Delineation of a coastal gray whale feeding area using opportunistic and systematic survey effort

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ABSTRACT: A seismic survey took place during June and July 2010 adjacent to the gray whale (Eschrichtius robustus) coastal feeding area on the northeast Sakhalin Shelf, Russia. Seismic surveys produce underwater sound that can cause hearing injury and behavioural disturbance in marine mammals. In addition to common mitigation measures to prevent acoustic injury, mitigation measures to avoid behavioural disturbance to gray whales within the feeding area were applied. This behavioural mitigation required delineation of the feeding area; however, no clear boundary was obvious because gray whale distribution within the feeding ground varies within and across years. We estimated the feeding area’s offshore boundary using a 1.0 km² gray whale relative density surface derived from systematic and opportunistic survey data collected during June and July 2005 to 2007. We calculated a separate surface for each of the systematic and opportunistic data sets, then calibrated and merged the 2 surfaces. We evaluated 3 geostatistical kriging methods (ordinary, simple, and co-kriging) that were applied to the merged surface to estimate a smoothed surface across areas with and without survey effort. Simple kriging was most suitable due to its ability to transition over sharp gradients in whale abundance and provide reasonable predictions in data-void areas. A 95% abundance contour of the kriged surface was used as an estimate of the feeding area boundary. Our approach provided an objective and quantitative basis to delineate the feeding area boundary to support measures taken to mitigate the potential impacts of the seismic survey on the whales.

KEY WORDS: Gray whale · Eschrichtius robustus · Feeding area · Critical habitat · Seismic survey mitigation · Abundance surface · Systematic surveys · Opportunistic surveys · Kriging

INTRODUCTION

The summer range of gray whales Eschrichtius robustus includes 2 feeding areas on the northeast Sakhalin Shelf, Russia (Meier et al. 2007). An approximately 120 km long and 10 km wide nearshore ‘Piltun’ feeding area is located adjacent to the coast of Sakhalin Island. A deeper ‘Offshore’ feeding area is situated ~45 km southeast of the Piltun feeding area in water ~40 to 50 m deep. Higher densities of gray whales are typically observed in the Piltun area than the Offshore area (Vladimirov et al. 2013). The Piltun

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area is also an important habitat for cow-calf pairs and for newly weaned calves (Sychenko 2011).

The Sakhalin Shelf contains extensive oil and gas reserves presently being developed. Sakhalin Energy Investment Company Ltd. (hereafter Sakhalin Energy) conducted a repeat 3-dimensional (4-D) seismic survey in June and July 2010 to map production-related changes in subsea oil and gas reserves. The 167 km$^2$ seismic survey area was located approximately 10 to 25 km offshore, adjacent to the southern portion of the Piltun feeding area. Seismic surveys produce underwater sound that can cause hearing injury and behavioural changes in marine mammals (Richardson et al. 1984, Richardson et al. 1986, 1999, Ljungblad et al. 1988, McAuley et al. 2000, Gailey et al. 2007a, Yazvenko et al. 2015). In particular, Malme et al. (1986) found ~10% of gray whales stopped feeding and moved away from pulsed airgun sounds exceeding 163 dB re µPa rms.

Preliminary acoustic modelling for the 4-D seismic survey indicated that parts of the Piltun feeding area could be exposed to sound levels above 163 dB re µPa rms (IUCN 2007). Consequently, a monitoring and mitigation plan (MMP) was developed that used an equivalent per-pulse sound exposure value of 156 dB re µPa$^2$-s (156 dB SEL) as a threshold level for disturbance of gray whale behaviour (Bröker et al. 2015). The MMP required estimation of the Piltun feeding area boundary to delineate areas (‘A-zones’) within the feeding ground where sound levels exceeded 156 dB SEL. A-zones were defined for each survey line sailed by the seismic vessel as the overlap between the 156 dB SEL contour generated by the seismic source when acquiring that line and the Piltun feeding area boundary. Each A-zone was required to be clear of gray whales when that A-zone’s line was sailed.

