



Importance of bacterivory by pigmented and heterotrophic nanoflagellates during the warm season in a subtropical western Pacific coastal ecosystem

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ABSTRACT: We investigated temporal variations in the effects of bacterivory by different sizes of heterotrophic nanoflagellates (HNF) and pigmented nanoflagellates (PNF) during a warm period (May to September) in oligotrophic coastal waters of the subtropical western Pacific. Short-term experiments with fluorescently labeled bacteria (FLB) demonstrated ingestion rates of 0.3 to 5.8 bacteria HNF⁻¹ h⁻¹ by HNF in the size range 3–6 µm—rates that were higher than observed for other sizes of HNF. Rates of ingestion by PNF ranged between 0.9 and 15.5 cells PNF⁻¹ h⁻¹, and, as for HNF, were greatest for PNF in the 3–6 µm size group. Nanoflagellates of size <6 µm removed about 98% of the total amount of bacteria consumed. The 3–6 µm PNF, 2–3 µm HNF, and 3–6 µm HNF were major consumers in the nanoflagellate community and were responsible for an average of 52, 28 and 16% of the total consumption of bacteria, respectively. The smallest PNF (2–3 µm) consumed only about 2% of the total and were considered to be primarily autotrophic. Despite ingestion rates in the range of those reported elsewhere, the low abundance of nanoflagellates observed resulted in relatively low grazing impacts (<10% of bacterial standing stock). We found a significant negative correlation between PO₄ concentrations and ingestion rates of the 3–6 µm PNF, suggesting that the PNF ingestion rate increased under nutrient-deficient conditions.

KEY WORDS: Pigmented nanoflagellates · Mixotrophy · Subtropical western Pacific · Fluorescently labeled bacteria · Ingestion rates

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INTRODUCTION

Bacteria are generally considered to be consumed mainly by heterotrophic nanoflagellates (HNF) and ciliates in both marine and freshwater environments (Berninger et al. 1991, Epstein & Shiaris 1992, Hall et al. 1993, Nakano et al. 1998, Almeida et al. 2001, Christaki et al. 2001, Cleven & Weisse 2001, Ichinotsuka et al. 2006). However, increasing evidence from natural environments suggests that pigmented nanoflagellates

(PNF) can also contribute greatly to bacterivory at certain times or at certain depths within the water column (Bird & Kalff 1986, Sanders et al. 1989, Berninger et al. 1992, Hall et al. 1993, Havskum & Riemann 1996, Safi & Hall 1999, Hitchman & Jones 2000, Medina-Sánchez et al. 2004, Unrein et al. 2007). For example, Havskum & Riemann (1996) reported that the PNF were responsible for 86% of the entire nanoflagellate bacterivory in the upper layer of the Bay of Aarhus, but accounted for only 19% below the pycnocline. Therefore, PNF

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can significantly affect the flow of organic matter in the microbial food web of some aquatic environments.

Water temperature and the abundance of prey are usually considered among the most important factors regulating the phagotrophic activity of HNF (Choi 1994, Vaqué et al. 1994). However, a range of environmental factors, including light, prey abundance and nutrients, can affect the feeding behavior of PNF and it is unlikely that there is a single explanation or a universal stimulus for the ingestion of particles by PNF (Bird & Kalff 1986, Salonen & Jokinen 1988, Caron et al. 1993, Nygaard & Tobiesen 1993, Holen 1999, Moorthi & Berninger 2006, Unrein et al. 2007). In some studies, PNF feeding rates have been found to increase in low-light conditions or in darkness (Hall et al. 1993, Holen 1999), but to decrease under limited light conditions in other studies (Caron et al. 1993, Jones & Rees 1994). One analysis found that light had no effect at all on the feeding rates of PNF (Sanders et al. 2001). Previous studies have reported that PNF are very important bacterial grazers in nutrient-insufficient conditions (Nygaard & Tobiesen 1993, Arenovski et al. 1995, Havskum & Riemann 1996). Unrein et al. (2007) suggested that PNF could rely on their phagotrophic capabilities to obtain phosphorus when this nutrient is limited. Conversely, Bird & Kalff (1986) showed that PNF were able to assimilate organic carbon and other nutrients, such as phosphorus, from the ingestion of prey. It also has been suggested that bacterivory by PNF could be a means of removing competitors for nutrients (Thingstad et al. 1996) and simultaneously obtaining phosphorus (Urabe et al. 1999). These suggestions are based on the fact that bacteria compete better for inorganic P than do most phytoplankton because they have a higher surface-to-volume ratio (Currie & Kalff 1984, Probyn et al. 1990) and they have a higher internal P:N ratio compared with phytoplankton (Jürgens & Güde 1990).

In prior investigations using fluorescent particles as *Synechococcus* spp. equivalents at our study site, evidence suggested that PNF were the main grazers of *Synechococcus* spp., and that HNF had little grazing impact on cyanobacteria in subtropical western Pacific coastal waters (Tsai et al. 2007, Chan et al. 2009). These authors hypothesized that while HNF are usually the main consumers of bacteria, the grazing impacts of HNF and PNF on bacteria are significantly different in subtropical western Pacific coastal waters.

The present study investigates the impact of different sizes of PNF and HNF (2–3, 3–6 and 6–10 μm) on bacteria during the warm season (May to September) in a coastal ecosystem of the subtropical western Pacific, and it examines factors that potentially regulate the grazing behavior of HNF and PNF.

