Reproduction, growth and connectivity among populations of *Girella tricuspidata* (Pisces: Girellidae)

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ABSTRACT: *Girella tricuspidata* is widely distributed and harvested by recreational and commercial fishers along the coastline of eastern Australia. The present study examined variability in the reproductive biology and growth of individuals within populations of *G. tricuspidata* across 3 estuaries (Clarence, Tuggerah and Tuross) and assessed population connectivity via a large-scale tag-recapture study in which fish were tagged across 9 estuaries. Spawning occurred predominantly between June and September in the Clarence River and between October and January in the Tuross River, suggesting that spawning occurs later in the year at higher latitudes. The recruitment of young to nursery grounds was spatially and temporally variable. *G. tricuspidata* are group-synchronous spawners, and the estimated batch fecundity was positively correlated with fish length. The estimated length and age at which 50% of *G. tricuspidata* attained reproductive maturity was similar for both sexes: ~286 mm fork length (FL) and 4.1 yr for males and 295 mm FL and 4.5 yr for females. *G. tricuspidata* were aged using otoliths to >26 yr, whereas the reading of scales consistently underestimated the age of fish older than 5 yr. Growth was flexible but varied significantly between sexes and among estuaries; females grew faster than males and attained a larger asymptotic length in the Clarence and Tuross Rivers but not in Tuggerah Lake. Growth was rapid for both sexes until 4 to 5 yr of age, after which it slowed. Of the 6871 *G. tricuspidata* tagged, 15% were recaptured, with 96% of these fish recaptured in the estuary in which they were initially tagged. The recaptured individuals that migrated between estuaries predominantly displayed a northward movement into the prevailing coastal current, which is probably a life-history tactic to facilitate wide dispersal of eggs and larvae along eastern Australia.

KEY WORDS: Fish · Life history · Maturity · Spawning · Otolith · Ageing · Movement · Tag · Australia

INTRODUCTION

Members of the teleost family Girellidae occur in shallow coastal and estuarine subtropical and temperate waters of the Pacific, Indian and Atlantic Oceans (Yagishita & Nakabo 2000, www.fishbase.org). The family comprises 18 species of *Girella* and the monotypic *Graus*. There is little widely published documentation of the fundamental life-history characteristics of the Girellidae, including reproductive mode, size and age at maturity, growth, longevity and mortality schedules. Such information is important for advancing our understanding of the population ecology and ecological role of these fishes and assisting in management for sustainable fisheries exploitation (Megrey 1989, Campana 2001, Begg et al. 2005, Ruttenberg et al. 2005).
Girella tricuspidata inhabits estuaries and nearshore coastal waters along eastern and southern Australia and around the north island of New Zealand (Kailola et al. 1993, Miskiewicz & Trnski 1998), where it is a significant component of several recreational and commercial fisheries (West & Gordon 1994, Gray & Kennelly 2003, Gray et al. 2005, 2010, Butcher et al. 2011). In New South Wales (NSW) alone, ~700 to 1000 t of G. tricuspidata are harvested in total by recreational and commercial anglers each year (Henry & Lyle 2003, Rowling et al. 2011). These fish can live for >25 yr but may be subject to high rates of fishing mortality that can vary spatially and temporally (Gray et al. 2010). At present, broad-scale management arrangements apply to the commercial and recreational fishing sectors and include fishing gear restrictions, minimum legal lengths (MLL) of fish that can be retained (270 mm total length [TL], 300 mm fork length [FL] in NSW) and recreational possession limits (20 individuals in NSW). These management initiatives were developed in the absence of information concerning the species reproductive biology, knowledge of which could help develop more appropriate management plans for the species (Jakobsen et al. 2009). Further, many of the estuaries in which G. tricuspidata occur in eastern Australia have different fishing management regimes (e.g. open or closed to commercial fishing, recreational only or marine reserve), and there is an increasing need to understand the effects of these differing management regimes on exploited populations. This requires an understanding of the linkages and connectivity of populations among these different estuaries (Meyer et al. 2007, Chateau & Wantiez 2009).

Fish display flexible life-history characteristics that can vary throughout their distribution (Meekan et al. 2001, Ruttenberg et al. 2005) and, for some species, among individual estuaries within a specific region (Sarre & Potter 1999, Bedee et al. 2002). In the present study, we examine aspects of the reproductive biology and growth of Girella tricuspidata across 3 estuaries in eastern Australia. We specifically investigate the time of reproductive activity, length and age at sexual maturity, mode of oocyte development and potential batch fecundity, and test for differences in growth parameters between sexes and among populations. We also build on previous otolith-based ageing work (Gray et al. 2010) by examining the timing when the first opaque growth zone on otoliths is deposited. The large-scale movements and population connectivity of G. tricuspidata among estuaries was investigated through a tag-recapture study. We discuss the life-history characteristics of G. tricuspidata in relation to current knowledge of the ecology and management plans for the species.

