



REVIEW

Climate change and aquaculture: considering adaptation potential

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ABSTRACT: Increases in global population and seafood demand are occurring simultaneously with fisheries decline in an era of rapid climate change. Aquaculture is well positioned to help meet the world's future seafood needs, but heavy reliance of most global aquaculture on the ambient environment and ecosystem services suggests inherent vulnerability to climate change effects. There are, however, opportunities for adaptation. Engineering and management solutions can reduce exposure to stressors or mitigate stressors through environmental control. Epigenetic adaptation may have the potential to improve stressor tolerance through parental or early life stage exposure. Stressor-resistant traits can be genetically selected for, and maintaining adequate population variability can improve resilience and overall fitness. Information at appropriate time scales is crucial for adaptive response, such as real-time data on stressor levels and/or species' responses, early warning of deleterious events, or prediction of longer-term change. Diet quality and quantity have the potential to meet increasing energetic and nutritional demands associated with mitigating the effects of abiotic and biotic climate change stressors. Research advancements in understanding how climate change affects aquaculture will benefit most from a combination of empirical studies, modelling approaches, and observations at the farm level. Research to support aquaculture adaptation requires an increasing amount of environmental data to guide biological response studies for regional applications. Increased experimental complexity, resources, and duration will be necessary to better understand the effects of multiple stressors. Ultimately, in order for aquaculture sectors to move beyond short-term coping responses, governance initiatives incorporating the changing needs of stakeholders, users, and culture ecosystems as a whole are required to facilitate planned climate change adaptation and mitigation.

KEY WORDS: Governance · Monitoring · Data needs · Mitigation · Prediction · Engineering · Research · Stressor

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1. INTRODUCTION

Seafood is one of the most highly traded foods internationally (Smith et al. 2010, FAO 2016), and demand is only expected to increase (World Bank 2013). The FAO reports that global capture fisheries have fluctuated between 80 and 90 Mt yr⁻¹ since the mid-1990s, with 50% capture decrease in the developed world, and reciprocal increase in undeveloped nations since the late 1980s (FAO 2018), with the proportion of globally exploited and overfished stocks increasing since the mid-1970s (FAO 2016). However, other sources have suggested that global catches may have actually peaked at 130 Mt yr⁻¹ in 1996 and have been strongly declining since (Pauly & Zeller 2016). Nevertheless, while some wild fisheries may experience short-term benefits from climate change, overall global landings are predicted to decrease 10% by 2050 (Barange et al. 2014). With projected increases in global population and seafood demand occurring simultaneously with ongoing fisheries decline, aquaculture will need to fill this gap.

Global aquaculture production has been steadily increasing for decades, reaching 110.2 Mt in 2016 (including aquatic plants), with the first-sale value estimated at US \$243.5 billion (FAO 2018). This production has largely been driven by China, accounting for 60% globally, followed by other major producer nations such as India, Vietnam, Bangladesh, and Egypt. Inland aquaculture systems account for the majority of global production (47 Mt), with most finfish culture occurring in earthen ponds (FAO 2016). Almost 600 different aquatic species are cultured (FAO 2018) in almost 200 countries, with almost a third produced without feed (e.g. bivalves and filter-feeding carps), along with 30.1 Mt yr⁻¹ of seaweeds and other algae (FAO 2016, 2018).

Most global aquaculture culture systems are therefore reliant to a greater or lesser extent on the ambient environment and ecosystem services. This suggests some inherent vulnerability to climate change effects. Nevertheless, due to the capacity to manage captive stock, aquaculture may have an adaptive advantage over wild species under a changing climate (Richards et al. 2015, Oyebola & Olatunde 2019). Capitalizing on this advantage requires recognizing and learning to respond to climate change effects in order to build adaptive capacity (Cinner et al. 2018). Understanding the effects of climate change on biological responses, resources, and economics in aquaculture is therefore a prerequisite. As climate change effects on aquaculture become better understood, innovative approaches for aquaculture

adaptation will help guide strategic planning, which will in turn define needs and expectations for research.

This review explores climate change adaptation strategies with potential application across a range of aquaculture species, regions, and environments. Consequently, the review is organized by strategy and approach, rather than culture system, country, or species. Potential effects of climate change stressors (biotic or abiotic drivers that can be altered by climate change with the potential for negative impacts) are only summarily discussed. This review aims to support pragmatic decision making by operators, managers, policy makers, and researchers, for planned adaptation. Expectations, needs, and practicalities of research implementation required to support meaningful aquaculture adaptation in a multi-stressor environment are explored.

2. CLIMATE CHANGE AND AQUACULTURE

2.1. Water temperature

The oceans have absorbed over 90% of the increase in energy in the climate system, causing the upper 75 m to warm by 0.11°C decade⁻¹ between 1971 and 2010 (Rhein et al. 2013). End-of-century model projections for average sea surface temperatures predict increases ranging from (mean ± SD) 0.71 ± 0.45°C (Representative Concentration Pathway [RCP] 4.5) to 2.73 ± 0.72°C (RCP 8.5) (Howes et al. 2015), with some regional 'hot spots' projected (RCP 8.5) in the Arctic, tropics, and North Pacific, in excess of 4°C increase (Bopp et al. 2013, Howes et al. 2015). Marine heat waves will be exacerbated as anthropogenic climate change raises mean temperatures (Schmidt & Boyd 2016). Global river water temperatures are projected to increase on average by 0.8–1.6°C by the end of the century (under the IPCC Special Report on Emission Scenarios B1–A2 scenario) compared to 1971–2000, with the greatest warming projected for the USA, Europe, eastern China, and parts of southern Africa and Australia (van Vliet et al. 2013).

Temperature is an aquaculture driver, and climate-driven temperature change has the potential to influence aquaculture in many ways. Some of these include effects on aerobic fitness (Pörtner 2008), oxygen demand (Fry & Hart 1948), hypoxia tolerance (Remen et al. 2013), reproductive performance (Pankhurst & Munday 2011), growth rate (Iwama & Tautz 1981, Reid et al. 2015), maturation (Wilkinson et al. 2010), nutrition and feeding (Britz et al. 1997,

Siikavuopio et al. 2012, Khan et al. 2014, Huguet et al. 2015, Remen et al. 2016), energy partitioning (Glencross & Bermudes 2012), immune response (Bowden et al. 2007, Buchtíková et al. 2011, Chiaramonte et al. 2016), disease proliferation (Burge et al. 2014), seasonal performance (Báez et al. 2011), pond stratification (Pickering et al. 2011), and dissolved oxygen (Timmons et al. 2002).

