 4b). Thus, the lowest value of 15.0 ‰ recorded in January 1976 was much greater than the lows of 3.9 ‰ in February and of 2.7 ‰ in early March of the following year. By

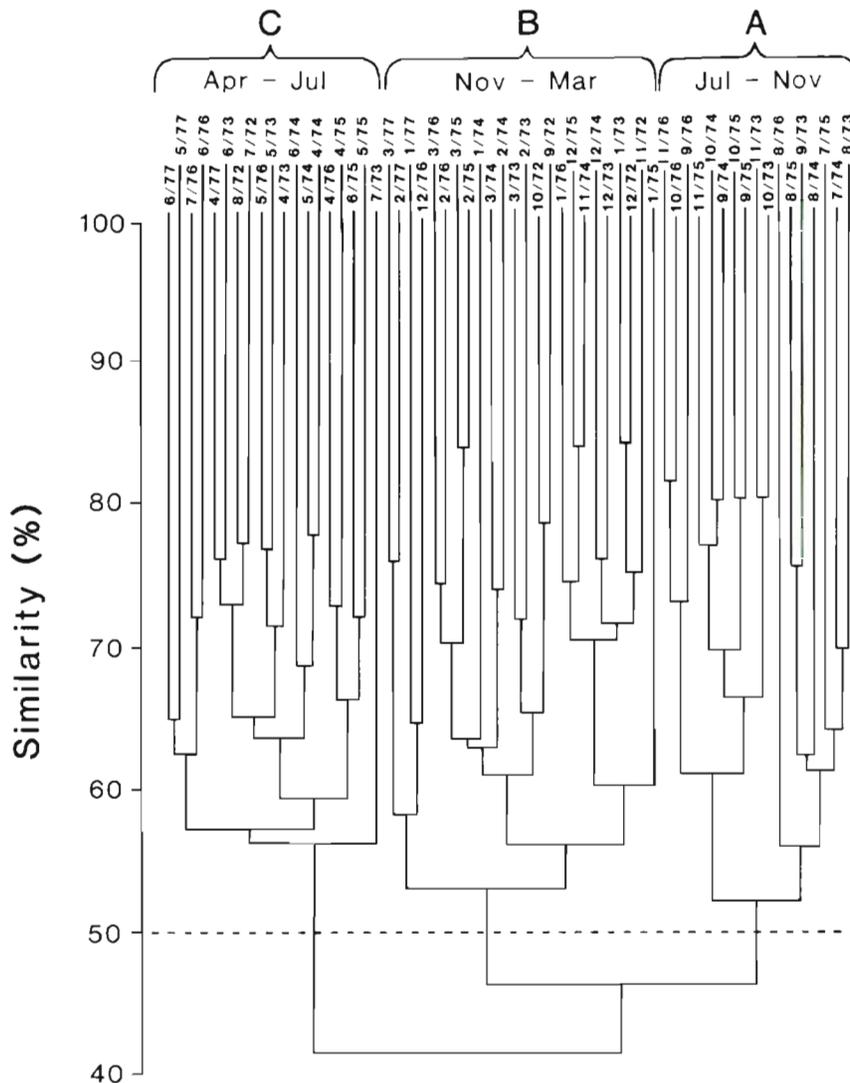


Fig. 1. Dendrogram showing classification of the samples of lampreys, elasmobranchs and teleosts collected from Oldbury, based on pooled data for each month between Jul 1972 and Jun 1977

contrast, the respective maximum values of 30.2 and 29.8% in the 2 yr, which were both recorded in September, were very similar. The trends shown by freshwater discharge bear a close inverse relation with salinity ($r = -0.971$; Fig. 4b, c). For example, in both years, the discharge reached a maximum in the same month that salinity fell to a minimum. Moreover, the greater maximum discharge in 1976/77 than in 1975/76 (519 and 130 $\text{m}^3 \text{s}^{-1}$ respectively) helped to account for the marked difference in the minimum salinities in 1975/76 and 1976/77.

