



# Interactive effects of oyster and seaweed on seawater dissolved inorganic carbon systems: implications for integrated multi-trophic aquaculture

Tingting Han<sup>1</sup>, Rongjun Shi<sup>1,2</sup>, Zhanhui Qi<sup>1,\*</sup>, Honghui Huang<sup>1</sup>, Qingyang Liang<sup>1</sup>, Huaxue Liu<sup>1</sup>

<sup>1</sup>Guangdong Provincial Key Laboratory of Fishery Ecology Environment and Key Laboratory of South China Sea Fishery Resources Exploitation and Utilization, Ministry of Agriculture, South China Sea Fisheries Research Institute, Chinese Academy of Fishery Sciences, Guangzhou 510300, PR China

<sup>2</sup>Guangdong Provincial Key Laboratory of Applied Marine Biology, South China Sea Institute of Oceanology, Chinese Academy of Sciences, Guangzhou 510301, PR China

**ABSTRACT:** We examined the separate effect of Portuguese oyster *Crassostrea angulata* and the interactive effects of oyster and red seaweed *Gracilaria lemaneiformis* on seawater dissolved inorganic carbon (DIC) systems and the air–sea CO<sub>2</sub> flux ( $F_{CO_2}$ ) in Daya Bay, southern China. Respiration and calcification rates of oysters were measured and the effects of oyster aquaculture on marine DIC systems were evaluated. The interactive effects on seawater DIC and air–sea  $F_{CO_2}$  were examined using mesocosms containing oyster and seaweed assemblages. Results showed populations of *C. angulata* cultured in Daya Bay sequestered ca. 258 g C m<sup>-2</sup> yr<sup>-1</sup> for shell formation, whereas the CO<sub>2</sub> released due to respiration and calcification was 349 and 153 g C m<sup>-2</sup> yr<sup>-1</sup>, respectively. This indicates that oyster cultivation in Daya Bay is a CO<sub>2</sub> generator, favoring the escape of CO<sub>2</sub> into the atmosphere. DIC, HCO<sub>3</sub><sup>-</sup> and CO<sub>2</sub> concentrations and the partial pressure of CO<sub>2</sub> in oyster–seaweed co-cultured mesocosms were significantly lower than the oyster monoculture mesocosm. These results indicated that *G. lemaneiformis* effectively absorbs the CO<sub>2</sub> released by oysters. The negative values of air–sea  $F_{CO_2}$  in the co-culture groups represent a CO<sub>2</sub> sink from the atmosphere to the sea. These results demonstrated that there could be an interspecies mutual benefit for both *C. angulata* and *G. lemaneiformis* in the integrated culture system. Considering that photosynthesis of seaweed is carbon limited, we suggest that the 2 species are co-cultured at a ratio of ca. 4:1 (based on fresh weight) for efficient utilization of DIC in seawater by *G. lemaneiformis*, and further to increase the ocean CO<sub>2</sub> sink.

**KEY WORDS:** *Crassostrea angulata* · *Gracilaria lemaneiformis* · Daya Bay · Dissolved inorganic carbon · Integrated multi-trophic aquaculture · IMTA · Air–sea CO<sub>2</sub> flux

## INTRODUCTION

Calcification of aquatic animals such as shellfish (e.g. oyster, scallop, and clam) is also a source of CO<sub>2</sub> (Chauvaud et al. 2003, Martin et al. 2006, Mistri & Munari 2013, Munari et al. 2013, Jiang et al. 2014). Shellfish utilize carbon in 2 ways. First, they consume organic carbon to sustain their growth and meta-

bolism, following the reaction  $CH_2O + O_2 \rightarrow CO_2 + H_2O$ . Second, they use HCO<sub>3</sub><sup>-</sup> from seawater to generate CaCO<sub>3</sub> shells, based on the reaction  $Ca^{2+} + 2HCO_3^- \leftrightarrow CaCO_3 + CO_2 + H_2O$ . These 2 processes both lead to net CO<sub>2</sub> production in ocean waters. Second, shellfish secrete calcium carbonate (CaCO<sub>3</sub>) to form their skeletal material. This process acts as a marine biological pump by removing CO<sub>2</sub> from circu-

\*Corresponding author: qizhanhui@scsfri.ac.cn

lation and storing carbon in the ocean (Lerman & Mackenzie 2005). In fact, the ratio of released CO<sub>2</sub>/precipitated CaCO<sub>3</sub> is largely dependent on the buffering capacity of the surrounding seawater, such that in some marine ecosystems, the ratio could be ca. 0.6 (Frankignoulle et al. 1995). It is reasonable that the buffering capacity of seawater might vary significantly among different waters with variations in pH, alkalinity, salinity, and temperature (Millero 1995, Lerman & Mackenzie 2005, Dickson 2010, Mackenzie & Andersson 2013)

Seaweeds (e.g. *Saccharina*, *Gracilaria*) are intricately involved as primary producers in coastal ecosystems. They assimilate inorganic carbon either via diffusion (for CO<sub>2</sub>), or active uptake of HCO<sub>3</sub><sup>-</sup> using carbon-concentration mechanisms. During photosynthesis, these mechanisms result in an increase in seawater pH and a drop in seawater CO<sub>2</sub> partial pressure (pCO<sub>2</sub>) (Han et al. 2013). Therefore, seaweed could induce a significant shift in seawater dissolved inorganic carbon (DIC) systems according to the following formula: CO<sub>2</sub> + H<sub>2</sub>O ↔ H<sub>2</sub>CO<sub>3</sub> ↔ H<sup>+</sup> + HCO<sub>3</sub><sup>-</sup> ↔ 2H<sup>+</sup> + CO<sub>3</sub><sup>2-</sup>. Therefore, seaweed may exert a significant impact on the DIC buffering capacity of seawater.

As mentioned above, both shellfish and seaweed can change the seawater DIC system and the buffering capacity. One implication is a complex interspecies interaction between shellfish and seaweed in co-cultured systems. For example, CO<sub>2</sub> or HCO<sub>3</sub><sup>-</sup> can become a major limiting factor affecting the photosynthetic rates and aquaculture production of seaweed, particularly when they are grown under conditions of high biomass densities and reduced seawater motion (Zou et al. 2004). Likewise, the alteration in DIC speciation can cause responses in calcifying organisms (e.g. oyster), thereby potentially affecting their growth and physiological functions (Ho & Carpenter 2017, Scanes et al. 2017). Thus, the interaction between shellfish and seaweed and their combined effect on DIC partitioning and cycling still needs to be investigated using an ecosystem approach. Similarly, few studies have been conducted to elucidate the influence of integrated aquaculture of shellfish and seaweed on variations in DIC systems, as well as the air–sea CO<sub>2</sub> flux. Furthermore, the optimum culture ratio in co-culture systems for obtaining the largest CO<sub>2</sub> sink is not known.