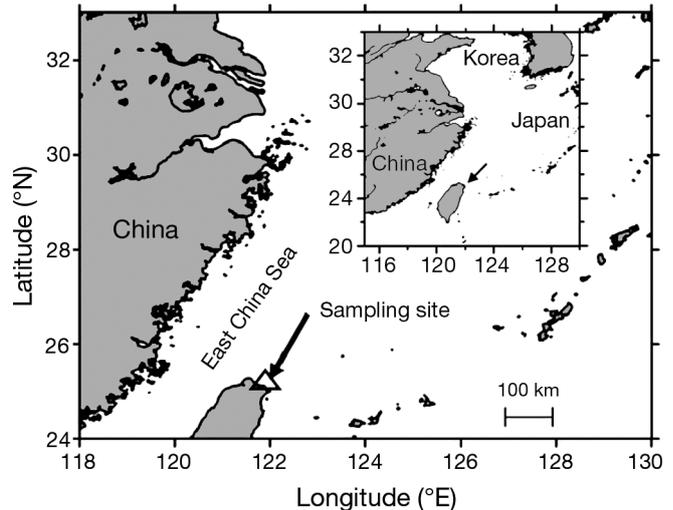


Fig. 1. The sampling station, northeastern Taiwan

MATERIALS AND METHODS

Sampling. Samples were collected weekly from May to September 2009 at an established coastal station (25°09.4' N, 121°46.3' E) along a rocky shore in northeastern Taiwan (Fig. 1). The environment at this site was previously described on the basis of data gathered from 1999 to 2001 (Tsai et al. 2005). Water temperature is constantly above 25°C between June and October, and daytime temperatures are generally 0.5 to 1.5°C higher than nighttime temperatures (Tsai et al. 2005). Annually, salinity ranges from 33.1 to 34.3, the lower salinity within this range likely reflecting the influence of rainfall runoff.

During this study in the warm season of 2009, surface seawater was collected between 09:00 and 10:00 h local time. Water temperature was measured immediately, and all samples were brought to the laboratory within 30 min. Nitrate, phosphate and silicate were measured according to Gong et al. (1995). Water samples were filtered (25 mm GF/F) and, after extraction, chlorophyll *a* (chl *a*) was measured with an *in vitro* fluorometer (Turner Design 10-AU-005) (Parsons et al. 1984).

Bacteria, HNF and PNF. For enumeration, 50 ml water samples were fixed with glutaraldehyde to a final concentration of 1% (Christaki et al. 1999, Sanders et al. 2000), and 1 to 2 ml or 20 ml were filtered onto black Nuclepore filters of pore size 0.2 μm or 0.8 μm for bacteria and nanoflagellates, respectively. Samples were stained with DAPI (4',6-diamidino-2-phenylindole) at a final concentration of 1 $\mu\text{g ml}^{-1}$ (Porter & Feig 1980). PNF and HNF were enumerated according to the presence or absence of chlorophyll autofluorescence using a separate filter set optimized

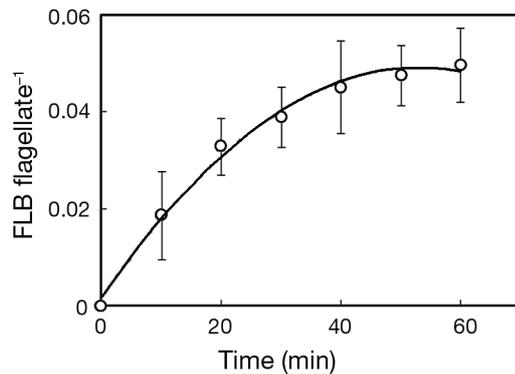


Fig. 2. Time course of uptake of fluorescently labeled bacteria (FLB) by flagellates from the study site

for chlorophyll or DAPI under a 1000× epifluorescence microscope (Nikon-Optiphot-2). Bacteria and HNF were identified by their blue fluorescence under ultraviolet (UV) illumination, and PNF were identified by their orange and red autofluorescence under blue-light excitation. To obtain reliable estimates of abundance we counted 30 and 50 fields of view for bacteria and nanoflagellates, respectively.

Grazing experiments. Grazing rates of HNF and PNF feeding on bacteria were determined using fluorescently labeled bacteria (FLB) generally following the methods of Sherr et al. (1987). FLB were prepared from natural bacterioplankton collected from our study site, stored frozen in 10 ml aliquots, thawed, and briefly sonicated prior to use. FLB were distinguished

from the natural bacteria by the bright green color of the FLB under blue-light excitation using an epifluorescence microscope.

In grazing experiments, which were done in triplicate, surface water was filtered through a 20 µm nylon mesh into 500 ml polycarbonate bottles and then FLB were added to the bottles at <10% of the *in situ* bacterial abundance (McManus & Okubo 1991). Bottles were incubated in a water bath at *in situ* temperature and light intensities for 1 h. Preliminary experiments indicated that ingestion of FLB by nanoflagellates was saturated after 40 to 60 min (Fig. 2). After the incubation, 20 ml subsamples were preserved with glutaraldehyde (to 1% v/v final conc.), and the number of FLB ingested by nanoflagellates was determined using epifluorescence microscopy (magnification 1000×). In order to understand the role played in feeding by HNF and PNF of different sizes, we divided HNF and PNF into 3 size categories: 2–3, 3–6, and 6–10 µm, and recorded their FLB content separately. Ingestion rates of FLB (FLB HNF⁻¹ h⁻¹; FLB PNF⁻¹ h⁻¹) were calculated according to Dolan & Šimek (1998). Ingestion rates (bacteria HNF⁻¹ h⁻¹; bacteria PNF⁻¹ h⁻¹) were then calculated by multiplying the ingestion rate of FLB by the ratio of bacteria to added FLB (Pace et al. 1990). Community consumption rates (bacteria ml⁻¹ h⁻¹) were estimated by multiplying average ingestion rates on bacteria and total HNF and PNF abundance.