The dendrogram produced by classification of the weekly samples for the 'dry' year of 1975/76 distinguished at the 45% level of similarity 3 groups which correspond to the following periods: July to mid-September; mid-September to April; May and June. By contrast the dendrogram for the 'wet' year of 1976/77 separated 4 groups at the same level, representing August and early September, mid-September to early

December, mid-December to March and April to June. July 1976 grouped with the April to June 1977 period.

Lines drawn through the points for the sequential weekly samples in the MDS array showed similar trends in 1975/76 and 1976/77 (Fig. 5a, b). However the lines joining the samples between December and March were considerably longer in 1976/77 than in the preceding year, indicating a greater change in community composition in this second winter. Superimposition of salinity values on the points for the weekly samples showed that the December to March period in 1976/77 was characterised by exceptionally low salinities and also, but to a lesser extent, low temperatures (Fig. 5c, d; see also Fig. 4b).

The cyclical trends shown by the mean daily catch were similar in 1975/76 and 1976/77, with a marked peak occurring in September and October respectively (Fig. 6). While the values in October and November were greater in 1975/76 than in 1976/77, the converse

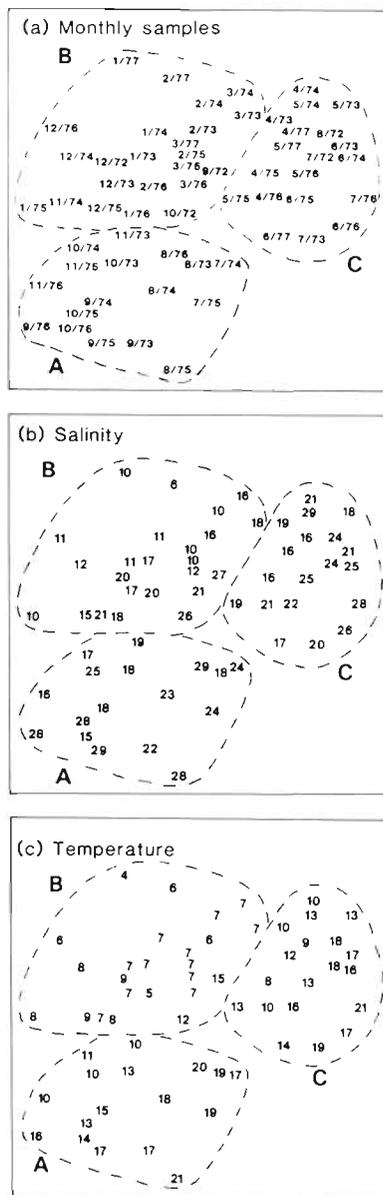


Fig. 2. MDS ordination of each monthly mean of fish from Oldbury between Jul 1972 and Jun 1977 (a), together with superimposition of the corresponding mean salinities (b) and temperatures (c). Dotted line around the 3 main groups separated by classification does not correspond to clearly defined clusters following ordination

was true for the immediately following months. Although the number of species also tended to follow a cyclical pattern, this was much less pronounced than with total abundance (cf. Fig. 6a, b). Species richness (D), the Shannon-Wiener Index (H') and Evenness (J) each showed similar seasonal trends in both years. Differences in these parameters in the December to March periods of the 2 yr only differed in the case of species richness. Thus, values for this index in these

months were generally greater than 3 in 1975/76 but usually below this in 1976/77.

The differences between the years with dry and wet winters were reflected by some conspicuous differences in the species composition of the fauna. Thus, 8 freshwater teleosts (carp, crucian carp, silver and bronze bream, rudd, chub, dace, roach \times bronze bream hybrid) collected in 1976/77 were not recorded in the previous year. Conversely, several rare, essentially marine species (lesser spotted dogfish, ling, fifteen-spined stickleback, scaldfish) were obtained in 1975/76 but not in 1976/77.