Shellfish and seaweed mariculture in the coastal waters of China has been growing rapidly over the past 3 decades; they are by far the largest and most well-known aquaculture industries in the world, with an annual production of ca. 13.6 × 10<sup>6</sup> t and 2.1 ×

10<sup>6</sup> t, accounting for ca. 72.4 % and 11.1 % of the total mariculture production in China, respectively (China Bureau of Fisheries 2016). In most coastal waters, shellfish and seaweed are co-cultured, using suspended longlines as the main cultivation method. In fact, they usually dominate an entire bay, such as in Sanggou Bay (Fang et al. 2016) and in Daya Bay (Yu et al. 2014) where this study was conducted.

In the present study, we conducted an *in situ* mesocosm experiment to measure the calcification and respiration rates of the Portuguese oyster *Crassostrea angulata* to evaluate its effect on marine DIC systems. In addition, the role of the red seaweed *Gracilaria lemaneiformis* was assayed for elimination of CO<sub>2</sub> in seawater. Subsequently, we investigated the impact of co-culture interactions on DIC systems and the air–sea CO<sub>2</sub> flux using different ratios. The results from this study will be useful in evaluating the effects of large-scale coastal aquaculture of oysters and seaweed on the marine CO<sub>2</sub> budget with the hope of finding methods of carbon removal from coastal waters.

## MATERIALS AND METHODS

### Study site

Daya Bay, located in Guangdong Province, southern China, is a 600 km<sup>2</sup> semi-enclosed embayment in the northeast South China Sea (Fig. 1). The average water depth is 10 m (range: 6–20 m). The annual mean air temperature is 22°C. The minimum sea surface temperature occurs in winter (15°C) and the maximum in summer and fall (30°C). The bay is one of the most intensive culture areas in China. The Portuguese oyster *Crassostrea angulata* is the main cultured bivalve species. Suspension aquaculture of *C. angulata* has been practiced for over 3 decades, with an estimated standing stock of 6.6 × 10<sup>4</sup> t in 2016. The seaweed *Gracilaria lemaneiformis* is another important cultured species, with a production of ca. 27 × 10<sup>4</sup> t in China in 2015 (China Bureau of Fisheries 2016).

### Estimation of calcification and respiration rate of *C. angulata*

Experimental *C. angulata* were collected from Daya Bay in April 2016 and taken to the laboratory in a temperature-controlled case within 1 h. After arrival, animals were disinfected, and any visible fouling organisms on shell surfaces were cleaned by

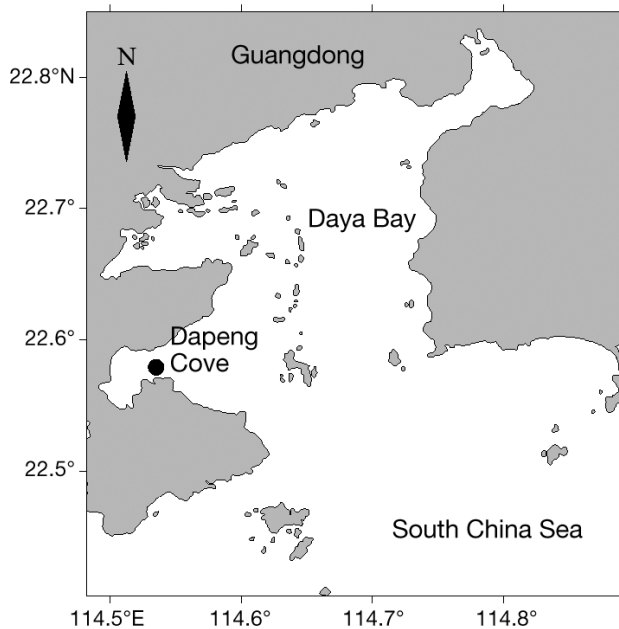


Fig. 1. Location of the study area (●) in Daya Bay

washing with filtered seawater. Approximately 30 oysters of similar sizes (7–8 cm shell height) were acclimatized to laboratory conditions for 1 wk in a 50 l tank with aerated seawater. During acclimation, seawater was changed once per day and oysters were fed daily with  $2 \times 10^4$  cell  $\text{ml}^{-1}$  *Chaetoceros* sp. at 08:00 h. During the trial, temperature ( $T$ ), salinity ( $S$ ), and pH were  $22 \pm 1^\circ\text{C}$ ,  $30.8 \pm 0.1$ , and  $8.03 \pm 0.02$ , respectively.

At the end of the acclimatization period, the oysters were placed in closed 20 l transparent polyethylene plastic mesocosms filled with seawater. The mesocosms were hung from a suspended longline so that the experimental oysters were at a depth of ca. 2 m, corresponding to the routine culture depth for oysters. A factorial design was used to test the effects of 3 stocking densities: i.e. low, medium, and high (ca. 1, 5, and 10 g (fresh weight, FW) oyster  $\text{l}^{-1}$ , respectively) on water pH, dissolved oxygen (DO), total alkalinity (TA), DIC, and carbonate ion ( $\text{CO}_3^{2-}$ ) concentrations. Water samples were taken at 0 and 4 h.  $T$  and  $S$  were measured using a multi-parameter water quality meter (YSI Professional Plus 6600, Yellow Springs Instrument Company). pH was measured using a pH meter (Thermo Scientific Orion 320P-01, Thermo Electron Corporation) calibrated on the US National Bureau of Standards scale. The precision of pH measurements was  $\pm 0.01$  pH units. Oxygen concentrations were determined by the Winkler method (Strickland & Parson 1972). TA was measured using an 848 Titrino plus automatic titrator (Metrohm) on

100 ml GF/F filtered samples. The accuracy of measurements was  $\pm 2 \mu\text{mol l}^{-1}$ . DIC and  $\text{CO}_3^{2-}$  concentrations were computed from  $T$ ,  $S$ , pH, and TA using the  $\text{CO}_2\_SYS\_XLS$  calculation program (Pierrot et al. 2006). The dissociation constants for carbonic acid ( $K_1$ ,  $K_2$ ) were from Mehrbach et al. (1973) as refitted by Dickson & Millero (1987), the dissociation constant for bisulfate ion ( $K_{\text{HSO}_4}$ ) was obtained from Dickson (1990), and the dissociation constant for boric acid ( $K_B$ ) was from Uppstrom (1974).

$\text{CO}_2$  respiratory rate ( $R$ ;  $\mu\text{mol FW g}^{-1} \text{h}^{-1}$ ) can be expressed as follows:

$$\Delta\text{DIC} = \text{DIC}_f - \text{DIC}_i \quad (1)$$

$$R = \frac{(2\Delta\text{DIC} + \text{TA}_i - \text{TA}_f) \times V}{2 \times t \times M} \quad (2)$$

where  $\Delta\text{DIC}$  is the net change in DIC concentration ( $\mu\text{mol l}^{-1}$ ), which was caused by the interaction between calcification and respiration.  $\text{DIC}_i$  and  $\text{DIC}_f$  were the initial and final DIC concentration ( $\mu\text{mol l}^{-1}$ ), respectively.  $V$  is the incubation chamber volume (l),  $t$  is the experimental time (h), and  $M$  is the fresh weight of experimental oyster (g).  $\text{TA}_i$  and  $\text{TA}_f$  were the initial and final TA concentration ( $\mu\text{mol l}^{-1}$ ), respectively.