2.2. Sea-level rise

Sea-level rise has the potential to affect coastal aquaculture operations through loss of culture area (Hargreaves 2014), greater and more distant salt intrusion into coastal groundwater (Ahmed 2013, Nguyen et al. 2014, Smajgl et al. 2015, Tully et al. 2019), and, in some areas, the augmentation of seasonal or episodic flooding via storm surges (Wassmann et al. 2004, Rhein et al. 2013). Global mean sea level rose 1.2 ± 0.2 mm yr⁻¹ between 1901 and 1990, with this rate increasing to 3.0 ± 0.7 mm yr⁻¹ between 1993 and 2010 (Hay et al. 2015). By the end of the 21st century, it is very likely that sea-level rise will have occurred in over 95% of the world's oceans (Stocker et al. 2013). Small, gradual increases in sea-level rise are not inconsequential, as increased flood inundation is not linear. An increase in sea level by 0.1 m will increase flood frequency by approximately 3 times (Church et al. 2006, Zhai et al. 2014). Most land-based aquaculture (e.g. pond, hatcheries) is located close to a water source and therefore susceptible to flooding. Flooding can result in escapes, introduction of invasive species (Oyebola & Olatunde 2019), and contamination of culture water (Adhikari et al. 2018, Kais & Islam 2018, Oyebola & Olatunde 2019). Flood-related fish kills may occur for several reasons, but are mainly due to low oxygen in flood waters (Idris et al. 2014). Flood-related mortalities are not uncommon with pond culture (e.g. Bell et al. 2010) and hatcheries (e.g. Lawrence 2016).

2.3. Changing weather

Pending El Niño events are expected to drive increased interannual variability in regional temperature extremes (Fasullo et al. 2018), and a near doubling in the frequency of both extreme La Niña and El Niño events is expected within this century (Cai et al. 2014, 2015). El Niño has been linked to massive harmful algal blooms in Chile (in 2015 and 2016), extreme heat, and low dissolved oxygen, damaging

Central American tilapia production (Soto et al. 2018), and slackening trade winds and reducing pond mixing, prompting stratification and hypoxia in Hawaiian fish ponds (McCoy et al. 2017). The effects of climate change on the overall severity and frequency of harmful algal blooms are uncertain (Wells & Trainer 2016).

Overall, global warming is likely to lead to overall drying of land surfaces due to increased evaporation (Sherwood & Fu 2014), although regions will vary. Large rainfall changes are projected for a considerable proportion of the tropics by mid-century (Chadwick et al. 2016). Decreases in precipitation will occur in some regions, such as South Asia during summer monsoons (Singh 2016), and increased drought is expected to increase in other areas such as eastern China (Sun et al. 2014), whereas precipitation-based flood frequency will increase in many regions such as Southeast Asia, peninsular India, eastern Africa, the northern half of the Andes (Hirabayashi et al. 2013), the central USA (Rahmani et al. 2016), and parts of Europe (Alfieri et al. 2015). Increased eutrophication due to changes in precipitation and agriculture expansion is expected over the 21st century (Tilman et al. 2001, Sinha et al. 2017). Intensified land run-off may further drive sewage or agricultural fertilizers into coastal production areas where shellfish can become contaminated (Cornelisen et al. 2011) or potentially increase algal blooms, which reduce dissolved oxygen and result in fish mortalities (Díaz et al. 2009).

Storm activity has increased in some ocean basins, such as increased wind stress in the Southern Ocean and increase in average winter wave heights in the North Atlantic (Rhein et al. 2013). There has been greater intensity of tropical cyclones in the North Atlantic since the 1970s (Rhein et al. 2013), and increased occurrences of tropical cyclones in the Caribbean and landfall typhoons in East Asia are expected (Stocker et al. 2013). Changing storminess has recently been identified as having potentially catastrophic impacts for global fisheries, largely due to destruction of vessels, infrastructure, and communities (Sainsbury et al. 2018). Such concerns can be extrapolated to coastal aquaculture, where farms are highly sensitive to storm events (Allison et al. 2009, Rahman & Hossain 2012, Luening 2013), as evident by the storm-driven destruction of almost entire land-based aquaculture sectors (Reid & Jackson 2014, Kais & Islam 2018). Open-water aquaculture is also susceptible to severe weather, with large-scale escapes from sea cages correlated with storm events (Jensen et al. 2010).

2.4. Ocean acidification

Atmospheric increases in anthropogenic carbon dioxide (CO₂) are raising the concentration of hydrogen ions in the Earth's oceans. The resultant decrease in pH is known as ocean acidification (Pörtner 2008). Predictions integrating the continued use of fossil fuel have estimated atmospheric CO₂ increases from 400 ppm (current levels) to 750 ppm (under the IPCC 1S92a scenario, where atmospheric CO₂ concentration rises 1% yr⁻¹ after 1990) in 2100, to more than 1500 ppm in the next century (IPCC 2007). The average global ocean pH has already declined by 0.1 unit (which is an increase in oxygen hydrogen ions of approximately 26%) compared with pre-industrial values (Orr et al. 2005) and is predicted to decrease by another 0.4 units by the end of this century (Caldeira & Wickett 2003). Ocean acidification is occurring more rapidly in coastal and estuarine areas (Waldbusser et al. 2011), which are key aquaculture production zones, with more variable pH and CO₂ levels than those in the open ocean, as local inputs and diel cycles of photosynthesis and respiration can further drive changes in pH beyond the influence of ocean acidification. Coastal and estuarine pH can also change due to natural variability, eutrophication, and net heterotrophy (Kemp et al. 2005, Borges & Gypens 2010), acid-forming compound deposition (Doney et al. 2012), regional changes in land use (Green et al. 2009), and watershed inputs (Dove & Sammut 2007, Salisbury et al. 2008). The increased partial pressures of carbon dioxide (pCO₂) diminish the seawater saturation states of aragonite and calcite, the 2 most commonly biomineralized forms of calcium carbonate (Miller et al. 2009) used for the shell formation of some marine organisms.

