REVIEW

Linkage between crustacean zooplankton and aquatic bacteria

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ABSTRACT: Bacteria and metazoan zooplankton (mainly crustaceans) are often viewed as 2 separate functional groups in the pelagic food webs indirectly linked via nutrient cycling and trophic cascades. Yet a zooplankton's body carries a high abundance of diverse bacteria, often at an equivalent concentration orders of magnitude higher than the ambient bacterial concentration. Zooplankton bodies are organic-rich micro-environments that support fast bacterial growth. Their physical-chemical conditions differ from those in the surrounding water and therefore select for different bacterial communities, including anaerobic bacteria that otherwise may not thrive in a well-oxygenated water column. The zooplankton body provides protection to the associated bacteria from environmental stresses similar to biofilms. Furthermore, migration by zooplankton enables rapid dispersal of bacteria over vast distances and across boundaries such as the pycnocline. In addition to live zooplankton, molts, fecal pellets, and carcasses of zooplankton all influence water column and benthic microbial communities in various ways. We review the recent advances in the study of (crustacean) zooplankton–bacteria interactions and discuss future research opportunities and challenges. Traditional aquatic microbial ecology emphasizes free-living bacteria, which represent only a fraction of the microbial world. By transcending disciplinary boundaries, microbial ecologists and zooplankton ecologists can work together to integrate the two disciplines and advance our understanding in aquatic microbial ecology.

KEY WORDS: Zooplankton · Bacteria · Zooplankton–bacteria interactions · Food web · Bacterial dispersal · Hotspots

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INTRODUCTION

Bacteria and metazoan zooplankton are important components of the pelagic food web and major contributors to pelagic biodiversity and biogeochemical processes. Although both inhabit the same environment, they are often treated as separate functional units only indirectly connected via nutrient cycling and trophic cascades (Azam & Malfatti 2007). The seeming lack of physical structures in the water column gives the impression that pelagic bacteria are living in a rather homogeneous environment, and respond only to pre-

dation, viral attack, and physical–chemical changes in the surrounding water. This 'free-living bacteria' point of view still dominates the primary literature and text books, and microbial ecologists often ignore zooplankton and other higher organisms as potential habitats for aquatic bacteria except for a few pathogenic species. Likewise, zooplankton ecologists tend to focus only on interactions between zooplankton and their prey or predators, and because bacteria do not usually fall within either category, they are often overlooked by zooplankton ecologists. Consequently, microbial ecology and zooplankton ecology are taught and prac-

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ticed as separate disciplines, and collaboration between the two is very limited.

In reality, however, bacteria and zooplankton can be closely linked in occurrence and ecological functions (Fig. 1). For example, the exoskeleton and gut lining of a copepod provide favorable surfaces for bacterial attachment (Nagasawa & Nemoto 1988, Pruzzo et al. 1996, Carman & Dobbs 1997; Fig. 2). Bacterial attachment is not limited to crustacean zooplankton, and symbiotic bacteria have been observed in non-crustacean zooplankton such as appendicularians (Flood 1991), rotifers (Selmi 2001), and jellyfish (Schuett & Doepke 2009). However, the biological characteristics of many of these bacteria remain obscure.

Readers who are familiar with current aggregate research will realize the importance of bacterial attachment and surface colonization in aquatic microbial ecology (Simon et al. 2002, Carrias & Sime- Ngando 2009). There is growing evidence that bacterial attachment to aggregates and phytoplankton has profound effects on nutrient cycling and microbial evolution (reviewed by Grossart 2010). However, there is a very important difference between aggregates and zooplankton: A detrital aggregate is composed of pri-

marily non-living particles, and it cannot regenerate itself or actively interact with bacteria. In contrast, a live zooplankton actively and continuously interacts with the bacteria that inhabit its body, leading to a dynamic and complex relationship between them. The few existing quantitative studies have shown that the equivalent bacterial concentrations associated with zooplankton, i.e. bacterial abundance per unit of zooplankton body volume, range between 10^7 and 10^{11} ml^{-1} (Table 1), which are orders of magnitude higher than that in the surrounding water. Zooplankton-associated bacteria include those that attach to the exterior and interior of zooplankton bodies (e.g. symbionts) and those that are temporarily present in the gut (ingested bacteria). Even at a population density of just 1 zooplankton l^{-1} , the abundance of these zooplankton-associated bacteria could rival that of free-living bacteria. Because of the heterogeneous distribution of zooplankton (Folt & Burns 1999), bacterial processes associated with zooplankton occur in spatial and temporal scales that cannot be properly characterized by conventional sampling techniques. Hence, what we have learned from traditional microbial ecological studies represents only the tip of an iceberg, and it is unset-

Fig. 1. Conceptual view of linkages between life cycles of zooplankton and aquatic bacteria. (a–d) Different types of zooplankton-derived organic matter which is subjected to bacterial colonization and degradation. Superscripts 1, 2, and 3 indicate the major processes by which zooplankton generate different types of organic matter. DOM: dissolved organic matter

Fig. 2. Bacteria attached to a zooplankton body surface can be directly observed by (a) phase contrast microscopy and (b) 4',6-diamidino-2-phenylindole (DAPI) epifluorescence microscopy. Labeling the bacteria with green fluorescent protein (GFP) also allows researchers to directly observe (c) attachment of live bacteria to the zooplankton body surface and (d) their presence inside the zooplankton gut

ting to realize how much of the aquatic microbial world remains unexplored by scientists when neglecting bacteria associated with zooplankton and other organisms.

Although earlier review papers on the subject do exist (e.g. Harris 1993, Carman & Dobbs 1997), they are largely limited to describing abundances and species compositions. Because earlier studies on zooplankton-associated bacteria relied on culturing techniques or biochemical assays, they likely have missed many of the bacterial phylotypes. Modern molecular techniques, in comparison, allow for phylogenetic investigation of these bacterial communities in greater detail, and experimental studies conducted in recent years have also shed light on the complex relationship between zooplankton and their associating bacteria. A more up-to-date review of the subject is therefore warranted.

In this review, we do not deal at length with pathogenic bacteria such as *Vibrio* species, for which a tight association with various zooplankton species has been discussed in detail by others (e.g. Huq et al. 1983, Cottingham et al. 2003). We also do not focus on the more obvious types of zooplankton–bacteria interactions such as predation, trophic cascades, or nutrient recy-

cling (including sloppy feeding), which have been extensively addressed in the literature. Instead, we focus on zooplankton as an important and unique microhabitat for many pelagic bacteria and the interactions between the two, and point to future challenges and opportunities in this research field. Much of the work described in this article is limited to crustacean zooplankton in coastal and estuarine environments and to a lesser extent limnetic systems, whereas comparable research on gelatinous zooplankton and microzooplankton in the open ocean environment remains scarce. Considering the ubiquity and large abundances of zooplankton in marine and freshwater environments, tight associations between zooplankton and bacteria can widely affect bacterial behavior, growth, and biogeochemical activities (Dattagupta et al. 2009). In times of rapid global change, there is an urgent need for advancing our understanding of zooplankton–bacteria interactions and how they would respond to future climate scenarios, a goal that can be achieved when scientists from both fields transcend conventional thinking and disciplinary boundaries, and begin to work together on the subject. It is our hope that this article serves as a catalyst for such collaboration.