For  $\text{CaCO}_3$  production, 30 oysters were sampled, and the dry weight (DW) of oyster shell was determined by drying at  $80^\circ\text{C}$  till constant weight ( $\pm 0.01$  g). Bivalve shells largely consist (95%) of  $\text{CaCO}_3$ , and the remaining 5% are made up by magnesium,  $\beta$ -chitin, and various glycoproteins (Gouletquer & Wolowicz 1989). Shell DWs were corrected accordingly. The calcimass ( $\text{g CaCO}_3 \text{m}^{-2}$ ) was estimated by the shell DWs per  $\text{m}^2$  and its  $\text{CaCO}_3$  content. Dry tissue weight was calculated for each individual using the ash-free dry weight (AFDW) method: oyster soft tissue was dried at  $80^\circ\text{C}$  (72 h) and then ashed at  $500^\circ\text{C}$  (4 h), with tissue weight computed as the difference between the 2 weights.

The  $\text{CO}_2$  released from *C. angulata* respiration was estimated using: (1) the relation established by Schwinghamer et al. (1986):  $\log_{10} R = 0.367 + 0.993 \log_{10} P$ , where  $P$  and  $R$  are production and respiration ( $\text{kcal m}^{-2} \text{yr}^{-1}$ ), respectively; (2) the  $22.79 \text{ J mg AFDW}^{-1}$  conversion factor for bivalve energetic content (Rumohr et al. 1987); (3) the conversion  $1 \text{ J} = 0.239 \text{ cal}$  (Peters 1983); and (4) the conversion  $1 \text{ g C} = 11.4 \text{ kcal}$  (Chauvaud et al. 2003).

The monthly ratio of  $\text{CO}_2$  released to  $\text{CaCO}_3$  precipitated ( $\Psi$ ) was estimated as a function of the water temperature, measured with a YSI meter.  $\text{CO}_2$  fluxes due to calcification were calculated using a  $\Psi$  value

estimated by computing the temperature according to:  $\Psi = 0.8-8.3 \times 10^{-3}T$ , where  $T$  is the water temperature ( $^{\circ}\text{C}$ ), and assuming  $p\text{CO}_2 = 350 \mu\text{atm}$  (Fankignoulle et al. 1994).

### Deployment of *in situ* mesocosm experiment

The *in situ* mesocosm experiment was carried out from 20 to 22 April 2016 in Daya Bay ( $22^{\circ}34' \text{N}$ ,  $114^{\circ}32' \text{E}$ ) (Fig. 1). The cylindrical mesocosms (1 m diameter  $\times$  1.5 m height) were made from transparent polyethylene plastic and were hung on suspended longlines with the top ca. 2.0 m below the water surface. Fifteen cylindrical mesocosms (450 l) were deployed over 24 h periods and consisted of 5 treatments each with 3 replicates (Table 1). One treatment with only seawater served as the control (C), the second treatment contained only oysters (oyster only, O), and the other 3 treatments were co-culture systems, with 3 oyster:seaweed ratios, i.e. 8:1, 4:1, and 2:1 (based on FW of oyster and seaweed, referred to as OS\_8:1, OS\_4:1 and OS\_2:1, respectively). After filling with natural seawater, the mouths of the mesocosms were tied using ropes. A pipe was placed in the tied site and maintained for 24 h to keep DIC concentration in equilibrium with air due to water mixing with air. Oysters and seaweed were coiled into 100-cm-long ropes and placed in the mesocosms. The ropes were suspended using thin ropes and tied to the mouth of the mesocosm, such that the seaweed thalli were positioned vertically around the oysters. The mesocosms were immobilized using a set of ropes connected top-side to a float and submerged under the water surface.

The experiment began at 10:00 h and lasted for 24 h. Water  $T$ ,  $S$ , DO, pH, and TA were measured at the beginning and end of the experiment. Parameters for the seawater DIC system and aqueous  $p\text{CO}_2$  were calculated by the CO2\_SYS\_XLS calculation program (Pierrot et al. 2006).

The sea-air  $\text{CO}_2$  fluxes ( $F_{\text{CO}_2}$ ) were calculated based on the following equation:  $F_{\text{CO}_2} = k \times \alpha \times \Delta p\text{CO}_2$ , where  $k$  ( $\text{cm h}^{-1}$ ) is the gas exchange coefficient of  $\text{CO}_2$ . We computed  $k$  using the parameterization given by Wanninkhof & McGillis (1999) that uses short-term winds  $k = 0.0283u_{10}^3 (Sc/660)^{-1/2}$ .  $u_{10}$  stands for the wind speed at a 10 m height from the water surface level ( $\text{m s}^{-1}$ ) and  $Sc$  is the Schmidt number calculated according to the relationship proposed by Wanninkhof (1992).  $\alpha$  ( $\text{mol kg}^{-1} \text{atm}^{-1}$ ) is the solubility coefficient of  $\text{CO}_2$  calculated after Weiss (1974).  $\Delta p\text{CO}_2$  is the  $p\text{CO}_2$  difference between

Table 1. Overview of the co-culture systems in the 5 treatments, with oysters and seaweed retained in mesocosms (g fresh weight per mesocosm, mean  $\pm$  SD)

Treatment (abbreviation)	Oyster	Seaweed
Control (C)	–	–
Oyster (O)	4402 $\pm$ 141	–
Oyster:seaweed 8:1 (OS_8:1)	4418 $\pm$ 126	583 $\pm$ 27
Oyster:seaweed 4:1 (OS_4:1)	4750 $\pm$ 102	1103 $\pm$ 68
Oyster:seaweed 2:1 (OS_2:1)	4631 $\pm$ 150	2210 $\pm$ 115

surface seawater and the atmosphere. In this study, the value of atmospheric  $p\text{CO}_2$  was downloaded from [www.cmdl.noaa.gov](http://www.cmdl.noaa.gov) (National Oceanographic and Atmospheric Administration, NOAA, Climate and Meteorological Diagnostics Laboratory) and corrected for water vapor pressure (Takahashi et al. 2002). Positive magnitudes of  $F_{\text{CO}_2}$  indicate a flux from water to air and vice versa.

The net oxygen production rate by *G. lemaneiformis* in co-culture systems was determined based on the DO concentrations in experimental mesocosms, as:

$$\text{Net oxygen production rate } (\mu\text{mol O}_2 \text{ g}^{-1} \text{ FW h}^{-1}) = (C_{\text{os}} - C_{\text{o}}) \times V/\text{FW}/t, \quad (3)$$

where  $C_{\text{os}}$  and  $C_{\text{o}}$  are the DO concentrations ( $\mu\text{mol l}^{-1}$ ) of oyster-seaweed co-culture mesocosms and oyster-only mesocosm, respectively, at the end of the experiment.  $V$  is the volume of mesocosms (l), FW is the fresh weight of *G. lemaneiformis* (g), and  $t$  is the duration of the experiment (h).

