Estimating the dispersal potential of polychaete species in the Southern California Bight: implications for designing marine reserves

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ABSTRACT: Using known and inferred life-history information, we estimated the dispersal potential of 501 polychaete species sampled during a 1998 monitoring study in the Southern California Bight. We tested the hypothesis that species having life-history traits that suggest long-distance dispersal will be encountered more frequently throughout the region than will species having life histories that suggest limited dispersal. When all 501 species and all 200 sampling sites were analyzed, occurrence frequency (percentage of sites at which a species was collected) was not significantly related to dispersal potential. When data from 53 shelf sites in the Channel Islands were analyzed separately from collections at 147 mainland-shelf sites, there was a significant positive relationship between dispersal potential and occurrence frequency at the island sites but not at the mainland sites. Of the 501 species, 119 were collected only at island sites, 98 were found only at mainland sites, and 284 were found at both island and mainland sites. The majority of the ‘island only’ species had life-history traits indicating low dispersal potential. In contrast, only 13% of the ‘mainland only’ species were categorized as having low dispersal potential. The ‘cosmopolitan’ species had a broad range of dispersal potential. Models indicate that efforts to conserve biodiversity by establishing Marine Protected Areas (MPAs) must consider species’ dispersal. In the shelf communities of the Southern California Bight, networks of small reserves that are located in existing areas of high diversity should succeed in the Channel Islands, where the majority of polychaete species tend to have limited dispersal potential. On the mainland shelf, however, designing effective MPAs will be more challenging due to the prevalence of species that have a greater potential for long-distance dispersal to or from unprotected sites.

KEY WORDS: Dispersal · Larval ecology · Life history · Marine Protected Area · Marine reserve · Polychaete

INTRODUCTION

The life histories of most marine organisms include a planktonic larva that has the potential to disperse away from the parental habitat. The ability of these larvae to disperse and successfully recruit to suitable habitats affects the demography (Strathmann 1985) and interrelatedness (Hellberg 1996, Bohonak 1999) of local populations. When patches of suitable habitat are discontinuous in time and space, the distance over which an organism is able to disperse will greatly affect its ability to colonize new habitats or recover from disturbance (Gunther 1992, Shull 1997, Whitlatch et al. 1998).

Directly determining the dispersal distances for marine organisms is difficult due to the vastness of the potential dispersal area and the large numbers and high mortality of microscopic larvae (Rumrill 1990, Thurrold et al. 2002). Until dispersal is tested directly in the field using recently developed tagging technology (reviewed by Thurrold et al. 2002), the dispersal of organisms must be estimated using other methods,
such as an examination of species’ life histories. For example, the duration of a planktonic larval period is known to correlate with dispersal distance (Shanks et al. 2003).

Previous studies have applied life-history information to estimate the dispersal potential of organisms in relation to specific habitats. Parker & Tunnicliffe (1994) assessed the dispersal potential of the species encountered on a seamount off the coast of Oregon, attempting to find a common dispersal strategy that resulted in those species’ accumulation in that isolated habitat. Grantham et al. (2003) examined 91 species lists of macroinvertebrates in 7 rocky and sandy habitats along the US Pacific Coast and demonstrated that dispersal profiles differed between the species inhabiting rocky and sandy substrates.

Polychaete worms are among the most abundant organisms in many soft-sediment communities. Polychaetes also show an amazing diversity of life-history strategies, from brooded direct development to swimming larvae that spend weeks in the plankton (Wilson 1991a). Polychaetes often are among the first colonists of disturbed habitats (e.g. Shull 1997) and affect later succession in soft-sediment communities (e.g. Gallagher et al. 1983, Thrush et al. 1996). Polychaetes also are common items in the diets of species that have greater socioeconomic value, such as commercial fishes (e.g. Stehlik & Meise 2000). Although polychaetes and other infauna do not receive much attention from lay audiences, ecologists have long recognized that these diverse worms are central to the functioning of marine ecosystems (e.g. Wilson 1991b, Waldbusser et al. 2004).

In recent years, strict ‘marine reserves’ (Lubchenco et al. 2003) have become a popular tool for managing some fisheries and are often designated to promote the conservation of marine biodiversity. Lubchenco et al. (2003) have applied existing knowledge of polychaete life histories to assess the dispersal potential of 501 polychaete species found in the mainland or island-shelf communities of the Southern California Bight (Ranasinghe et al. 2003). We found that 43% of the polychaete species were collected at island or island-shelf communities of the Southern California Bight and designating which sites within existing MPAs should become more restrictive marine reserves (e.g. Airame et al. 2003, Roberts et al. 2003), a network of MPAs is unlikely to protect the region’s biodiversity unless the size, spacing, and location of various MPAs are based, in part, on knowledge about the dispersal of species in the region.