MATERIALS AND METHODS

Sampling

Girella tricuspidata were sampled from the Clarence River (29.427°S, 153.372°E), Tuggerah Lake (33.345°S, 151.504°E) and the Tuross River (36.067°S, 150.135°E) in NSW, Australia, at regular intervals (generally monthly) between September 2003 and March 2005. Fish from the Clarence River and Tuggerah Lake were obtained mainly from commercial gillnet (80 to 100 mm stretched mesh) catches, whereas fish from the Tuross River were collected from research multi-mesh gillnets with 50 to 100 mm stretched mesh and a 140 m beach-seine with 25 mm mesh. A total of 2571 G. tricuspidata were collected for reproduction and age examination. The FL (mm), total weight (0.1 g) and sex were determined for each fish sampled.

Reproductive staging and activity

Gonads were removed, weighed (0.1 g), macroscopically examined and assigned a reproductive stage according to a developmental criteria based on the size, colour and visibility of oocytes (adapted from Scott & Pankhurst 1992) (Table 1). Female and male Stages I and II were considered immature, female Stages III to V and male Stages III to IV were considered mature, and female Stage VI and male Stage V were spent. The macroscopic staging of females was validated by microscopic examination of a small subset of individuals: Stage II immature ovaries contained unyolked oocytes of a variety of sizes; Stage III ovaries contained a mixture of unyolked, partially yolked and advanced yolk stage oocytes; Stage IV ovaries contained hydrated oocytes and oocytes in each of the previous stages of development.

The timing of reproductive activity at each location was assessed by examining trends in the peaks and troughs in the mean male and female gonadosomatic indices (GSI) and proportions of macroscopic gonad stages among months. GSI were calculated for all mature fish: GSI = (Wg/Wt) × 100, where Wg is the gonad weight, and Wt is the total weight of the fish. This index is commonly used in fish reproduction studies; however, we acknowledge its potential limitations as discussed by Ebert et al. (2011).
Mode of spawning and batch fecundity

Gonad tissue from a subset of females was preserved in a solution of formaldehyde, acetic acid and calcium carbonate to determine the mode of spawning and estimate potential batch fecundity. The patterns of oocyte development and mode of spawning were determined by examining the oocyte diameters of 5 individuals with Stage III ovaries and 5 individuals with Stage IV ovaries collected in the Tuross River during the identified spawning period. Each ovary was blotted dry and weighed (within 0.0001 g), after which 3 replicate sub-samples were taken, blotted dry, weighed (within 0.0001 g) and placed in a sealed sample jar containing a 70% alcohol solution. Each sub-sample was then placed in a sonic bath (Unisonics FXP4) for a period no longer than 20 min to dislodge individual oocytes from surrounding connective tissue. Oocytes from each sub-sample were transferred into a Petri dish and separated from each other, with all remaining connective tissue removed. The Petri dish was scanned (Canon CanoScan 8600F) at 1200 dpi resolution, and image analysis software (ImageJ Version 1.38i) was used to determine the size and number of oocytes in each sub-sample. Individual size-frequency plots of oocyte diameters were produced for each of the 2 gonad stages.

Table 1. Macroscopic appearance and corresponding histological condition used to stage gonads and testes of *Girella tricuspidata*. Adapted from Scott & Pankhurst (1992)

<table>
<thead>
<tr>
<th>Stage</th>
<th>Classification</th>
<th>Macroscopic appearance</th>
<th>Histological characteristics</th>
</tr>
</thead>
<tbody>
<tr>
<td>Female</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>I</td>
<td>Immature</td>
<td>Ovary: clear, thread-like, pink in colour</td>
<td>Previtellogenic oocytes present</td>
</tr>
<tr>
<td>II</td>
<td>Regressed</td>
<td>Ovary: small, clear and pink in colour</td>
<td>Cortical alveoli stage oocytes present</td>
</tr>
<tr>
<td>III</td>
<td>Vitellogenic</td>
<td>Ovary: colour orange, opaque oocytes visible through epithelium</td>
<td>Oocytes in exogenous vitellogenesis</td>
</tr>
<tr>
<td>IV</td>
<td>Hydrated</td>
<td>Ovary: hydrated oocytes visible through epithelium</td>
<td>Final oocyte maturation, hydrated oocytes present</td>
</tr>
<tr>
<td>V</td>
<td>Ovulated</td>
<td>Eggs: in the oviduct and can be extruded with gentle pressure</td>
<td>Hydrated oocytes and post-ovulatory follicles present</td>
</tr>
<tr>
<td>VI</td>
<td>Spent</td>
<td>Ovary: bloody and flaccid</td>
<td>Atretic vitellogenic oocytes and previtellogenic oocytes present</td>
</tr>
<tr>
<td>Male</td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>I</td>
<td>Immature</td>
<td>Testis: white, thread-like</td>
<td>Spermatogonia</td>
</tr>
<tr>
<td>II</td>
<td>Spermatogenic</td>
<td>Testis: firm and ivory white in colour</td>
<td>Secondary spermatocytes and spermatozoa present</td>
</tr>
<tr>
<td>III</td>
<td>Partially spermiated</td>
<td>Testis: firm, ivory white in colour with viscous milt in sperm duct</td>
<td>Spermatozoa predominate</td>
</tr>
<tr>
<td>IV</td>
<td>Fully spermiated</td>
<td>Testis: firm, ivory white in colour with free-flowing milt in sperm duct</td>
<td>Spermatozoa predominate</td>
</tr>
<tr>
<td>V</td>
<td>Spent</td>
<td>Testis: grey to bloody in colour and flaccid</td>
<td>Residual spermatozoa, reduced spermocytes and increased connective tissue</td>
</tr>
</tbody>
</table>