### Statistical analysis

Data were analyzed by 1-way ANOVA. All data were graphically assessed for normality and homogeneity of residuals (Faraway 2002). When overall differences were significant at the 0.05 level, Tukey's HSD multiple range test was used to compare the mean values of individual groups. Data are reported as means  $\pm$  SE ( $n = 3$ ). All statistical tests were performed using SPSS 17.0 for Windows.

## RESULTS

### Calcification and respiration rates of *Crassostrea angulata*

As shown in Fig. 2, after a 4 h incubation, the seawater pH, TA, and  $\text{CO}_3^{2-}$  concentrations in the closed

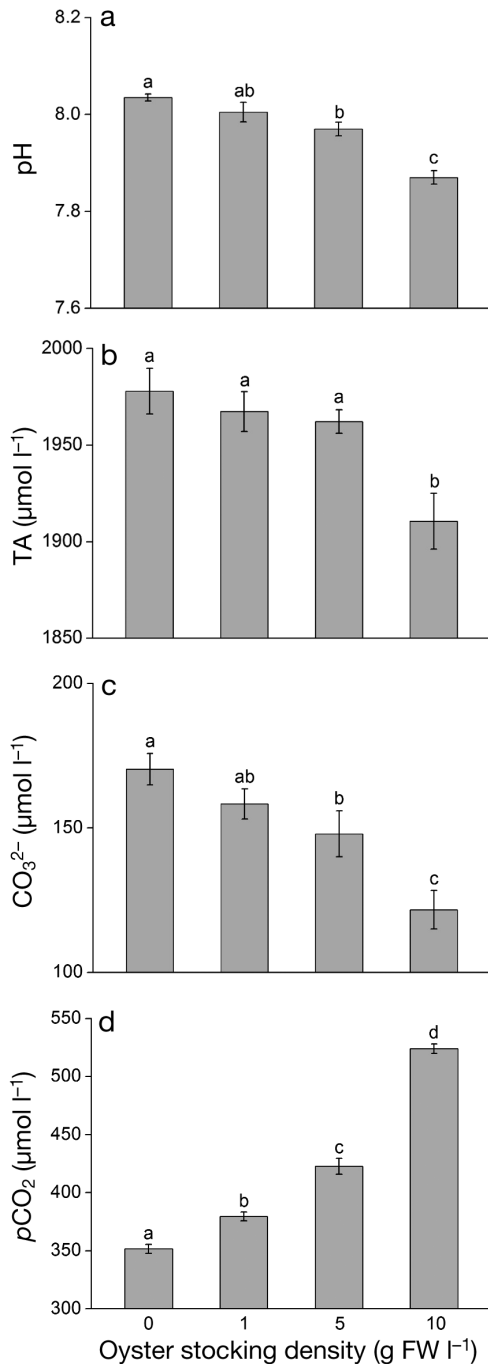


Fig. 2. Effects of Portuguese oyster *Crassostrea angulata* stocking density on (a) pH, (b) total alkalinity (TA), (c)  $\text{CO}_3^{2-}$ , and (d)  $p\text{CO}_2$  of seawater in the closed mesocosm systems. Bars with different lowercase letters are significantly different (ANOVA,  $p < 0.05$ )

mesocosms gradually decreased with increasing oyster stocking density. Values in the highest stocking density group were significantly lower than the other groups ( $p < 0.05$ ) (Fig. 2a–c). The  $p\text{CO}_2$  followed the converse general pattern, and values were signifi-

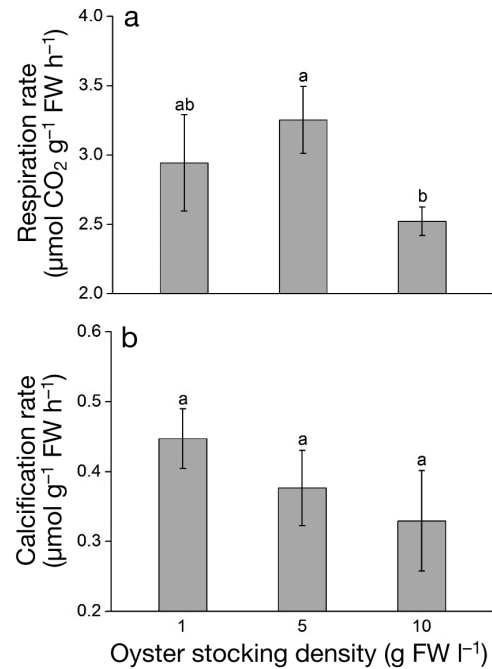


Fig. 3. (a) Respiration and (b) calcification rate of the Portuguese oyster *Crassostrea angulata* at different stocking densities. Bars with different lowercase letters are significantly different (ANOVA,  $p < 0.05$ )

cantly different between each treatment ( $p < 0.05$ ) (Fig. 2d).

The lowest and highest respiration rates of oysters were found in the high- and medium-density groups, respectively, and were significantly different ( $p < 0.05$ ) (Fig. 3a). Calcification rates of oysters decreased with increasing stocking density, but no significant difference occurred ( $p > 0.05$ ) (Fig. 3b).

The total  $\text{CaCO}_3$  production by the *C. angulata* population in Daya Bay was estimated to be ca.  $2150 \text{ g CaCO}_3 \text{ m}^{-2} \text{ yr}^{-1}$ . *C. angulata* sequestered  $258 \text{ g C m}^{-2} \text{ yr}^{-1}$  for shell formation. The ratio of  $\text{CO}_2$  released to  $\text{CaCO}_3$  precipitated ( $\Psi$ ) ranged from 0.54 to 0.65 and varied monthly with water temperature variation in Daya Bay (Table 2).  $\text{CO}_2$  released due to calcification and respiration was 153 and  $349 \text{ g C m}^{-2} \text{ yr}^{-1}$ , respectively.

#### Variations of seawater pH and TA in different mesocosms

As shown in Fig. 4a, the seawater pH differed significantly among different mesocosms ( $p < 0.05$ ). pH was lowest in the oyster-only treatment, significantly lower than that of the control and the oyster–seaweed co-culture groups ( $p < 0.05$ ). pH gradually increased

Table 2. Mean monthly water temperature ( $T$ , °C) and corresponding molar ratio ( $\Psi$ ) in Daya Bay

Month	$T$	$\Psi$
January	18.3	0.65
February	18.2	0.65
March	18.7	0.64
April	22.4	0.61
May	24.5	0.59
June	29.0	0.56
July	31.3	0.54
August	29.7	0.55
September	29.5	0.55
October	28.3	0.56
November	27.6	0.57
December	19.2	0.64
Mean $\pm$ SD	24.7 $\pm$ 5.1	0.59 $\pm$ 0.04

with increasing seaweed density in co-culture treatments, and was significantly higher than that of the control ( $p < 0.05$ ). TA was highest in the control, significantly higher than that of the other groups ( $p < 0.05$ ) (Fig. 4b). TA values in OS\_4:1 and OS\_2:1 co-culture groups were significantly lower than that of the other groups ( $p < 0.05$ ) (Fig. 4b).