In this paper, we apply existing knowledge of polychaete life histories to assess the dispersal potential of 501 polychaete species found in the mainland or island-shelf communities of the Southern California Bight (San Miguel, Santa Rosa, Santa Cruz, Anacapa and Santa Barbara Islands. Together with Santa Catalina Island, which lies outside the Sanctuary, the 6 island shelves have a total of 2 139 km2 of benthic habitat (Ranasinghe et al. 2003). These islands straddle an important ecological boundary where currents flowing southeast from the cold-water, northern marine communities meet the warmer currents flowing northwest from southern California (Hewatt 1946). This may contribute to greater species diversity in the island shelves, as eddies from the 2 water masses mix pelagic larvae from both regions. Hewatt (1946) found that about 20% of the species he collected on Santa Cruz Island were at the approximate northern or southern extent of their range. Although many factors must be considered when selecting locations for additional MPAs in the Southern California Bight and designating which sites within existing MPAs should become more restrictive marine reserves (e.g. Airame et al. 2003, Roberts et al. 2003), a network of MPAs is unlikely to protect the region’s biodiversity unless the size, spacing, and location of various MPAs are based, in part, on knowledge about the dispersal of species in the region.

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MATERIALS AND METHODS

The Southern California Bight Regional Monitoring Project sampled 147 soft-sediment benthic sites from the continental shelf off the southern California coast between Point Conception and the US–Mexico border in 1998 (Fig. 1; Ranasinghe et al. 2003). An additional 53 samples were taken from the shelves off the 5 Channel Islands and Santa Catalina Island (Fig. 1). This project also sampled benthic communities in Mexican waters and embayments of the US mainland, but these samples were excluded from our analysis. Ranasinghe et al. (2003) describe the experimental design, sampling, and laboratory procedures in detail. Sites were chosen using a random tessellation stratified design, and organisms collected using a 0.1 m–2 Van Veen grab sieved through a 1 mm mesh. It should be noted that this mesh size probably excluded several species of polychaetes from the analysis. Sample depths ranged from 10 to 120 m depth. Collected organisms were preserved and stored for laboratory identification. Workers keyed the samples to the lowest taxonomic level possible, using the taxonomic scheme of the Southern California Association of Marine Invertebrate Taxonomists.

We divided the resulting list of 501 polychaete species (Ranasinghe et al. 2003) into 3 dispersal potential categories: ‘high’, ‘medium’, and ‘low’ (Table 1) based on these species’ known life-history parameters or inferences based on the life histories of closely related species (Appendix 1; available at: www.int-res.com/articles/suppl/m316p105_app.pdf). The dispersal potential of some species was listed as ‘undetermined’ either because we could find little or no life-history information about the species and closely related taxa, or because closely related species exhibited a wide range of life histories that prevented reasonable extrapolation.

The high-dispersal category includes species that can potentially disperse tens of kilometers and exchange larvae with distant populations. These species can be considered to have ‘open’ marine populations (Roughgarden et al. 1985, Caley et al. 1996). The medium-dispersal category implies dispersal on the scale of kilometers. These demographic populations will normally be contained on 1 island or island group. Large barriers to dispersal, such as extended stretches of unsuitable habitat, may divide the mainland shelf into demographically distinct units in this group as well. The low-dispersal category, most likely less than 1 km, denotes species that have relatively ‘closed’ populations (Roughgarden et al. 1985, Caley et al. 1996). Because the larvae of low-dispersing populations are retained close to their parents, the shelf of a single island or stretch of mainland coast may contain several...
Table 1. Polychaete species in the Southern California Bight. Mean occurrence frequencies (±SD) of all species that were assigned to each dispersal-potential category based on known or inferred life-history traits. Occurrence frequency is the percentage of all 200 sampling sites at which a given species was collected (see Appendix 1; available at: www.int-res.com/articles suppl/m316p105_app.pdf).

