



User community preferences for climate change mitigation and adaptation measures around Hainich National Park, Germany

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ABSTRACT: In contemporary media discourse, suggestions for publicly mandated climate change mitigation or adaptation measures are frequently challenged from a cost perspective. However, empirical data on the actual economic appreciation of local mitigation and adaptation measures expressed as citizen willingness-to-pay (WTP) are scarce. In this paper, we report results of a preference survey using a choice experiment (CE) that quantifies economic preferences for biodiversity-based climate change mitigation and adaptation in the region surrounding Hainich National Park (Thuringia, Germany). A representative sample of 302 respondents — the majority of them frequent Hainich forest visitors — was interviewed in autumn 2006. Nested logit analysis showed that respondents state a substantially positive WTP for climate change mitigation by afforestation ($p < 0.001$). If converted to WTP for an additional sequestration of carbon that average German citizens emit as CO_2 , a monetary value of $7.34 \text{ € yr}^{-1} \text{ t C}^{-1}$ is obtained. For increasing forest resistance and resilience against insect pests and storms (climate change adaptation) a WTP of 27.54 € yr^{-1} ($p < 0.001$) is obtained, and 16.83 € yr^{-1} ($p < 0.001$) is obtained for increasing the general resilience and resistance of forest ecosystems to environmental stressors. Respondents support moderate programs to eradicate invasive plants when compared to more aggressive eradication measures. Due to the lack of comparable studies, it can only be conservatively assumed that WTP would be lower if mitigation and adaptation measures were to be implemented in forests not, or only rarely, used by respondents. As all proposed means for climate change mitigation and adaptation contribute to local forest ecosystem biodiversity, the results of the study advocate the realization of measures that potentially benefit both climate policy and regional conservation concerns.

KEY WORDS: Economic valuation · Choice experiment · CE · Willingness-to-pay · WTP · Afforestation · Insect pests · Climatic extremes · Resilience · Ecological biodiversity

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1. INTRODUCTION

A very substantial body of scientific evidence indicates that the Earth's climate is rapidly changing, mainly as a result of increasing anthropogenic greenhouse gas emissions (e.g. Levitus et al. 2000, Ruddiman et al. 2005, IPCC 2007). Influential political actors frequently question the expediency of the Kyoto Protocol and other suggested climate change mitigation targets that require substantial reductions of greenhouse gas (GHG) emissions. Usually economic reasons are cited, e.g. more ambitious

measures would 'cause serious harm to the U.S. economy' (Bush 2001, EIA 2002). In contrast, Stern (2006) claims that the costs of action need not be prohibitive, and would be much smaller (1% of GDP) than the damage caused by climate change (5 to 20% GDP) in high emission scenarios in several decades from now. The Stern review has also been criticized several times (e.g. Cole 2007, Nordhaus 2007, Tol 2007). In view of conflicting macroeconomic projections, we may safely assume that the willingness of political decision-makers to initiate costly climate protection measures is influenced by

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evidence of the willingness of consumers and citizens (voters) to pay for such measures. Surprisingly few willingness-to-pay (WTP) studies are available (Table 1). However, an improved integration of economic and social science expertise in climate research is critically called for (Msangi et al. 2006).

In this paper, we present results of a survey conducted in several villages and 2 smaller towns around Hainich National Park (Thuringia, Germany). The results document a substantial WTP for differing types of regional climate change mitigation and adaptation measures. We first explain the main features of the choice experiment (CE) method and its analysis (Sections 2 & 3). Empirical results and WTP estimates follow in Section 4; these are discussed in Section 5.

2. METHODS

2.1. Choice experiment

The choice experiment (CE) method is based on the characteristics theory of value (Lancaster 1966) and on random utility theory (McFadden 1973, Adamowicz et al. 1998a). CEs were initially developed in marketing research to estimate economic preferences for characteristics (attributes) of new products. They have proven to be a versatile method for the estimation of economic values regarding the environment (Hensher et al. 2005). The CE method relies on social science surveys for data collection, as does the related contingent valuation method (Bateman et al. 2002, Perrings 1995a). Instead of asking questions such as ‘Are you willing to pay X € to prevent Y effect of climate change?’, CEs focus on trade-offs among different policy scenarios with different environmental outcomes. CEs have been recognized to be particularly suitable for analyzing different options for mitigation of and adaptation to climate change (Australian Greenhouse Office 2004, p. 19).

The choice scenarios (options) of a CE are differentiated through the attribute levels (characteristics) possible in a policy scenario. In CE face-to-face applications, the scenarios are often printed on cards and presented to respondents. We presented 3 of these cards (the choice set; Fig. 1) to respondents, and asked them to choose the option they wished to see implemented. Two cards characterised the generic policy options A and B, and the third card represented the current situation (status quo). The price (cost) of each policy option was included as one attribute on the choice cards, and operationalised as a mandatory tax payment. For the status quo option, the cost was always zero. The specific combination of attribute levels on the cards was determined by experimental design rules⁴. Some more details on the econometric estimation of CE models are given in Section 3.1.