Table 1. Equivalent bacterial concentrations associated with crustacean zooplankton based on literature data. Bacterial abundances are averaged total bacteria. Only data for whole animals are included. Equivalent bacterial concentration is calculated as (bacterial abundance zooplankton⁻¹) / zooplankton body volume. Ambient bacterial concentrations are included for comparison. NA: not available. See Table 2 for further details

Zooplankton type	Zooplankton body volume (ml)	Bacterial abundance (cells ind. ⁻¹)	Equivalent bacterial concentration (cells ml ⁻¹)	Ambient bacterial concentration (cells ml ⁻¹)	Source
<i>Acartia tonsa</i>	^a 2.5 × 10 ⁻⁵	^b 2.0 × 10 ⁵	8 × 10 ⁹	^c 1 × 10 ⁷	Hansen & Bech (1996)
^d <i>Artemia franciscana</i>	^e 3.1 × 10 ⁻⁵	^f 1.7 × 10 ⁴	5.6 × 10 ⁸	NA	Olsen et al. (2000)
^g <i>A. franciscana</i>	^e 3.1 × 10 ⁻⁵	^f 3.3 × 10 ³	9.7 × 10 ⁷	NA	Olsen et al. (2000)
>202 µm	^h 2.5 × 10 ⁻⁵	ⁱ 9.6 × 10 ⁶	3.8 × 10 ¹¹	NA	Heidelberg et al. (2002)
64–202 µm	^j 8.3 × 10 ⁻⁶	^k 2.9 × 10 ⁵	3.4 × 10 ¹⁰	NA	Heidelberg et al. (2002)
<i>A. tonsa</i>	^a 2.5 × 10 ⁻⁵	2.0 × 10 ³ – 4.5 × 10 ⁵	8 × 10 ⁷ – 1.8 × 10 ¹⁰	NA	Tang (2005)
<i>Calanus helgolandicus</i> + <i>C. finmarchicus</i>	^l 5.5 × 10 ⁻⁴	1.9 × 10 ⁵	3.5 × 10 ⁸	^m 1.4 × 10 ⁵ – 5 × 10 ⁵	Møller et al. (2007)
<i>Daphnia cucullata</i>	ⁿ 1.2 × 10 ⁻⁴	1.0 × 10 ⁵ – 3.9 × 10 ⁵	0.8 × 10 ⁹ – 3.2 × 10 ⁹	^o 3.4 × 10 ⁶	Tang et al. (2009a)
<i>Eudiaptomus gracilis</i>	^p 1.2 × 10 ⁻⁴	1.7 × 10 ⁵ – 4.3 × 10 ⁵	2.0 × 10 ⁹ – 5.1 × 10 ⁹	3.4 × 10 ⁶	Tang et al. (2009a)
<i>Diaphanosoma brachyurum</i>	^q 2.5 × 10 ⁻⁵	3.3 × 10 ⁵	1.3 × 10 ¹⁰	3.4 × 10 ⁶	Tang et al. (2009a)

^aBody volume of *A. tonsa* was estimated by Tang et al. (2001)
^bBacterial abundance is from body surface and intestine combined based on acridine orange direct count
^cAmbient bacterial concentration is based on acridine orange direct count
^dData are for 2 d old starved nauplii with mean body length 1.07 mm
^eBody volume is estimated from body length as $V = a L^3$ using scaling factor (*a*) for *A. tonsa* ($V = 2.5 \times 10^{-5}$ ml, $L = 1$ mm; Tang et al. 2001)
^fBacterial abundance is for 'Bacteria'
^gData are for 2 d old fed nauplii with mean body length 1.07 mm
^hBody volume per zooplankton is assumed to be same as for *A. tonsa*
ⁱBacterial abundance is average 'Bacteria' abundance for zooplankton >202 µm
^jBody volume per zooplankton is assumed to be one-third of that for *A. tonsa*
^kBacterial abundance is average 'Bacteria' abundance for zooplankton 64–202 µm
^lBody volume of *C. helgolandicus* and *C. finmarchicus* is estimated from body length (~2.8 mm) using scaling factor for *A. tonsa*
^mAmbient bacterial concentration is based on flow cytometry
ⁿBody volume of *D. cucullata* is estimated from body length (~1.7 mm) using scaling factor for *A. tonsa*
^oAmbient bacterial concentration is based on 4',6-diamidino-2-phenylindole (DAPI) direct count
^pBody volume of *E. gracilis* is estimated from body length (~1.5 mm) using scaling factor for *A. tonsa*
^qBody volume of *D. brachyurum* (body length ~1 mm) is assumed to be same as for *A. tonsa*

BACTERIAL ASSOCIATION WITH LIVE ZOOPLANKTON

Chemotactic bacteria actively attach and colonize surfaces (e.g. Kjørboe et al. 2002 for aggregates; Carman & Dobbs 1997 for metazoans) by following chemical and perhaps hydrodynamic cues. Several studies have demonstrated that various zooplankton species, especially crustaceans, from different aquatic environments are densely colonized by bacteria (Table 2). However, many of the early studies did not give information on the phylogenetic affiliation of these attached bacteria, were biased by the use of culture-dependent methods, or were restricted to specific bacterial phylotypes.

Recent studies using culture independent methods have revealed much more diverse bacterial communities associated with both marine and freshwater zooplankton. Using CARD-FISH, Peter & Sommaruga (2008) found 4 groups of bacteria in freshwater crustacean zooplankton guts: *Cytophaga-Flavobacteria* and *Alpha*-, *Beta*-, and *Gammaproteobacteria*. Using a

combination of culturing techniques and 16S rRNA sequencing, Schuett & Doepke (2009) identified 21 bacterial species from 4 cnidarian species, including 2 planktonic forms. These bacteria formed 4 closely related groups of *Gammaproteobacteria* (*Pseudoalteromonas tetraodonis*/*P. elyakovii*/*P. haloplanctis*; *Shewanella sairae*/*S. marinintestina*; *S. waksmanii*/*S. surugaensis*/*S. kaireiae*; and *Vibrio splendidus*/*V. lentus*/*V. tasmaniensis*/*V. kanaloae*), in addition to species belonging to *Bacillus subtilis* (*Firmicutes*), *Ilyobacter psychrophilus* (*Fusobacterium*), and *Arcobacter butzleri* (*Epsilonproteobacteria*). Eleven of these isolates have been recently described as novel species. Additionally, 4 of the sequenced 16S rDNA fragments from 2 *Cyanea* species had extremely low relationships to their next relatives and hence represented members of the endobiotic 'terra incognita.' Based on 16S rRNA, Grossart et al. (2009) recovered 36 discrete phylogenetic units from 2 freshwater crustacean zooplankton species that belonged to 6 major bacterial groups (*Actinobacteria*, *Firmicutes*, *Bacteroidetes*,

Table 2. Bacterial abundances and taxa associated with zooplankton. Numbers of bacteria zooplankton⁻¹ are mean values ($\pm$ SD where available) either directly taken from the text or recalculated based on information in the cited references. ND: not determined, WAH: whole animal homogenate, DMSP: dimethylsulfoniopropionate, TCBS: thiosulfate citrate bile sucrose agar, SEM: scanning electron microscopy, CFU: colony-forming units, AO: acridine orange, MPN: most probable number, DAPI: 4',6-diamidino-2-phenylindole. Methods used to quantify bacteria are also included