To assess the differences in PNF and HNF ingestion and consumption rates, *t*-tests were undertaken, while

Table 1. Temperature, salinity, chlorophyll *a* (chl *a*) and nutrient (NO₃, NO₂ and PO₄) concentrations in the coastal ecosystem of subtropical western Pacific during the summer period in 2009. nd = no data

Date	Temperature (°C)	Salinity	Chl <i>a</i> (mg m ⁻³)	NO ₃ (µmol l ⁻¹)	NO ₂ (µmol l ⁻¹)	PO ₄ (µmol l ⁻¹)
May 8	23.0	nd	nd	nd	nd	nd
May 15	23.0	34.01	0.35	1.26	0.18	0.19
May 23	26.0	34.06	0.51	0.61	0.17	0.10
May 29	27.0	33.99	0.38	0.44	0.10	0.11
Jun 5	27.0	33.89	0.48	3.16	0.19	0.17
Jun 12	27.0	33.82	0.46	0.44	0.08	0.06
Jun 19	28.0	33.71	0.37	0.56	0.11	0.09
Jun 26	29.0	33.90	0.56	0.43	0.13	0.14
Jul 10	29.0	33.48	0.39	0.15	0.05	0.06
Jul 17	29.5	33.31	0.65	0.19	0.12	0.12
Jul 21	28.0	33.41	0.63	0.41	0.17	0.22
Jul 23	28.5	nd	nd	nd	nd	nd
Jul 30	28.0	33.12	0.57	3.09	0.47	0.42
Aug 8	28.0	33.54	0.78	1.80	0.19	0.23
Aug 13	28.0	33.10	1.06	1.70	0.15	0.06
Aug 20	28.0	33.10	0.58	1.40	0.20	0.11
Aug 27	28.0	33.29	0.96	6.60	0.11	0.30
Sep 3	27.0	33.32	0.15	2.00	0.34	0.35
Sep 10	28.5	33.55	0.28	nd	0.13	0.32
Sep 17	27.0	32.66	0.42	nd	0.84	nd
Sep 24	25.0	nd	nd	nd	nd	nd

1-way analysis of variance (ANOVA) and Tukey tests were applied to assess significant differences in ingestion and consumption rates for different sizes of PNF and HNF. All statistical analyses were performed using SPSS version 11.0.

RESULTS

Physical and chemical environment

Surface water temperature at our site fluctuated between 23.0°C (May 8) and 29.5°C (July 17) throughout the study period (Table 1). Salinity ranged from 32.66 to 34.06, with the lowest salinity reflecting the influence of heavy rainfall runoff before sampling (Table 1). Typically, concentrations of nutrients were low during the warm season (May to September), with NO_3^- , NO_2^- and PO_4^{3-} varying from 0.15 to 6.60 $\mu\text{mol l}^{-1}$, 0.08 to 0.84 $\mu\text{mol l}^{-1}$ and 0.06 to 0.42 $\mu\text{mol l}^{-1}$, respectively (Table 1).

Abundance of bacteria and nanoflagellates

Concentrations of chl *a* ranged from 0.15 to 1.06 mg m^{-3} , and bacterial abundance followed a pattern similar to that of chl *a* (Fig. 3A). Pearson correlation analysis showed a positive and significant correlation be-

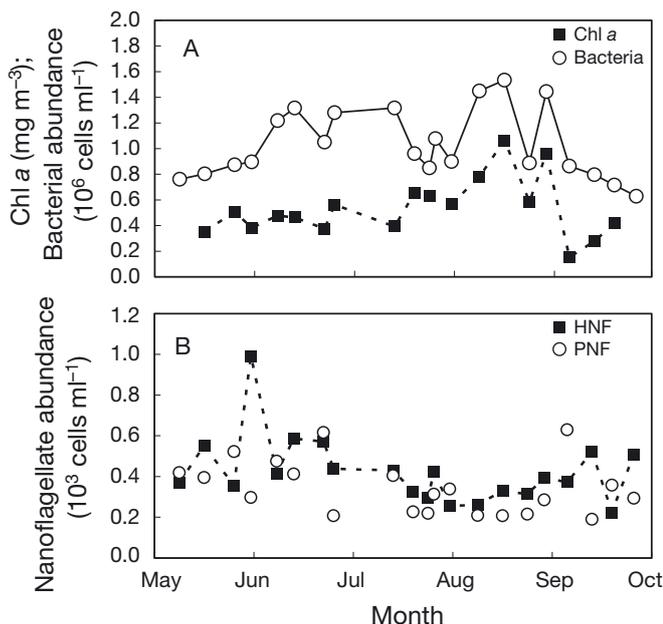


Fig. 3. (A) Temporal variations in concentrations of chlorophyll *a* (Chl *a*) and bacterial abundance, and (B) the abundance of heterotrophic nanoflagellates (HNF) and pigmented nanoflagellates (PNF) from May to September 2009