<table>
<thead>
<tr>
<th>Dispersal potential</th>
<th>No. of species</th>
<th>Occurrence frequency (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>High</td>
<td>128</td>
<td>11.2 (±16.5)</td>
</tr>
<tr>
<td>Medium</td>
<td>93</td>
<td>9.4 (±13.6)</td>
</tr>
<tr>
<td>Low</td>
<td>123</td>
<td>7.1 (±9.8)</td>
</tr>
<tr>
<td>Undetermined</td>
<td>157</td>
<td>5.6 (±8.0)</td>
</tr>
<tr>
<td>Total</td>
<td>501</td>
<td>8.1 (±12.3)</td>
</tr>
</tbody>
</table>

Several life-history characteristics were used in our analysis: method of fertilization, extent of brooding (either in or on the body or in egg masses), method of dispersal, feeding behavior of the larvae, average time before settlement, mobility of adults, and presence or absence of epitokous reproduction. Not all characteristics were given equal weight. Some life-history characteristics provided very clear indications of dispersal potential. For example, species that have crawling larvae or direct development were always assigned a low dispersal potential. For example, species that have crawling larvae or direct development were always assigned a low dispersal potential (Hellberg 1996). For most taxa, however, a suite of known life-history information was considered.

The most important estimator of dispersal for pelagic developers was time before settlement. If this is known, oceanographic information can be incorporated to estimate dispersal in terms of actual distances (e.g. Gaines et al. 2003). Unfortunately, the duration of the pelagic larval period was known for very few species (22 of 501). For most taxa, however, a suite of known life-history information was considered.

For most taxa in our analysis, time before settlement was inferred from the feeding mode of larvae. Planktotrophic larvae, which feed during their pelagic development, tend to have greater dispersal distances relative to lecithotrophic larvae, which metabolize stored energy reserves (Strathmann 1985, Todd 1998). In Grantham et al. (2003), for example, average time in the plankton was known for 5 polychaete planktrotrophs and 17 polychaete lecithotrophs. The planktrotrophs averaged 57 d in the plankton (range: 13 to 150 d), which allows for dispersal of tens or hundreds of kilometers. Lecithotrophs averaged just 7 d before settlement (range: 1 to 25 d), suggesting more modest dispersal.

Because the dispersal distances of mobile adults are usually several orders of magnitude less than the distances planktonic larvae can travel (reviewed by Gunther 1992), the mobility of adult polychaetes was given less weight when estimating dispersal potential. Epitoky, the budding of a sexual stage to search for a suitable mate, can be another source of dispersal, but also was given less weight because epitokes and adults usually swim for brief periods at night and travel on the order of 1 km or less (e.g. Hobson & Chess 1976, Dean 1978). A final method of dispersal, rafting, is used by some otherwise immobile species to disperse. Eggs, larvae, or adults attached to drifting plants or other material are transported by currents. Rafting potential has been identified for some species in the family Nereididae (Grantham et al. 2003).

In general, species with pelagic, planktotrophic development were assigned to the high-dispersal category. Species with pelagic but lecithotrophic larvae or evidence of a pelagic period of <10 d were assigned to the medium-dispersal category. All other species for which key life-history characteristics were known were considered to have a low dispersal potential.

All relevant life-history characteristics, including method of fertilization, location of larval development, and average time spent in the plankton were known for a very small percentage of species listed in the survey. Our analysis was further complicated because not all polychaetes were keyed to the level of species (Ranasinghe et al. 2003). Often, however, 1 or more of these characteristics was known for a species in the same genera or family as the species in question. Only when at least 80% of known members of a taxon shared the same life-history characteristics were those characteristics assigned to an unknown species in that taxon. Otherwise, a diversity of life histories within the group relegated the unknown species to the undetermined category.

To test the hypothesis that polychaete species having a greater dispersal potential should occur more frequently across the region than do species having a lower dispersal potential, we calculated the occurrence frequency for each species by dividing the number of sites in which a species was collected by the total number of sites sampled (200). We also divided the sites into mainland (147) and island (53) shelves and recalculated the occurrence frequencies separately for each subset of sampling sites. Because the distribution of occurrence frequencies was heavily skewed toward low values, we tested for differences among dispersal categories using a nonparametric Kruskal-Wallis test (Zar 1999).

We also tested the hypothesis that species restricted to either the island or the mainland sites would have different dispersal potentials. We determined which species were collected at either island or mainland sites, but not both. The numbers of high, medium, and low dispersing species for the ‘island-only’ and ‘mainland-only’ groups were compiled in a 2 × 3 contin-
gency table and compared with a chi-square test (Zar 1999). We also used chi-square tests to compare the dispersal potentials of the species that were collected at both island and mainland sites to those of either the 'island only' or the 'mainland only' species.