Zooplankton taxon	Environment	Sample type	Number of bacteria ind. ⁻¹	Bacterial taxa	Method used	Source
<i>Acartia</i> sp.	Anclote River, FL	Body surface	60	Marine heterotrophs	Marine agar 2216	Sochard et al. (1979)
<i>Acartia</i> sp.	28° 35' N, 89° 00' W	Body surface	100	Marine heterotrophs	Marine agar 2216	Sochard et al. (1979)
<i>Pontellopsis regalis</i>	27° 00' N, 86° 00' W	Body surface	1000	Marine heterotrophs	Marine agar 2216	Sochard et al. (1979)
<i>Labidocera aestiva</i>	27° 30' N, 88° 30' W	Body surface	120	Marine heterotrophs	Marine agar 2216	Sochard et al. (1979)
<i>L. aestiva</i>	28° 35' N, 89° 00' W	Body surface	20–300	Marine heterotrophs	Marine agar 2216	Sochard et al. (1979)
<i>L. aestiva</i>	29° 19' N, 88° 30' W	Body surface	3000–30000	Marine heterotrophs	Marine agar 2216	Sochard et al. (1979)
<i>Centropages furcatus</i>	29° 19' N, 88° 30' W	Body surface	>100	Marine heterotrophs	Marine agar 2216	Sochard et al. (1979)
<i>Acartia</i> sp.	Anclote River, FL	Body surface	3	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>Acartia</i> sp.	28° 35' N, 89° 00' W	Body surface	<1	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>Acartia</i> sp.	29° 19' N, 88° 30' W	Body surface	25	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>Pontellopsis regalis</i>	27° 00' N, 86° 00' W	Body surface	13	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>Labidocera aestiva</i>	27° 30' N, 88° 30' W	Body surface	120	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>L. aestiva</i>	28° 35' N, 89° 00' W	Body surface	<1	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>L. aestiva</i>	29° 19' N, 88° 30' W	Body surface	30–2900	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>Centropages furcatus</i>	29° 19' N, 88° 30' W	Body surface	25	<i>Vibrio</i> -like	TCBS agar	Sochard et al. (1979)
<i>Diaptomus nevadensis</i>	Goose Lake, NE	Body surface	<20 – >200	ND	SEM	Holland & Hergenrader (1981)
<i>D. sicilis</i>	Goose Lake, NE	Body surface	<20–200	ND	SEM	Holland & Hergenrader (1981)
<i>D. siciloides</i>	Branched Oak Lake, NE	Body surface	20 – >200	ND	SEM	Holland & Hergenrader (1981)
<i>Diaptomus</i> sp.	Roundup Lake, NE	Body surface	20 – >200	ND	SEM	Holland & Hergenrader (1981)
<i>Artemia</i> sp.	Lab culture	Internal body	50	Aerobic heterotrophs	Spread plate technique	Austin & Allen (1982)
<i>Tigriopus fulvus</i>	Tyrrhenian Sea	Body surface (male)	9.6 $\pm$ 15.2 $\times$ 10 ³	Heterotrophic bacteria	Marine agar 2216	Carli et al. (1993)
<i>T. fulvus</i>	Tyrrhenian Sea	Body surface (female)	1.7 $\pm$ 2.5 $\times$ 10 ⁴	Heterotrophic bacteria	Marine agar 2216	Carli et al. (1993)
<i>T. fulvus</i>	Tyrrhenian Sea	Surface (egg sac)	3.2 $\pm$ 4.0 $\times$ 10 ⁴	Heterotrophic bacteria	Marine agar 2216	Carli et al. (1993)
<i>T. fulvus</i>	Tyrrhenian Sea	Body surface (male)	1.5 $\pm$ 2.2 $\times$ 10 ³	<i>Vibrio alginolyticus</i>	TCBS agar	Carli et al. (1993)
<i>T. fulvus</i>	Tyrrhenian Sea	Body surface (female)	1.4 $\pm$ 2.3 $\times$ 10 ³	<i>V. alginolyticus</i>	TCBS agar	Carli et al. (1993)
<i>T. fulvus</i>	Tyrrhenian Sea	Surface (egg sac)	742 $\pm$ 1103	<i>V. alginolyticus</i>	TCBS agar	Carli et al. (1993)
<i>Acartia tonsa</i>	Lab culture	Body surface	1422 $\pm$ 108	<i>Vibrio</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Body surface	553 $\pm$ 42	<i>Alteromonas</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Body surface	2291 $\pm$ 174	<i>Pseudomonas</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Body surface	76 $\pm$ 6	<i>Photobacterium</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Body surface	790 $\pm$ 60	<i>Aeromonas</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Body surface	2765 $\pm$ 210	<i>Cytophaga/ Flavobacterium</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	275 $\pm$ 100	<i>Vibrio</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	132 $\pm$ 48	<i>Alteromonas</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	55 $\pm$ 20	<i>Pseudomonas</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	220 $\pm$ 80	<i>Photobacterium</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	198 $\pm$ 72	<i>Aeromonas</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	33 $\pm$ 12	<i>Cytophaga/ Flavobacterium</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	165 $\pm$ 60	<i>Corynebacterium</i>	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	22 $\pm$ 8	Other	CFU	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Body surface	7.9 $\times$ 10 ⁴	ND	AO direct count	Hansen & Bech (1996)
<i>A. tonsa</i>	Lab culture	Intestine	12.1 $\times$ 10 ⁴	ND	AO direct count	Hansen & Bech (1996)
<i>Artemia franciscana</i> nauplii	Lab culture (starved)	WAH	1.7 $\pm$ 1.1 $\times$ 10 ⁴	Bacteria	M65 agar	Olsen et al. (2000)

(Table continued on next page)

Table 2 (continued)