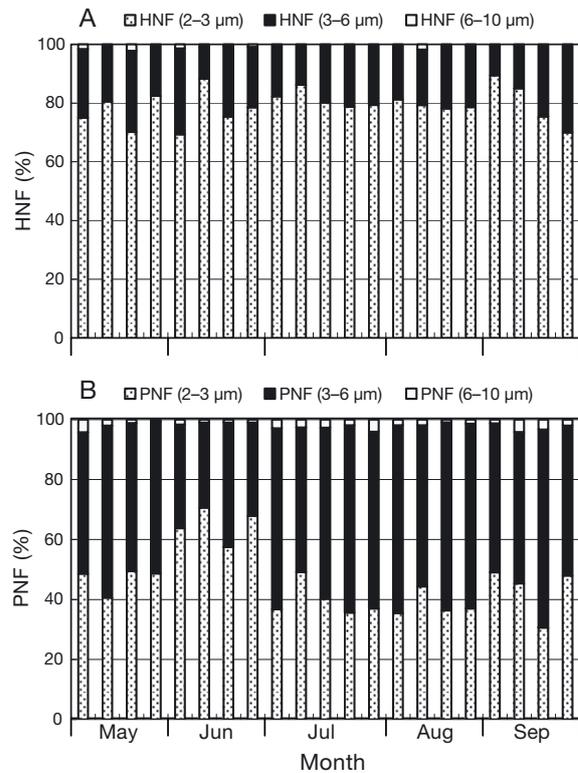


Fig. 4. Percentage contributions of 3 size classes (2–3, 3–6 and 6–10 μm) to the total abundance of (A) heterotrophic nanoflagellates (HNF) and (B) pigmented nanoflagellates (PNF)

tween chl *a* concentration and bacterial abundance ($r = 0.65$, $p < 0.05$) (data not shown).

Nanoflagellate abundance showed no consistent trend during the study period (Fig. 3B). The abundance of HNF and PNF generally ranged from 200 to 600 cells ml^{-1} , though the highest abundance of HNF observed at the end of May was 990 cells ml^{-1} (Fig. 3B). No significant difference was detected between HNF and PNF abundance over the entire investigation period (t -test, $p > 0.05$).

HNF and PNF $> 6 \mu\text{m}$ in size were always a minor component of the total nanoflagellate community. HNF in the 2–3 μm size class dominated the HNF community, ranging from 69 to 89% of total HNF abundance (average 79%) (Fig. 4A). Abundance of the 2–3 μm PNF accounted for 31 to 71% (average 46%) of total PNF, while 28 to 66% of the PNF were between 3 and 6 μm (average 52%) (Fig. 4B).

Grazing on bacteria by the HNF and PNF assemblages

The grazing rates of HNF and PNF, estimated on the basis of direct uptake of FLB, ranged from 0.2 to 2.0

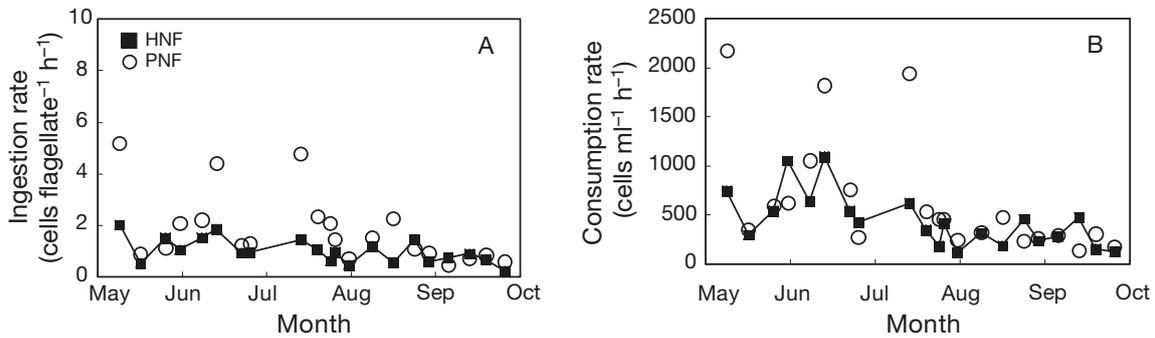


Fig. 5. Temporal variations in (A) ingestion and (B) community consumption rates of heterotrophic nanoflagellates (HNF) and pigmented nanoflagellates (PNF)

cells HNF⁻¹ h⁻¹ and from 0.5 to 5.2 cells PNF⁻¹ h⁻¹, respectively (Fig. 5A). The highest grazing rates for PNF (>4 cells PNF⁻¹ h⁻¹) were observed between May and July, although more moderate rates were also found during these months (Fig. 5A). During most grazing experiments, the range of ingestion rates of HNF and PNF on bacteria were narrower (0.2 to 2.3 cells flagellate⁻¹ h⁻¹, Fig. 5A).

The grazing impact (community consumption) on bacteria was generally greater in May through July, especially when PNF had high ingestion rates (Fig. 5). The effect of PNF grazing (>1500 cells ml⁻¹ h⁻¹) was 3 times greater than that of HNF on 3 dates from May to July (Fig. 5B). Otherwise, hourly consumption of bacteria by both HNF and PNF was lower, varying from 120 to 1100 cells ml⁻¹ h⁻¹, respectively, and there was no significant difference in consumption rates between HNF and PNF (*t*-test, *p* > 0.05) across the complete data set (Fig. 5B).