Estimates of potential batch fecundity were determined by counting the number of hydrated eggs present within Stage IV ovaries from individuals ranging in size from 290 to 420 mm FL collected from the Tuross River during the identified spawning period. The number of hydrated eggs in a weighed (~0.1 g) sub-sample of ovarian tissue was counted, and the potential total batch fecundity for each fish was estimated by scaling this number up to the total gonad weight. A linear regression analysis was used to determine the significance of the relationship between estimated potential batch fecundity and FL.

Length and age at maturity

The estimated FL and age at which 50% ($L_{50}$ and $A_{50}$ respectively) of males and females attained reproductive maturity in the Tuross River was determined in each case by fitting a logistic regression model using the binomial general linear model function in the R statistical environment to the proportions of mature (Stage III and above) and immature (Stage I and II) fish in each 1 cm length class and each age class. No useful estimates of $L_{50}$ and $A_{50}$ could be derived for fish collected in the Clarence River because of a lack of immature fish sampled.
and, for Tuggerah Lake, because of a lack of mature fish sampled. Only fish caught during the identified spawning period were used for these analyses, and differences between sexes in the estimated \( L_{50} \) and \( A_{50} \) values in the Tuross River were tested using the 2-sampled \( Z \) technique (Gunderson 1977).

**Age estimation: otoliths and scales**

Discrepancies exist in estimates of growth and longevity for *Girella tricuspidata*; previous work provided vastly different maximum ages, but these differences were most likely attributable to the structures (scales versus otoliths) used for ageing (Gray et al. 2010). We tested this hypothesis by using both sectioned sagittal otoliths and whole scales to estimate the age of *G. tricuspidata*. Otoliths from each fish were cleaned and dried, and one was embedded in resin, after which a 300 µm transverse section was made through the core perpendicular to the longest axis. Scales were cleaned, dried and briefly examined to select the most readable scale from each fish for further examination. When no interpretable scales were collected (i.e. they had regenerated or were deformed; a total of 113 fish), those fish were excluded from further analyses. Whole scales and otolith sections were mounted on slides and viewed using a microscope with reflected white light against a black background. The scales and otolith sections displayed alternating narrow opaque and broad translucent zones, and the numbers of completed opaque zones along a radius from the core to the outer edge of the structure were counted. All structures were read without knowledge of sample details (i.e. length, sex, location and date of capture), and 25% of otoliths haphazardly drawn from each estuary and 25% of scales from the Tuross River were re-read by the same reader. The repeatability of counts of opaque zones (combined across all age classes and samples) was assessed using the coefficient of variation (CV) (Campana 2001). Age bias plots were used to determine whether there was systematic variation in the estimated ages of individual *G. tricuspidata* in the Tuross River derived from sectioned otoliths and whole scales (Campana et al. 1995).

**Identification of first opaque zone on otoliths**

The position and timing of deposition of the first opaque zone on otoliths were determined by keeping young-of-the-year fish in an aquarium facility. A total of 25 juveniles (20 to 80 mm FL) were captured in Botany Bay (34.02° S, 151.32° E) in March 2005, transferred and housed in a 5000 l tank at the nearby Cronulla Fisheries Research Centre (34.17° S, 151.35° E). Fish were collected from Botany Bay due to its closeness to the research facility to minimise the logistic constraints of transporting fish from distant locations. It was not possible to replicate the environmental conditions from each of the 3 study estuaries in the laboratory. Fish were therefore maintained at local ambient water and air temperatures (water pumped directly from the Port Hacking estuary), exposed to natural cycles of light and fed commercially available pellets (Ridley Aqua Feed) once a day. Every 3 mo between September 2005 and December 2006, 5 fish were captured, euthanised and had their otoliths removed. One otolith from each pair was processed for age estimation as described above. The distance from the otolith core to the otolith edge and to the first opaque zone was measured using a microscope-mounted camera interfaced with digital image analysis software (ImageJ Version 1.38i). All measurements were made along the ventral margin of the sulcus acusticus.