Higher densities of gray whales in the Piltun feeding area have been observed in water depths of <25 m and between the latitudes of approximately 52° 20’ N to 53° 30’ N (Fadeev et al. 2012), but no clear boundary for the feeding area was obvious because distribution varies within and among years (Meier et al. 2007, Vladimirov et al. 2013). Amphipods and isopods are preferred prey in the Piltun feeding area; however, bivalve molluscs, worms, and sand lance Ammodytes hexapterus are also eaten opportunistically (Fadeev 2013). These prey have a patchy spatial distribution, and locations of high biomass vary across years, although higher biomasses are generally found in water 5 to 20 m deep that is typically located within 5 km of shore (Fadeev 2013).

The MMP defined the feeding area boundary as the line incorporating 95% of estimated gray whale abundance in the Piltun feeding area during the planned June–July time frame of the seismic survey (IUCN 2008). A 95% abundance contour is commonly used to delineate individual and species’ geographic range boundaries (Worton 1987, Laver & Kelly 2008, Kie et al. 2010). This paper presents the methods we used to estimate the boundary based on data from both systematic and opportunistic survey effort. While considerable systematic shore-based effort was available, spatial coverage of the feeding ground was constrained in some areas by low-elevation observation stations. Systematic vessel surveys provided some additional coverage in this vicinity, but effort was limited. The vessels used for systematic surveys spent 3 to 4 mo each season conducting other research (e.g. benthos sampling or photo-identification) on the northeast Sakhalin shelf. Marine mammal observers (MMOs) were on watch during daylight hours when environmental conditions permitted, thus providing considerable opportunistic effort that filled in some of the temporal and spatial gaps in systematic survey effort within the Piltun feeding ground and surrounding areas. Our approach calibrated and combined the 2 data sets to produce a smoothed surface for which a 95% abundance contour was estimated as a proxy of the Piltun feeding area boundary.

**MATERIALS AND METHODS**

**Survey effort**

Data from systematic and opportunistic survey effort during June and July 2005 to 2007 were used to characterize recent gray whale distribution patterns in the study area. Shore-based and vessel-based surveys systematically sampled gray whale distribution and abundance within the Piltun feeding area and its
surroundings (Fig. 1). Shore-based scan surveys were conducted daily throughout each field season, weather permitting, from several permanent observation stations along the coast (our Tables 1 & 2; Gailey et al. 2007b, 2008; Vladimirov et al. 2007, 2008). MMOs used Fujinon FMTRC-SX 7 × 50 reticle binoculars to scan the nearshore waters surrounding a station at a constant rate. Bearing, reticle estimate, and number of individuals were recorded for each sighting. MMOs also recorded environmental conditions for each scan at an observation station (Beaufort wind force scale [hereafter Beaufort Scale], visibility in km, wind speed and direction, presence and location of glare, and swell height).

Systematic vessel surveys were conducted up to a few times monthly using large vessels (mean length 74 m) in 3 areas of the northeast Sakhalin shelf (Vladimirov et al. 2007, 2008; Fig. 1, Table 3). Transects in the Piltun feeding area (‘Piltun survey’) were located parallel to shore. A single transect was sailed at distances of 2.5 km (2005) and 4 km (2007) from shore. The 2006 survey sampled 2 transects at 2.5 and 6.5 km from shore. Vessel surveys of the Offshore feeding area were conducted in 2005 along 11 east–west transects 28 km in length spaced 6.5 km apart. The ‘Offshore survey’ was redesigned in 2006 when the survey grid was shifted south by 2 km and the 8 southerly transects were retained and shortened to 23 km. A new ‘Arkutun-Dagi survey’ in 2006 sampled the deeper-water Arkutun-Dagi licence block adjacent to the southern part of the Piltun feeding area. The Arkutun-Dagi survey consisted of the remaining 3 northerly transects from the 2005 Offshore survey, with the addition of 4 transects to the north that were also spaced at 6.5 km intervals. Vessels sailed at 10 knots (Piltun) or 11 knots (Offshore, Arkutun-Dagi). Surveys were conducted only in good visibility (>5 km) and Beaufort Scale <4. Two MMOs were on watch during the systematic surveys in 2005. One or two MMOs were on watch in 2006, and a single MMO was on watch in 2007. In 2005, MMOs estimated the distance to a marine mammal sighting by eye when the animal was abeam of the vessel. Protocols were amended in 2006 to record distance and bearing at first detection. Distance was estimated using Fujinon FMTRC-SX 7 × 50 reticle binoculars, and the ship’s gyrocompass was used to estimate the azimuth to the sighting.
Vessels involved in work other than systematic whale surveys were required to remain at least 4 km from shore (approximately the 20 m isobath) when operating in the Piltun feeding area, unless conducting research activities that required the vessel to enter shallower water. In such cases, the vessel draft restricted it to waters >10 m deep. Two MMOs were on watch in 2005; a single MMO was on duty during 2006 and 2007. The opportunistic MMO protocols recorded the same attributes for sightings and individual gray whales recorded during each survey is shown. Arkutun-Dagi surveys began in 2006.
environmental conditions as those recorded for systematic surveys. Vessel GPS tracks were also recorded. Opportunistic surveys therefore provided considerable and complementary effort in deeper waters of the Piltun feeding area and on the Sakhalin shelf, although vessel speeds and environmental conditions varied considerably.