Grazing by different size classes of nanoflagellates

Nanoflagellate size appeared to affect cell-specific ingestion rates of bacteria. Among the bacterivorous HNF community, HNF of 2–3 μm were considerably less efficient feeders with lower maximum ingestion rates (<1.5 bacteria HNF⁻¹ h⁻¹) than the larger HNF (Fig. 6A). On the other hand, the higher grazing rates of 6–10 μm HNF did not translate into large feeding impacts on the bacteria because of their

generally low abundance (0 to 10 cells ml⁻¹, data not shown). Overall, the 3–6 μm HNF had higher ingestion rates (0.3 to 5.8 bacteria HNF⁻¹ h⁻¹) than did the other size classes of HNF (ANOVA, Tukey, *p* < 0.05) (Fig. 6A). The ingestion rates of the 2–3 μm PNF on bacteria were very low (0–0.7 cells PNF⁻¹ h⁻¹) during the whole study

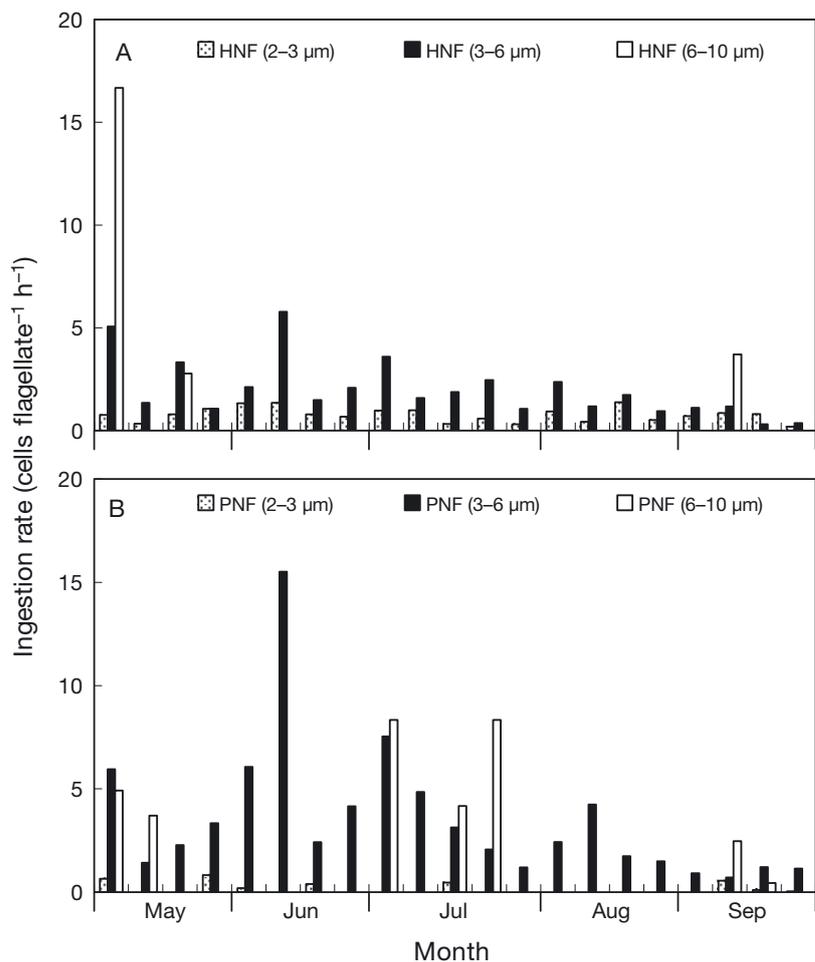


Fig. 6. Temporal variations in ingestion rates of 3 size classes (2–3, 3–6 and 6–10 μm) of (A) heterotrophic nanoflagellates (HNF) and (B) pigmented nanoflagellates (PNF)

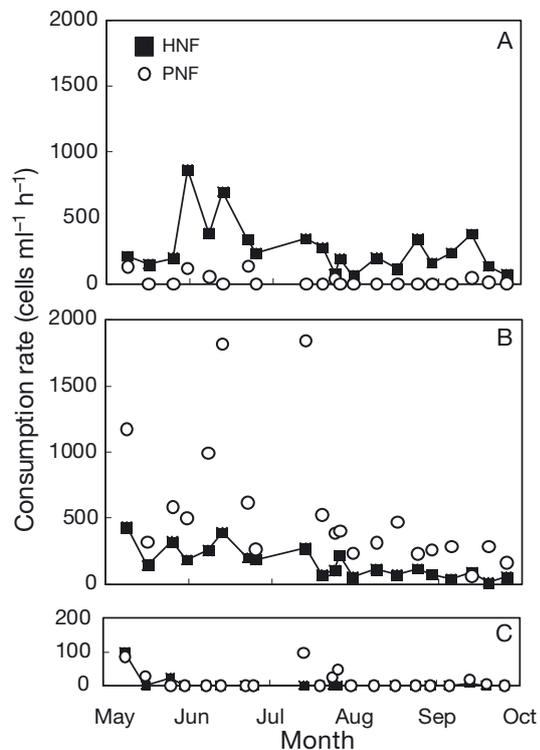


Fig. 7. Temporal variations in community consumption rates of (A) 2–3 μm , (B) 3–6 μm , and (C) 6–10 μm heterotrophic nanoflagellates (HNF) and pigmented nanoflagellates (PNF)

period, while the ingestion rates of 3–6 μm PNF fluctuated from 0.9 to 15.5 cells PNF⁻¹ h⁻¹ (Fig. 6B), with the highest value recorded in June.