Table 1. Case studies attempting to estimate the societal value of global warming, climate change mitigation or adaptation

Source	Object of valuation	Monetary value estimation ^a (\$ or €) ^b	Remarks
Zepek & Shively (2003)	Marginal costs for carbon storage	\$3.30 (2.25 €) to \$3.90 (2.65 €) [tC] (depending on site quality and term between 122 and 437 € t C yr ⁻¹)	Marginal costs for C storage on fallow land used for forestry or agroforestry
Berk & Fovell (1999)	Prevention of significant climate change	\$13.70 (9.33 €) [WTP person ⁻¹ yr ⁻¹]	Winter and summer climate scenarios were presented to respondents
Reiner et al. (2006)	‘Solving’ global warming	\$50 (34 €) [WTP person ⁻¹ mo ⁻¹] (408 € person ⁻¹ yr ⁻¹)	Study was conducted in 4 different countries (Japan, USA, Britain, Sweden)
Viscusi & Zeckhauser (2006)	Avoidance of global warming for a 30 yr period	\$0.50 (0.32 €) [WTP gallon ⁻¹] (140 € t C ⁻¹ or 170 € person ⁻¹ yr ⁻¹)	Respondents were asked to suppose that scientists develop technologies that eliminate all climate change problems by the year 2024
Alavalapati et al. (2004)	Additional C sequestration by silvopasture	\$58.05 (39.51 €) to \$62.72 (42.69 €) [WTP household ⁻¹ yr ⁻¹ for a total of 5 yr]	Also valuated: water quality and wildlife habitats associated with silvopasture
Present study	Climate change mitigation by afforestation	7.34 € [WTP t C emitted person ⁻¹ yr ⁻¹]	Calculated from 20.01 € person ⁻¹ yr ⁻¹ tax increase for one-level increase in afforestation (cf. Table 2)

^aIncluding comparative values; ^bAll \$ values are in US \$, and were converted to euros at the exchange rate \$1 = 0.68 €

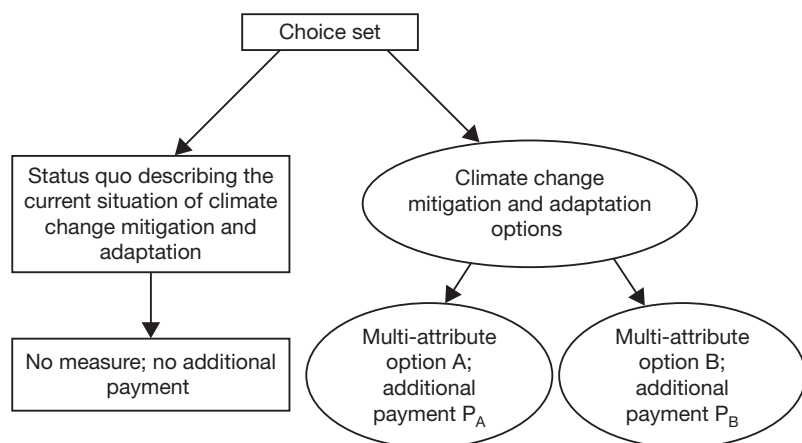


Fig. 1. Structure of the choice experiment: a choice set with one status quo option and 2 multi-attribute climate change mitigation and adaptation options are presented to respondents

2.2. Study site

All mitigation and adaptation scenarios in this study were based upon the Hainich area in Thuringia, comprising the Hainich National Park (Hainich NP, 7600 ha), and the wider Hainich forest ecosystem (~16 000 ha, including the NP). It is the largest coherent deciduous forest in Germany. The national park is used for recreation, tourism and partly just as natural habitat. Hainich NP is dominated by *Fagus sylvatica* L. (European beech), mostly in uneven-aged stands that have not been harvested, thinned or used at least since 1997. From 1964 to 1997, the forest was part of a military training area where only a few single trees of high value were cut (Mund 2004).

Nevertheless, the unmanaged sites (90% of Hainich NP) do not represent a primary forest because nearly all forests in central Europe are influenced by historical forest use and management. For at least 140 yr (equivalent to one rotation period), all parts of the Hainich forest were managed at varying intensities as shelterwood forests or with a selection system (Mund 2004). For the Hainich forest ecosystem, a yearly average net carbon uptake as high as 480 to 580 g C m⁻² yr⁻¹ was calculated (Knohl et al. 2003). Only a fraction of this can be attributed to long-term soil storage. Timber carbon pools are similar to other values for primary beech stands under favourable conditions (Korpel

1995, Meyer et al. 2003). Generally, regenerated forests store less carbon than natural forests, even at maturity (Buchmann & Schulze 1999).

2.3. Measures of climate change mitigation and adaptation

To ensure that respondents properly understood the climate change mitigation and adaptation policy scenarios, we employed 2 strategies. First, we conducted 16 semi-structured in-depth interviews to elucidate the perceptions held by typical respondents on issues such as climate change, risks from increasing weather anomalies, invasive species and environmental risks in general (Rajmis & Barkmann 2007). Second, we made sure that any mitigation or adaptation attribute made sense to respondents in terms of their personal values and interests. To explicate the benefits that respondents would obtain from an implementation of the proposed mitigation and adaptation options, we used an ecosystem service approach developed specifically for the valuation of 'unfamiliar' environmental attributes (Barkmann et al. 2008). In this approach, the focus shifts from a valuation of technically described changes in ecosystem structure or functioning (or the measures that bring about these changes) to the valuation of ecosystem service effects, i.e. benefits to people.

2.3.1. Mitigation: additional carbon sequestration by afforestation

Pre-study interviews revealed that respondents did not consider the unit 'tons of sequestered carbon per year' as a meaningful quantification of a climate change mitigation attribute brought about by a forest ecosystem. If explained in terms of their own annual CO₂ emissions (~10 t C yr⁻¹ person⁻¹ in Germany), however, respondents could relate much more easily to a quantitative aspect of climate change mitigation measures. This result prompted us to focus explicitly on individual contributions towards climate change mitigation. Thus, we decided to elicit preferences quantified in terms of per person CO₂ emission equivalents. Although climate change is a global issue, the regional scope of our study prompted us to search for Hainich forest-based measures that could serve as 'vehicles' to effectuate climate change mitigation. In the following we describe the scientific basis for the quan-

¹This procedure results in a few choice cards that request an additional tax payment even though the environmental attributes appear worse than the status quo (see Table 3). However, because respondents are always free to choose the status quo option at zero cost, the choice set always offers meaningful choices

tification of the respective CE attribute levels on regional afforestation measures.