DISCUSSION

The most striking feature of the ordination of the data for samples of fish collected from Oldbury between 1972 and 1977 was the consistency of the clockwise progression of the monthly points in the MDS arrays for each of the 5 successive 12 mo periods. This demonstrates that in each year a similar progressive and cyclical change took place in the composition and abundance of the assemblage of fish in the shallows of the inner Severn Estuary. The lack of correspondence between the trends shown by the monthly samples in the MDS arrays and those exhibited by temperature and by salinity and therefore freshwater discharge shows that the seasonal changes in faunal structure were not closely related with concomitant changes in these environmental variables.

The basis for the sequential seasonal change in the abundance and composition of the fish fauna of the inner Severn Estuary is revealed by examining the times of immigration, residency and emigration of the more abundant species at Oldbury shown in Table 1 (see also Claridge et al. 1986). These data demonstrate that, while the common species were generally present throughout the year, the majority tended to appear at Oldbury at varying times and often remained in large numbers within this region for only a few weeks. Since this group of species was not usually again numerous until the following year, this resulted in a continual change in the relative abundance of the most abundant species, with none of these species becoming dominant in more than one period.

The sequential pattern of recruitment of the various species responsible for initiation of the annual change in faunal structure is well illustrated by considering the data on time of immigration of juveniles of the 10 estuarine-dependent marine species most frequently caught during the study, all of which ranked amongst the 15 most abundant species overall (Table 2). Thus, the period when these reached peak abundance ranged from late June to August in the flounder