Experimental outcomes of ocean acidification exposure studies on shellfish can be influenced by numerous variables such as species, acute vs. chronic exposure, animal size, parental exposure, the driver investigated (i.e. CO₂, calcium carbonate saturation, pH), magnitude of change, diet availability, and presence of other stressors (e.g. Ries et al. 2009, Thomsen et al. 2010, 2017, Waldbusser et al. 2010, 2015, Gobler et al. 2014, Zhao et al. 2018). To date, negative responses to ocean acidification have been reported in the majority of shelled mollusc larval studies on aquaculture and commercial fisheries species (Gazeau et al. 2013). Most shellfish aquaculture relies either on naturally occurring wild or hatchery-derived juveniles (seed) as the initial stage of production, and therefore decreases in seed abundance will

have a significant effect on industry sustainability. Ocean acidification has already had a significant impact on shellfish hatchery production in some regions, such as the US Pacific coast (Ekstrom et al. 2015), whereas later stage (post-larval) shellfish appear to be more resilient to ocean acidification (Green et al. 2009, Waldbusser et al. 2010, Range et al. 2011, Talmage & Gobler 2011). The impacts of ocean acidification on physiology and reproduction are likely to have significant carry-over effects on the sustainability of future populations of cultured or wild shellfish. In wild finfish, impairments in olfactory responses under elevated pCO₂ have been linked to receptor impairment of GABA-A, the primary inhibitory neurotransmitter receptor in the vertebrate brain (Nilsson et al. 2012, Chivers et al. 2014, Chung et al. 2014, Hamilton et al. 2014), but it is unclear how ocean acidification could ultimately affect finfish aquaculture. Marine macroalgal species (seaweeds), which are currently CO₂-limited, are expected to benefit from increases in atmospheric CO₂ and dissolved CO₂ in water (Beardall et al. 1998). In contrast, calcified seaweeds (like crustose coralline algae) will be more susceptible to ocean acidification (Kroeker et al. 2010).

2.5. Diet

Approximately 70% of total global aquaculture production by weight is dependent upon the supply of external feed inputs (Tacon & Metian 2015), and climate change will affect ingredient sourcing for aquafeeds (De Silva & Soto 2009, Brugère 2015) and feed management (Shelton 2014). Despite great strides in identification and development of alternative ingredients (e.g. meals derived from soy or corn), small pelagic fish (De Silva & Soto 2009) are still heavily used for aquafeeds, and reduction fisheries accounted for about 17% of the global fisheries landings in 2014 (Tacon & Metian 2015). Some reduction fisheries have been identified as climate-sensitive (Merino et al. 2010, Lindegren et al. 2013, Buchheister et al. 2016). It is therefore expected that climate change may further complicate the relationship between aquaculture and the capture fisheries.

Climate change effects on microalgae (Doney 2006) may also have major implications for some types of aquaculture, since phytoplankton comprise common diets for cultured finfish larvae (through zooplankton), bivalve molluscs, and crustaceans (Wikfors & Ohno 2001). Warmer and more stratified oceans are expected to shift phytoplankton communities towards

smaller size with reduced adaptive capacity (Acevedo-Trejos et al. 2014) and shift cold-water species ranges poleward (Hallegraeff 2010). Impacts will vary with location (Doney 2006), possibly even benefitting filter-feeding species of shellfish aquaculture in currently nutrient-limited areas (De Silva & Soto 2009). Climate warming is also expected to decrease the omega-3 fatty acid (FA) content in phytoplankton, the main source of omega-3 FAs in aquatic systems (Hixson & Arts 2016), and this could have implications for crucial omega-3 FA content in aquafeeds.

2.6. Simultaneous stressors

It has been suggested that the greatest threat for sustainable aquaculture development is the co-occurrence and interaction of multiple environmental stressors (Sarà et al. 2018). Simultaneous changes in carbon dioxide, temperature, phytoplankton (community and abundance), salinity, and oxygen are occurring to various degrees through the world's waters (Boyd & Hutchins 2012, Stramma et al. 2012, P. W. Boyd et al. 2015, Schmidtko et al. 2017). Studies on aquatic species' responses to multiple stressors are still relatively limited and are typically short term and laboratory-based. Research to date has encompassed a variety of aquatic species' responses to simultaneous exposure to changes in ocean acidification and temperature (Byrne et al. 2009, 2013, Matozzo et al. 2012, Brisolin De Souza et al. 2014, Gräns et al. 2014, Graiff et al. 2015, Miller et al. 2015, Suckling et al. 2015), ocean acidification and dissolved oxygen (Gobler et al. 2014, DePasquale et al. 2015, Miller et al. 2016, Sui et al. 2016), salinity and temperature (Choi et al. 2006), and salinity and dissolved oxygen (Wang et al. 2011). Few studies have examined biological responses to more than 2 simultaneous stressors (Catalán et al. 2019). Those that have, reported that multiple stressors that are negatively synergistic may not always be fully additive (i.e. there is interaction), with the greatest biological response often being attributed to a single dominant driver (Brennan & Collins 2015).

3. ADAPTATION STRATEGIES

The wide range of global aquaculture species, regions, and environments that are increasingly exposed to climate change stressors prohibits the examination of specific species–stressor combinations to explore options for adaptation in this discussion.

Nevertheless, commonalities among approaches and strategies can have general applicability across multiple aquaculture sectors. These approaches can be loosely categorized into diet and nutrition; management and engineering solutions; warning of occurring, pending, or future deleterious effects; genetics and biotechnology; and governance. Herein we explore research literature, industry publications, and governance reports that provide insights for aquaculture adaptation to climate change.

3.1. Diet and nutrition

3.1.1. Diet quantity and quality

There is an energetic cost for organisms acclimating under environmental or biological stressors. Nutrition can therefore influence adaptation potential. For climate change stressors, this has been documented for several mollusc species in response to ocean acidification exposure, where energy intake and assimilation may be unaffected in conditions of plentiful food (Parker et al. 2013, Timmins-Schiffman et al. 2013). Abundant food supply can increase tolerance to high pCO₂ and low pH for blue mussels *Mytilus edulis* (Melzner et al. 2011, Thomsen et al. 2013) and Pacific oyster *Crassostrea (Magallana) gigas* larvae (Timmins-Schiffman et al. 2013). A similar compensatory response upon warm water exposure has also been reported. Several shellfish have demonstrated better survival and conditioning from heat shock and thermal stress with abundant diet, such as California mussels *M. californianus* (Fitzgerald-Dehoog et al. 2012) and juvenile South African abalone *Haliotis midae* (Vosloo et al. 2013).

Increased demand for dietary energy in response to changing metabolic and physiological demands is consistent with nutritional bioenergetic theory, and existing nutritional strategies may already apply under a changing climate. Nutritional bioenergetic models have been developed for numerous aquaculture species, such as rainbow trout *Oncorhynchus mykiss* (Hua & Bureau 2009), Nile tilapia *Oreochromis niloticus* (Chowdhury et al. 2013), and barramundi *Lates calcarifer* (Glencross & Bermudes 2012), to estimate optimal digestible protein and energy requirements under different temperatures.

Diet quality can also affect tolerance to stressors. Providing a high antioxidant diet of grape seed extract and the macroalga *Ulva lactuca* improved high temperature survival of the greenlip abalone *H. laevigata* (Lange et al. 2014). Experimental evidence

suggests that purple sea urchins *Paracentrotus lividus* can improve adaptation to acidification through modulation of the dietary Mg:Ca ratio (Asnaghi et al. 2014). Increased dietary protein at elevated temperatures improved blood serum immune parameters, antioxidant enzymes, and heat shock protein gene expression in juvenile mirror (common) carp *Cyprinus carpio* (Huang et al. 2015). Propolis, a resinous antioxidant honeybee product, improved sea bass resistance to low-temperature stress (Šegvić-Bubić et al. 2013). There may also be a role for dietary immunostimulants to enhance environmental protection (Wang et al. 2017) under a changing climate.