In terms of grazing impact on bacteria, the 2–3 μm HNF and 3–6 μm HNF were the major consumers of bacteria in the HNF community (60 to 860 cells ml⁻¹ h⁻¹

and 20 to 440 cells ml⁻¹ h⁻¹, respectively) (Fig. 7A,B). The grazing impact of 3–6 μm PNF was greater than the other size categories of PNF (ANOVA, Tukey, $p < 0.05$) and often greater than that of the HNF community (Fig. 7B). The grazing impact of both HNF and PNF in the 6–10 μm size range was negligible (0 to 98 cells ml⁻¹ h⁻¹, Fig. 7C) because of the low abundance of flagellates in this size category (Fig. 4). Regardless of the presence of a chloroplast, nanoflagellates <6 μm accounted for about 98% of the total bacterivory (Fig. 8). PNF of 3–6 μm in size were responsible for 12 to 72% of the bacterivory (average 52%), 2–3 μm HNF for 9.8 to 61.9% and 3–6 μm HNF for 3.6 to 28.7% (Fig. 8).

DISCUSSION

It is now commonly accepted that nanoflagellates are the most important grazers of bacteria in most aquatic environments (Epstein & Shiaris 1992, Nakano et al. 1998, Safi & Hall 1999, Sanders et al. 2000, Tsai et al. 2005, 2008, Ichinotsuka et al. 2006), though some planktonic ciliates also can be significant consumers of bacteria (Rassoulzadegan et al. 1988, Jürgens & Šimek 2000, Kisand & Zingel 2000). On an individual basis, a bacterivorous ciliate may graze an order of magnitude more bacterial cells than a nanoflagellate (Šimek & Straškrabová 1992, James et al. 1996, Ichinotsuka et al. 2006). However, ciliate abundance is usually less than that of nanoflagellates in marine environments, so total grazing pressure on bacteria by ciliates is usually lower than that of nanoflagellates. In the present study, the low abundance of ciliates (~1 cell ml⁻¹) (A.-Y. Tsai unpubl. data), resulted in negligible consumption of

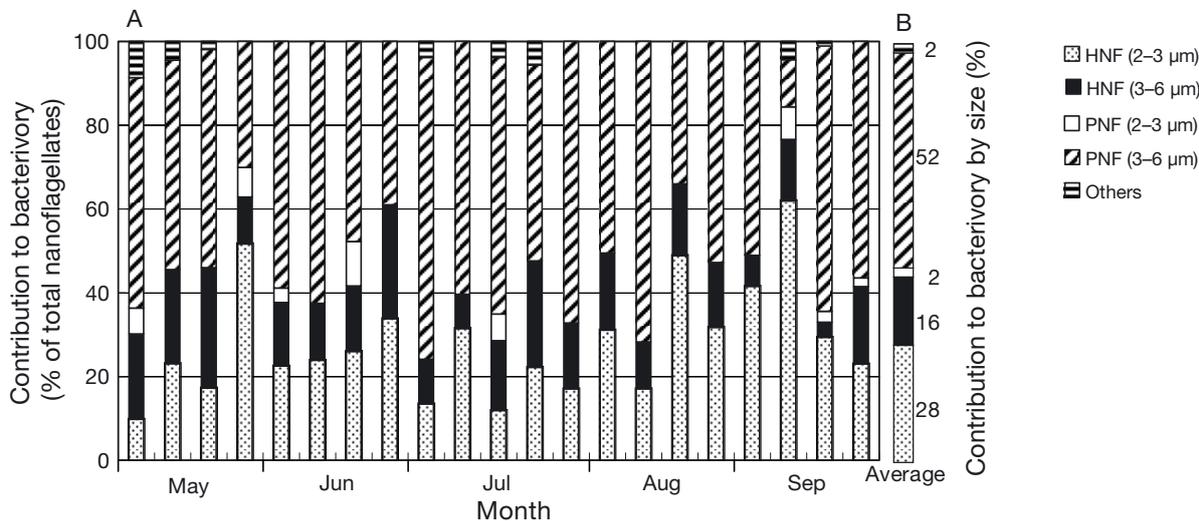


Fig. 8. Proportion of bacterivory (in %) contributed by each size class of heterotrophic nanoflagellates (HNF) and pigmented nanoflagellates (PNF) (A) throughout the study period and (B) as average values

bacteria by ciliates relative to that by HNF or PNF. Furthermore, in a previous study in the same general area, ciliates removed only ~3% of the *Synechococcus* spp. production, despite high ingestion rates by individual ciliates (Tsai et al. 2007). In that study, the grazing impact by PNF was 20 times that of ciliates, and PNF were clearly the key warm-season grazers of *Synechococcus* spp. (Tsai et al. 2007). Taken together, these results confirm a report by Tsai et al. (2005) that nanoflagellates were the most important grazers of picoplankton (bacteria and *Synechococcus* spp.) at our study site.