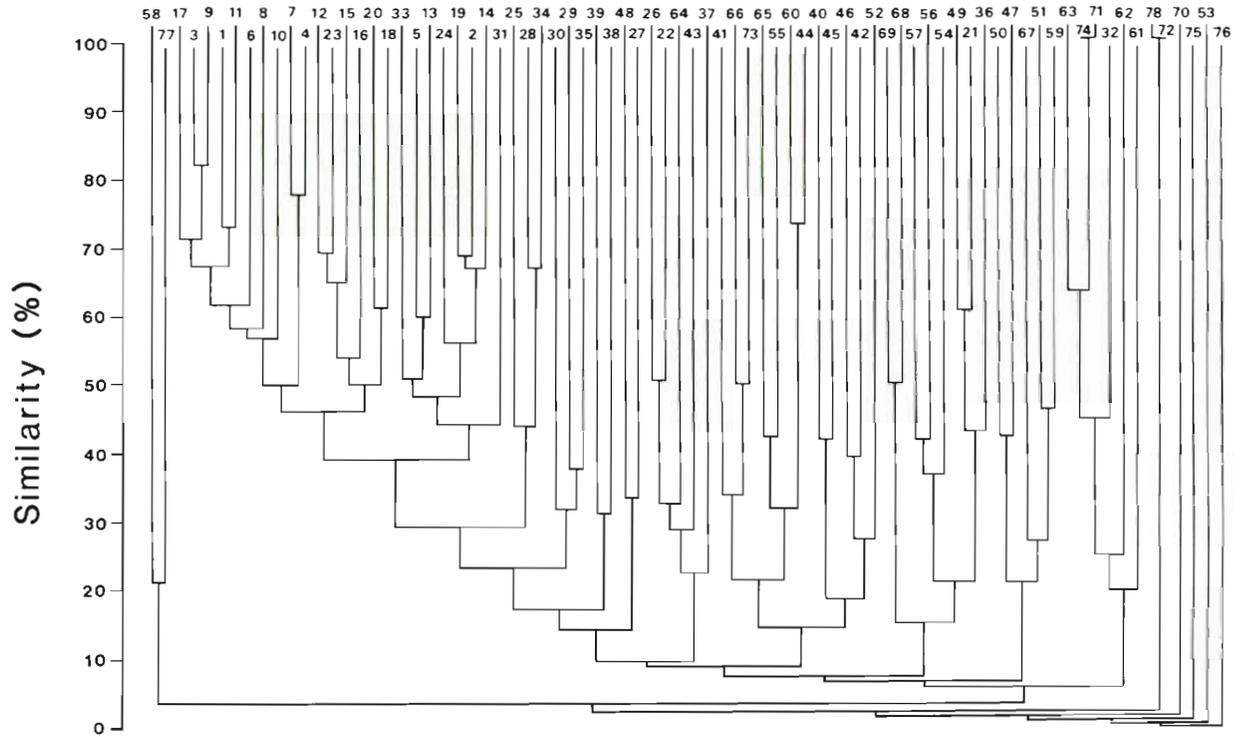


Fig. 3. Dendrogram showing classification of the 78 species of lampreys, elasmobranchs and teleosts collected from Oldbury based on pooled data for each month between Jul 1972 and Jun 1977. Numbers refer to those given for each species in Table 1. List includes a species complex and 1 hybrid

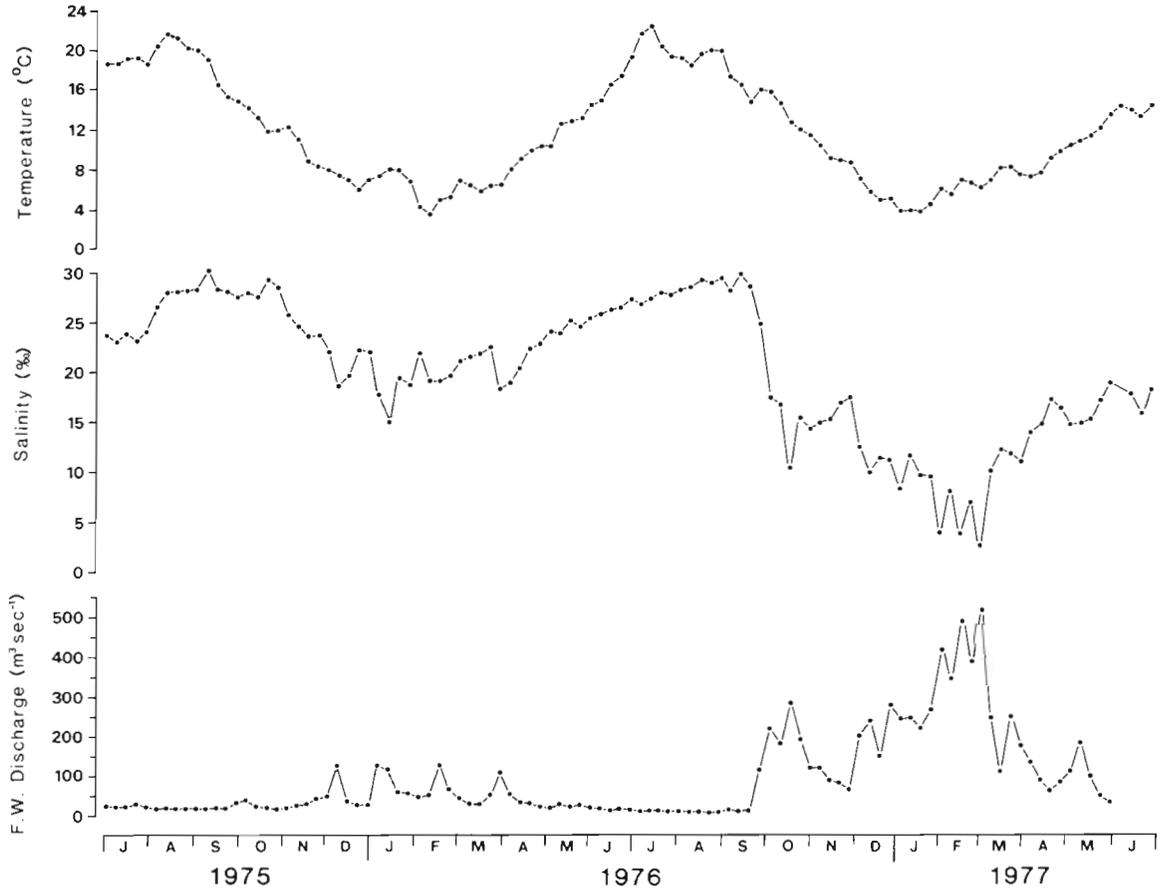


Fig. 4. Weekly values for (a) temperature and (b) salinity at Oldbury and for (c) freshwater discharge at Gloucester during the Jul-Jun period of the years with the driest (1975/76) and wettest winters (1976/77)