**Growth**

Growth was modelled separately for each sex in each estuary by fitting the von Bertalanffy growth function (VBGF) to the length-at-age data from sectioned otoliths. Because of variations in the proportions of small and young fish obtained in samples (Table 2), the analyses for the Clarence River and Tuggerah Lake were constrained with \( t_0 = -0.3 \). This was based on unconstrained results from the Tuross River, which had the widest spread of size and age groups in samples (Table 2). The best-fit VBGF for each sex in each estuary was determined using non-linear least-squares procedures, and the growth curves between sexes and among estuaries were compared using likelihood ratio (LR) tests (Kimura 1980, Cerrato 1990).

**Population connectivity**

Between 1988 and 1995, *Girella tricuspidata* in 9 estuaries in NSW were captured in beach-seine nets, measured (FL to the nearest cm), tagged with a unique numbered plastic T-bar (Hallprint TBA) between the dorsal fin rays and released in situ. A total of 6871 *G. tricuspidata* were tagged, and their FL ranged between 127 mm and 562 mm (mean ±
standard error [SE]: 260.1 ± 0.46 mm). Tagging took place year-round in association with commercial fishers, so the time of tagging in a particular estuary was often dependent on the fisher’s activity. The tag-release program was highly publicised, with fishers offered a reward upon the return of tagged fish and supply of information including where and when the fish was recaptured and, where applicable, re-released. A $\chi^2$ goodness-of-fit test was used to determine whether there was a significant difference in the numbers of fish that relocated north or south. Because only 4 fish relocated south, meaningful statistical analyses comparing the distances relocated either north or south could not be conducted. The extent of movements between the points of release and subsequent capture were plotted on a map.

**RESULTS**

**Reproductive period**

Mean male and female GSI peaked between June and September in the Clarence River and between October and January (and in February 2005) in the Tuross River (Fig. 1). In contrast, the mean GSI values changed little throughout the year for each sex in Tuggerah Lake, although slightly higher mean GSI values were evident in August and September 2003 and in July and November 2004 (Fig. 1). The greatest mean GSI for females was 8.62 in the Clarence River and 7.18 in the Tuross River, but only 1.83 in Tuggerah Lake. Based on the macroscopic staging of gonads, the greatest proportion of mature female and male fish (Stage III to Stage VI) occurred between May and September in the Clarence River and between October and March in the Tuross River (Fig. 2). No Stage V males occurred in samples. Very few reproductively active fish were sampled in Tuggerah Lake in any month, precluding the identification of any temporal reproductive pattern.

**Mode of spawning and batch fecundity**

The frequency distribution of oocyte diameters differed between individuals with Stage III ovaries and those with Stage IV ovaries (Fig. 3). The diameters of the largest cohort of oocytes ranged between 0.40 and 0.55 mm for individuals with Stage III ovaries and between 0.60 and 0.75 mm for those with Stage IV

![Fig. 1. *Girella tricuspidata*. Mean gonadosomatic indices (GSI) for (a) females and (b) males in Clarence River, Tuggerah Lake and Tuross River](image-url)
Fig. 2. *Girella tricuspidata*. Proportion of each assigned gonad stage (see Table 1) for females and males in Clarence River, Tuggerah Lake and Tuross River.
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A cohort of oocytes with diameters of 0.40 to 0.55 mm was also obvious in some individuals with Stage IV ovaries (e.g. Fig. 3d,e). For Stage III and IV gonads, there was a scattering of developing oocytes with diameters ranging from 0.20 to 0.40 mm (Fig. 3). The frequency distribution of oocyte diameters within mature ovaries is indicative of asynchronous oocyte development, which is consistent with batch spawning.

Estimates of potential batch fecundity ranged between ~64 000 and 834 000 eggs for *Girella tricuspidata* between 296 and 426 mm FL (495 to 1312 g total weight) respectively (Fig. 4). There was a significant positive relationship between the relative batch fecundity and FL ($p = 0.00057$, $r^2 = 0.2215$).

### Length and age at maturity

The smallest observed mature female and male in the Tuross River were 273 mm and 264 mm FL respectively, whereas the estimated $L_{50}$ and $A_{50}$ (±SE) were 294.6 ± 4.0 mm (n = 418) and 4.52 ± 0.1 yr (n = 412) for females and 286.2 ± 6.7 mm (n = 242) and 4.16 ± 0.2 yr for males.

![Fig. 3. *Girella tricuspidata*. Oocyte size frequency distributions within (a,b,c) Stage III and (d,e,f) Stage IV gonads of individuals sampled in Tuross River](image-url)
Precision of age estimation and comparisons between otolith and scale readings

The within-reader agreement for counting opaque zones on sectioned otoliths was 93.4%, with a corresponding CV value of 10.6. The within-reader variation for scale interpretation was 90.1% (CV = 12.8).

The age-bias plots of the comparison of the counts of completed opaque zones on otoliths and scales showed similar values for fish aged ≤5 yr, after which the agreement between counts decreased because there was a systematic underestimation of ageing based on scales (Fig. 6). The maximum discrepancy between counts of opaque zones on otoliths and scales was 13 zones, and this was for a fish with 17 opaque zones on the otolith (Fig. 6).