First, the C-sequestration of the existing forest had to be assessed. Robertson & Vitousek (1981) found a long-term sequestration rate of $9.4 \text{ g C m}^{-2} \text{ yr}^{-1}$ for the soil compartment of a 250 yr old oak forest. Much lower values ($2.4 \text{ g C m}^{-2} \text{ yr}^{-1}$) for a 200 yr old oak-hickory-pine stand (Switzer et al. 1979) as well as much higher values (60 to $80 \text{ g C m}^{-2} \text{ yr}^{-1}$) for an 80 yr old stand (Jenkinson 1971) have been measured in an old field succession (long-term agriculture to oak forest). In absence of more detailed long-term C-sequestration rates, we adopted a rate of $9.2 \text{ g C m}^{-2} \text{ yr}^{-1}$ for Hainich forest, i.e. slightly lower than the rate determined by Robertson & Vitousek (1981). For roughly 16 000 ha of Hainich forest, this results in C-sequestration equivalent to the emissions of 540 average German citizens (for calculation, see Appendix 1).

As alternatives to the status quo, we offered respondents 2 different local afforestation measures: one that results in the sequestration of the CO_2 emitted by an additional 90 individuals (Level 1; resulting sequestration of 630 individuals), and one that sequesters twice as much (Level 2; resulting sequestration of 720 individuals). At C-sequestration rates typical for young aggrading forests ($6.2 \text{ t CO}_2 \text{ ha}^{-1} \text{ yr}^{-1}$; Deckmyn et al. 2004), around 145 ha of afforestation area would be required for a Level 1 measure, and 290 ha for Level 2 (for calculation, see Appendix 1). The size of these 2 afforestation measures seems somewhat arbitrary. The measures were chosen (1) to convey the idea that a substantial improvement of C-sequestration of Hainich forest could be achieved, but (2) without requiring areas of land so extensive as to cause negative emotional reactions.

To calculate the shadow price of C-sequestration, we first calculated the marginal WTP (mWTP) for a 1-level change in the afforestation attribute per respondent. Because of the strategy to relate preferences for C-sequestration to personal emissions, i.e. to individual responsibility for climate change, this mWTP is then attributed to the individual average annual C emission of about 2.7 t yr^{-1} .

2.3.2. Adaptation: removal of potentially invasive plants

Several invasive species are likely to benefit from expected global climate change trends (Rejmanek 2000). Invasive plants are regarded as the second largest threat to biodiversity in Central Europe (Kowarik 2003). Besides competing with native species for habitats and resources, invasive species can harm forestry and agriculture. For instance, the giant hog-

weed *Heracleum mantegazzianum* is a host for the crop pathogen *Sclerotinia sclerotina* (Williamson & Forbes 1982), can reduce species richness of invaded habitats by 40% (Pyšek & Pyšek 1995), and is known to produce human-toxic substances in its sap (Jaspersen-Schib et al. 1996).

Our pre-study showed that respondents felt threatened by non-native plants, particularly by the giant hogweed. Thus, we decided to use *Heracleum mantegazzianum* to explain risks of invasive plants that are known to be harmful. We used 2 attribute levels—Level 1 represented the current practice (status quo), i.e. removal of invasive plants in particular cases where negative effects are known. Level 2 represented the alternative measure, i.e. large-scale removal of invasive plants regardless of whether they have negative effects or not.

We had no *a priori* hypothesis that Level 2 constituted, on average, a utility gain compared with the status quo.

2.3.3. Adaptation: increasing forest resistance and resilience to pests and storms

Shifts in life cycles of pests (e.g. beetles) can result from climate change (Parmesan 2006). One example is the mountain pine beetle *Dendroctonus ponderosae* in the Rocky mountains in the USA which now only takes one year per generation rather than its previous 2 yr life cycle (Logan et al. 2003). This, in turn, can increase infestation rates with the pine blister rust fungus *Cronartium ribicola*. Also wind storm related risks rise for German forests (Zebisch et al. 2005, p. 6). Stochastic dynamic models provide evidence that asynchronicity of species responses can ensure that species-rich forest ecosystems provide climate change adaptation services because they are able to adjust better to the changing environmental conditions (cf. Yachi & Loreau 1999).

Our qualitative interviews revealed that respondents perceived forests as threatened by insect pest outbreaks, particularly from bark beetles *Scolytida* spp. and through damage by caterpillars, and by storm events. We bundled these rather well-known threats into one attribute on forest resistance and resilience (*Widerstandskraft*) to pests and storms. Indicating that this is a simplification of a more complex ecological reality, we explained that better resistance and resilience could be effectuated by a higher diversity of site-adapted tree species.

We offered 3 attribute levels ranging from 'high' resistance and resilience to insect pests and storms by planting additional site-adapted tree species, to 'low' resistance and resilience characterised by low diver-

sity of site-adapted tree species. In the Hainich project region, several monoculture forest tree plantations exist in which respective 'diversification measures' could be implemented.

2.3.4. Adaptation: increasing general forest ecosystem resistance and resilience

Because not all climate-change related effects are apparent today, biological insurance services in the face of new, fundamentally uncertain or unknown risks (Barkmann & Marggraf 2007, Rajmis et al. 2008) are called for. Thus, we included an attribute on increasing general forest ecosystem resistance and resilience. The WBGU (German Advisory Council) on Global Change (WBGU 2000) suggests a precautionary risk management strategy to handle as-yet undiscovered risks. This argument is often implicitly used when ecologists and environmental economists suggest preserving as many species as possible, thereby hoping for an ecological insurance effect (Perrings 1995b, Gitay et al. 1996, Myers 1996, Edwards & Abivardi 1998, Yachi & Loreau 1999). From an ecosystem science perspective, species diversity provides key information on the processing of matter and energy within an ecosystem, and is an indispensable factor of the capacity of the ecosystem to self-organise in the face of internal or external changes (Kutsch et al. 2001).