**Approach used to estimate the feeding area boundary**

Our approach to use systematic and opportunistic effort to estimate a 95% gray whale abundance contour of the Piltun feeding area consisted of 4 main steps (Fig. 2) that (1) separately estimated a relative density surface (whales per km², hereafter WPKM²) and whales per unit effort (WPUE) surface for the systematic and opportunistic data sets respectively, (2) calibrated and merged the 2 surfaces, (3) applied geostatistical kriging methods to the merged surface to estimate a smoothed surface across areas with and without survey effort, and (4) delineated a contour enclosing 95% of the estimated abundance in the resulting surface to represent the Piltun feeding area boundary. The spatial extent used for analysis was a bounding box of spatial coverage by both data sets, with exclusion of the Offshore feeding area so that its high gray whale abundance would not influence the boundary estimation. Details for each analysis stage are presented below.

**Systematic WPKM² surface estimation**

The study area was overlaid with a grid of 1.0 km² cells and a gray whale density estimated for each grid cell sampled by each systematic survey from a vessel transect or onshore station during June–July 2005–2007. Gray whale sightings were corrected for availability and detection biases that typically underestimate abundance (Marsh & Sinclair 1989). These correction factors were estimated separately for each platform. The correction for availability, i.e. the probability that a gray whale was on the ocean surface and available to be detected (a), was estimated for each year and survey platform based on McLaren (1961) using mean gray whale dive cycle times measured in the field (Gailey et al. 2007b, 2008). Distance sampling (Buckland et al. 2001, 2004) was used to analyse the

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**Fig. 2. Schematic showing the analysis steps for estimation of the Piltun feeding area boundary and determination of the acoustic monitoring line. Separate gray whale surfaces were estimated using systematic shore-based and vessel survey effort (top left) and opportunistic vessel survey effort (top right). The surfaces were calibrated and combined, and the resultant surface kriged to produce a smoothed surface from which the 95% abundance contour for the Piltun feeding area was estimated.**
effects of distance and other factors (whale group size, Beaufort Scale, glare, visibility, number of observers, and year) on detection probability of available gray whales for vessel surveys (p). Due to low sample sizes, Beaufort Scale was categorized only into ‘Low’ (0, 1) and ‘Moderate’ (2, 3), and glare was categorized into ‘Light’ and ‘Severe’. Gray whale sightings made during the Arkutun-Dagi and Offshore vessel surveys throughout June to October from 2006 to 2007 were used to estimate the vessel detection function to increase the sample size and precision of estimates. Surveys in 2005 were excluded because MMOs did not record sighting distance and bearing at time of first detection (see ‘Survey effort’). Sightings made during Piltun surveys were also excluded because there is a gray whale density gradient with respect to shore in the Piltun feeding area. The Piltun transects were positioned parallel to shore and thus were parallel to the gray whale density gradient, which violates one of the main assumptions of distance sampling, i.e. that objects are distributed uniformly with respect to distance in any direction from the sampling point (Buckland et al. 2001). Similarly, the gray whale density gradient prevented conventional distance sampling from being used to estimate the shore-based detection function. Instead, we used a double platform (vessel- and shore-based) experiment to estimate the shore-based detection function, which was flat (i.e. detection did not decrease with increasing distance from the observer) up to the 8 km distance tested, i.e. \( \hat{p}_{ij} = 1 \) (E. Rexstad & D. Borchers unpubl.). A gray whale density \( (\hat{D}_{ij}) \) was estimated in the \( j \)th grid cell that was sampled during a particular survey \( i \) as follows:

\[
\hat{D}_{ij} = \frac{1}{\hat{a}_{ij} A_{ij}} \sum_{k=1}^{S_{ij}} \frac{\hat{p}_{ij,k}}{\hat{p}_{ij}} 
\]

where \( \hat{a}_{ij} \) is the estimated availability correction for the platform that sampled grid cell \( j \) during survey \( i \); \( A_{ij} \) is the area covered by survey \( i \) in the \( j \)th cell; \( S_{ij} \) is the number of sightings by survey \( i \) in cell \( j \); \( n_{ij,k} \) is the number of gray whales observed in the \( k \)th sighting by survey \( i \) in cell \( j \); and \( \hat{p}_{ij,k} \) was set to 1 as described above for shore-based surveys or estimated for the sighting’s distance from the transect using the vessel detection function.

A density of zero was assigned to a sampled grid cell if no gray whales were sighted within it. No density was estimated for unsurveyed grid cells. The estimated survey-level densities across all platforms with effort during June–July 2005–2007 were used to calculate a weighted average within each grid cell, with weights proportional to the amount of area covered in the cell by a survey \( (A_{ij}) \), to produce the systematic survey density surface. Further details of the density estimation methods may be found in Vladimirov et al. (2011).

Opportunistic WPUE surface estimation

We used 3 main steps to create a WPUE surface for opportunistic June–July 2005–2007 vessel data: (1) effort segments were constructed from adjacent vessel positions with similar speeds and environmental conditions, (2) distance sampling (Buckland et al. 2001, 2004) was used to model a detection function for gray whale sightings associated with the effort segments, and (3) WPUE was estimated for each sampled cell of the 1 km² spatial grid using the effort segments, the right truncation distance of the detection function to represent the approximate width of effort for each segment, and the gray whale sightings. These steps are described in detail below. We included data from August 2005–2007 in the first 2 steps to increase sighting sample size and precision in the estimate of the detection function. However, these data were not used when estimating the WPUE surface.

Segment creation. Vessel effort was mainly provided by GPS tracks. Tracks were not available for 1 of 2 vessels that operated in 2007. We inferred this vessel’s locations from the associated MMO records that noted the vessel location upon record entry. MMOs did not record watch start and end times; they did, however, enter a record at least every 30 min when on watch. We assumed track locations were off effort if they occurred (1) before the first MMO record of the day, (2) >30 min after the final MMO record of the day, or (3) were within the first and final MMO records of the day but not within 1 h of a record, i.e. we assumed MMOs had gone off watch due to poor environmental conditions. All ‘off effort’ track locations were excluded. Vessel locations associated with reported Beaufort Scale >4 (opportunistic effort occurred at Beaufort Scale ≤5) were also removed because sighting detection typically deteriorates at higher sea states (Barlow et al. 2001).

Successive retained vessel positions were aggregated into initial segments with similar sighting conditions and speed to ensure similar whale detection conditions within a segment. Only segments with speeds >5 km h⁻¹ were retained because this speed likely exceeded that of feeding gray whales (i.e. mean 1.1 km h⁻¹, SD 0.55; Gailey et al. 2009). Slow vessel speeds can inflate encounter rates and associated density estimates because an animal can travel alongside the vessel, or new animals can move into
the search area (Buckland et al. 2001). Buckland et al. (2001) recommended that average observer speeds be at least 2- to 3-fold greater than that of the animals. Adjacent initial segments were merged with the constraint that the combined segment’s total time was <1.2 h and total length did not exceed 9 km. Adjacent initial segments were only combined if their difference in speed was <5 km h\(^{-1}\) or if the mean speed was 15 km h\(^{-1}\) with the difference in the segments’ speed no greater than 10 km h\(^{-1}\).