#### Variations of seawater DIC systems and $p\text{CO}_2$ in different mesocosms

After 24 h incubation, DIC,  $\text{HCO}_3^-$ , and  $\text{CO}_2$  concentrations and  $p\text{CO}_2$  showed similar trends among treatments (Fig. 5a,b,d,e). The highest values were found in the oyster-only group, and then continuously decreased with increasing seaweed density in co-culture groups. Co-culture with seaweed lead to a significant decrease of  $\text{CO}_2$  concentration and  $p\text{CO}_2$  ( $p < 0.05$ ). The degree of reduction was positively correlated with the seaweed density (Fig. 5d,e).  $\text{CO}_3^{2-}$  concentration followed the converse general pattern to  $\text{HCO}_3^-$  concentrations (Fig. 5c).

#### Variations of air–sea $\text{CO}_2$ flux in different mesocosms

The air–sea  $\text{CO}_2$  flux ( $F_{\text{CO}_2}$ ) in the oyster-only treatment group had a high and positive value ( $110.4 \pm 10.5 \text{ mmol m}^{-2} \text{ d}^{-1}$ ), representing a  $\text{CO}_2$  source to the atmosphere, and was significantly higher than in the other groups ( $p < 0.05$ ). In contrast, the negative values (from  $-[8.4 \pm 0.7]$  to  $-[33.6 \pm 4.0] \text{ mmol m}^{-2} \text{ d}^{-1}$ ) in the control and co-culture groups represent a  $\text{CO}_2$  sink from the atmosphere to the sea, where the

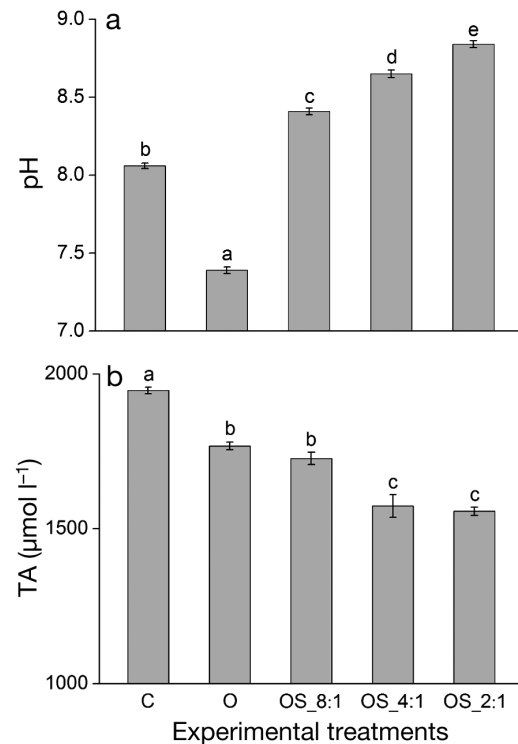


Fig. 4 (a) Seawater pH and (b) total alkalinity (TA) of oyster *Crassostrea angulata* monoculture and oyster–seaweed *Gracilaria lemaneiformis* co-culture mesocosms. C: control; O: oyster only; OS\_8:1, OS\_4:1, and OS\_2:1: oyster co-cultured with seaweed at ratios of 8:1, 4:1, and 2:1, respectively. Bars with different lowercase letters are significantly different (ANOVA,  $p < 0.05$ )

degree of  $\text{CO}_2$  sink was proportional to seaweed stocking density; there was no significant difference in  $F_{\text{CO}_2}$  between the oyster–seaweed (OS\_8:1) co-culture and the control ( $p < 0.05$ ), but the  $F_{\text{CO}_2}$  values in OS\_4:1 and OS\_2:1 groups were significantly lower than that in the control ( $p > 0.05$ ) (Fig. 6).

#### Oxygen production rate of *Gracilaria lemaneiformis* in co-culture mesocosms

In oyster–seaweed co-culture mesocosms, although the DO was mainly produced by *G. lemaneiformis*, the phytoplankton also produced some oxygen. Therefore, the oxygen concentration in oyster–seaweed co-culture mesocosms (Fig. 7) was refitted by the oxygen concentration in the control and oyster-only mesocosms. The net oxygen production rates in the low- (OS\_8:1) and medium- (OS\_4:1) seaweed-density treatments were significantly ( $p < 0.05$ ) higher than that in the high-seaweed-density group (OS\_2:1).

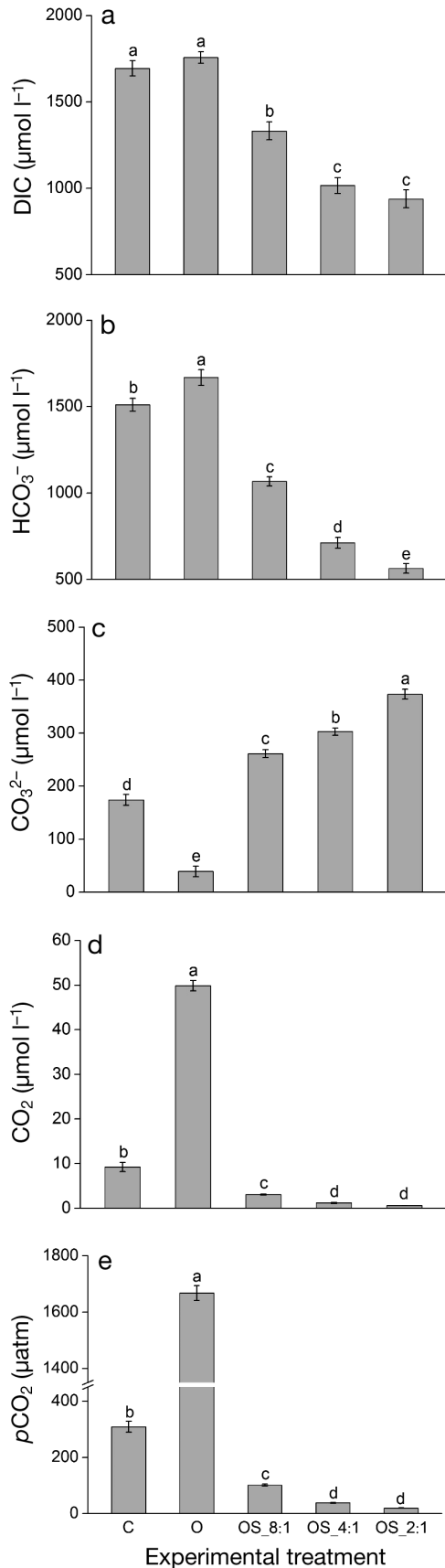


Fig. 5 Seawater concentrations of (a) dissolved inorganic carbon (DIC), (b)  $\text{HCO}_3^-$ , (c)  $\text{CO}_3^{2-}$ , (d) aqueous  $\text{CO}_2$ , and (e)  $p\text{CO}_2$  in oyster *Crassostrea angulata* monocultures and oyster–seaweed *Gracilaria lemaneiformis* co-culture mesocosms (see Table 1 for treatments). Bars with different lowercase letters are significantly different (ANOVA,  $p < 0.05$ )

## DISCUSSION

The results of the present study indicated that the oyster *Crassostrea angulata* cultivated in Daya Bay seems to be a  $\text{CO}_2$  generator, as  $p\text{CO}_2$  increased in oyster-only culture mesocosms. Oyster harvesting sequesters ca.  $258 \text{ g C m}^{-2} \text{ yr}^{-1}$  due to shell formation in Daya Bay. In contrast, the  $\text{CO}_2$  fluxes due to respiration and calcification were ca.  $349$  and  $153 \text{ g C m}^{-2} \text{ yr}^{-1}$ , respectively, accounting for 69.5% and 30.5% of the total  $\text{CO}_2$  fluxes ( $502 \text{ g C m}^{-2} \text{ yr}^{-1}$ ), respectively.