Zooplankton taxon	Environment	Sample type	Number of bacteria ind. ⁻¹	Bacterial taxa	Method used	Source
<i>A. franciscana</i> nauplii	Lab culture (fed)	WAH	$3.3 \pm 3.7 \times 10^3$	Bacteria	M65 agar	Olsen et al. (2000)
<i>A. franciscana</i> nauplii	Lab culture (starved)	WAH	2550 ± 71	Hemolytic bacteria	Blood agar	Olsen et al. (2000)
<i>A. franciscana</i> nauplii	Lab culture (fed)	WAH	960 ± 1117	Hemolytic bacteria	Blood agar	Olsen et al. (2000)
<i>A. franciscana</i> nauplii	Lab culture (starved)	WAH	$8.3 \pm 8.1 \times 10^3$	<i>Vibrio</i>	TCBS agar	Olsen et al. (2000)
<i>A. franciscana</i> nauplii	Lab culture (fed)	WAH	$2.2 \pm 2.7 \times 10^3$	<i>Vibrio</i>	TCBS agar	Olsen et al. (2000)
<i>Acartia tonsa</i>	Lab culture	WAH	$0.9 \times 10^3 - 2.8 \times 10^4$	DMSP-consuming bacteria	MPN	Tang et al. (2001)
Zooplankton >202 µm	Choptank River, MD	WAH	$9.6 \pm 23.3 \times 10^6$	Bacteria	Flow cytometry	Heidelberg et al. (2002)
Zooplankton >202 µm	Choptank River, MD	WAH	$1.3 \pm 3.6 \times 10^6$	Gammaproteobacteria	Flow cytometry	Heidelberg et al. (2002)
Zooplankton >202 µm	Choptank River, MD	WAH	$6.2 \pm 19.1 \times 10^5$	<i>Vibrio-Photobacterium</i>	Flow cytometry	Heidelberg et al. (2002)
Zooplankton >202 µm	Choptank River, MD	WAH	$1.3 \pm 2.3 \times 10^3$	<i>V. vulnificus</i>	Flow cytometry	Heidelberg et al. (2002)
Zooplankton >202 µm	Choptank River, MD	WAH	577 ± 714	<i>V. cholerae - V. mimicus</i>	Flow cytometry	Heidelberg et al. (2002)
Zooplankton >202 µm	Choptank River, MD	WAH	28 ± 38	<i>V. cincinnatiensis</i>	Flow cytometry	Heidelberg et al. (2002)
Zooplankton 64–202 µm	Choptank River, MD	WAH	$2.9 \pm 5.12 \times 10^5$	Bacteria	Flow cytometry	Heidelberg et al. (2002)
Zooplankton 64–202 µm	Choptank River, MD	WAH	$1.5 \pm 2.3 \times 10^4$	Gammaproteobacteria	Flow cytometry	Heidelberg et al. (2002)
Zooplankton 64–202 µm	Choptank River, MD	WAH	$6.4 \pm 10.5 \times 10^3$	<i>Vibrio-Photobacterium</i>	Flow cytometry	Heidelberg et al. (2002)
Zooplankton 64–202 µm	Choptank River, MD	WAH	22.6 ± 33.2	<i>V. vulnificus</i>	Flow cytometry	Heidelberg et al. (2002)
Zooplankton 64–202 µm	Choptank River, MD	WAH	29.7 ± 38.6	<i>V. cholerae - V. mimicus</i>	Flow cytometry	Heidelberg et al. (2002)
Zooplankton 64–202 µm	Choptank River, MD	WAH	1.0 ± 1.9	<i>V. cincinnatiensis</i>	Flow cytometry	Heidelberg et al. (2002)
<i>Acartia tonsa</i>	Lab culture	WAH	$2.0 \times 10^3 - 4.5 \times 10^5$	ND	MPN	Tang (2005)
<i>Calanus helgolandicus</i> and <i>C. finmarchicus</i>	North Sea	WAH	$1.9 \pm 0.2 \times 10^5$	Alphaproteobacteria	DAPI	Møller et al. (2007)
<i>Daphnia cucullata</i>	Lake Stechlin	WAH	$1.0 \pm 0.2 \times 10^5$	ND	SybrGold	Tang et al. (2009a)
<i>D. cucullata</i>	Lake Stechlin	WAH	$3.9 \pm 1.5 \times 10^5$	ND	SybrGold	Tang et al. (2009a)
<i>Eudiaptomus gracilis</i>	Lake Stechlin	WAH	$1.7 \pm 0.7 \times 10^5$	ND	SybrGold	Tang et al. (2009a)
<i>E. gracilis</i>	Lake Stechlin	WAH	$4.3 \pm 1.4 \times 10^5$	ND	SybrGold	Tang et al. (2009a)
<i>Diaphanosoma brachyurum</i>	Lake Stechlin	WAH	$3.3 \pm 0.2 \times 10^5$	ND	SybrGold	Tang et al. (2009a)

Alpha-, Beta-, and Gammaproteobacteria). Using the same technique, Tang et al. (2009b) identified 37 discrete phylogenetic units from the marine copepod *Acartia tonsa* that belonged to 3 major bacteria groups (Alpha- and Gammaproteobacteria, Bacteroidetes). Surprisingly, Møller et al. (2007) found only 3 different *Roseobacter* species on *Calanus* spp. in the North Sea using similar methods. These findings indicate that Alphaproteobacteria are an important component of the zooplankton-associated bacterial community, which can vary greatly in richness and phylotype composition among zooplankton species, sampling location, and time.

Zooplankton-associated bacteria must also exist in a free-living stage if they are to maintain a viable population beyond the lifetime of the host. For many symbiotic bacteria such as *Vibrio fischeri* on the squid *Euprymna scolopes* (Koropatnick et al. 2004), it has been shown that the bacterial symbiont is frequently released into the surrounding water to allow for colonization of other hosts. Earlier studies have shown that zooplankton bodies and the surrounding water share similar bacterial groups but in different proportions (Sochard et al. 1979, Delille & Razouls 1994, Hansen & Bech 1996), suggesting an active exchange of bacteria between the 2 compartments, but that their different physical-chemical conditions tend to favor different bacterial groups.

Separation of epibionts and gut microflora is difficult in practice, but the latter are expected to be more directly affected by the host's feeding activity. Harris (1993) proposed a conceptual model that differentiates between transient and resident gut microflora associated with invertebrates: Resident bacteria are persistently present in the gut, whereas transient bacteria do not form stable populations within the gut. Empirical evidence in support of this view is provided by 2 recent studies. When the copepod *Acartia tonsa* was feeding on different axenic phytoplankton diets, the bacterial diversity associated with the copepod decreased and converged, indicating the presence of a rather stable resident bacterial community (Tang et al. 2009b). In

contrast, the bacterial diversity greatly diverged and several species such as *Pseudoalteromonas*, *Sulfitobacter*, and *Roseobacter* only appeared when the copepod was feeding on xenic phytoplankton diets, indicating that many of the transient bacteria either actively attached to the animal's body surfaces or were passively ingested (Tang et al. 2009b). Similarly, in a field study, Grossart et al. (2009) observed that the bacterial diversity associated with freshwater zooplankton (*Thermocyclops oithonoides* and *Bosmina coregoni*) decreased after gut evacuation. In addition, these authors found that the magnitude of change in bacterial diversity induced by gut evacuation was smaller for zooplankton from an oligotrophic lake versus those from a eutrophic lake. This suggests that the resident bacterial communities behave differently even for the same zooplankton species when living in different environments.

ZOOPLANKTON AS MICROBIAL REACTORS

Both resident and transient bacteria may take advantage of the rich organic environments provided by zooplankton and attain higher growth rates than free-living bacteria. For example, Carman (1994) demonstrated that nitrogen excretion by copepods stimulates the growth of epibiotic bacteria. Tang (2005) further developed the concept that live copepods act as a reactor for bacterial colonization and growth. In a nutritionally dilute environment, copepods through feeding concentrate organic matter into their guts, which their gut microflora can exploit to attain high growth rates. For gut bacteria (transient and resident) to exploit the ingested materials, they must first survive digestion by the host. Several studies (e.g. King et al. 1991, Plante & Shriver 1998) have shown that many bacteria, including cyanobacteria (Friedland et al. 2005), can survive zooplankton gut passage, and may then gain access to specific resources that are otherwise limiting in the surrounding water. For example, dimethylsulfoniopropionate (DMSP) is produced by many marine phytoplankton species (Keller & Korjeff-Bellows 1996) and is a substrate for DMSP-consuming bacteria (DCB; e.g. Diaz et al. 1992, Visscher et al. 1992). However, direct release of DMSP from actively growing phytoplankton is negligible (Keller & Korjeff-Bellows 1996), and the half-saturation constant for free-living DCB in coastal to oligotrophic waters (24 to >500 nM) is 5- to 50-fold higher than the dissolved DMSP concentrations (Ledyard & Dacey 1996), indicating substrate limitation in ambient water. Through feeding, zooplankton can liberate DMSP from phytoplankton cells and accumulate it in their guts to μM to mM levels (Tang et al. 1999), which would allow for

maximum turnover by DCB. DCB have indeed been isolated from copepod bodies, and their abundance increased 17 to 30 times when the copepods were fed DMSP-containing food (Tang et al. 2001).