Soil organisms are essential for the fundamental functioning of terrestrial ecosystems (e.g. Balsler et al. 2002). One group of species whose diversity has been linked to the stability properties of forest are the mycorrhizae (Bot & Benites 2005). Mycorrhizal fungi play a key role in the uptake of nutrients and water (Galvan et al. 2007), and contribute to the maintenance of soil porosity and organic matter content (Dighton 2003).

We chose a measure of soil organism diversity to indicate the capacity of the Hainich forests to self-organise, and thus to provide 'general' forest ecosystem resistance and resilience over other biota, as soil organisms are not visible. Changes in tree or herb diversity, for example, may have triggered the statement of additional aesthetic preferences in our non-expert respondent sample. This would have made it more difficult to interpret WTP statements for an attribute on general forest resilience. Interviewers briefly explained the importance of mycorrhizae, and their role in generalised ecosystem resilience and resistance in the face of as yet unknown environmental threats. Furthermore, it was explained that forest thinning may improve mycorrhizae numbers, and without being specific about improvement measures, explained that stands with lower mycorrhizal diversity exist in the Hainich project region in which mycorrhizal diversity could be fostered.

We offered 3 levels to respondents: 'high', 'medium' and 'low' resistance and resilience to unknown threats. The attribute level description for the option only included this quantification. It did not mention the mycorrhizal indicator nor did it hint at potential forestry measures (see Fig. 2).

Changes between levels are difficult to bring about by traditional forestry measures for many forests. Still, mycorrhizal diversity responds to changes in forest structure (Kernaghan et al. 2003), at least partly to inoculation in afforested stands (Menkis et al. 2007), and changes in atmospheric carbon inputs (Talbot et al. 2008). Because we restrict the interpretation of the attribute to adaptive ecosystem insurance services (S. Rajmis et al. unpubl.), the fact of mycorrhizal diversity can, in principle, be managed suffices to justify the utilisation of mycorrhizal diversity as a plausible indicator of positive biodiversity effects on general forest resistance and resilience. Although few respondents were knowledgeable about the importance of tree-mycorrhizae interactions, our extensive preliminary studies clearly showed that the role of functional diversity and functional redundancy for ecological insurance (McCann 2000) was familiar to respondents if described in non-technical terms (Rajmis & Barkmann 2007).

2.4. Preparation and administration of the CE study

From attributes and attribute levels (Table 2), we generated an orthogonal main effects design (Hensher et al. 2005, p. 115) with 64 different choice cards using the mix and match procedure by Chrzan & Orme (2000). The 64 choice cards were, in turn, randomly assigned to 4 blocks of 8-choice sets. Interviewers alternated blocks systematically with each consecutive respondent being assigned to the choice sets of 1 block.

The CE main survey ($n = 302$) was preceded by qualitative preliminary studies ($n = 16$; Rajmis & Barkmann 2007), pre-test interviews ($n = 57$), and a pilot study ($n = 106$). Pilot study and main study were conducted as a household survey using face-to-face interviews in 19 of the 30 villages adjacent to Hainich National Park, in the administrative districts Wartburg and Unstrut-Hainich, and in the neighbouring towns of Eisenach (pop. 43 300) and Bad Langensalza (pop. 18 390) (Thuringia). To sample private households, we used the random route method (Kromrey 2006). Interviewers were given exact points of departure in a certain street beginning with a defined starting point, and had rules about the direction to walk, which houses to call on and other details on establishing contacts to survey participants.

Table 2. Climate change mitigation and adaptation measures/indicators used in the choice experiment, translated from the German version of the questionnaire. Coding in parentheses if not directly given; ^{SQ}: status quo

Attribute	Measure/Indicator	Level of change of mitigation or adaptation measure/indicator		
Climate change mitigation				
Afforestation	Quantified by carbon sequestration/emission equivalents per capita	Sequestration/emission equivalents of 540 persons (540) ^{SQ}	Sequestration/emission equivalents of 630 persons (630)	Sequestration/emission equivalents of 720 persons (720)
Climate change adaptation				
Invasive plants	Removal of invasive non-native plants	Only if harmful and in particular cases (1) ^{SQ}	Large scale removal if harmful or not (2)	
Pests and storms	Planting of better site-adapted tree species	Low resistance and resilience (1)	Medium resistance and resilience (2) ^{SQ}	High resistance and resilience (3)
General forest eco system resilience	Changes in diversity of mycorrhizal fungi ^a	Low resistance and resilience (1)	Medium resistance and resilience (2) ^{SQ}	High resistance and resilience (3)
Cost (price)	Income change per month per person by a dedicated tax	0 € ^{SQ} , 5 €, 10 €, 20 €, 35 €, 50 €, 60 €, 80 € (= coding)		

^aChanges in the diversity of mycorrhizal fungi are an indicator of attribute level changes rather than a measure themselves, see Section 2.1

Target respondents were all German-speaking individuals who were at least 18 yr old. Target persons were identified by the 'last-birthday' method, i.e. the person with the most recent birthday was sampled (Rizzo et al. 2004). All interviews (pilot and main study) were conducted by 5 trained university students.

Of the respondents contacted, 471 declined to give an interview. Of the successfully contacted 302 households, 282 respondents completed the choice task. 20 respondents (6.6%) protested the choice experiment (see Adamowicz et al. 1998b). Eleven of these protested against the cost attribute (rejection of any extra tax, extremely skeptical attitude towards implementation of the proposed measures); 3 respondents refused any measure in general (strongly disliked) involving Hainich National Park, while another 2 respondents had a strict non-interference attitude towards the national park. Only 2 respondents protested the proposed mitigation and adaptation measures. Finally, 2 individuals clearly did not understand the choice task. An average interview took 35 min. In addition to a suite of attitudinal questions not included in this paper (Rajmis et al. 2008), we asked respondents to rate the importance of the CE attributes explicitly ('most important = 5' to 'completely unimportant = 1'). A sample of choice cards is shown in Fig. 2. See Fig. 4 for details on rating.