3.1.2. Feed and diet sourcing

A decline in reduction fisheries, climate-driven or not, may not necessarily limit global aquaculture expansion. The World Bank, together with the FAO, projected that a steady rise in the price of fish-derived feed ingredients will drive technological change and increase conversion efficiency, and consequently, aquaculture growth is expected to continue (World Bank 2013). This is plausible, as the largest consumers of commercial aquaculture feeds have been herbivorous and omnivorous species, namely, the carps, tilapia, shrimp, and catfish species (Tacon & Metian 2015). The dietary flexibility of these species enables substantially greater opportunities for the use of alternative ingredients compared to carnivorous species (Olsen & Hasan 2012). Many marine aquaculture species in China, Indonesia, Taiwan, Malaysia, and Thailand are fed almost exclusively with low-value fish, resulting in poor feed conversion efficiency (output:input; fish weight:feed weight) (FAO 2014). Good feed conversion efficiency using low-value fish diets can be achieved with some culture systems (Bunlipatanon et al. 2014), suggesting there could be a more efficient use of this resource. Nevertheless, the FAO (2014) suggested that transitioning to compound feeds would not only increase feed efficiency, but would require only a third of the fish source inputs compared to low-value fish feed. Global aquaculture production will continue to increase, without greatly increased fish meal usage (Olsen & Hasan 2012), and there have been significant advancements to reduce fish-based ingredients in some aquaculture sectors with cultured carnivorous species. For example, 1 kg of protein in the edible portion of Atlantic salmon can now be produced from 0.7 kg of marine protein, resulting in a

net production of marine protein (Ytrestøyl et al. 2015).

Several nutritional sourcing strategies have been advocated to meet current challenges while promoting climate change adaptation. These include sourcing nutritionally sound feed ingredients at the local level (Tacon et al. 2011); pursuing ‘underutilized’ alternative crops, resistant to drought and temperature, with good nutritional properties (Hall 2015); promoting rice–fish farming systems (J. Xie et al. 2011, B. Xie et al. 2013, Shelton 2014); and implementing integrated multi-trophic aquaculture (IMTA) to augment diet through increased use of waste organic and dissolved (inorganic) nutrients (Chopin et al. 2012, Reid et al. 2013). While sourcing aquaculture feed ingredients from terrestrial crops may relieve some resource pressure from the marine environment, there is a concern that environmental impacts could shift to those associated with increased crop demand (Fry et al. 2016), especially when global crop output for 2050 is projected to be insufficient for global demand (Ray et al. 2013, Asseng et al. 2015). It has been further suggested that terrestrial crops for aquafeeds may have negative implications for human health such as the reduced omega-3 FA content associated with terrestrial ingredients (Fry et al. 2016). Nevertheless, marine sources of omega-3 FAs may become less reliable. New supplies of omega-3 FAs will need to be found, and this is an area of ongoing investigation as reviewed by Tocher (2015).

Enabling good nutritional access for species that extract their diet from the environment under a changing climate is likely to present a suite of challenges as well as opportunities. Dietary control will be a much greater challenge for species that extract their nutrition from the ambient environment such as shellfish and seaweeds, where changing climatic conditions will directly influence the quality and quantity of available nutrition. Farm relocation is one potential solution (Soto et al. 2018). Rearing systems such as land-based aquaculture could also enable more dietary control with extractive species, but this may come at an increased cost to other resources such as infrastructure requirements, land usage, and water pumping.

3.2. Genetics and biotechnology

The epigenetic response potential of fishes (Pittman et al. 2013) and marine invertebrates (Sanford & Kelly 2011) suggests some level of adaptive capacity to climate change, but there are significant

knowledge gaps. For example, the capacity for marine populations to adapt to increased acidity is largely unknown, as few studies have considered acclimation times of more than a few months (Harley et al. 2006, Kurihara et al. 2007, Doney et al. 2009, Thomsen et al. 2010, Gazeau et al. 2013) or with relevant exposure levels. Aquaculture-based populations are less likely to be impacted due to the potential to manipulate culture conditions (Richards et al. 2015). Plastic responses, particularly in early life stages, suggest that greater environmental control during early rearing may help direct adaptive epigenetic responses. Hatcheries are already well positioned to use this strategy.

Selective breeding programmes for desirable traits are already common in aquaculture, and this may provide additional options for climate change adaptation. Generational improvements have the potential to occur rapidly for some traits. Some of the most globally produced culture species, such as carps (*Puntius gonionotus*, *Labeo rohita*, *Cyprinus carpio*) and prawns (*Macrobrachium rosenbergii*), have reported productivity gains ranging from 7 to 12% per generation, through application of conventional selective breeding approaches (Nguyen 2016). There is also compelling anecdotal evidence that environmental stressor tolerance traits may 'evolve' under culture conditions. Ellis et al. (2017) reported that finfish in recirculated aquaculture systems are routinely raised in high-CO₂ environments, often in excess of end-of-century predictions, sometimes >10 000 µatm, with no apparent ill effects. While this apparent discrepancy from many studies on wild species could be for several reasons, the authors suggested one possibility may be that aquaculture species selection and breeding under intensive culture environments may have conferred greater tolerance to high CO₂. Oyster breeding programmes that have experienced declines in survival and yield related to changing ocean conditions (de Melo et al. 2016) are now selecting for increased calcification rates (Waldbusser et al. 2010). Organisms adapted to local carbonate chemistry indicate that spatially varying selection could provide adaptive traits and maintain genetic variation (Kelly et al. 2013), and given variable sensitivities to ocean acidification between strains of Sydney rock oysters *Saccostrea glomerata*, selective breeding is being advocated (Parker et al. 2011). Salinity-tolerant strains of shark catfish *Pangasianodon hypophthalmus* are now being developed to adapt to saline water intrusion from floods in the Mekong Delta, Vietnam (Nguyen et al. 2014). Adaptation through domestication of a variety of species tolerant to climatic varia-

tions may also reduce dependency on wild-caught seed (Oyebola & Olatunde 2019).