Most copepods have a short gut passage time such that ingestion is tightly coupled with egestion (Dam & Peterson 1988, Besiktepe & Dam 2002). Hence gut bacteria also face the challenge of being flushed out during egestion. The presence of gut bacteria in zooplankton even after gut evacuation has been reported (Peter & Sommaruga 2008). Kirn et al. (2005) identified a single bacterial protein that mediates intestinal colonization and attachment to zooplankton and other organisms by binding to a surface sugar on epithelial cells. Nevertheless, the number of habitable sites in the gut is presumably limited, and the amount of enteric bacteria cannot increase indefinitely. Tang (2005) hypothesized that in the absence of an input of foreign bacteria, the amount of enteric bacteria should follow a predictable relationship with a zooplankter's ingestion rate determining the balance between growth as stimulated by substrate input and loss due to mortality and egestion. This prediction was subsequently confirmed in an experimental study using the copepod *Acartia tonsa* feeding on various axenic algal diets (Fig. 3)

In the same study, Tang (2005) conservatively estimated a growth rate of 0.89 d^{-1} for the enteric bacteria, which is much higher than the global average growth rate of 0.05 to 0.3 d^{-1} for pelagic bacteria (Ducklow 1999). The author also observed that bacteria were released together with other fecal materials by the copepods. Because the copepod was fed axenic food in

Fig. 3. Bacterial abundance associated with the copepod *Acartia tonsa* as a function of the copepod's ingestion rate on an axenic diet of *Rhodomonas salina*, *Dunaliella tertiolecta*, or *Thalassiosira weissflogii*. Different symbols represent replicate experiments. Curves are overall fit to the data for each diet type. Without an input of foreign bacteria, feeding by the copepod introduces organic nutrients to its gut, where bacteria can exploit the nutrients and increase in abundance until their growth is balanced or exceeded by loss due to mortality and egestion. Data from Tang (2005)

the experiments, the release of bacteria via defecation suggests that actively feeding zooplankton function as microincubators and contribute new bacteria to the water column. Although this contribution appears to be small (<0.1%) relative to the ambient bacterial standing stock, defecation and detachment may serve as important mechanisms for inoculating different parts of the water column (Grossart et al. 2010).

ZOOPLANKTON AS A SELECTIVE FORCE FOR BACTERIA

The exterior as well as the interior of a zooplankter provide specific habitats for aquatic bacteria. For example, bacteria attached to the surface of a zooplankter may experience a changing hydrodynamic environment due to the host's movement (Lawrence & Caldwell 1987) and must also adapt to the specific surface property (Dunne 2002). The reactive dissolved organic carbon (DOC) pool in pelagic systems is mainly composed of molecules such as carbohydrates, amino acids, fatty acids, hydrocarbons, and steroids, and they usually account for a small fraction of total DOC in the water column (<15%; Benner 2002). In contrast, reactive DOC (including amino acids and carbohydrates) in the vicinity of zooplankton occurs in much higher amounts due to active excretion and sloppy feeding of the zooplankton, and this could become available for attached and even free-living bacteria (Peduzzi & Herndl 1992, Hansson & Norrman 1995, Møller et al. 2003, Møller 2005). In an experimental mesocosm study, however, Kragh et al. (2006) demonstrated that the presence of zooplankton leads to an accumulation of less labile aldoses, indicating that zooplankton affect the quantity and quality of the DOC pool in the surrounding water. As a result, release and accumulation of specific DOC compounds in the presence of zooplankton potentially selects for specific bacterial phylotypes, e.g. opportunistic bacteria (Eilers et al. 2000, Cottrell & Kirchman 2003).

Enteric bacteria, on the other hand, will experience high organic matter and nutrient concentrations and large changes in pH and oxygen availability during gut passage. The organic-rich, low oxygen environment inside a zooplankter's gut and fecal pellets may favor anaerobic bacterial processes that are otherwise not feasible in the oxygenated water column. Indeed, an important observation in the past decades is the presence of strict anaerobes inside the zooplankton gut and fecal pellets (Bianchi et al. 1992, Marty 1993, Proctor 1997, Braun et al. 1999). Globally it is estimated that 12% of the primary production is consumed, directly or indirectly, by metazoan zooplankton (Calbet 2001). Due to the possible variety of bacterial metabolism

associated with zooplankton, these microenvironments represent a platform allowing for inter-linkages of microbial processes in processing and degrading a large fraction of the organic matters in marine and freshwater environments, and support anaerobic processes in an otherwise aerobic environments (Stief et al. 2009). As such they have potentially significant effects on global biogeochemical cycles.

The chemical composition of the zooplankton body as a potential source for bacterial substrates substantially differs from that of phytoplankton biomass. Due to the lack of large amounts of carbohydrate storage products, carbon content of zooplankton is usually lower in relation to nitrogen and phosphorus (Beers 1966). Hence, the zooplankton body itself is a preferential bacterial substrate (see below under 'Zooplankton carcasses'). The chemical composition and availability of zooplankton-derived particulate and dissolved organic matters therefore has major implications for bacterial dynamics and phylotype selection not only on the zooplankton but also in the surrounding water.

ZOOPLANKTON AS A REFUGE FOR BACTERIA

Previous studies have shown that potentially pathogenic bacteria, e.g. *Vibrio* spp., find suitable growth conditions in zooplankton bodies (Colwell 1996). Free-living bacteria are constantly subject to environmental hazards such as predation, viral lysis, and harmful radiation and chemicals, and a zooplankton body may provide a refuge for bacteria against these external hazards. Endosymbiotic bacteria that are commonly found in protozoa have been shown to survive disinfectants that normally kill free-living bacteria, hence posing a great threat to public health (Barker & Brown 1994, Bichai et al. 2008).

Various disinfection technologies are used commercially to treat ballast water in an attempt to curb the global spread of harmful organisms. Recent experiments by K. W. Tang et al. (unpubl.), however, suggest that bacteria associated with cladocerans and copepods could survive conventional ballast water treatments such as heat, UV radiation, and ozonation even though free-living bacteria and zooplankton themselves did not (Fig. 4). Subsequent discharge of treated ballast water into coastal zones may unknowingly inoculate the local water with potential pathogens carried by zooplankton. Protection by zooplankton may also affect the evolution of local bacterial communities under the influence of other environmental stresses.