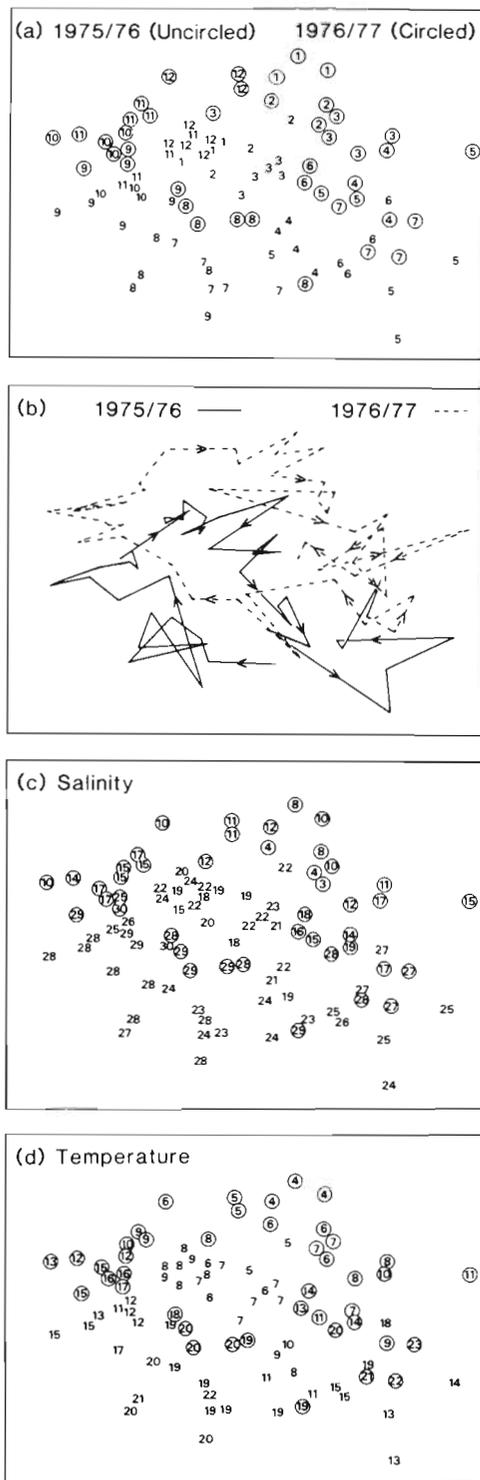


Fig. 5. MDS ordination of weekly samples for the Jul–Jun period in the years with the driest (1975/76) and wettest winters (1976/77) (a and b), together with superimposition of the corresponding mean salinities (c) and temperatures (d). In (a) number = month of each weekly sample. In (a, c and d), points for samples in the wet year are encircled

through September in the bass to December to mid-January in the sea snail. Since the interspecific variability in the period between spawning and arrival at Oldbury is greater than between peak spawning times (Table 2), this variability can be related to differences in the methods by which these species are transported from their spawning grounds to upstream nursery areas in the estuary. Species such as herring, sprat and bass, which appear early and at a relatively small size, probably use passive and selective tidal transport to enter and pass up the estuary (Fortier & Leggett 1982, Norcross & Shaw 1984, Aprahamian & Barr 1985, Dando & Demir 1985, Claridge et al. 1986). By contrast, species such as whiting, which are represented by individuals with a larger modal length at Oldbury, have been described as entering their nursery areas through 'an active migration of juveniles rather than a passive denatant drift of planktonic larvae' (Cooper 1980, 1983).