Some previous studies showed that the smallest HNF were primarily bacterivorous (Sherr & Sherr 2002, Unrein et al. 2007). In our study, HNF 2–3 μm in size grazed fewer bacterial cells on an individual basis than did larger HNF (e.g. 3–6 μm HNF) (Fig. 6A). However, the abundance of 2–3 μm HNF was usually greater than that of other HNF, and grazing pressure on bacteria by 2–3 μm HNF was thus higher than that by other HNF. In the present study, 28% of the total measured bacterivory was attributed to HNF 2–3 μm in size (Fig. 8B), a significantly higher grazing effect than the larger HNF (3–6 μm or 6–10 μm HNF). The grazing impact of >6 μm HNF and PNF on bacteria was considerably lower than for other nanoflagellates (2% grazing effect, Fig. 8B). This suggests that they obtain most of their carbon by feeding on larger prey (or by photosynthesis). Havskum & Riemann (1996) reported that larger nanoflagellates typically fed on prey larger than

bacteria, and as noted above for PNF, HNF can also feed on phototrophic picoplankton (Sherr et al. 1991). In agreement with this view of larger flagellates ingesting larger prey, Vaqu e et al. (2008) found that there were 2 trophic chains: one from bacteria to <5 μm HNF and another beginning with prey larger than bacteria, such as small algae, and leading to >5 μm HNF. Grazing by larger nanoflagellates on the smaller size fractions of nanoflagellates could complicate the interpretation of microbial food webs. Lin et al. (2009) observed an indirect trophic cascade grazing pattern at our study site, in which nanoflagellates 2–5 μm in size were the principal predators of picoplankton, and these small nanoflagellates were, in turn, consumed by 5–20 μm nanoflagellates. Feeding on small flagellates by larger ones may partly explain their consequently relatively low impact on the bacterial standing stock (Table 2).

A maximum of ~9% of bacterial standing stock (~7 $\times 10^4$ cells $\text{ml}^{-1} \text{d}^{-1}$) was calculated to be consumed by nanoflagellates in the present study (Table 2). Running the feeding experiments for 1 h may have underestimated the ingestion rates, if the number of FLB ingested had begun to plateau earlier in the experiments as in Fig. 2; the rate determined at 1 h was only 58% of the rate determined from a linear fit to the first 30 min of the time course. This suggests that ingestion rates and impact could be nearly double what we report. It is also possible that grazing rates in the present study could be underestimated due to feeding selectivity

Table 2. Bacterial abundance and consumption of bacteria by heterotrophic (HNF) and pigmented (PNF) nanoflagellates

Date (2009)	Bacterial abundance (10^6 cells ml^{-1})	Consumption rates (10^4 cells $\text{ml}^{-1} \text{d}^{-1}$)		Bacterial standing stock ingested (%)	
		HNF	PNF	HNF	PNF
May 8	0.76	1.78	5.22	2.34	6.86
May 15	0.80	0.71	0.85	0.88	1.05
May 23	0.87	1.30	1.41	1.49	1.62
May 29	0.90	2.52	1.49	2.81	1.66
Jun 5	1.22	1.53	2.53	1.25	2.08
Jun 12	1.32	2.62	4.37	1.99	3.32
Jun 19	1.05	1.30	1.82	1.23	1.73
Jun 26	1.28	1.01	0.64	0.79	0.50
Jul 10	1.32	1.48	4.66	1.13	3.54
Jul 17	0.96	0.84	1.27	0.87	1.33
Jul 21	0.85	0.44	1.10	0.51	1.29
Jul 23	1.08	0.99	1.09	0.92	1.01
Jul 30	0.90	0.28	0.57	0.31	0.64
Aug 8	1.45	0.75	0.76	0.52	0.53
Aug 13	1.53	0.45	1.14	0.29	0.74
Aug 20	0.89	1.09	0.56	1.23	0.64
Aug 27	1.44	0.56	0.63	0.39	0.44
Sep 3	0.86	0.66	0.69	0.77	0.80
Sep 10	0.80	1.13	0.32	1.43	0.41
Sep 17	0.72	0.36	0.73	0.50	1.02
Sep 24	0.63	0.29	0.42	0.47	0.66

(Landry et al. 1991). Nonetheless, ingestion rates reported here are within the range determined in other environments with greater grazing impact by nanoflagellates (Christaki et al. 2001, Cleven & Weisse 2001, Jezbera et al. 2003). Thus, the low grazing in the present study seems to be mostly a consequence of the low abundances of nanoflagellates present on most of the sampling dates. Flagellates, as expected, seem to be the major protistan bacterivores in this coastal system. Their low overall impact on bacteria (~20% of the bacterial production expected in the area based on other cruises, A.-Y. Tsai unpubl. data), indicates that alternative sources of bacterial mortality, such as lysis by viruses and bacterivory by other organisms (Sanders et al. 1989, Vaque & Pace 1992, Fuhrman & Noble 1995, Hwang & Heath 1999, Almeida et al. 2001, Taira et al. 2009) may also be important in this environment.

Grazing on bacteria by HNF has been estimated using the FLB method in numerous ocean and coastal sea environments (James et al. 1996, González et al. 1998, Safi & Hall 1999, Sime-Ngando et al. 1999, Ichinotsuka et al. 2006). However, only a few studies have made comparisons between bacterivory rates by HNF and PNF, and information on the importance of these groups of grazers in oceans and coastal seas is still limited (Epstein & Shiaris 1992, Hall et al. 1993, Safi & Hall 1999, Sanders et al. 2000, Unrein et al. 2007, Moorthi et al. 2009). The present study found no significant difference between the abundance of PNF and HNF in the coastal waters off Taiwan. Furthermore, PNF and HNF were responsible for approximately equal consumption of bacteria (56 and 44% of the total ingestion, respectively).