Specific climate change performance traits may warrant consideration in broodstock selection programmes. Genomic approaches may be necessary to fully exploit these traits or improve traits at a rate sufficient to match the rate of climate change. When there is no possibility to include climate change-related traits into a broodstock programme, retaining as much genetic variability as possible in the breeding nucleus should be a priority. The success of the genetic improvement programmes for carp, tilapia, and prawns are in part due to the establishment of respective base populations with ample genetic variability (Nguyen 2016). Maintenance of genetic diversity could ensure preservation of rare alleles that might be associated with resistance to a future disease or increased survival to environmental stressors (Gurney-Smith et al. 2017). While there is immense potential for breeding programmes to improve resistance to climate change stressors, experience suggests that domestication does not always result in improved fitness compared to wild counterparts (Araki et al. 2008), and there are numerous challenges to overcome, including environment effects, like climate change, on genotype stability (Nguyen 2016). Breeding programmes also require a medium- to long-term investment (Nguyen et al. 2016) and may not be easily attainable for subsistence farmers.

There are also encouraging biotechnological options. If the rate of climate change supersedes breeding schemes based on recorded phenotypic data, including molecular selection techniques (e.g. genomic marker-assisted selection) with phenotypic selection programmes may be necessary. Advancements in commercial-scale cryopreservation enable 'gene banking' or retention of genetics from one parent (male) regardless of continued inclusion in a brood stock programme, and viable larval preservation protocols have been and are being developed for some marine shellfish (e.g. Adams et al. 2004), depending on species. Consequently, gene or germplasm banking approaches have been advocated as a biological insurance for future needs of aquaculture breeding and stock selection (Hulata 2001, Barrento et al. 2016).

3.3. Management and engineering solutions

The global aquaculture sector has demonstrated great ingenuity and adaptability, as reflected by the immense diversity of species cultured in vastly

different environments. This capacity for creative problem-solving will ultimately be required for climate change adaptation. At their simplest, adaptation strategies are likely to be based around management and husbandry practices (McCoy et al. 2017, Adhikari et al. 2018). These practices are often accompanied by engineering solutions, so both will be considered together.

3.3.1. Flooding and storm protection

Protection against floods will be a combination of management strategies and age-old engineering approaches, such as increasing physical barriers or use of tanks and inland enclosures (Kais & Islam 2018, Oyebola & Olatunde 2019). The scope of response will be a function of damage potential. An increase in sea level of 0.1 m requires coastal infrastructure such as wharfs to be raised by >0.1 m (an allowance) in order for the same historical flood frequency to be maintained (Zhai et al. 2014). Some flood-response management strategies are already routine in regions with predictable seasonal flooding. For example, in Malaysia, fish are harvested prior to known flood periods in areas prone to episodic flooding (Idris et al. 2014). Likewise in Taiwan, pond water levels are dropped prior to flooding (Chang et al. 2013), and excess water volume is pumped out of ponds in some Indian regions (Adhikari et al. 2018). Nets are placed around shrimp farms to prevent escapes during heavy rains and flooding in some regions of Bangladesh (Kais & Islam 2018) and fish ponds in India (Adhikari et al. 2018). Fish evacuation or movement to safer culture areas may be necessary where flooding can overtop tanks or ponds (Dodd 2011, Adhikari et al. 2018).

Protection of land-based coastal aquaculture against storm surges can be improved with natural barriers, such as mangroves (Ahmed & Glaser 2016, Chow 2018) reefs, and coastal vegetation (Arkema et al. 2013). Giving up some coastal culture ponds in Java has been advocated to protect some residual mangroves (Bosma et al. 2017), and the translocation of shrimp culture from mangrove swamps to offshore aquaculture has been suggested as a way to reduce mangrove loss while increasing carbon sequestration (Ahmed et al. 2017). Some initiatives are increasing the percent coverage of mangroves in integrated mangrove–shrimp farming areas (Ahmed et al. 2018) and restoring coastal mangroves in abandoned shrimp ponds (Friess et al. 2016). Seaweed aquaculture has also been suggested as a strategy to

dampen incoming wave energy and protect shorelines (Duarte et al. 2017).

In some instances, flooding could expand aquaculture opportunities. Some aquaculture fisheries rely on flooding to fill water bodies with wild fish for culture, such as floodplain ponds in Kenya (Kipkemboi et al. 2010) or Whedo aquaculture systems in west Africa (Hauber et al. 2011). Flooded land which is no longer suitable for agriculture could provide new opportunities for aquaculture (IFAD 2014). Areas rendered unsuitable for traditional agriculture (e.g. rice farming) as a result of saltwater intrusion could be repurposed for shrimp farming (De Silva & Soto 2009).

Weather extremes associated with climate change may increase escapes and negatively affect pathogens and hosts, emphasizing the need for good biosecurity under a changing climate (Bondad-Reantaso et al. 2018). Uncertainties in storm extremes have introduced additional considerations for the design of aquaculture facilities (Alvarez-Lajonchère & Pérez-Roa 2012). Storm intensity duration frequency (IDF) curves are used by engineers to guide design specifications; these are changing and being updated for new climate scenarios and predictions (Liew et al. 2014). Aquaculture infrastructure design will likewise need to account for changes in storm IDF, or redesign efforts may become necessary due to structural failure (Can & Tuan 2012). Fish cages need to withstand powerful waves under a changing climate (Binh et al. 2017). Developments in off-shore aquaculture, such as submersible systems for both finfish (Shainee et al. 2014) and shellfish (Kim et al. 2014), have already spurred technologies capable of withstanding high energy exposure for open-water aquaculture (Shainee et al. 2013).

3.3.2. Relocation

Choosing farm locations less impacted by climate change effects is an intuitive adaptation strategy, and the inclusion of climate change and other risks into spatial planning and aquaculture zoning has been advocated as an approach to reduce impact exposure (FAO 2016). Some shellfish hatcheries have already relocated to less acidic waters (Welch 2012). GIS or remote sensing tools have been used for some time to select appropriate aquaculture locations (Nath et al. 2000, Perez et al. 2005, Hossain et al. 2007, Radiarta et al. 2008, Hossain & Das 2010, Mamat et al. 2014, Brigolin et al. 2015, Dapuetto et al. 2015, Ottinger et al. 2016), and these are now being used to identify

culture areas less prone to sea-level rise, drought, thermal stress, and flooding (Handisyde et al. 2008, 2014, Hossain & Das 2010, Khan et al. 2012, Aura et al. 2017). Deeper ponds can provide a thermal refuge and greater dissolved oxygen reserves and are less sensitive to environmental factors in dry seasons (Soto et al. 2018).

A locational adaptation strategy could also reduce disease, limiting pathogen or parasite exchange between wild and cultivated species (Peeler & Feist 2011, Lafferty et al. 2015). A potential complicating factor with relocating to areas of optimal water quality is that conditions may not be optimal for all life stages. Báez et al. (2011) reported that for rainbow trout, the fastest growth rates occurred at temperatures which are detrimental to reproductive performance, and have advocated separate site locations for broodstock and grow-out to ensure optimal temperatures across the full life cycle. This has implications for the required resources, infrastructure, and expertise to facilitate these adaptive goals.