The increases in abundance of the representatives of estuarine-dependent marine species between late summer and early winter was augmented by a marked elevation in abundance of the catadromous twaite shad in August and September and by a less precipitous rise in the catches of anadromous river lamprey between September and November (Table 2, Claridge et al. 1986). The recruitment of estuarine-dependent marine species, and twaite shad and river lamprey, into the inner estuary largely accounts for the rise in fish abundance between late summer and early winter. The sequential replacement of species during this period of high abundance could be of value in reducing the likelihood of competition for space and even food in the shallows of the inner estuary at this time. A similar conclusion was reached by Weinstein et al. (1980) for a fish community in North Carolina where species also entered the estuary at different times and sizes.

The progressive change in the composition of the fish community between late summer and winter is continued in subsequent weeks through the appearance of common goby and freshwater species such as three-spined stickleback. While the advent of the freshwater species occurs at the time when freshwater discharge in the rivers is greatest and salinity in the inner estuary is lowest, the movements of the more common of the estuarine-dependent species of gadoids in the estuary are not closely related to either of these environmental variables or to temperature. The latter conclusion is based on the observation that the changes in weekly numbers of these species was not significantly correlated with changes in these environmental variables (Claridge & Potter 1984). Such a feature also helps explain the lack of correlation between seasonal changes in faunal structure and environmen-