RESULTS

The data set included 49 polychaete families, 226 genera, and 501 different polychaete species (Appendix 1; Ranasinghe et al. 2003). Most of the species (302 of 501) were identified to described species. Of the remaining 199 species, 110 were denoted as different species within a genus (Appendix 1: ‘Species A, B, C’, etc.) and 89 were keyed to distinct genera that were each represented by what appeared to be a single species (Ranasinghe et al. 2003). A total of 403 of the 501 different species in the data set were collected from sites on the island shelves, and 382 species were collected from mainland-shelf sites. Samples from the 53 island sites included 119 species that were not collected at any of the 147 mainland sites. Conversely, 98 species were collected at mainland sites but not at island sites. A slight majority of the species (284 of 501) were found on both island and mainland shelves. We refer to these 3 groups as 119 ‘island-only’ species, 98 ‘mainland-only’ species, and 284 ‘cosmopolitan’ species. Species-accumulation curves further show that the island sites contained a greater number of polychaete species, and the lack of an asymptote in the island data suggests that increasing the sampling effort on the island shelf probably would result in collecting additional ‘island-only’ species (Fig. 2).

Based on literature accounts of polychaete life histories, we were able to know or infer the dispersal potential of 344 of the 501 species (Appendix 1). Of the 344 species assigned to the ‘high’, ‘medium’ or ‘low’ dispersal-potential categories, 92 had known life-history traits or belonged to families in which all members that have been studied share the same life-history traits. Inferences about the dispersal potential of the remaining 252 species are less certain due to some variability of the known life-history traits among species within a family or genus (Appendix 1). Of the 49 families, 13 had no published information on reproduction and development: Acoetidae (2 spp.), Acrocirridae (3 spp.), Cossuridae (3 spp.), Faunvelopsidae (2 spp.), Lacydoniidae (1 sp.), Longosomatiidae (1 sp.), Oenionidae (8 spp.), Paraonidae (27 spp.), Pholoidae (2 spp.), Pilargidae (9 spp.), Questidae (1 sp.), Saccocirridae (1 sp.) and Sternapsidae (1 sp.). Appendix 1 summarizes the known life-history characteristics, inferred dispersal potentials, frequencies of occurrence, and literature references for each of the 501 species.

Of the 344 species that could be assigned to dispersal-potential categories based on published life-history data, 128 were designated as ‘high’ dispersers, 93 as ‘medium’ dispersers, and 123 as ‘low’ dispersers (Table 1). On average, the 128 species in the high dispersal-potential category occurred at 11.2% of the 200 sampling sites (Table 1). The 93 species in the medium category and the 123 species in the low category were found at fewer sites on average (9.4 and 7.1%, respectively; Table 1). As predicted, the average occurrence frequencies tended to decline as the inferred dispersal potential decreased, but the high, medium, and low categories did not have significantly different occurrence frequencies (p = 0.272, Kruskal-Wallis test). The 157 species that could not be assigned to dispersal-potential categories occurred at 5.6% of the 200 sites; the lack of published information on these species’ life histories can be explained largely because these species are relatively uncommon.

Because there are many oceanographic and ecological differences between the 53 island and 147 mainland locations, we decided a priori to analyze each of the 2 sampling areas separately. Of the 403 species collected at island sites, 96 were designated as high dispersers, 64 as medium dispersers, and 114 as low dispersers (Table 2). Species in the high, medium, and low dispersal-potential categories had significantly different occurrence frequencies on the island shelf (p = 0.031, Kruskal-Wallis test). On average, species in the high dispersal-potential category occurred at
14.6% of the 53 island sites. Species in the medium and low dispersal categories occurred at 12.1 and 9.0% of the 53 island sites, respectively (Table 2). Of the 382 species collected at mainland sites, 114 were designated as high dispersers, 78 as medium dispersers, and 82 as low dispersers (Table 2). Although species in the high dispersal-potential category tended to occur more frequently than medium and low dispersers at the 147 mainland sites (Table 2), the slight differences among the 3 dispersal-potential categories were not significant at the mainland sites (p = 0.987, Kruskal-Wallis test).

The occurrence frequencies also differed between the ‘island-only’ species, ‘mainland-only’ species, and ‘cosmopolitan’ species (Table 3). On average, the 284 ‘cosmopolitan’ species occurred at 13.3% of the 200 sites. This occurrence frequency is significantly greater than that of either the 119 ‘island-only’ species, which occurred at 3.7% of the 53 island sites (p < 0.001, Mann-Whitney U-test) or the 98 ‘mainland-only’ species, which occurred at 2.6% of the 147 mainland sites (p < 0.001, Mann-Whitney U-test). The occurrence frequencies of the ‘island-only’ and ‘mainland-only’ species also differed significantly (p < 0.001, Mann-Whitney U-test).