Validation of first opaque zone formation on otoliths

Juvenile fish held in captivity were first sampled in September 2005, with all 5 sampled fish displaying a newly deposited opaque zone on their otoliths (Fig. 7). No further opaque zones were observed in fish sampled between December 2005 and June.
2006, whereas a second opaque zone was visible on the remaining 5 fish sampled in December 2006. These observations indicate that the first opaque zone was formed between July and September and the second between July and December. The mean (±SE) distance from the otolith core to the first opaque zone of these captive fish was 0.361 ± 0.008 mm, whereas for wild-caught *Girella tricuspidata* (n = 567), the distance was 0.373 ± 0.005 mm.

**Length-at-age and growth**

The length of *Girella tricuspidata* at any given age varied considerably within and among estuaries (Fig. 8). The oldest (and largest) fish sampled was a 24 yr old female (426 mm FL) from the Tuross River. The oldest fish sampled in the Clarence River was a female aged 18 yr, and in Tuggerah Lake, the oldest fish was a male aged 19 yr. Few fish older than 10 yr were sampled from all 3 estuaries (Table 2).

The von Bertalanffy growth curves differed significantly between males and females in the Clarence (p < 0.0001) and Tuross (p = 0.012) Rivers, where females grew faster and attained a greater mean estimated maximum length ($L_\infty$) than males (Table 3, 61).

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**Fig. 6. Girella tricuspidata.** Relationship between estimated ages of fish from sectioned otoliths and whole scales. Straight line shows 1:1 relationship; error bars represent 95% confidence limits (n = 1934)

**Fig. 7. Girella tricuspidata.** Distances (mm) from the core to the edge of each otolith of juvenile fish kept in captivity and sampled at different times between September 2005 and December 2006

**Fig. 8. Girella tricuspidata.** Variation in fork length-at-age and von Bertalanffy growth curves for males and females in (a) Clarence River, (b) Tuggerah Lake and (c) Tuross River. Von Bertalanffy growth parameters are provided in Table 3.
There was no significant difference ($p = 0.167$) between the growth curves of males and females in Tuggerah Lake. The growth of *Girella tricuspidata* also differed significantly (in all cases, $p < 0.05$) among estuaries for fish of the same sex (Table 3, Fig. 8).

**Population connectivity**

Of the 6871 *Girella tricuspidata* tagged, 988 (14.4%) were recaptured at least once. A total of 939 fish were recaptured once, 44 twice, 3 thrice, and two 4 times, providing a total of 1043 records of recapture (Table 4). The location of the first recapture was within the estuary in which the fish was tagged in all but 36 cases (including all 49 multiple-recapture fish), with the duration at large before the first recapture ranging between 1 and 1462 d (mean ± SE = 125.9 ± 3.9 d; $n = 988$) (Table 4). The direction along the coast that emigrating *G. tricuspidata* had relocated from the estuary in which they were tagged differed significantly ($H_0$: north = south; $\chi^2 = 24.641$, df = 1, $p < 0.0001$) and was north in 35 cases and south in 4 cases (Fig. 9, Table 4). The distances relocated ranged between 11 and 455 km for northward-relocated fish (mean ± SE = 160.2 ± 21.2 km; $n = 35$), with the furthest relocation being from the Shoalhaven River to the Hastings River. Southward-relocated fish travelled between 26 and 78 km (mean ± SE = 46.3 ± 11.1 km; $n = 4$), with the furthest relocation being from the Shoalhaven River to Lake Conjola (Fig. 9). The small number of southward-relocated fish prevented meaningful analysis comparing distances moved between the 2 groups.

The mean (±SE) duration at large before the first recapture for all recaptured *Girella tricuspidata* was 125.8 ± 3.9 d ($n = 986$). Of the 49 multiple-recapture fish, 41 were tagged in estuaries on the north coast. For all but 2 of the multiple-recapture fish, the second and subsequent recaptures were, like the first, within the estuary in which the fish was initially tagged, with the cumulative duration at large prior to