While the zooplankton hosts may provide refuge for the associated bacteria, the hosts themselves may also benefit from these bacteria. For example, Rico-Mora &

Fig. 4. Zooplankton as a refuge for bacteria. Seawater was exposed to UV for 2 h on Day 0 with or without the copepod *Acartia tonsa*. The copepod died after the exposure. Free-living bacteria in the absence of copepod (O) declined significantly after the initial exposure. However, free-living bacteria in the presence of copepod (●) decreased only slightly and recovered by Day 3. Attached bacteria (bars) also recovered after 3 d. Data from K. W. Tang et al. (unpubl.)

Voltolina (1995) showed that inoculation of the brine shrimp *Artemia* sp. with bacterial isolates from diatom cultures could significantly improve the survivorship of the animals relative to the bacteria-free control and those challenged with pathogenic *Vibrio* spp. In addition to being a potential food source or aiding the host's digestion (Fong & Mann 1980, Wainwright & Mann

1982), attached benign bacteria may have a probiotic effect and protect the host against pathogenic bacteria (Verschuere et al. 1999, 2000).

BACTERIAL DISPERSAL BY ZOOPLANKTON

Many zooplankton species perform strong diel, seasonal, or ontogenetic vertical migration, sometimes up to thousands of meters (Lampert 1989, Visser & Jónasdóttir 1999, Kobari & Ikeda 2001). In a stratified water column, a migrating zooplankter will be exposed to and subsequently acquire different bacteria between different water layers, and may act as a conveyor belt for dispersing bacteria over vast distances and across boundaries such as the pycnocline. Sinking aggregates perform a similar function (Turley & Mackie 1995) but primarily in the downward direction (in addition to horizontal advection). In comparison, a zooplankter can migrate both upward and downward multiple times during its lifetime. Therefore, this conveyor belt works much longer than an aggregate and facilitates bacterial exchange in both directions. In a recent experimental study, Grossart et al. (2010) used stratified migration columns and confirmed that migrating zooplankton (*Daphnia magna*) dispersed bacteria between the separate water layers in both upward and downward directions (Fig. 5), hence confirming the validity of the conveyor belt hypothesis.

Fig. 5. Vertical dispersal of bacteria by migrating zooplankton. Experiments were conducted in clear plexiglass columns with a stable salinity stratification (2 psu). Green fluorescent protein (GFP)-labeled bacteria (*Pimelobacter* sp.) were initially added to either (a) the lower or (b) the upper layer. Forty individuals of *Daphnia magna*, which were positively phototactic, were guided to complete 1 migration cycle (up + down) every 2 h for 8 h by a focused light source. Bacterial concentration (cells ml⁻¹; mean ± SE) in the receiving end of the column increased significantly faster when *Daphnia* were present in both experiments, confirming that migrating *Daphnia* facilitate vertical dispersal of bacteria in both upward and downward directions. Data from Grossart et al. (2010)

Research on aggregates has shown that bacteria attached to surfaces are capable of quorum sensing via homoserine lactone communication (Gram et al. 2002), which allows them to synchronize specific metabolic processes to maximize their metabolic rate and growth efficiency. In the case of probiotic as well as pathogenic bacteria, this will have important consequences for the environment. Multidrug resistance can frequently be found in bacteria on fish, and it is believed that the increased tolerance to antibiotics and heavy metals is related to the close proximity of resistant and non-resistant bacteria, which allows for increased horizontal gene exchange (e.g. Pathak & Gopal 2005). Long-range migration by zooplankton and other animal hosts may help the spread of drug-resistant genotypes across a broad geographical range even among otherwise isolated water masses, posing a hidden threat to ecosystem and human health.

OTHER MODES OF ZOOPLANKTON–BACTERIA ASSOCIATION

Zooplankton feces

Approximately a third of the food ingested by pelagic organisms is egested as feces (Parsons et al. 1977). Zooplankton such as copepods and salps produce fast-sinking fecal pellets, from tens to thousands m d^{-1} (Bruland & Silver 1981, Komar et al. 1981, Caron et al. 1989), and thus contribute substantially to the 'biological pump' (Cherry et al. 1978). On the other hand, fecal matter may be remineralized by bacteria within the water column (Gowing & Silver 1983, Jacobsen & Azam 1984, Caron et al. 1989), reducing the overall vertical flux. This process can be further enhanced by physical breakage of the pellets by zooplankton (Lampitt et al. 1990). Differences in the physical structure of the pellets may also lead to different remineralization rates. For example, a copepod fecal pellet is bound by a peritrophic membrane and may remain intact for a longer time, whereas a salp fecal pellet lacks a membrane and tends to be more brittle.

Fecal matter released by zooplankton often already carries large amounts of bacteria (Turner 1979). For example, the reported equivalent total bacterial concentrations in copepod fecal pellets are 10^{10} to 10^{11} cells ml^{-1} for *Acartia tonsa* (Hansen & Bech 1996, Thor et al. 2003) and 10^3 to 10^5 cells ml^{-1} for *Temora stylifera* (Delille & Razouls 1994). The latter study also showed that the frequency of dividing bacterial cells and mean bacterial cell volumes increased, respectively, from 3% and $0.12 \mu\text{m}^3$ in seawater to >6% and $0.24 \mu\text{m}^3$ in fecal pellets. Previously, it had been thought that bacterial colonization of zooplankton fecal pellets is

mainly initiated from the outside by chemotactic bacteria in the surrounding water (Honjo & Roman 1978, Turner 1979, Jacobsen & Azam 1984). However, Gowing & Silver (1983) demonstrated that most of the bacterial colonization is from the inside by enteric and digestion-resistant bacteria. The presence of bacteria in fecal materials produced by copepods feeding on axenic food also confirmed that enteric bacteria can be passed onto fecal pellets (Tang 2005). On the other hand, a study with deposit-feeders showed that bacterial proliferation on egesta was faster than could be explained by growth alone (Plante & Wilde 2001), indicating that colonization by external bacteria must have also occurred at the same time.

Zooplankton molts

Chitin is the second-most abundant biogenic polysaccharide on earth after cellulose and hence is among the most abundant and important sources of nutrients and energy in the marine environment (Gooday 1990). Chitin is a major component of the cuticles and exoskeletons of worms, mollusks, and arthropods (Jeuniaux 1982). The annual biosynthesis of chitin has been estimated at 1.4 billion tons in all aquatic environments combined (Cauchie 2002). The dry mass of chitin produced by the planktonic crustacean *Euphausia pacifica* through molting alone is ca. 5 to 12 million tons annually (Jerde & Lasker 1966, Jeuniaux 1971). Yu et al. (1991) asserted that chitin production by copepods and other organisms leads to a continuous 'rain' of particulate organic matter. Furthermore, molts (exuviae) of crustacean zooplankton significantly contribute to the formation of macroscopic organic aggregates in marine (Gardner 1977) and freshwater (Grossart et al. 1997, Grossart & Simon 1998) systems. From an ecological point of view, chitin plays a key role in the biogeochemical cycles of both C and N, and the rates of chitin production and degradation influence C and N pools and their availability (Poulicek et al. 1998, Pruzzo et al. 2008).