3. ECONOMETRIC ANALYSIS

We assume an additive utility function linear in parameters with respect to the attribute levels as coded in

Table 2 (see also Section 2.1). The utility function can be separated into an observable component V_{in} and unobservable (error) component ε_{in} :

$$U_{in} = V_{in} + \varepsilon_{in} \quad (1)$$

where U_{in} is the total utility of option i for individual n . The probability (Pr) that individual n will choose option i over option j within the complete choice set C is given by:

$$\Pr_{in} = \Pr (V_{in} + \varepsilon_{in} > V_{jn} + \varepsilon_{jn}, \text{ all } j \in C) \quad (2)$$

If a deterministic utility component V_1 is hypothesized to be a linear function of attribute Z_1 itself, plus an interaction term of the attribute Z_1 with an individually varying socio-demographic variable A , V_1 can be expressed as:

$$V_1 (Z_1, A) = c_A Z_1 A + c_1 Z_1 \quad (3)$$

where c_A is the utility coefficient of the interaction term (Barkmann et al. 2007). In the econometrically estimated utility models, a positive coefficient c indicates a positive influence of the respective term on choices, and thus on utility. To reduce collinearity between the interaction term and the non-interacted attribute term, the socio-demographic variable A can be standardized before being multiplied by Z_1 . The vector of the utility coefficient is usually estimated with maximum likelihood estimation techniques. Usually the estimated choice models include an alternative specific constant (ASC) that picks up systematic differences in choice patterns between the 3 choice cards. The ASC was coded 'zero' for cards A and B, and '1' for the status quo option (status quo-ASC).











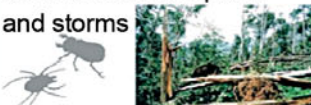
















1/6	Measure A	1/6	Measure B
Risks from climate change 	Carbon sequestration 720 persons 	Risks from climate change 	Carbon sequestration 630 persons 
Risks from non native plants  	Removal of non-native plants over large areas if harmful or not	Risks from non native plants  	Removal of harmful non-native plants in particular cases
Risks from insect pests and storms 	High resistibility 	Risks from insect pests and storms 	Low resistibility 
Threats of unknown risks 	High resistibility 	Threats of unknown risks 	Resistibility 
Extra duty per year 	+ 60 €	Extra duty per year 	+ 35 €

Fig. 2. A sample set of choice cards (translated from the original German version). Forest ecosystem resistance and resilience (*Widerstandskraft*) is translated as ‘resistibility’

Preliminary analyses unveiled a risk of violation of the independence from irrelevant alternatives (IIA) assumption necessary for the application of the (simple) conditional logit model. Thus, nested logit models (NL) were used that partially relax the IIA assumption (Train 1998, Hensher et al. 2005, p. 518). Suitable NL model structures were identified, and the corresponding models estimated with NLOGIT 3.08 (www.limdep.com). The inclusive value was set to 1.0 for the degenerated branch, and the model initiated with starting values obtained from a non-nested NL model (Hensher et al. 2005, p. 530). All scale parameters were normalized at the lowest level (random utility level 1, RU1). As model statistics we report adjusted pseudo-R² values in relation to a ‘constants only’ model. Values between 0.06 and 0.07 correspond to R² values of 0.19 and 0.22 in relation to their ordinary least squares equivalents (Hensher et al. 2005, p. 338).

To test for influences of socio-demographic variables on choices, interaction terms of the ASC and of the climate change mitigation attribute were generated (see Table 5). The improved model was generated stepwise by initially including all statistically ‘significant’ interaction terms from single interaction models and then

	Present situation
Risks from climate change 	Carbon sequestration 540 persons 
Risks from non native plants  	Removal of harmful non-native plants in particular cases
Risks from insect pests and storms 	Resistibility 
Threats of unknown risks 	Resistibility 
Extra duty per year 	+ 0 €

deleting non-significant interaction terms one by one. Significant interaction terms of socio-demographic variables and the climate change mitigation attribute are presented in the improved model (see Table 5).

Statistically significant attribute coefficients allow for the calculation of respondent WTP for attribute level changes. For attributes linear in parameters, the mWTP equals the negative ratio of the respective attribute coefficient c_z and the coefficient of the monetary attribute c_y :

$$\text{mWTP} = -\frac{c_z}{c_y} \quad (4)$$

All WTP values given in Table 4 refer to a marginal one-level change in the attributes. For the 20 respondents (6.62% of interview respondents) protesting the choice experiment, mWTP = 0 was assumed.

4. EMPIRICAL RESULTS

4.1. Socio-demographic characteristics of respondents

The gender ratio of respondents was 49:51 (female: male; Table 3). The youngest respondent was 18 yr old, the oldest was 92 yr. The majority of respondents (45%) had a secondary school certificate and received a salary (47 to 52.2%). Average monthly net income was 1125 € per person. Socio-demographic key variables were close to state and national averages (Fig. 3;