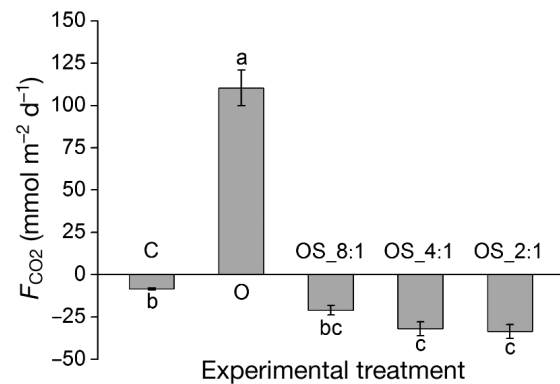


Fig. 6. Variations in air-sea  $\text{CO}_2$  flux ( $F_{\text{CO}_2}$ ) among different experimental mesocosms. C: control; O: oyster only; OS\_8:1, OS\_4:1, and OS\_2:1: oyster co-cultured with seaweed at ratios of 8:1, 4:1, and 2:1, respectively. Bars with different lowercase letters are significantly different (ANOVA,  $p < 0.05$ )

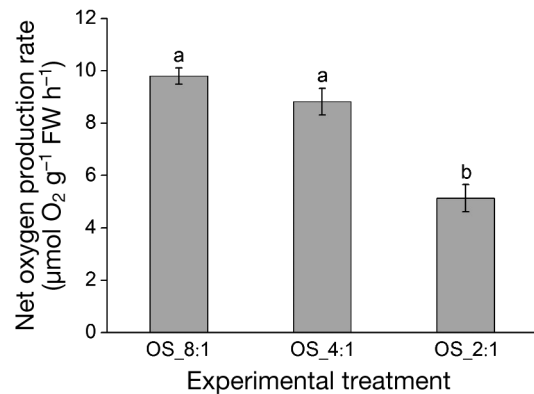


Fig. 7. Net oxygen production rates of seaweed *Gracilaria lemaneiformis* in oyster *Crassostrea angulata*-seaweed co-culture mesocosms (see Table 1 for treatments). Bars with different lowercase letters are significantly different (ANOVA,  $p < 0.05$ )

This result indicated that total carbon fluxes were mainly influenced by respiration, but the contribution of calcification was not negligible. Based on the balance between  $\text{CaCO}_3$  sequestration and  $\text{CO}_2$  release, the *C. angulata* populations in Daya Bay increase  $\text{CO}_2$  release to the atmosphere in coastal ecosystems. Moreover, our measurements may have underestimated the overall contribution of *C. angulata* to  $\text{CO}_2$  fluxes, since we have not considered the rate of carbonate dissolution of shells that remained in the system after oyster death.

During the 4 h incubation, the DO concentrations in all mesocosms were above  $4 \text{ mg l}^{-1}$ . This level was not likely to induce stress to the oysters (Diaz & Rosenberg 2008). The reduced respiration rate by oysters in the high-density group (Fig. 3a) might be a strategy to cope with variability in seawater pH and the ability to adapt to seawater acidification (Guppy & Withers 1999, Langenbuch & Pörtner 2004).

Seawater  $\text{CO}_3^{2-}$  can affect the ability of calcifying organisms to precipitate  $\text{CaCO}_3$  (Gazeau et al. 2007, Zhang et al. 2011, Dineshram et al. 2013, Li et al. 2013, Mos et al. 2015, McGrath et al. 2016). However, in the present experiment, no significant differences in calcification rate were found among the different density treatments (Fig. 3b), although the  $\text{CO}_3^{2-}$  was lower in the high-density group (Fig. 2c). This might indicate that the  $\text{CO}_3^{2-}$  deficiency stress was not severe enough to depress calcification. Therefore, further studies with longer incubation times and/or larger biomass of oysters are needed to produce more severe acidification stress.

According to the calcification rate and culture density of oysters, the mean  $\text{CaCO}_3$  production by *C. angulata* population in Daya Bay is ca.  $2150 \text{ g CaCO}_3 \text{ m}^{-2} \text{ yr}^{-1}$ . This is higher than that of the oyster *Crassostrea gigas* ( $134 \text{ g CaCO}_3 \text{ m}^{-2} \text{ yr}^{-1}$ ) in Brest Bay (Lejart et al. 2012). Varying results could be due to species-specific differences. Lejart et al. (2012) studied the natural populations of *C. gigas*, which inhabit the intertidal zone going through 14 h of underwater respiration and calcification, and 10 h aerial respiration each day with the changing tides. However, *C. angulata* in Daya Bay were cultured under constant immersion conditions. Therefore, *C. angulata* has a longer period of calcification to produce higher amounts of  $\text{CaCO}_3$ .

$\text{CO}_2$  released during  $\text{CaCO}_3$  precipitation of oysters in Daya Bay represented about 30.5% of the total  $\text{CO}_2$  production. This result was consistent with previous findings, e.g. 30% for *Ophiothrix fragilis* (Migné et al. 1998), 33% for *Potamocorbula amurensis* (Chauvaud et al. 2003), and 23–26% for *Acroc-*

*nida brachiata* (Davoult et al. 2009) in the eastern English Channel. Therefore, although total carbon fluxes were mainly influenced by underwater respiration, there is a contribution from calcification that should not be neglected.

*Gracilaria lemaneiformis* can use both  $\text{CO}_2$  and  $\text{HCO}_3^-$  for photosynthesis. In the oyster–seaweed co-culture mesocosm, the  $\text{HCO}_3^-$  concentration was significantly higher in the OS\_4:1 group than in the OS\_2:1 group, but there was no significant difference in  $\text{CO}_2$  concentration (Fig. 5). This phenomenon was consistent with the findings of Raven et al. (2014) and Axelsson et al. (2000), who reported that in seawater of pH 8.0 and above, the principal species of DIC in the medium is  $\text{HCO}_3^-$ , but the active transport of  $\text{HCO}_3^-$  needs higher energy than passive  $\text{CO}_2$  diffusion. Hence, it is reasonable that *G. lemaneiformis* has a higher affinity for  $\text{CO}_2$  than  $\text{HCO}_3^-$ , which lead to a preferential  $\text{CO}_2$  exhaust over  $\text{HCO}_3^-$ .