Most importantly, the ‘island-only’, ‘mainland-only’, and ‘cosmopolitan’ species showed different patterns of dispersal potential. Of the 119 ‘island-only’ species, 70 were assigned to dispersal-potential categories; the majority of these species had life-history traits that suggested a limited potential for larval dispersal (Table 3). In contrast, the 98 ‘mainland-only’ species tended to have high or medium dispersal potential; only 9 of the 70 categorized ‘mainland-only’ species were assigned to the low dispersal-potential category (Table 3). The frequency distribution of species in the high, medium, and low dispersal-potential categories differed significantly between the ‘island-only’ and ‘mainland-only’ groups (p < 0.001, chi-square = 31.978). Of the 284 ‘cosmopolitan’ species, 204 could be assigned to dispersal-potential categories. As one would expect, the majority of these ‘cosmopolitan’ species were in the high and medium dispersal-potential categories, but surprisingly 73 of the 204 species were designated as ‘low’ dispersers (Table 3). The frequency distribution of species in the high, medium, and low dispersal-potential categories differed significantly between the ‘cosmopolitan’ and ‘island-only’ groups (p = 0.003, chi-square = 12.722) and between the ‘cosmopolitan’ and ‘mainland-only’ groups (p < 0.001, chi-square = 15.085).

### DISCUSSION

The dispersal of larvae poses one of the greatest challenges to designing MPAs as effective tools for conserving biodiversity (Gaines et al. 2003, Grantham et al. 2003, Kinlan & Gaines 2003, Palumbi 2003, Shanks et al. 2003). Our analysis of soft-sediment polychaetes in the Southern California Bight reveals that a regional network of MPAs must include both mainland and island sites. More importantly, our analysis suggests that the sizes and spacing of MPAs should be designed and managed somewhat differently in island and mainland shelves in the region.

Samples from sites in the Channel Islands included 119 polychaete species that were not collected from any of the mainland sites. The majority of these ‘island-only’ species had life-history traits that suggest a limited potential for larval dispersal. Sustainable populations of many low-dispersing species are likely to be protected in the Channel Islands if relatively small MPAs are located in sites that presently have high biodiversity. Because much of the island shelf already lies within the CINMS, in which oil and mineral exploration are prohibited, the major question associated with MPAs in the Channel Islands is whether certain sites should be managed as more restrictive marine reserves in which additional activities such as bottom trawling are prohibited (CINMS 2005). Gaining broad support for small reserves is often easier and requires fewer political compromises than establishing larger reserves (Roberts et al. 2003). Presently, 20% of the CINMS has been designated as marine reserves (CINMS 2005). Our analysis suggests that a network of small reserves in the Channel Islands is likely to protect many of the infaunal species that have limited dispersal.

In contrast, our analysis suggests that MPAs are much less likely to be effective at conserving infaunal diversity along the mainland shelf. The 98 polychaete species that were collected from main-
land sites but were not found at any of the island sites.

Table 3. Polychaete species in the Southern California Bight. Mean occurrence frequencies (±SD) of species that were collected from island sites but not any of the mainland sites (‘island-only’ species), species collected at mainland sites but not any of the island sites (‘mainland-only’ species), and species collected from both island and mainland sites (‘cosmopolitan’ species). Occurrence frequency is the percentage of the total sampling sites at which a given species was collected (see Appendix 1; available at: www.int-res.com/articles/supp/m316p105_app.pdf). The occurrence frequencies of ‘island-only’ and ‘mainland-only’ species were calculated conservatively at a total of 53 island sites or 147 mainland sites, respectively. Occurrence frequencies of the ‘cosmopolitan’ species were calculated at all 200 sites in the region.