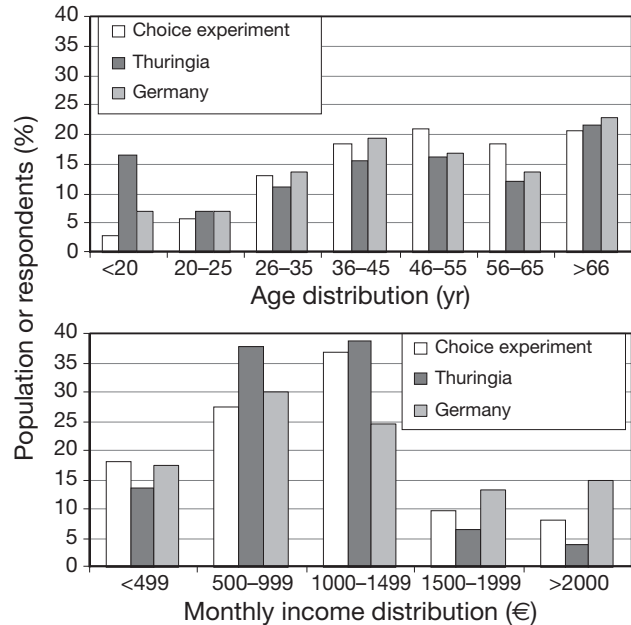


Fig. 3. Age and income distribution of choice experiment respondents (white bars, $n = 282$; this study), and of the populations of Thuringia (dark grey bars; Thüringer Landesamt für Statistik 2008), and Germany (light grey bars; Statistisches Bundesamt 2008)

Thüringer Landesamt für Statistik 2008, Statistisches Bundesamt 2008). For park usage, 86% of respondents had visited Hainich NP at least once. On average, respondents visited the park 14 times per year.

Table 3. Socio-demographic characteristics of respondents

Variable	Description	Results (SD)
Gender	Dummy variables describing gender of respondents [0: female; 1: male]	49% women, 51% men
Age	Age in years	50.27 yr (16.95) [between 18 and 92 yr]
Education	5 point score to describe formal educational level attained [1: None; 2: general education school level; 3: secondary school level; 4: A-level; 5: university of applied sciences ^a degree or university degree]	19% general education school level; 45% secondary school level; 12% A-level; 24% university of applied sciences degree or university degree
Occupation	Dummy variables to describe occupation status [1: employed; 1: self-employed; 0: unemployed, 0: in training/student, 0: retired; 0: homemaker]	Active work force (employed, self-employed) 52.2%; other (unemployed, in training/student, retired, homemaker) 47.8%
Income	7 categories from <500 to >4000 € mo ⁻¹ person ⁻¹	1125 € mo ⁻¹ person ⁻¹ (760.5 €)
Residence	Dummy variable describing residence of respondents [0: rural; 1: urban]	31.6% urban, 68.4% rural resident
NP user status	Dummy variable if respondent was at least once in Hainich NP [0: no; 1: yes]	86% visited Hainich NP at least once
Frequency of NP visits	Frequency of visiting Hainich NP yr ⁻¹	14.01 visits yr ⁻¹ (38.2)

^aIn the German system of higher education, universities of applied sciences (*Fachhochschulen*) have lower entry requirements than full universities and usually grant BSc or BA degrees

4.2. Importance of attributes for respondents

According to the results of our survey, 91.4% of respondents rated the attribute climate change mitigation by carbon sequestration as very important to rather important. 80.5% of respondents rated the attribute climate change adaptation by pest and storm resilience as very important or rather important. General ecosystem resilience was rated by 68.5% of respondents as very important to rather important. 55.3% of respondents rated climate change adaptation towards invasive plants as very important to rather important.

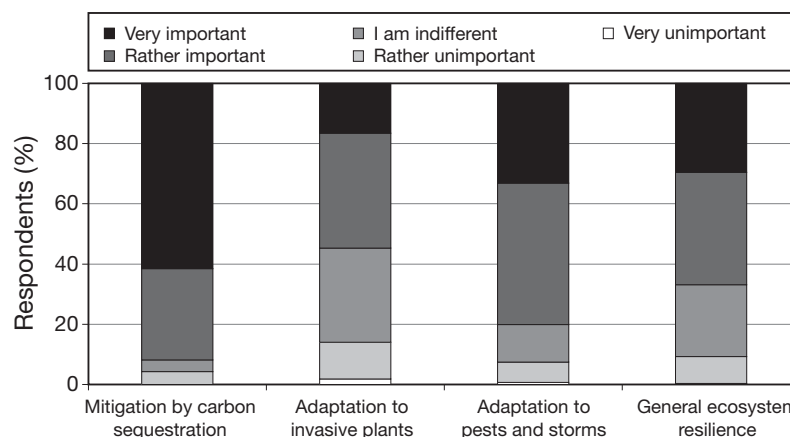


Fig. 4. Importance of climate change mitigation and adaptation rated by respondents

4.3 Base model results

The model is overall most highly significant (p chi-squared < 0.0001 ; $df = 7$). All attributes are significant determinants of choice; 4 attributes (including cost) at $p < 0.001$ and 1 attribute at $p < 0.05$ (Table 4). Before correction for protest responses, mWTP values for a 1-level change in attributes range from 27.54 to -9.59 €. While respondents supported afforestation, increasing protection levels against pest and storms, and better general ecosystem resilience, more aggressive measures to reduce potential threats from invasive plants were disliked (negative WTP). Climate change adaptation to the risks from pests and storms had the highest mWTP (27.54 €). For climate change mitigation, mWTP was 21.43 €, for general ecosystem resilience it was almost 17 €. Compared to the current practice of selective removal of harmful invasive

plants, more aggressive removal schemes including removal of plants without known dangers were valued less (mWTP = -9.59 €). When including protest responses (WTP = 0 €), slightly lower mWTP values are obtained (6.62% of respondents protested).

Taking WTP values corrected for protest responses into account, a shadow price for C-sequestration related to personal emissions can be calculated. Based on the mWTP value for a 1-level change of 20.01 € yr^{-1} divided by ~ 2.7 t C emissions per year, a shadow price of 7.34 € $\text{t}^{-1} \text{yr}^{-1}$ is calculated.