The smallest size category of PNF (2–3 μm) was relatively very abundant during the summer period (averaged 46% of total PNF, Fig. 4B), but never ingested bacteria at a significant rate; HNF 2–3 μm in size had a significantly higher grazing effect than did the 2–3 μm PNF (Fig. 7). These results suggest that most of the 2–3 μm PNF were primarily autotrophic. This finding is consistent with a previous study (Unrein et al. 2007) that observed no ingestion of bacteria by PNF <3 μm in size, but differs from another (Zubkov & Tarran 2008) in which the smallest PNF were strongly bacterivorous. During the summer period of our study, the consumption rates of PNF 3–6 μm in size were considerably higher than those of HNF of the same size (Fig. 7B), and they made up the majority of the bacterivorous PNF community (Fig. 4B).

In several studies, the impact of PNF grazing was important and accounted for up to 86% of the total bacterivory (Berninger et al. 1992, Arenovski et al. 1995, Havskum & Riemann 1996, Sanders et al. 1989, 2000). Reviews of PNF bacterivory in both freshwater and marine environments report highly variable clear-

ance rates among different PNF populations (Jones 1994, Safi & Hall 1999, Sanders et al. 2000, Unrein et al. 2007), but still suggest that PNF can be important grazers on bacteria and picophytoplankton, especially in oligotrophic waters. To date, there are not enough data to support a clear explanation for PNF ingestion of bacteria at a given site. A range of environmental factors (nutrients, light, prey abundance) may all affect the abundance and feeding behavior of PNF (Bird & Kalff 1986, Salonen & Jokinen 1988, Caron et al. 1993, Nygaard & Tobiesen 1993, Holen 1999, Sanders et al. 2000, Moorthi & Berninger 2006, Unrein et al. 2007). In most cases, photosynthesis would presumably be the primary energy source for these PNF, while phagotrophy would allow them to compete successfully with non-phagotrophic algae for growth-limiting nutrients. Such a strategy for growth might be particularly useful in oligotrophic waters (Arenovski et al. 1995).

In the present study, we did not evaluate the importance of light availability influencing the feeding behavior of PNF, as all the experiments were performed under lighted conditions. Furthermore, we did not find any relationship between the ingestion rates of PNF and bacterial abundance. However, there was a significantly negative correlation between PO_4 concentrations and ingestion rates by the 3–6 μm PNF (Fig. 9). This indicates a tendency toward increased ingestion at low PO_4 concentrations, which is consistent with nutrient uptake via phagotrophy permitting mixotrophic phytoplankton to outcompete other phytoplankton when dissolved nutrients are restricted. Support for this explanation is found in a study by Nygaard & Tobiesen (1993) who found that marine mixotrophic nanoflagellates have higher ingestion rates in P-limited conditions. Another nutrient enrichment experiment with Sargasso Sea populations also produced marked declines in phagotrophically active PNF after the addition of phosphorus (Arenovski et al. 1995). Havskum & Riemann (1996) also noted that PNF were responsible for 86% of the entire nanoflagellate bac-

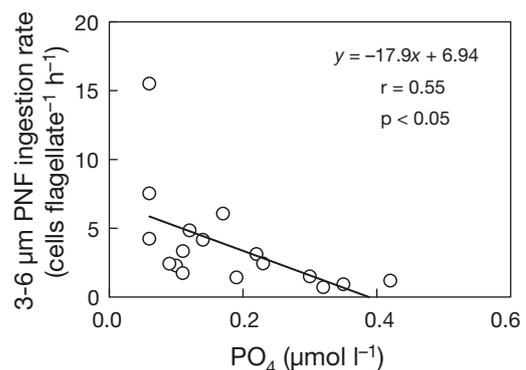


Fig. 9. Relationship between PO_4 concentrations and ingestion rates of 3–6 μm pigmented nanoflagellates (PNF)

terivory in surface water with low inorganic nutrient concentration ($<0.1 \mu\text{mol l}^{-1}$), while below the pycnocline, where nutrients were present at higher concentrations, PNF accounted for only 19% of the nanoflagellate bacterivory. Together, these findings strongly support our results that the mixotrophic PNF use their feeding ability to supply themselves with phosphorus when concentrations of dissolved P are low during the warmer season. However, it is noteworthy that acquisition of nutrients is only one of several possible inducements of phagotrophic behavior among PNF species. Mixotrophy can be observed even when nutrients are high. For example, Hall et al. (1993) measured high rates of bacterivory by PNF in coastal waters during an upwelling event when nitrogen and phosphorus concentrations were high. All of these results advocate for a conceptual integration of PNF into the discussion of plankton nutrient limitation.

Information about grazing on bacteria by different sizes of HNF and PNF is still limited for marine systems, and our data contribute to furthering the understanding of the structure and functioning of microbial food webs in coastal marine environments. This study identified $<6 \mu\text{m}$ HNF (44% grazing effect) and $3\text{--}6 \mu\text{m}$ PNF (52% grazing effect) as the major grazers of bacteria in the coastal water of the subtropical western Pacific. Thus, mixotrophic grazing on bacteria may play a critical role in the utilization of energy and nutrients within the microbial food web in these coastal waters.

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