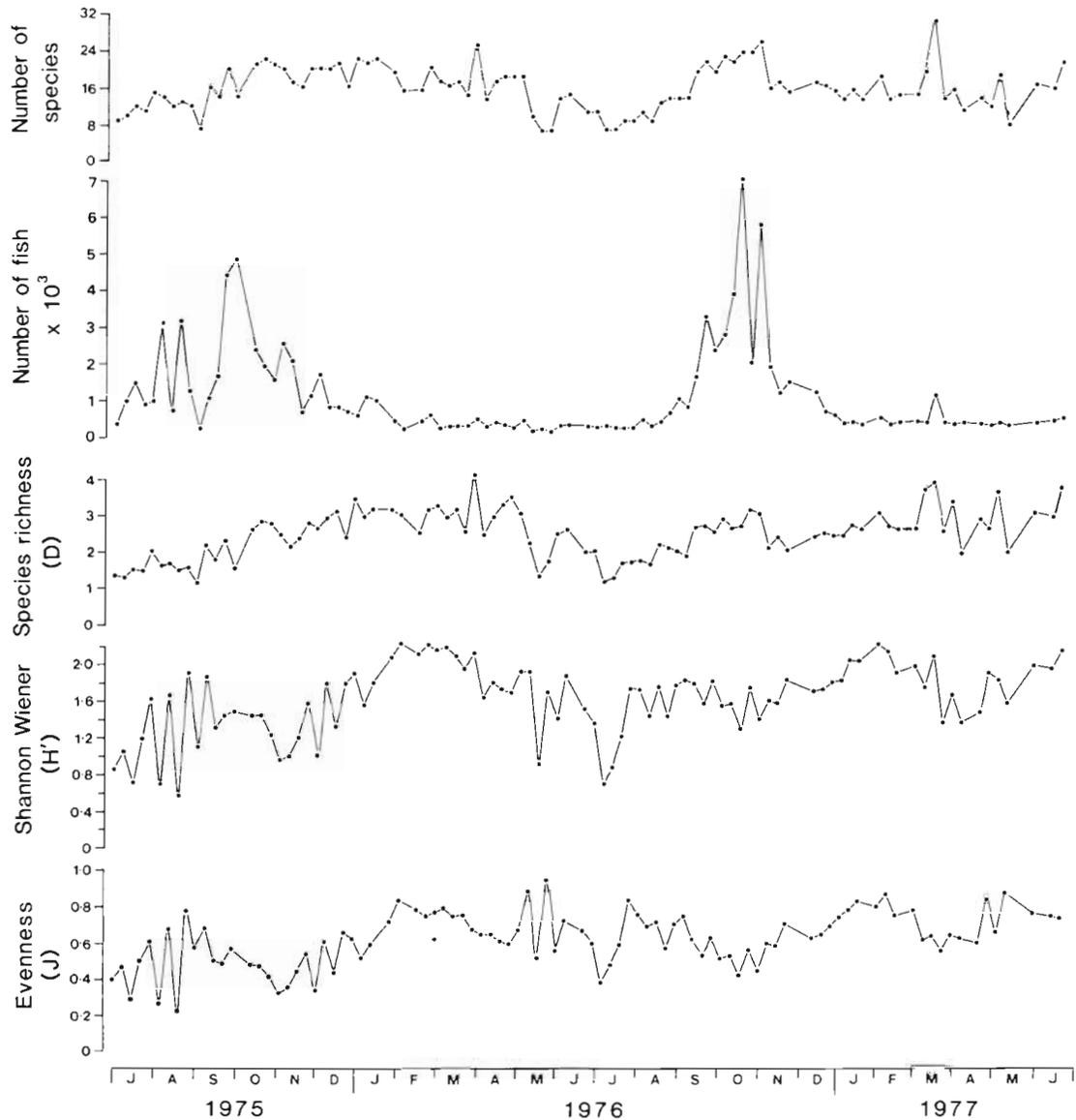


Fig. 6. Species number (S), total abundance (N), species richness (D), Shannon-Wiener index (H') and evenness index (J) for the weekly samples between Jul 1975 and Jun 1977

tal variables. In this context, it is also probably relevant that there was no conspicuous tendency for groups of the more abundant species to cluster together.

Despite the conclusion that to a large degree the annual cyclical change in faunal structure reflects differences in the times of recruitment of the more abundant species, there is also good evidence that these changes can be modified by extremes in environmental variables either in the estuary or in the freshwater regions of tributary rivers. This view is based partly on the observation that, in contrast to the dry winter of 1975/76, classification separated the samples in the wet winter of 1976/77 into 2 groups, with the separation occurring in December when salinity and,

less conspicuously, temperature fell to an exceptionally low level and then remained depressed for several weeks. Support for this opinion is provided by the far greater separation of weekly samples on the MDS arrays at the same time of the year in 1976/77 than in 1975/76, which implies that a greater change took place in the community in the winter of the wet year. It is also relevant that the points for December 1976 and January and February of 1977 in the MDS array for the whole 5 yr period lay outside those for the same months in other years. Since salinity was more closely correlated with changes in faunal structure than was temperature, it would seem to be a better candidate for causing such changes.

The view that salinity can affect the abundance of fish is supported by a comparison between the presence of higher numbers in October and November of 1976/77 than in 1975/76, with the converse situation between December and March when salinities in the winter of 1976/77 had dropped to far lower levels. The effect of low salinities on the fauna is also illustrated by the lower species richness in the very low salinity winter of 1976/77 and by the greater numbers of freshwater species in that year.

In summary, the magnitude and cyclical pattern of change in the fish community structure of the shallows of the inner Severn Estuary was similar in all years. These cyclical patterns are not driven directly by temperature, salinity or freshwater discharge but are a reflection of sequential immigrations of different species, and particularly those of the juveniles of those marine species which are frequently referred to as estuarine-dependent. The composition and abundance of the fish fauna of the inner estuary were modified to some degree, however, by extremes in environmental variables. The observation that salinity has only a moderating influence on the annual pattern of sequential change at Oldbury would be consistent with the view that since this locality is in the inner estuary, it would be expected to be characterized by a relatively euryhaline fauna. Although salinity does not have a pronounced effect on the species composition of the fish fauna at Oldbury in the inner Severn Estuary, studies on other estuaries have shown that differences in salinity regimes throughout an estuary can influence the composition of the fish fauna in the different regions (see e.g. Weinstein et al. 1980, Loneragan et al. in press).

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