3.3.3. Localized mitigation

In some circumstances, direct mitigation of the localized environment is possible, and this is achievable with the most common method of global aquaculture production, pond culture. While ponds are exposed to the open environment, a high degree of control over environmental variables through water management, aeration, protective cover, and water treatment is possible (Lorenzen et al. 2017). Oxygen tablets and aeration can augment dissolved oxygen, pumping in freshwater can supplement water during the dry season, and along with shading and increased depth (>1 m), this can reduce pond temperatures (Binh et al. 2017, Adhikari et al. 2018). For infaunal marine species in detrimental sediment saturation states, sediment buffering using crushed shell has been tested to increase sediment alkalinity, pH, and aragonite saturation states, thereby decreasing shell dissolution and/or promoting larval recruitment (Green et al. 2009). Seaweed or macrophyte culture has been suggested as a method to provide localized mitigation, act as a net producer of oxygen, sequester carbon dioxide, and increase pH (De Silva & Soto 2009, Han et al. 2013, Clements & Chopin 2017, Duarte et al. 2017, Wahl et al. 2018), although effectiveness will depend upon many factors such as species, culture system, production scale, nutrient load, water flow, latitude, irradiance, and production

timing (Broch et al. 2013, Reid et al. 2013, Hurd 2015). Land-based or closed containment rearing strategies could be options to either protect sensitive life stages from environmental stressors or control rearing conditions to enable adaptive responses (see Section 3.2). These strategies enable water quality control through either recirculation approaches (Timmons et al. 2002), strategic water intake to avoid periodic stressors such as CO₂ upwelling (Barton et al. 2012), or through bio-buffering.

3.3.4. Diversification

Culture diversification is a common historical approach as a contingency to spread the risk against losses. Globally, aquaculture species produced are highly diverse, covering many regions, environmental conditions, and trophic levels. Such diversity has the potential to enable selection of alternative species should one species become non-viable under regional climate change. Aquaculture practices, such as polyculture and IMTA, at the farm level are options that can enable continued production if one crop fails (Chopin et al. 2012, Oyebola & Olatunde 2019) and are advocated for climate change adaptation (Binh et al. 2017). Co-culture not only serves the dual purpose of expanding diversity and enhancing the use of nutrients, but can also reduce competition for other resources such as water, through integrated aquaculture and agriculture systems, in addition to improving water quality (IFAD 2014, Shelton 2014, Oyebola & Olatunde 2019). In developing countries where there is intense competition for space and resources, there may also be considerable future expansion of culture-based fisheries, an extensive aquaculture practice based on the principle of stock and recapture (De Silva 2016, Oyebola & Olatunde 2019). This approach has the potential to address climate change-related issues of wild stock recruitment requiring minimal feed use and care (Beveridge et al. 2018). There is increasing interest in culturing species which have the capacity to breathe air, and this could be a viable option where water quality is poor (Lefevre et al. 2014).

3.4. Improving information

Timely information can enable management responses and strategic planning, from real-time, to forecasting early warning, to prediction over long time scales. Initiatives to promote local-level adapta-

tion through training, data collection, analysis, and sharing, are advocated as an approach to connect environmental data with broader forecasts to support decision making by local aquaculture stakeholders (FAO 2016, Oyebola & Olatunde 2019).

3.4.1. Real-time monitoring

Real-time monitoring can alert farmers to the presence of deleterious conditions that may not be obvious until the onset of behavioural or clinical symptoms in the stock. Monitoring is routine for parameters such as oxygen and temperature in many aquaculture sectors, but recent developments in technology and networking have greatly expanded breadth and capability. Near real-time water quality is available through some ocean condition monitoring networks (e.g. Integrated Ocean Observing System [IOOS, <https://ioos.noaa.gov/>] and regional nodes such as the Northwest Association of Networked Ocean Observing Systems [NANOOS; <http://nvs.nanoos.org/ShellfishGrowers>]) which allow aquaculturists to track crucial water quality parameters online for informed management decision-making (e.g. IOOS). Some monitoring stations are located directly at aquaculture sites with industry participation (e.g. NANOOS). In the absence of detailed water quality data, animal behavioural cues are often indicators of environmental stress. Observation is difficult with some species, like many shellfish, that exhibit limited behavioural cues upon the onset of environmental stress. This has spurred recent developments in microsensor technology which now enable shellfish heart rates to be monitored in sentinel animals, as a means to assess real-time response to environmental or biological stressors (Hellicar et al. 2015). New metrics are now being considered such as shell thickness and condition index of mussels to assess thermal impacts on the shellfish aquaculture sector (Martinez et al. 2018).

3.4.2. Early warning

Early warning of acute, deleterious events can improve response times of farmers (Nguyen et al. 2015). Aquaculturists can be alerted to pending weather conditions in some Chinese regions, and seasonal predictions of water temperature are now available in Australia (Spillman & Hobday 2014). Avoiding the use of low aragonite-saturated waters under strong upwelling conditions has enabled sig-

nificant restoration of oyster hatchery production in the US Pacific Northwest (Barton et al. 2012). Adequate warning of stressful environmental events may also enable intervention time to bolster the immune system at times when species are known to be immunosuppressed, through changes such as photoperiod manipulation or the application of immunomodulators (Bowden et al. 2007). Early warning systems have been strongly advocated in order to reduce aquatic food safety risk posed by climate change-related natural disasters, such as contamination (e.g. pathogens) from extreme weather events (Cornelisen et al. 2011, Bondad-Reantaso et al. 2018).

3.4.3. Prediction

Aquaculture sectors can clearly benefit from long-term climate predictions of environmental variables. For example, templates of Tasmanian sea surface temperature projections can be applied by managers to identify onset years when some regions will become unsuitable for the culture of certain species (Hobday et al. 2018). However, one of the most pressing predictive needs is arguably predicting disease outbreak. Limitations in monitoring and prediction of disease outbreaks could be major hurdles for climate change adaptation for some aquaculture sectors. Early detection followed by quick targeted responses can reduce the impact of disease (Peeler & Feist 2011, Groner et al. 2016). Tracking serious and potentially reoccurring infections is a common practice for some developed regions, but baseline data are also required for determining prevalence and distribution of pathogens and parasites in wild species before establishing operations to prevent transfer (Lafferty et al. 2015). There are some encouraging developments such as web-based management programmes that maintain an epidemiological database, tracking temporal and spatial trends of farm infections, and environmental conditions for decision support (e.g. Harris 2015). However, at the global scale, most disease outbreaks are unreported. There is a lack of standardized reporting of aquaculture-based epizootics and conditions surrounding outbreaks, which is needed to help formulate adaptation strategies under climate extremes (Leung & Bates 2013). Enhanced reporting on disease prevalence is necessary for better understanding of climate influences. While effective disease surveillance needs to start with farmers (Brugere et al. 2017), epidemiological modelling has the potential to predict the spread of some aquaculture diseases under a changing climate

(Lafferty 2009). The use of GIS-based statistical models that enable spatially distributed determinants of aquatic health and disease for risk mapping have been encouraged (Thrush et al. 2011). However, these models are data-driven, which further emphasizes the need for detailed reporting, interdisciplinary approaches to mitigate disease impacts in aquatic systems (Adlard et al. 2015), and additional research on effects of environmental change on disease (Lafferty et al. 2015).