In his pioneering *in situ* study of chitin degradation (practical grade chitin from crustacean shells) using the mesh bag technique, Kirchner (1995) revealed that 90% of the initial substrate was lost after 3 mo of incubation in seawater at 10 to 18°C. Other studies also found that several marine bacteria are capable of degrading chitin (Bassler et al. 1991, Keyhani & Roseman 1999, Ramaiah et al. 2000, Tilly et al. 2001, Itoi et al. 2006), but none of them used natural zooplankton molts as a substrate. Studying bacterial colonization of lake snow in Lake Constance, Weiss et al. (1996) showed that aggregates containing a high fraction of zooplankton exoskeleton during the clear water phase were mainly

colonized by *Proteobacteria* of the *Alpha* (5 to 32%), *Beta* (20 to 60%), and *Gamma* subclasses (6 to 42%). In the Delaware estuary, Cottrell & Kirchman (2003) found chitinase genes in a handful of bacterial groups including *Enterobacteriaceae*, *Alteromonadaceae*, *Vibrionaceae*, and *Alphaproteobacteria* of the *Roseobacter* clade. However, except for *Vibrio* spp., almost nothing is known about bacterial colonization of zooplankton molts. Tamplin et al. (1990) showed that 4 of 5 clinical *V. cholerae* O1 strains and endogenous bacteria preferentially attached to zooplankton molts rather than whole animals. *V. cholerae* growing on natural chitin surfaces show coordinated gene expression involved with chitin chemotaxis and adherence as well as the transport and assimilation of N-acetylglucosamine (GlcNAc; Meibom et al. 2004), suggesting that this species is well adapted to using this highly abundant carbon and nitrogen source in aquatic environments.

Using methylumbelliferyl-N, N'-diacetyl-chitobioside, and ¹⁴C-chitin, Kirchman & White (1999) showed that bacterial hydrolysis of chitin was always faster than mineralization, resulting in the release of GlcNAc to the surrounding water. On average, however, chitin hydrolysis and uptake accounts for only 10% of the total bacterial production (Kirchman & White 1999). Interestingly, Riemann & Azam (2002) showed not only that roughly a third of all marine bacteria take up GlcNAc, but also that these bacteria are predominantly facultative anaerobes. The capability of fermentative metabolism could be an important adaptation to low oxygen microenvironments in the pelagial associated with molts, fecal pellets, and zooplankton gut.

Zooplankton carcasses

Traditional zooplankton sampling tends to ignore the presence of carcasses due to methodological difficulties in separating live and dead animals. Nevertheless, zooplankton carcasses have been observed worldwide (Tang et al. 2009a), and simple staining methods are now available for distinguishing between live and dead crustacean zooplankton in freshwater and marine systems (Bickel et al. 2009, Elliott & Tang 2009). Using these staining methods, it has been shown that up to 40% of the zooplankton (copepods and daphnids) can be dead *in situ* (Bickel et al. 2009, Elliott & Tang 2009). This is of particular interest since carcasses of both crustacean and non-crustacean zooplankton species represent concentrated reservoirs of labile organic substrates for water column bacteria (Tang et al. 2006a, Titelman et al. 2006, Bickel & Tang in press). Previous experiments have shown that zooplankton carcasses are rapidly colonized and decomposed by bacteria mainly from the outside (Harding 1973, Bickel & Tang

in press), with the caveat that some of the epibionts could have been killed by the treatment that produced the zooplankton carcasses, leading to an underestimation of the contribution to the decomposition process by epibionts. The initial increase in bacterial abundance during the carcass decomposition process is temperature-dependent but is less dependent on aerobic versus anaerobic conditions (Tang et al. 2006b). Tang et al. (2009a) observed that cladoceran carcasses were colonized by bacteria faster than copepod carcasses both in the laboratory and in the field. In addition, carcasses suspended in a eutrophic lake had a higher average carbon loss rate than those suspended in an oligotrophic lake. These differences suggest that exploitation of zooplankton carcasses by bacteria depends on carcass type and environmental conditions.

In addition to bacteria, fungi are also able to colonize and decompose zooplankton carcasses, especially when bacterial activities are suppressed (Tang et al. 2006b). The exoskeleton of zooplankton contains a large fraction of chitin and can be densely colonized by aquatic fungi. Chitin is often used for isolation and cultivation of a variety of fungi, and fungal chitinases are well known to efficiently degrade chitin (reviewed by Wurzbacher et al. 2010). Hence zooplankton molts and carcasses are important organic carbon sources not solely for bacteria but also for fungi.

Although these carcass-colonizing microbes exhibit both very high protease and lipase activities, carcass protein decomposition is faster than lipid decomposition (Bickel & Tang in press), resulting in differential remineralization and preservation of different biochemical components of the carcasses. Analysis of similarity (ANOSIM) of denaturing gradient gel electrophoresis (DGGE) banding patterns revealed that bacterial communities on decomposing zooplankton carcasses rapidly diverged from those in the surrounding water, but remained similar among different types of zooplankton carcasses (Tang et al. 2009a). These observations demonstrate that zooplankton carcasses are decomposed by similar bacterial groups and serve as important microbial microenvironments where rapid and efficient local selection takes place.

In many lakes, crustacean zooplankton populations can suffer high and abrupt mortality due to starvation and diseases resulting in the phenomenon of 'mid-summer decline' (Gries & Güde 1999, Hülsmann & Weiler 2000, Hülsmann & Voigt 2002). Massive die-off of zooplankton will produce abundant zooplankton carcasses as microbial hotspots to fuel water column and benthic bacterial production. Unfortunately, in most traditional microbial studies, these hotspots have been excluded by pre-filtration of water samples.

There are very few published studies on the potentially important aspect of jellyfish blooms, i.e. the fate

of jellyfish-derived organic matter. Jellyfish carcasses can impact microbial processes, causing significant changes to nitrogen and oxygen dynamics in the surrounding environment (Pitt et al. 2009, West et al. 2009) as well as the composition and activity of the ambient bacterial community (Riemann et al. 2006, Titelman et al. 2006, Tinta et al. 2010). Some preliminary results of the bacterial community structure as measured by DGGE indicated differences in the bacterial community response between ecosystems where medusae occur throughout the year and areas where they occur only seasonally (Tinta et al. 2010). Mass deposition of jellyfish and salp carcasses has also been hypothesized to fuel the microbial food web in the seabed (Billet et al. 2006, Lebrato & Jones 2009).

FUTURE CHALLENGES AND OPPORTUNITIES

The available, albeit limited, quantitative data in the literature (Table 2) indicate large variability in zooplankton-associated bacterial abundance in space and in time, as well as among zooplankton species. Understanding what factors contribute to this variability will be an important task for researchers. Because zooplankton likely acquire bacteria via direct bacterial attachment and ingestion, zooplankton-associated bacterial abundance could be dependent on ambient free-living and particle-attached bacterial concentrations. Differences between zooplankton species may be related to the zooplankton feeding strategy (e.g. non-selective versus selective feeders) and body size. Because both bacteria and zooplankton reside within a viscous environment, their encounter and interactions will also be influenced by ambient flow. The use of new *in situ* flow visualization experiments (Dabiri et al. 2005, Costello et al. 2008) and new mechanical models based on hydrodynamic patterns might explain the relationship between flow, zooplankton, and microbial community ecology. Simple and readily testable hypotheses as such may unveil some general patterns in zooplankton–bacteria associations.