<table>
<thead>
<tr>
<th>Group</th>
<th>Location</th>
<th>Sex</th>
<th>$L_\infty$ (±SE)</th>
<th>$k$ (±SE)</th>
<th>$t_0$</th>
<th>RSS</th>
<th>n</th>
<th>LR test and significance</th>
</tr>
</thead>
<tbody>
<tr>
<td>1</td>
<td>Clarence</td>
<td>Male</td>
<td>307.753 (3.628)</td>
<td>0.5935 (0.36)</td>
<td>0.3</td>
<td>104539.33</td>
<td>175</td>
<td>1 vs. 2**, 1 vs. 3**, 1 vs. 5***</td>
</tr>
<tr>
<td>2</td>
<td>Clarence</td>
<td>Female</td>
<td>338.838 (3.642)</td>
<td>0.4786 (0.021)</td>
<td>0.3</td>
<td>82189.53</td>
<td>169</td>
<td>2 vs. 1**, 3 vs. 4**, 4 vs. 6***</td>
</tr>
<tr>
<td>3</td>
<td>Tuggerah</td>
<td>Male</td>
<td>385.533 (3.447)</td>
<td>0.3944 (0.01)</td>
<td>0.3</td>
<td>256177.5</td>
<td>416</td>
<td>3 vs. 1**, 3 vs. 4**, 4 vs. 5***</td>
</tr>
<tr>
<td>4</td>
<td>Tuggerah</td>
<td>Female</td>
<td>375.025 (3.156)</td>
<td>0.4222 (0.011)</td>
<td>0.3</td>
<td>245584.08</td>
<td>410</td>
<td>4 vs. 2**, 4 vs. 3**, 4 vs. 6***</td>
</tr>
<tr>
<td>5</td>
<td>Tuross</td>
<td>Male</td>
<td>363.03 (4.11)</td>
<td>0.3309 (0.01)</td>
<td>0.3</td>
<td>482169.34</td>
<td>434</td>
<td>5 vs. 1**, 3 vs. 4**, 5 vs. 6***</td>
</tr>
<tr>
<td>6</td>
<td>Tuross</td>
<td>Female</td>
<td>370.41 (3.44)</td>
<td>0.3351 (0.009)</td>
<td>0.3</td>
<td>585605.89</td>
<td>622</td>
<td>6 vs. 2**, 4 vs. 3**, 6 vs. 5*</td>
</tr>
</tbody>
</table>

*Fig. 9. Girella tricuspidata. Locations of released and recaptured fish tagged in estuaries throughout New South Wales. Number of fish movements northward (solid line) or southward (dashed line) is shown adjacent to each line. The numbers in parentheses associated with some estuary names represent the total number of all recaptures (i.e. single and multiple combined) of fish tagged and recaptured in that estuary.*
Table 4. Time period, numbers and fork length (FL) ranges of fish tagged and released in each estuary and a summary of the subsequent reported numbers of recaptures in estuarine and coastal waters in eastern Australia

<table>
<thead>
<tr>
<th>Estuary</th>
<th>Latitude (°S), longitude (°E)</th>
<th>Date of tagging (mm/yy)</th>
<th>FL range (mm)</th>
<th>No. tagged</th>
<th>Percent recaptured</th>
<th>Total recaptures</th>
<th>Percent recaptured</th>
</tr>
</thead>
<tbody>
<tr>
<td>Richmond River</td>
<td>28.874, 153.591</td>
<td>02/88−01/90</td>
<td>170−388</td>
<td>845</td>
<td>167 (%)</td>
<td>167</td>
<td>19.8</td>
</tr>
<tr>
<td>Clarence River</td>
<td>29.427, 153.372</td>
<td>02/88−01/90</td>
<td>127−390</td>
<td>907</td>
<td>166 (%)</td>
<td>170</td>
<td>18.7</td>
</tr>
<tr>
<td>Bellinger/Kalang River</td>
<td>30.502, 153.032</td>
<td>11/90−09/91</td>
<td>187−400</td>
<td>597</td>
<td>61 (%)</td>
<td>120</td>
<td>20.4</td>
</tr>
<tr>
<td>Nambucca River</td>
<td>30.650, 153.014</td>
<td>10/90−08/91</td>
<td>183−372</td>
<td>1115</td>
<td>184 (%)</td>
<td>190</td>
<td>17.0</td>
</tr>
<tr>
<td>Macleay River</td>
<td>30.874, 153.025</td>
<td>10/90−08/91</td>
<td>167−408</td>
<td>742</td>
<td>116 (%)</td>
<td>118</td>
<td>15.9</td>
</tr>
<tr>
<td>Shoalhaven River</td>
<td>34.901, 150.765</td>
<td>09/92−01/95</td>
<td>153−432</td>
<td>811</td>
<td>73 (%)</td>
<td>81</td>
<td>10.0</td>
</tr>
<tr>
<td>St Georges Basin</td>
<td>35.185, 150.594</td>
<td>05/93−06/94</td>
<td>178−300</td>
<td>389</td>
<td>55 (%)</td>
<td>56</td>
<td>14.4</td>
</tr>
<tr>
<td>Lake Conjola</td>
<td>35.268, 150.508</td>
<td>10/92−12/94</td>
<td>175−380</td>
<td>763</td>
<td>89 (%)</td>
<td>91</td>
<td>11.9</td>
</tr>
<tr>
<td>Burrill Lake</td>
<td>35.395, 150.447</td>
<td>06/93−12/94</td>
<td>178−562</td>
<td>702</td>
<td>93 (%)</td>
<td>100</td>
<td>14.2</td>
</tr>
<tr>
<td>Total</td>
<td></td>
<td></td>
<td>127−562</td>
<td>6871</td>
<td>1004 (%)</td>
<td>1043</td>
<td>15.2</td>
</tr>
</tbody>
</table>