In the present study, calcification and respiration by oysters occurred over the duration of 24 h in the closed mesocosm system, while photosynthesis by *G. lemaneiformis* occurred only during the daytime when there is light. The  $\text{CO}_2:\text{CO}_3^{2-}$  ratio and pH of the seawater in the co-culture system would depend on the balance between the photosynthesis rate by the seaweed and the respiration rate and calcification rate of the oysters (Menéndez et al. 2001, Zhang et al. 2012). Seawater  $p\text{CO}_2$  and  $\text{CO}_2:\text{CO}_3^{2-}$  ratios decreased in all oyster–seaweed co-culture systems, indicating that there was stronger  $\text{CO}_2$  uptake by *G. lemaneiformis* than  $\text{CO}_2$  release from *C. angulata*, leading to a net uptake of  $\text{CO}_2$  from the atmosphere into the seawater. Meanwhile, we found that the net oxygen production rate of *G. lemaneiformis* in the OS\_2:1 treatment was significantly decreased compared with that in the OS\_8:1 and OS\_4:1 groups (Fig. 7). As the primary production of seaweed is carbon limited, the carbon-saturated maximum photosynthesis of *G. lemaneiformis* would drastically reduce when it was 'starved' of DIC (Han et al. 2013). Since the numbers of oysters among the 3 treatments were almost the same, the decreased oxygen production rate i.e. the photosynthesis rate of *G. lemaneiformis* in the OS\_2:1 group was probably due to a carbon limitation. Thus, there could be an evident interspecies mutual benefit for both *C. angulata* and *G. lemaneiformis* in the co-culture system. Based on the results of the present study, we suggest that the 2 species are co-cultured at a ratio of ca. 4:1 (fresh weight) for efficient utilization of seawater DIC by *G. lemaneiformis*, and further to increase the ocean  $\text{CO}_2$  sink.



In conclusion, the physiological activities of *C. angulata* lead to a shift in the seawater DIC system equilibria towards higher CO<sub>2</sub>, lower pH, and lower CO<sub>3</sub><sup>2-</sup> concentration, and subsequently are affected by this shift. Seaweed *G. lemaneiformis* could act as an efficient sink for CO<sub>2</sub>. Incorporation of seaweed into oyster aquaculture can be helpful in eliminating DIC release from *C. angulata*. There could be complex interspecies effects between *C. angulata* and *G. lemaneiformis*. The beneficial effects of an integrated multi-trophic aquaculture system on seawater carbon budget and air–sea CO<sub>2</sub> fluxes should be determined based on an ecosystem approach.

**Acknowledgements.** This study was supported by Major State Basic Research Development Program of China (973 Program, 2015CB452904, 2015CB4529001), National Natural Science Foundation of China (31602183, 41106088), Special Fund of Basic Research for Central non-profit Scientific Research Institutes (2016YD02, 2014A01YY03), Guangdong Natural Science Foundation (2014A030310331), Project of Science and Technology of Guangdong Province (2016A020222024, 2014B030301064), the Key Laboratory of South China Sea Fishery Resources Development and Utilization, Ministry of Agriculture (LSF2014-05).

#### LITERATURE CITED

- ✦ Axelsson L, Mercado JM, Figueroa FL (2000) Utilization of HCO<sub>3</sub><sup>-</sup> at high pH by the brown macroalga *Laminaria saccharina*. *Eur J Phycol* 35:53–59
- ✦ Chauvaud L, Thompson JK, Cloern JE, Thouzeau G (2003) Clams as CO<sub>2</sub> generators: the *Potamocorbula amurensis* example in San Francisco Bay. *Limnol Oceanogr* 48: 2086–2092
- China Bureau of Fisheries (2016) China Fisheries Yearbook in 2016. Agricultural Press of China, Beijing
- ✦ Davoult D, Harlay J, Gentil F (2009) Contribution of a dense population of the brittle star *Acrocnida brachiata* (Montagu) to the biogeochemical fluxes of CO<sub>2</sub> in a temperate coastal ecosystem. *Estuaries Coasts* 32:1103–1110
- ✦ Diaz RJ, Rosenberg R (2008) Spreading dead zones and consequences for marine ecosystems. *Science* 321:926–929
- ✦ Dickson AG (1990) Standard potential of the reaction: AgCl(s) + 1/2 H<sub>2</sub>(g) = Ag(s) + HCl(aq), and the standard acidity constant of the ion HSO<sub>4</sub><sup>-</sup> in synthetic seawater from 273.15 to 318.15 K. *J Chem Thermodyn* 22:113–127
- Dickson AG (2010) The carbon dioxide system in seawater: equilibrium chemistry and measurements. In: Riebesell U, Fabry VJ, Hansson L, Gattuso JP (eds) Guide to best practices for ocean acidification research and data reporting. Publications Office of the European Union, Luxembourg, p 17–52
- ✦ Dickson AG, Millero FJ (1987) A comparison of the equilibrium constants for the dissociation of carbonic acid in seawater media. *Deep-Sea Res* 34:1733–1743
- ✦ Dineshram R, Thiyagarajan V, Lane A, Yu ZN, Shu X, Leung PTY (2013) Elevated CO<sub>2</sub> alters larval proteome and its phosphorylation status in the commercial oyster, *Crassostrea hongkongensis*. *Mar Biol* 160:2189–2205
- ✦ Fang JG, Zhang J, Xiao T, Huang DJ, Liu SM (2016) Integrated multi-trophic aquaculture (IMTA) in Sanggou Bay, China. *Aquacult Environ Interact* 8:201–205
- Faraway JJ (2002) Practical regression and ANOVA using R. <http://cran.r-project.org/doc/contrib/Faraway-PRA.pdf> (accessed on 20 January 2014)
- ✦ Frankignoulle M, Canon C, Gattuso JP (1994) Marine calcification as a source of carbon dioxide: positive feedback of increasing atmospheric CO<sub>2</sub>. *Limnol Oceanogr* 39:458–462
- ✦ Frankignoulle M, Pichon M, Gattuso JP (1995) Aquatic calcification as a source of carbon dioxide. In: Beran MA (ed) Carbon sequestration in the biosphere. NATO ASI Series (Series I: Global Environmental Change) Vol 33. Springer, Berlin, p 265–271
- ✦ Gazeau F, Quiblier C, Jansen JM, Gattuso J, Middelburg JJ, Heip CHR (2007) Impact of elevated CO<sub>2</sub> on shellfish calcification. *Geophys Res Lett* 34:L07603
- ✦ Gouilletquer P, Wolowicz M (1989) The shell of *Cardium edule*, *Cardium glaucum* and *Ruditapes philippinarum*: organic content, composition and energy value, as determined by different methods. *J Mar Biol Assoc UK* 69: 563–572
- Guppy M, Withers P (1999) Metabolic depression in animals: physiological perspectives and biochemical generalizations. *Biol Rev Camb Philos Soc* 74:1–40
- ✦ Han TT, Jiang ZJ, Fang JG, Zhang JH and others (2013) Carbon dioxide fixation by the seaweed *Gracilaria lemaneiformis* in integrated multi-trophic aquaculture with the scallop *Chlamys farreri* in Sanggou Bay, China. *Aquacult Int* 21:1035–1043
- ✦ Ho M, Carpenter RC (2017) Differential growth responses to water flow and reduced pH in tropical marine macroalgae. *J Exp Mar Biol Ecol* 491:58–65
- ✦ Jiang ZJ, Fang JG, Han TT, Mao YZ, Li JQ, Du MR (2014) The role of *Gracilaria lemaneiformis* in eliminating the dissolved inorganic carbon released from calcification and respiration process of *Chlamys farreri*. *J Appl Phycol* 26:545–550
- ✦ Langenbuch M, Pörtner HO (2004) High sensitivity to chronically elevated CO<sub>2</sub> levels in a eurybathic marine sipunculid. *Aquat Toxicol* 70:55–61
- ✦ Lejart M, Clavier J, Chauvaud L, Hily C (2012) Respiration and calcification of *Crassostrea gigas*: contribution of an intertidal invasive species to coastal ecosystem CO<sub>2</sub> fluxes. *Estuaries Coasts* 35:622–632
- ✦ Lerman A, Mackenzie FT (2005) CO<sub>2</sub> air–sea exchange due to calcium carbonate and organic matter storage, and its implications for the global carbon cycle. *Aquat Geochem* 11:345–390
- ✦ Li J, Jiang Z, Zhang J, Qiu JW, Du M, Bian D, Fang J (2013) Detrimental effects of reduced seawater pH on the early development of the Pacific abalone. *Mar Pollut Bull* 74: 320–324
- ✦ Mackenzie FT, Andersson AJ (2013) The marine carbon system and ocean acidification during Phanerozoic time. *Geochem Perspect* 2:1–3
- ✦ Martin S, Thouzeau G, Chauvaud L, Jean F, Guerin L, Clavier J (2006) Respiration, calcification, and excretion of the invasive slipper limpet, *Crepidula fornicata* L.: implications for carbon, carbonate, and nitrogen fluxes in affected areas. *Limnol Oceanogr* 51:1996–2007
- ✦ McGrath T, McGovern E, Cave RR, Kivimäe C (2016) The inorganic carbon chemistry in coastal and shelf waters around Ireland. *Estuaries Coasts* 39:27–39