4.4 Model interactions: influences of socio-demographic characteristics of respondents

Age, gender and education were concurrently interacted with the afforestation attribute (climate change mitigation) and the ASC; 4 interactions were signifi-

Table 4. Nested logit estimation results of the base model. Willingness-to-pay (WTP) values refer to a 1-level shift of each attribute; $n = 282$; for attribute measures/indicators and codings, see Table 2. Marginal WTP: WTP of a marginal unit of the attribute, i.e. of one attribute level difference as coded in Table 2; Corrected marginal WTP: corrected for protest responses counted as stating a WTP of 0 €; Status-quo-ASC: alternative Specific Constant with respect to the status quo option (coded 1). ** $p < 0.05$; *** $p < 0.001$

Attribute	Coefficient	SE	Marginal WTP (€ yr^{-1} person $^{-1}$)	Corrected marginal WTP (€ yr^{-1} person $^{-1}$)
Climate change mitigation				
Afforestation/100	0.3537***	0.0466	21.43	20.01
Climate change adaptation				
Invasive plants	-0.1582^{**}	0.0711	-9.59	-8.96
Pests and storms	0.4545^{***}	0.0476	27.54	25.72
General ecosystem resilience	0.2777^{***}	0.0462	16.83	15.72
Status quo-ASC	-2.2740^{***}	0.5068	-137.80	-128.68
Cost	-0.0165^{***}	0.0014	na	na
Log-likelihood	-2257.510			
Chi-squared	1041.248			
Pseudo- R^2	0.0621			
Observations (n)	2256			

Table 5. Nested logit estimation results of the improved model. * $p < 0.05$; ** $p < 0.01$; *** $p < 0.001$; $n = 282$

Variable	Coefficient	p
Climate change mitigation		
Afforestation/100	0.3297***	0.0000
Afforestation \times age	0.1111**	0.0088
Afforestation \times gender (male = 1)	-0.1538***	0.0004
Climate change adaptation		
Invasive plants	-0.1586*	0.0186
Pests and storms	0.4191***	0.0000
General ecosystem resilience	0.2481***	0.0000
Status quo-ASC	-1.2635***	0.0000
Status quo-ASC \times age	0.3008***	0.0000
Status quo-ASC \times education	-0.1238*	0.0174
Cost attribute	-0.0157***	0.0000
Log-likelihood	-2230.352	
Chi-squared	1095.564	
Pseudo-R ²	0.07252	
Observations	2256	

cant at least at the 5% level. The interaction between age and climate change mitigation ($p < 0.01$; Table 5) shows that age (coded in years between 18 and 92) has a positive effect on higher preferences for the proposed climate change mitigation measures (if an improvement measure is chosen). The interaction between gender and climate change mitigation shows that women (women coded '0'; men coded '1') have stronger preferences for the proposed climate change mitigation measures ($p < 0.001$). For status-quo-ASC \times age, age has a positive effect on the probability for the current situation (status quo) being chosen ($p < 0.001$). For status quo-ASC \times education, the lower the education level, the higher the probability that the current situation (status quo) will be the choice ($p < 0.05$).

5. DISCUSSION

We conducted a CE on one climate change mitigation and 3 climate change adaptation attributes focusing on forest ecosystem biodiversity. Extensive qualitative preliminary studies (Rajmis & Barkmann 2007) had shown that respondents are sufficiently aware of the topics covered by our choice experiment attributes. To ensure understanding, attribute level details and wording were designed to reflect respondent previous knowledge on climate change risks as well as forest-based adaptation and mitigation measures. Furthermore, the CE was conducted as a face-to-face survey based on sampling procedures that achieved a good

socio-demographic match of respondents with population means (Fig. 3). The low number of respondents who protested the CE ($n = 20$; 6.62%) further supports our positive interpretation of the validity of the instrument with regard to the critical issue of respondent familiarity with ecosystem service-type CE attributes (cf. Barkmann et al. 2008).

Supporting content validity, all attributes are significant determinants of choice and display meaningful (signs of the) attribute utility coefficients (see Bateman et al. 2002, p. 363). WTP for climate change mitigation and adaptation ranges between -8.96 and 25.72 € yr^{-1} respondent⁻¹ for a one-level change (cf. Table 2). The strongest WTP of 25.72 € yr^{-1} respondent⁻¹ was identified for a 1-level change of climate change adaptation towards pests and storms. An independent importance rating of the attributes by respondents before actually making the choice shows that mWTP follows the importance rating for most attributes (cf. Table 4 & Fig. 4).

The ASC value regarding climate change measures in the Hainich forest area indicates that respondents have strong preferences for improving the status quo beyond what is depicted by the proposed measures displayed in the choice cards (negative ASC coefficient). Such a positive view is stronger with younger respondents (positive sign of interaction term ASC \times age) and for better-educated respondents (negative sign of interaction term ASC \times education). These results do not give any indication that choosing the status quo option was used simply to avoid making a decision at zero cost by respondents. This would have constituted a so-called status quo 'bias' casting doubts on the results. The positive valuation of the proposed scenarios beyond explicitly depicted attributes also shows that respondents do associate additional, implicit benefits with the proposed measures. Certainly, we cannot categorically exclude that such implicit benefit assignments also influenced the valuation of the explicit mitigation and adoption attributes. Still, the material and statistical significance of the coefficient of the ASC terms supports the notion that the majority of such benefits were actually captured by the ASC terms, and consequently do not interfere with the interpretation of the calculated mWTP for the climate change mitigation and adaptation attributes.