3.5. Governance

Climate change has been described as a complex and diabolical policy problem that poses difficult challenges for contemporary political systems (Steffen et al. 2011). It is an inherently global problem that has been the subject of international negotiations for over 20 yr. While parties to the United Nations Framework Convention on Climate Change (UNFCCC) meeting at COP21 (also known as the 2015 Paris Climate Conference) have reached an accord on emission limits, global annual emissions are expected to rise to 2030. The anticipated increase in emissions is likely to be inconsistent with the international goal of limiting the rise in global mean surface temperature to no more than 2°C above preindustrial levels (Rogelj et al. 2013, R. Boyd et al. 2015). This has raised concerns that the current climate policy promises, based on voluntary intended nationally determined contributions, will do little to stabilize the climate, and that their impact will be undetectable for many decades (Lomborg 2016).

In response to the challenges and opportunities posed by climate change, many countries have developed national comprehensive climate change adaptation strategies (e.g. Fransen et al. 2009, DEE 2015, SEMARNAT-INECC 2016). These plans acknowledge the wide-ranging impacts of climate change, and set out plans and targets for reducing greenhouse gas emissions. However, they seldom include specific adaptation and/or mitigation strategies for aquaculture. In the near term, climate change will have many direct impacts on aquatic farming systems, which will bring new challenges to maintaining sustainability. As noted by the FAO (2015) and others (Paprocki & Huq 2018), aquaculture will need to be integrated into national and regional adaptation plans, as otherwise it could suffer as a result of adaptation measures applied to other sectors.

The type and intensity of the impacts climate change will have on aquaculture will vary between

continents, countries, climatic regimes, and production systems. In practical terms, there is a need for local, regional, and national governments to establish regulatory policy and frameworks that will help direct aquaculture onto the most climate change-adaptive and resilient paths possible (Craig 2015). Decision-makers should examine their adaptation strategies/policies/activities over a wide range of plausible futures (e.g. RCP scenarios) to choose a strategy that is sufficiently robust to not only account for uncertainties of climate prediction but other economic, political, and cultural factors as well (Dessai et al. 2009). Although adaptation strategies will be influenced by sector and regional specifics, a number of actions can be applied to most aquaculture. In an FAO series of international case studies on regional climate change adaptive capacity of fisheries and aquaculture, Brugère (2015) concluded that ecosystem resilience and human adaptive capacity were the 2 major determinants of vulnerability, and that governance was itself a determinant of adaptive capacity. Fisheries and aquaculture will require a variety of policy and governance initiatives to facilitate planned climate change adaptation, which typically necessitates more effort and cost, and are focused on much longer time scales, compared to shorter-term coping responses (Shelton 2014). Owing to the cross-cutting nature of climate change effects, policies that are designed to conserve biodiversity, reduce external stressors on aquatic systems, and/or protect valuable natural areas as well as endangered species will all benefit aquaculture.

A consistent theme echoed by academics, NGOs, and national and international government agencies is that in the face of rising global demand for food, aquatic farming systems must not only adapt to ongoing and pending environmental changes in order to sustain production levels, but also increase production significantly. There is great potential for aquaculture to adapt, innovate, and expand. However, regulatory frameworks are complex, expensive for farmers to navigate, and slow to respond to technological innovations (Alexander et al. 2015, Bostock et al. 2016). In both developed and developing nations, there are calls for regulatory reforms to reduce red tape, improve certainties for operators so that they can obtain access to space and water (e.g. Watson 2015, James 2016), and provide trained specialists to analyse and implement adaptation strategies (Binh et al. 2017). While there is a need to maintain and, in some instances, improve environmental regulation within the industry, streamlining of regulatory regimes would go far to help the sector adapt and diversify.

4. RESEARCH CONSIDERATIONS TO SUPPORT AQUACULTURE ADAPTATION

Many farmers are currently dealing with climate change problems through short-term coping measures which ultimately need scientific improvements to support long-term solutions (Dubey et al. 2017). Investments in research and technology transfer are needed to develop sustainability and support climate change adaptation in aquaculture (Uppanunchai et al. 2018). The FAO has recognized the need to build capacity of institutions to integrate research, management, and policy while enabling partnerships between science and policy institutions, so that research is developed at relevant scales for decision-making (FAO 2018). While there is acknowledgement of strategic research needs required to develop adaptation, there are also tactical research challenges for both single and multiple stressors that require exploration to support informed research planning.

4.1. Environmental data needs

Climate information complements and rounds out adaptation and mitigation strategies (Trenberth et al. 2016). While some progress has been made in understanding aquaculture vulnerability to climate change, more research is needed to quantify the driving processes and develop alternative aquaculture approaches and practices accordingly (FAO 2016). Consequently, environmental data limitations and accessibility can be major impediments for conducting biological response research.

A formidable research challenge is accurately and precisely determining how environmental changes will affect the physical culture environment and therefore what conditions need to be simulated to run valid experiments. One example are the knowledge gaps regarding changes in coastal marine culture environments. Biological studies often employ conditions of future scenarios (e.g. from the IPCC) based upon stable atmospheric conditions (Riebesell et al. 2010) rather than more variable coastal *in situ* environments where most marine aquaculture occurs. Ocean carbonate system parameters are highly variable in response to changes in temperature and salinity (Riebesell et al. 2010), and quantifying ocean acidification at regional scales requires knowledge of the natural variability of ocean carbonate ion concentrations at seasonal, annual, and even greater timescales (Friedrich et al. 2012). Currently,

there is little spatial and temporal resolution of pH, pCO₂, salinity, alkalinity, oxygen, and temperature data at biologically relevant scales and locations, which also restricts our ability to identify areas where contemporary climate change may have already occurred. A better understanding of true coastal environmental conditions and how these will change under climate change scenarios would allow more realistic biological stressor studies (Reum et al. 2016), while enabling greater predictive capacity of adaptation potential (Riebesell & Gattuso 2015).