A consequence of the variable and protracted spawning period of *Girella tricuspidata* that occurs over at least 6 mo throughout the study area is that recruitment of young-of-the-year to inshore nursery grounds is spatially and temporally variable. For example, larval *G. tricuspidata* have been recorded in coastal waters adjacent to Sydney, central NSW, between August and May (Gray & Miskiewicz 2000), the second recapture ranging between 18 and 392 d (mean ± SE = 169.8 ± 14.0 d; n = 49) (Table 4). The 2 exceptions were a fish initially tagged in Lake Conjola and recaptured the second time (31 d after initially being tagged) in St Georges Basin (27 km north) and a fish tagged in the Nambucca River and recaptured a second time (130 d after initially being tagged) in the Clarence River (152 km north). The mean (±SE) cumulative duration at large prior to the third and fourth recaptures for the applicable multiple-recapture fish was 249.0 ± 21.4 d (n = 5) and 292.0 ± 13.0 d (n = 2) respectively, with only 1 of those 5 fish eventually leaving the estuary (Macleay River) in which it was tagged to be recaptured ~50 km north in the Bellinger River.

**DISCUSSION**

**Reproduction**

Populations of *Girella tricuspidata* displayed protracted periods of reproductive activity that varied between the 2 most distant estuaries, occurring mainly between June to September (winter) in the Clarence River and October to January (spring/summer) in the Tuross River. This suggests that spawning of *G. tricuspidata* potentially varies with latitude throughout eastern Australia, a pattern observed for other species, including *Hyporhamphus australis* (Hughes & Stewart 2006) and *Macquaria colonorum* (Walsh et al. 2011). Latitudinal and longitudinal variations in the duration and timing of peak reproductive activity are common in species of fish that are widely distributed (Bye 1990, Fowler et al. 1999, Sarre & Potter 1999, DeVries et al. 2002). We acknowledge, however, that because we only had samples from 2 distant estuaries, the patterns observed here might not be indicative of all estuaries within each region. Further sampling in other estuaries is required to determine the extent of variation in reproductive activity across the species distribution prior to any implementation of a management strategy that incorporates temporal fishing closures to protect spawning fish.
and different cohorts of newly settled juveniles have been observed in seagrasses in the Sydney region between July and March (Smith & Sinerchia 2004). Sampling of juveniles in seagrass beds in several estuaries throughout NSW identified that small post-settlement individuals (<20 mm FL) were most predominant between July and November in northern NSW (Clarence and Sandon Rivers) but between July and April in southern NSW (Shoalhaven River and Sussex Inlet) (Gray et al. 2000). We hypothesise that the northern recruits originated from local northern spawning events, but the recruits captured in the southern estuaries originated from both distant northern spawnings (early season) and local spawnings (later season). This may explain why recruits of G. tricuspidata were observed in the southern estuaries prior to local (spring) spawning. Transportation of eggs and larvae from northern waters to southern waters is likely facilitated by the southward-flowing East Australian Current.

Girella tricuspidata with hydrated oocytes were captured only in the lower marine-dominated reaches of the Clarence and Tuross Rivers. Although no direct observations of spawning were made, these observations indicate that G. tricuspidata could potentially spawn in these areas. It is generally assumed that G. tricuspidata spawns in nearshore coastal waters adjacent to surf beaches and estuary mouths (Kailola et al. 1993), and we acknowledge that fish could swim out of these estuaries into coastal waters within a diel or tidal cycle prior to actual spawning (Walsh et al. 2011). Moreover, the actual spawning locations of G. tricuspidata could be flexible and vary depending on local hydrographic conditions at the time of spawning. This may prohibit the use of rigid spatial closures to protect spawning adults as a management strategy for this species.

The oocytes of Girella tricuspidata developed in an asynchronous pattern, with mature Stage IV fish exhibiting oocytes at varying stages of development. These data indicate that G. tricuspidata have an indeterminate fecundity and likely spawn multiple times during the spawning period (De Vlaming 1983, West 1990). This type of spawning behaviour is common among many species of fish and is a strategy that maximises egg production in a spawning period and can help buffer adverse biotic and abiotic conditions that may affect the survival of eggs and larvae (Lambert & Ware 1984, Burt et al. 1988, Sarre & Potter 1999, Kendall & Gray 2009). We could not determine the frequency and number of actual spawning events by an individual throughout a spawning season. Hence, the total number of eggs that each individual produced in a spawning season, and therefore their total overall annual fecundity, could not be estimated within the present study. Nevertheless, the estimated batch fecundity was significantly related to fork length, indicating that reproductive output is potentially greater in larger individuals and that the total collective reproductive output of a population could be enhanced by having greater proportions of larger fish (Parker 1992).

The estimated mean length and age of Girella tricuspidata at reproductive maturity in the Tuross River was similar for males and females, ~286 and 294 mm FL and age 4.1 and 4.5 yr respectively. These mean lengths and ages at maturity corresponded to ~50 and 60% of the observed maximum lengths of males and females respectively and 17% of the observed maximum age of G. tricuspidata sampled. Although differences among estuaries in the L50 and A30 of G. tricuspidata could not be tested due to sample limitations, the smallest mature male and female were much smaller in the Clarence River than in the other 2 estuaries, indicating that length at maturity could potentially vary among distant estuaries, as observed in other species of fish (Sarre & Potter 1999, Silberschneider et al. 2009). This must be examined in greater detail to advise on an appropriate MLL for the species; the current MLL of 250 mm FL in NSW might need to be increased to allow a greater proportion of fish to attain sexual maturity and contribute to spawning prior to harvesting.