Viscusi & Zeckhauser (2006) found a mean WTP of 0.32 € per gallon of gas to eliminate the risks of climate change for a 30 yr period. Similarly, Reiner et al. (2006) asked how much respondents would pay on top of their existing electric bill to 'solve global warming'; 90% of respondents were willing to pay a maximum of 32 € per month, (<400 € yr^{-1}). When the Viscusi & Zeckhauser (2006) result is converted into WTP per t C in fuel, this represents about 140 € t^{-1} and about 170 € per average fuel consumption 2005 per vehicle (calculation based

on US Department of Transportation 2008). This is much higher than the shadow price for the sequestration of 1 t C emitted in our results (7.34 € t C emitted person⁻¹ yr⁻¹). These values cannot be directly compared, however, because the first value constitutes WTP for a fuel tax that eliminates the risks of climate change (and its potentially catastrophic negative effects) while our study proposes the much less ambitious effect of reducing one's own net carbon footprint. If we interpret the total documented WTP in our study for a 1-level improvement of climate change mitigation and adaptation measures including the status quo-ASC (see Table 4), WTP = 190 € person⁻¹ yr⁻¹. As an equivalent to 'stopping climate change' or 'solving global warming', this value is similar to the one implied by Viscusi & Zeckhauser (2006). In any case, our WTP-values are much lower than those in the study by Reiner et al. (2006).

Alavalapati et al. (2004) investigated benefits for 3 environmental attributes (C-sequestration, water quality, wildlife habitats in silvopasture). Average WTP for moderate improvements in all 3 environmental attributes was ~94 € yr⁻¹ per household for a 5 yr program. For our 4 attributes, a 1-level improvement results in a total WTP of ~60 €. Annual WTP for C-sequestration alone is estimated at between ~39.5 and ~42.7 € by Alavalapati et al. (2004), and between ~20 and ~40 € in our study (1- or 2-level improvement). Thus, WTP for C-sequestration alone appears lower in our study, but for all mitigation and adoption options it is higher, however, again lower than Alavalapati et al.'s (2004) more inclusive bundle of environmental improvements (including carbon sequestration, water quality and wildlife habitat in silvopasture). Additionally, some of the unexplained preferences for improvements over the status quo in our data (see Section 4.3) may have been attributable to specific attributes if we had also used attributes with a wider scope, including e.g. water quality and wildlife habitat. As actual values at each of these levels of comparison differ by less than a factor of 2, we interpret our data as generally supporting the findings of Alavalapati et al. (2004).

In summary, several studies including this one, document substantial citizen WTP for additional climate change mitigation and/or adaptation measures. The paucity of empirical data and the diversity of approaches do not allow for a definite conclusion over the absolute level of citizen WTP for climate change mitigation and/or adaptation. Even much more elaborate attempts to compare studies than accomplished here, e.g. accounting for temporal differences in real

purchasing power and socio-economic differences between studies and samples, will suffer from the incompatibility of the units by which C-sequestration is quantified. While this may not be completely avoidable, we recommend that future studies include a best estimate of WTP per t C additionally sequestered per person per year.

Most importantly, when comparing the few data sets available it must be kept in mind that the specific benefits associated with mitigation or adaptation options to be valued may influence value statements. We have discussed several examples above. Furthermore, our study differs from existing work in its community focus. Respondents are faced with mitigation and adaptation options to be implemented in the Hainich forest region, a local leisure and tourism area frequently visited by 86% of respondents. This means that respondents directly benefit from the proposed adaptation options. If we had proposed to implement them further away, lower stated preferences would be expected for other forests in Germany or abroad².

Finally, the regional setting of our study has an interesting side-effect in relation to nature conservation concerns. Because all mitigation and adoption options presented rest on improvement of forest biodiversity, our results highlight potential regional synergies between climate change and conservation policy. These synergies most likely do not exist for alternative climate change mitigation measures such as the production of agricultural biofuels (Reijnders 2009, Pineiro et al. 2009).

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Appendix 1. Background and calculations on WTP for additional sequestration of one ton self-emitted carbon being sequestered by afforestation measures (see Table 2 for measures)

Description of Status Quo calculation

Hainich NP has around 7600 ha; the whole Hainich forest area is about 16000 ha. In the current situation (status quo) long-term C accumulation of forest soils is estimated in a range from 2.4 to 450 g m⁻² yr⁻¹ (see e.g. Robertson & Vitousek 1981, Switzer 1979). For a beech stand in Hainich NP, high net carbon uptake values of up to 580 g m⁻² yr⁻¹ have been measured by Eddy covariance (Knohl et al. 2003). However, if we suggest smaller annual C sequestration rates (e.g. about 9.2 g m⁻² yr⁻¹), the whole Hainich area (16000 ha) would display an annual C uptake of roughly 1470 t. Converting C uptake into CO₂ equivalents (conversion factor of 1/0.2727 = 3.6 according due to the relation C/CO₂ and the molecular masses C:12 and O₂:32 = 12/44 = 0.2727), the carbon of about 5400 t CO₂ could be sequestered. With an annual CO₂ emission of about 10 t person⁻¹ yr⁻¹ in Germany, this is equivalent to the CO₂ emissions of about 540 citizens.

Description of afforestation measures (Level 1 and Level 2 of the climate change mitigation attribute)

Deckmyn et al. (2004) found by means of a carbon accounting model that an oak-beech forest reduces CO₂ emissions between 6.2 and 7.1 t CO₂ ha⁻¹ yr⁻¹ (Deckmyn et al. 2004). Considering the CO₂/C relation, this is a sequestration rate between 6.2/3.6 = 1.72 t C ha⁻¹ yr⁻¹ and 1.97 t C ha⁻¹ yr⁻¹. Sequestration rates by a young beech forest in southern France were estimated at max. 1.46 t C ha⁻¹ yr⁻¹ (Granier et al. 2000).

Calculating with minimum sequestration rates from Deckmyn et al. (2004) of 6.2 t CO₂ ha⁻¹ yr⁻¹ (900/6.2), 145 ha would be needed to be afforested to sequester annually 245 additional t C. Converted to German citizen CO₂ emission equivalents, this equals about 90 average citizen CO₂ emissions (difference between status quo and Level 1 and between Level 1 and Level 2). Taking a two level shift in account (from status quo to Level 2), about 290 ha would be needed to be afforested to compensate for 180 persons each emitting 10 t yr⁻¹