<table>
<thead>
<tr>
<th>Dispersal potential</th>
<th>‘Island-only’ species (53 island sites)</th>
<th>‘Mainland-only’ species (147 mainland sites)</th>
<th>Cosmopolitan species (all 200 sites)</th>
</tr>
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<tbody>
<tr>
<td></td>
<td>No. of species</td>
<td>Occurrence freq. (%) at island shelf</td>
<td>No. of species</td>
</tr>
<tr>
<td>High</td>
<td>14</td>
<td>4.7 (±3.1)</td>
<td>32</td>
</tr>
<tr>
<td>Medium</td>
<td>15</td>
<td>2.9 (±2.4)</td>
<td>29</td>
</tr>
<tr>
<td>Low</td>
<td>41</td>
<td>3.7 (±2.5)</td>
<td>9</td>
</tr>
<tr>
<td>Undetermined</td>
<td>49</td>
<td>3.6 (±2.7)</td>
<td>28</td>
</tr>
<tr>
<td>Total</td>
<td>119</td>
<td>3.7 (±2.6)</td>
<td>98</td>
</tr>
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Because ocean currents have large effects on pelagic dispersal, efforts to locate and manage MPAs must interpret life-history trends in the context of oceanographic features on a variety of scales. Large-scale oceanographic trends suggest a high degree of larval exchange across the Southern California Bight. At different times of the year water moves north, south, seaward and landward past the Channel Islands, depending on the flow of the southward California Current, associated upwelling and relaxation events, and the formation of the northward-flowing Southern California Countercurrent (Lynn & Simpson 1987). It is not surprising, therefore, that over half of all polychaete species in our analysis were found on both island and mainland shelves.

The presence of many rare species in each of the dispersal-potential categories suggests that a variety of small-scale oceanographic features and species-specific interactions with oceanographic conditions also must be understood and considered in the design and management of a regional MPA network. Recirculating currents and coastal boundary layers may help retain larvae near their point of origin even during long planktonic durations (Largier 2003). When Parker & Tunnicliffe (1994) found that many species inhabiting an isolated seamount had a high potential for dispersal, they hypothesized that oceanographic forces must play a role in the persistence of these species at the seamount. Several other studies have provided evidence of self-recruitment despite a population’s potential for longer-range dispersal (e.g. Todd 1998,
In general, diverse life histories in a community and oceanographic variability on a variety of spatial and temporal scales probably will doom any regional network of MPAs that promises to conserve biodiversity. Furthermore, MPAs designed to protect a subset of the community, will not necessarily protect the majority of species in a region. For example, Sala et al. (2002) developed a comprehensive model for locating marine reserves in the Gulf of California. Their design criteria focused on commercial reef fishes and explicitly assumed that a reserve network designed to protect these prominent species would also protect invertebrates and algae that have different life histories and dispersal distances than the targeted fishes. Our analysis and similar analyses (e.g. Grantham et al. 2003) suggest that such assumptions might be too simplistic, especially if the objective of an MPA network is to preserve biodiversity. Including a wide mix of habitat types within a reserve network (e.g. Sala et al. 2002) can be an effective bet-hedging strategy, but only if the habitat requirements of rare species are well understood.

Shelf-sediment habitats have only a few variables that can be measured routinely (e.g. grain size, organic content, depth, flow regime), and the ecological niches of most infauna are less clear to ecologists than those of many hard-substrate species (Wilson 1991b, Snelgrove & Butman 1994). While considerable progress has been made toward understanding the spatial and temporal nature of source-sink dynamics in populations of commercially important marine species (e.g. Stockhausen et al. 2000, Taylor 2004), similar efforts aimed at less conspicuous infaunal species that play key roles in the functioning of ecosystems have yet to occur.

Estimating dispersal potential from an incomplete set of life-history characteristics is a difficult task. The diverse life histories of polychaetes magnify the problem. Most all families have a variety of radically different life histories; the species within many genera have similarly diverse modes of reproduction and dispersal potentials. Furthermore, some polychaete species have demonstrated varying reproductive and developmental behavior across geographic ranges and within a single site (e.g. Levin & Huggett 1990). Assigning characteristics to a poorly studied species using knowledge from other species in the family or genus, therefore, amounts to an educated guess at best.

Much work is left to do to understand the dispersal of polychaetes and other infaunal taxa. First, the reproductive and developmental biology of more species must be described. This is especially critical for infaunal species that play key roles in communities such as creating biogenic structures like tubes and mounds, providing major food items for other species, and affecting sediment chemistry. Secondly, new advances in larval tagging technology (review in Thurrold et al. 2002) must be incorporated into field trials to test hypothesized dispersal potentials by measuring larval dispersal directly. Lastly, assays of population genetics must be performed on many more infaunal species to place the exchange of larvae between groups in a greater temporal context. Only when we understand the minimum size of sustainable populations and the ability of populations to supply larvae to other areas can conservation efforts, such as the design of MPAs, be effective at maintaining biodiversity.

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