Ageing

Interpretation of scales was more difficult and less precise compared to sectioned otoliths and consistently led to underestimations of the ages of Girella tricuspidata that were > 5 yr old. These results concur with the general paradigm that, for most species of fish, sectioned otoliths provide a more accurate estimate of age than scales, particularly for larger and older individuals (Casselman 1990, Campana 2001). We therefore assume that the previous scale-based ageing of G. tricuspidata (Pollock 1981) underestimated the true ages of older fish. This most likely explains why the oldest fish examined here (24 yr) and by Gray et al. (2010) (26 yr) was double the maximum age (12 yr) recorded by Pollock (1981). Underestimation of actual ages can have significant impacts on population dynamics modelling and fisheries assessments (Lai & Gunderson 1987, Campana 2001).

The first opaque zone on the otoliths of captive young-of-the-year fish formed over the austral winter/
early spring, which is in general agreement with that observed in similar field- and aquaria-based validation studies of young-of-the year of other species of fish in southeastern Australia (Hughes et al. 2008, Kendall et al. 2009). Based on the assumption that the first opaque zone forms throughout winter/spring across all estuaries, we hypothesise that the age of the first opaque zone forms throughout winter/spring (Kendall et al. 2009). Based on the assumption that fish in southeastern Australia (Hughes et al. 2008, Bedee et al. 2002, Sala-Bozono & Mariani 2011, Stocks et al. 2011). Such observed variations in growth parameters among populations could, however, simply reflect sample differences between sexes and among estuaries (Sainsbury 1980, Kritzer et al. 2001).

The observed wide range of lengths within each age class in each estuary is probably a manifestation of a combination of biotic and abiotic influences on the composition of populations. For example, each age class within an estuary could contain different cohorts of fish spawned several months apart as a result of the dispersal of eggs and larvae from local and distant spawning events. In addition, interspersion of individuals among meta-populations and genetic variation coupled with spatial and temporal differences in food availability and environmental conditions all potentially influence the observed variability in the length-at-age and growth of individuals (Blanck & Lamouroux 2007, Searcy et al. 2007, Hoxmeier et al. 2009). The consequence of such length-at-age variation could potentially mask or lead to distortions of estuary-specific differences in the growth characteristics of Girella tricuspidata and other species.

Population connectivity

The tag-recapture data presented here and by Thomson (1959) show that Girella tricuspidata are capable of moving between distant estuaries (up to 450 km) and that these movements are predominantly in a northerly direction. These movements, along with the potential relocation of eggs and larvae southwards via the East Australian Current, suggest that there is considerable population mixing of G. tricuspidata along eastern Australia. Although many individuals were recaptured in the same estuary in which they were tagged, the majority of these recaptures were within a short time period and may not indicate permanent residency. Novel acoustic tags have since been used successfully to determine the site fidelity and residency of fish in estuaries (Taylor...
et al. 2006, Heupel et al. 2010, Walsh et al. 2012), and similar studies are required to determine such attributes for *G. tricuspidata*.

Some movements of *Girella tricuspidata* between estuaries and coastal waters may be related to spawning. Throughout late autumn and winter, large schools of *G. tricuspidata* are often observed to congregate (presumably to spawn) near estuary mouths and coastal headlands and travel (mostly northward) along open surf beaches of central and northern NSW. During this period, *G. tricuspidata* are targeted and captured in large quantities by the coastal beach-seine fishery (Gray et al. 2000). The extent of the movements of these presumed 'spawning aggregations' has not been quantified and would require tracking individuals with more advanced acoustic telemetry technologies than used in the present study (e.g. Heupel & Simpfendorfer 2008, Danylchuk et al. 2011). Nevertheless, we hypothesise that the predominant northerly movement displayed by *G. tricuspidata*, along with several other estuarine and coastal teleosts (*Mugil cephalus*, *Acanthopagrus australis*, *Pagrus auratus* and *Pomatomus saltatrix*) (Thomson 1959) and crustaceans (*Penaeus plebejus*) (Montgomery 1990) along the east Australian coast, is a life-history tactic to facilitate the wide dispersal of eggs and larvae by the southward-flowing East Australian Current to coastal and estuarine nursery areas. Essentially, fish move upstream into the prevailing current to spawn, after which eggs and larvae are transported and dispersed by longshore currents to nursery habitats, a tactic displayed by coastal fishes inhabiting other coastal boundary current systems (Hare & Cowen 1993). Little is known, however, of the effects of the vagaries of the East Australian Current and associated oceanographic processes on rates of delivery and survival of larvae to inshore nursery grounds.

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