- Mehrbach C, Culberson CH, Hawley JE, Pytkowicz RM (1973) Measurement of the apparent dissociation constants of carbonic acid in seawater at atmospheric pressure. *Limnol Oceanogr* 18:897–907
- Menéndez M, Martínez M, Comín FA (2001) A comparative study of the effect of pH and inorganic carbon resources on the photosynthesis of three floating macroalgae species of a Mediterranean coastal lagoon. *J Exp Mar Biol Ecol* 256:123–136
- Migné A, Davoult D, Gattuso JP (1998) Calcium carbonate production of a dense population of the brittle star *Ophiothrix fragilis* (Echinodermata: Ophiuroidea): role in the carbon cycle of a temperate coastal ecosystem. *Mar Ecol Prog Ser* 173:305–308
- Millero FJ (1995) Thermodynamics of the carbon dioxide system in the oceans. *Geochim Cosmochim Acta* 59: 661–677
- Mistri M, Munari C (2013) The invasive bag mussel *Arcuatula senhousia* is a CO<sub>2</sub> generator in near-shore coastal ecosystems. *J Exp Mar Biol Ecol* 440:164–168
- Mos B, Byrne M, Cowden KL, Dworjanyn SA (2015) Biogenic acidification drives density-dependent growth of a calcifying invertebrate in culture. *Mar Biol* 162: 1541–1558
- Munari C, Rossetti E, Mistri M (2013) Shell formation in cultivated bivalves cannot be part of carbon trading systems: a study case with *Mytilus galloprovincialis*. *Mar Environ Res* 92:264–267
- Peters RH (1983) *The ecological implications of body size*. Cambridge University Press, New York, NY
- Pierrot D, Lewis E, Wallace DWR (2006) MS Excel program developed for CO<sub>2</sub> system calculations. ORNL/CDIAC-105 [R]. Carbon dioxide Information Analysis Center, Oak Ridge National Laboratory, US Department of Energy, Oak Ridge, TN
- Raven JA, Beardall J, Giordano M (2014) Energy costs of carbon dioxide concentrating mechanisms in aquatic organisms. *Photosynth Res* 121:111–124
- Rumohr H, Brey T, Ahkar S (1987) A compilation of biometric conversion factors for benthic invertebrates of the Baltic Sea. *The Baltic Marine Biologists Publications* Vol 9
- Scanes E, Parker LM, O'Connor WA, Stapp LS, Ross PM (2017) Intertidal oysters reach their physiological limit in a future high-CO<sub>2</sub> world. *J Exp Biol* 220:765–774
- Schwinghamer P, Hargrave B, Peer D, Hawkins CM (1986) Partitioning of production and respiration among size groups of organisms in an intertidal benthic community. *Mar Ecol Prog Ser* 31:131–142
- Strickland JD, Parson TR (1972) *A practical handbook of seawater analysis*, Vol 167, 2nd edn. *Bulletin of the Fish Research Board of Canada*, Ottawa, p 310
- Takahashi T, Sutherland SC, Sweeney C, Poisson A and others (2002) Global sea–air CO<sub>2</sub> flux based on climatological surface ocean pCO<sub>2</sub>, and seasonal biological and temperature effects. *Deep Sea Res II* 49:1601–1622
- Uppstrom LR (1974) The boron/chlorinity ratio of deep-sea water from the Pacific Ocean. *Deep-Sea Res* 21:161–162
- Veron JEN (2011) Ocean acidification and coral reefs: an emerging big picture. *Diversity* 3:262–274
- Wanninkhof R (1992) Relationship between wind speed and gas exchange over the ocean. *J Geophys Res* 97: 7373–7382
- Wanninkhof R, McGillis WM (1999) A cubic relationship between gas transfer and wind speed. *Geophys Res Lett* 26:1889–1892
- Weiss RF (1974) Carbon dioxide in water and seawater: the solubility of a non-ideal gas. *Mar Chem* 2:203–215
- Yu ZH, Jiang T, Xia JJ, Ma YE, Zhang T (2014) Ecosystem service value assessment for an oyster farm in Dapeng Cove. *Shuichan Xuebao* 38:853–860 (in Chinese with English abstract)
- Zhang ML, Fang JG, Zhang JH, Li B, Ren SM, Mao YZ, Gao YP (2011) Effect of marine acidification on calcification and respiration of *Chlamys farreri*. *J Shellfish Res* 30: 267–271
- Zhang N, Song J, Cao C, Ren R, Wu F, Zhang S, Sun X (2012) The influence of macronitrogen (NO<sub>3</sub><sup>-</sup> and NH<sub>4</sub><sup>+</sup>) addition with *Ulva pertusa* on dissolved inorganic carbon system. *Acta Oceanol Sin* 31:73–82
- Zou DH, Xia JR, Yang YF (2004) Photosynthetic use of exogenous inorganic carbon in the agarophyte *Gracilaria lemaneiformis* (Rhodophyta). *Aquaculture* 237:421–431

Editorial responsibility: Symon Dworjanyn,  
Coffs Harbour, New South Wales, Australia

Submitted: May 29, 2017; Accepted: October 4, 2017  
Proofs received from author(s): November 22, 2017