Past research on zooplankton–bacteria associations is largely limited to crustacean zooplankton, and extension of sampling effort to gelatinous zooplankton and microzooplankton is much needed. Many investigators used whole animals or homogenates, and therefore their measurements do not separate epibionts and gut microflora (Tables 1 & 2). Because the 2 communities are exposed to very different environments, separation between the two, for example by microdissection of the zooplankton gut (Peter & Sommaruga 2008), will be necessary for proper understanding of their respective physiology and ecology.

While previous work has tended to rely on culture-dependent methods and focused on a few pathogenic species (most notably *Vibrio* spp.), modern culture-independent molecular techniques have revealed very diverse assemblages of bacteria associated with marine and freshwater zooplankton, many of whose pathogenicity and other characteristics remain unexplored. Techniques such as metagenome sequencing, bioinformatics, systems biology, and new statistical approaches should allow for detailed analyses of changes in bacterial community structure in relation to the life cycle and behavior of the zooplankton host. Study of the metatranscriptome and proteome of these complex bacterial communities will also help us understand their ecological roles on the hosts as well as in the surrounding water, and when combined with single cell techniques such as Raman microscopy and NanoSims (secondary ion mass spectrometry), it will even be possible to quantify microbial processes in and on live zooplankton. Microautoradiography in combination with fluorescence *in situ* hybridization (MAR-FISH) can be used to more accurately measure specific substrate uptake and growth of bacteria associated with the different zooplankton body parts.

Although the presence of strict anaerobes inside the zooplankton gut has been known for some time (e.g. Proctor 1997), the physical-chemical characteristics of a zooplankton's gut remain poorly known. These environments may vary according to the feeding activities and physiology of the zooplankton and subsequently select for different bacterial communities. The use of pH- or oxygen-sensitive chemicals provides only a crude characterization of the gut environments (e.g. Pond et al. 1995). An alternative is to use microsenors to generate detailed profiles of the gut for pH, oxygen, nitrous oxides, sulfide, and other chemical species, which has been done successfully with terrestrial insects (e.g. Bignell & Anderson 1980). Detailed characterization of the gut environment in combination with molecular analysis of the gut microflora would allow researchers to study the selection and adaption of different bacteria in this unique micro-environment.

In addition to live zooplankton, other microenvironments such as zooplankton fecal pellets, molts, and carcasses also warrant further study. Although bacterial activities on zooplankton fecal pellets have been intensively studied in the past decades, they are largely limited to copepod fecal pellets. Fecal pellets produced by other zooplankton groups have different sizes and textures, but information on their bacterial activities is scarce. The reported occurrence of strict anaerobes in zooplankton fecal pellets (e.g. Bianchi et al. 1992, Marty 1993) suggests that, similar to the gut, fecal pellets provide a semi-enclosed environment

where microbial activities proceed very differently than in the well oxygenated surface water. Microbial activities in fecal pellets can contribute to water column denitrification and methane production (Oremland 1979, de Angelis & Lee 1994, Michotey & Bonin 1997), but the internal oxygen environment of fecal pellets remains largely unknown, a gap that can be filled by microsensor profiling combined with advanced molecular analysis.

To our knowledge, bacterial colonization of crustacean zooplankton molts and carcasses has not been studied in detail using modern molecular techniques such as metagenomic and metatranscriptomic approaches. This is surprising, considering the global abundance of chitin and that chitin degradation by aquatic bacteria is essential for returning carbon and nitrogen to aquatic nutrient cycles (Kirchner 1995, Flintoft 2004). Zooplankton carcasses may be of particular importance at times when other bacterial substrates are limited, e.g. during the clear water phase which follows the phytoplankton spring bloom in many temperate lakes. Besides bacteria, aquatic fungi are also capable of degrading chitin (Tang et al. 2006b, Wurzbacher et al. 2010), although their ecological role in remineralizing zooplankton molts remains to be explored. The combination of both molecular and ecological methods will provide a better understanding of how zooplankton molts and carcasses can affect bacterial community composition, microbial organic matter transformation, and biogeochemical cycling in the pelagic zone.

Comparative studies of different carcass types are also needed. For example, gelatinous zooplankton carcasses lack a chitinous covering, and presumably decompose more readily than crustacean carcasses, yet the limited data available have suggested otherwise (Bickel & Tang in press). Bacterial decomposition of jellyfish is also of particular interest because of the perceived increase in occurrence of jellyfish 'blooms' worldwide (Mills 2001, Purcell et al. 2007). Because most jellyfish species have few natural predators, their carcasses at the termination of a bloom will likely have large impact on the water column and benthic microbial communities (Titelman et al. 2006, Tinta et al. 2010).

The probiotic versus harmful effects of different bacterial species on the zooplankton hosts have far-reaching implications for aquaculture practices as well as zooplankton ecophysiology and ecosystem health. So far, research on this topic is largely restricted to brine shrimp because of the relative ease with which commercially available encysted eggs may be disinfected to produce bacteria-free animals for exposure experiments (e.g. Verschuere et al. 1999, 2000). Although copepods and daphnids, the most representative of

metazoan zooplankton in aquatic environments, do produce encysted resting eggs (e.g. Uye 1985, Alekseev & Lampert 2001), these resting eggs are not easily available, and the feasibility of using them to produce bacteria-free animals has not been carefully evaluated. The difficulty in producing bacteria-free zooplankton cultures will limit our ability to study how different bacterial species or assemblages affect the host. To partially circumvent this problem, comparative studies of the physiology and ecology of the same zooplankton species but colonized by different bacterial assemblages may provide helpful insights.

CONCLUDING REMARKS

Increasing specialization in science has accelerated the advances of many research disciplines, but at the same time has encouraged an increasingly fragmented and incomplete understanding of the natural world. This is evident in the traditional separation of 'microbial ecology' and 'zooplankton ecology.' As a consequence, we may have missed a large portion of the microbial world and hence are still far from fully understanding biodiversity and functioning of the ecosystem. Bridging these disciplines requires an unconventional way of thinking and approach to research, and recent studies on this topic have laid the necessary foundation for it.

Our review shows that bacteria colonize and interact with zooplankton in many ways with far-reaching implications for microbial production, dispersal, diversity maintenance, evolution, and biogeochemical fluxes (Fig. 6). Nevertheless, available information is still limited to only a few zooplankton groups and a few environments, in particular copepods in coastal and estuarine waters. Because different zooplankton taxa can have very different life history traits that can influence the form of zooplankton-bacteria interactions, extension of research efforts to other zooplankton taxa is much needed. In the oligotrophic open ocean where nutrients are limiting in the water column, direct association with zooplankton may become more important for bacteria (cf. Grossart et al. 2009). Recent and continuous methodological advances in microbiology and microbial ecology will allow researchers to study these interactions in unprecedented detail. As the complexity of zooplankton–bacteria interactions continues to unfold, new ecological concepts and models will be needed to help researchers integrate the new information into our current understanding of the ecosystem. We hope that this article will encourage more collaboration between zooplankton ecologists and microbial ecologists to advance this exciting and promising research topic.

Fig. 6. Aquatic biogeochemical processes as influenced by zooplankton. Zooplankton have been shown to increase particulate organic matter removal, nutrient regeneration, and vertical carbon flux. These processes will be modulated by bacteria associated with the zooplankton and fecal matter. Zooplankton and fecal matter also provide a microenvironment where bacterial production, diversity maintenance, and evolution may proceed differently than in the surrounding water

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