Data deficiency or limited data accessibility is particularly acute for certain culture regions. Over half of the world's inland fish culture occurs in China (FAO 2014), primarily in ponds and river delta areas of southeast China (Wang et al. 2015). Lack of accessible relevant environmental data and only a few published regional climate predictions (e.g. Zhang et al. 2006) makes planning for climate change for the largest portion of global aquaculture production particularly daunting. In other regions, such as North America, high-resolution water quality data (e.g. temperature, oxygen) have been collected at culture sites for decades by industry, but there has been no concerted effort to compile and analyse these data to disentangle climate change effects which could assist in aquaculture adaptation (Reid & Gurney-Smith 2016).

Environmental data sharing is one option to maximize resources, but this requires continuity and significant coordinated effort. While day-to-day operational water quality data are often collected by farmers, longer-term data sets are needed to help quantify timelines of regional effects. Although long-term temporal data collection programmes have been quantitatively demonstrated to improve ecosystem models and forecasts (Giron-Nava et al. 2017), such datasets are rare. In the ocean environment, this type of data collection has often been the purview of marine stations, but many stations throughout the world are consistently in danger of closing, observations are not financially supported, and long-term data collection requires significant investment (Boero et al. 2015). In an era of fiscal restraint, resources and support from multiple-user communities, environmental stewards and data collection initiatives, including citizen science and regional partnerships (Chambers et al. 2017), are needed to ensure essential data collection and effective research. Such commitments are not trivial. Successful data-sharing initiatives have typically been backed by significant government support, resources, and coastal user communities (e.g. IOOS).

4.2. Expectations of increased research complexity

A proactive, multidisciplinary approach is required to reconcile multiple stressors in setting sustainable aquaculture development standards and designing adaptive management solutions (Sarà et al. 2018). Interdisciplinary research is not only crucial for the growth and development of aquaculture (Engle 2016), but also for understanding how multiple climate change stressors affect biological responses. Considerable complexity can therefore be expected in both aquaculture and climate change research, as they are inherently multidisciplinary. Disentangling the interplay of multiple stressors over relevant production timeframes introduces even further complexity, because it is generally not possible to extrapolate biological responses from single to multiple stressors (Riebesell & Gattuso 2015). Stressor combinations that are negatively synergistic may not always be fully additive; they may buffer (Humphreys 2017) or interact, and the greatest biological response may be attributed to a single dominant driver (Brennan & Collins 2015). Not only can stressor interaction confound biological responses, but a poor understanding of interactive stressor effects on water chemistry (Kroeker et al. 2014) can lead to misinterpretation of environmental impacts. The duration of experimental trials is another important consideration, as it has the potential to confound results. Selecting plants, algae, or animals from the wild or farms for short-term laboratory exposure to single stressor magnitudes does not allow good opportunity to assess adaptive responses, account for interaction between stressors, or reflect rates of change. Exploring longer-term biological responses and reconciling water chemistry changes under combined stressors will produce more meaningful research results.

The increased complexity of long-term multi-stressor research will result in an increased need for research resources and logistics. Increasing the number of stressors examined in environmentally relevant climate combinations results in a non-linear increase in the number of research culture units required. Even a simple multi-factorial study applying 3 stressors at 3 different magnitudes with 3 replicates per treatment requires 81 experimental units to be scientifically robust. As experimental needs for identical experimental culture units increase, the number of research facilities or culture sites that can accommodate large numbers of identical culture units becomes fewer. If the research involves 'large'

animals, proportionate culture size units (e.g. tanks) will require even greater space along with other resource pressures associated with the infrastructure (e.g. water, electricity, maintenance). Experiments across a full production cycle could be on the order of years and require prolonged occupation of culture units, reducing availability for other research, which would be further compounded in considerations of transgenerational research.

A comparison of environmental variation of historical data sets with present-day variation is needed to assess and help predict changes in frequency, scale of extremes, and phenology important to aquaculture stakeholders. Changes in environmental variation at different scales may affect aquaculture species through increased energy costs (Mangan et al. 2017), changes to infection dynamics (Rohr et al. 2011, Altizer et al. 2013, Groner et al. 2014), and physiological responses (Hori et al. 2013). It has been advocated that forthcoming IPCC reports include variability forecasts of surface temperatures and pH of both natural and anthropogenic drivers to better infer how key organisms and ecosystems respond to climate change (Schmidt & Boyd 2016).

This logical progression towards data sharing and the exponentially increasing number of climate change studies (Haunschild et al. 2016, Pedersen et al. 2016, Xu et al. 2016) suggest a need for the standardization of some research methodologies. Such standardization may require accepted data collection methodologies, best practices for experimental designs, standardized apparatus and perturbations, biological reference organisms for *in situ* studies, standardized reporting, consideration of natural fluctuations, and more realistic simulation of regional environmental conditions to facilitate research progress and data interpretation (Riebesell et al. 2010, Boyd 2013, Riebesell & Gattuso 2015, Boyd et al. 2016, Cornwall & Hurd 2016, Ellis et al. 2017). Progress in this respect has occurred in some fields such as ocean acidification research (Riebesell et al. 2010). However, standardization is not without its complications either. Concerns have been expressed that experiments run at standardized levels may not account for local adaptation, and adaptive plasticity may confound interpretation of inter-population differential responses (Vargas et al. 2017). Nevertheless, these considerations require wider discussion. An increase in data requirements, research complexity, and scope of biological response research are to be expected, which will ultimately benefit from collaborative approaches to facilitate and optimize research outcomes.

5. CONCLUSIONS

Reconciling current and pending effects of climate change on aquaculture may be daunting, but the potential for a range of adaptation options provides some encouragement. There are many options for aquaculture adaptation to climate change, from simple management changes to complex engineering or biotechnology solutions. These can be applied at the farm management level or be driven by wider governance initiatives. A range of adaptation success is to be expected. Some adaptation will be planned and initiated to precede impacts, while other adaptations will be reactive. Some aquaculture sectors may be unable to adapt, while there will be opportunities for new sectors.

The wide range of climate change adaptation options suggests the need for research engagement at all levels. Adaptation to climate change is apt to be most effective when supported by research; whether through traditional subsidised science (government, academia, environmental NGOs), industry-driven research and development, or trial and error solutions by farmers. Effective advancements in aquaculture adaptation will benefit from research teams comprising diverse expertise and a combination of empirical studies, modelling approaches, and observations at the farm level. When only complex adaptation solutions are viable, increased research complexity and expertise will translate into increased costs. As research becomes more expensive, it will become less accessible to small-scale farmers that comprise the bulk of global aquaculture production, and therefore research avenues are likely to require ongoing support by subsidised science.

Changes to our climate are being documented around the globe, daily. Collectively, these changes have profound implications for the aquaculture sector and its ability to contribute to long-term global nutrition and food security. Governments and aquaculture producers have little choice but to develop measures to adapt to changes in environmental conditions that are being thrust upon them. Adaptation is not mitigation, however, and ultimately, the long-term solution to climate change is the reduction of green-house gases along with other deleterious anthropogenic impacts.

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