**Detection function modelling.** Distance v6.0 (Thomas et al. 2010) was used to model a gray whale detection function using perpendicular distances of sightings from their associated segments. North-south segments and associated sightings within 5 km of shore were excluded due to the whale density gradient with respect to shore and resultant violation of a distance sampling assumption (as described above for the systematic survey detection function modelling). The effects of covariates (visibility, Beaufort Scale, group size, vessel and observer) on detection were tested (Marques & Buckland 2004). A stepwise forward selection procedure was used (starting with a model containing perpendicular distance only) to decide which covariates to retain, with a minimum Akaike’s information criterion (AIC) inclusion criterion.

**WPUE surface creation.** June and July 2005–2007 ‘on effort’ segments were used to allocate effort, expressed as times visited, to each 1.0 km\(^2\) grid cell in the surface. We adopted a simple approach by using the right truncation distance of the opportunistic sightings detection function as an estimate of the width of effort coverage. Each segment’s effort coverage was then created spatially using the ArcGIS v9.2 Buffer tool (ESRI 2007) with the right truncation distance of the detection function as the buffer width. We considered a grid cell to have been effectively covered if that segment’s buffer overlapped with that cell’s centroid. We applied Hawth’s tool ‘Enumerate Intersecting Feature’ (Beyer 2004) to count the segment buffers that covered each grid cell centroid as an estimate of effort for that cell. The number of gray whales sighted by on-effort segments was summed within each sampled grid cell and divided by that cell’s number of segments to derive an estimate of WPUE.

Calibration, merging and smoothing of the systematic and opportunistic surfaces

Estimates in each surface were right-skewed and log-transformed with a constant of 1 added to allow for the log transform of 0 and to reduce the influence of left-skewed outliers. We assessed comparability in surface estimates by evaluating the associated Pearson’s correlation coefficient across grid cells sampled at least twice in each surface. These cells were used in a linear regression of the 2 logged datasets assuming an intercept of zero. The logged opportunistic estimates plus constant i.e. \(\log(WPUE + 1)\), were multiplied by the regression coefficient (\(\beta\)) to calibrate them to the logged systematic estimates plus constant, i.e. \(\log(WPKM^2 + 1)\).

The 2 surfaces, now both in units of \(\log WPKM^2\), were merged by calculating a weighted mean log value for cells with effort in both data sets, with weight proportional to the cell’s effort. The log value from the data set with effort was retained in grid cells with no overlap between the surfaces.

Geostatistical kriging methods were used to estimate a smooth surface across areas with and without survey effort of the merged logged surface. Kriging is a robust and widely used statistical method that uses a variogram-based weighting algorithm to estimate values at missing locations from spatially correlated samples in the neighbourhood (Cressie 1993). We tested 3 kriging approaches (ordinary, simple, and co-kriging). Ordinary kriging assumes strict stationarity (constant spatial mean and variance) but makes no assumptions about the magnitude of the mean (Cressie 1993). Simple kriging assumes both strict stationarity and an *a priori* known spatial mean over the entire domain including the large offshore region with no survey effort. We assumed a zero mean to stabilize predictions in deeper data-void areas in which no gray whales were observed during aerial surveys in 2001–2003 (Meier et al. 2007). Finally, we assessed the utility of co-kriging that includes a predictive covariate (Matheron 1970). Gray whales have been observed mainly between 5 and 15 m depth in the Pilut feeding area (Vladimirov et al. 2008). We built a co-kriging model using 1 m bathymetry that was available over the spatial extent of the analysis.

We selected the best variogram structure for each kriging method from a set of candidate variogram functions (exponential, spherical, Gaussian, and Matern) using least squares selection criteria. The variogram function parameters were selected using an ordinary (unweighted) least square optimization function in R using starting values attained by fitting an empirical variogram to the merged logged surface (fit.variogram, gstat package; R Development Core Team 2012). Potential anisotropy of the spatial autocorrelation was investigated. The spatial distribution of the kriging standard deviation was plotted and visually assessed.
Feeding area boundary estimation

We determined the upper 95% of the cumulative frequency in the kriged surface and used the contour function in R (base package; R Development Core Team 2012) to create the contour line in the surface corresponding to this value.

RESULTS

Systematic WPKM² surface estimation

The majority of 2005–2007 systematic survey effort was after the 3rd week of June, with most effort provided by shore-based scans (Table 2). Only 6 systematic vessel surveys were conducted within the spatial extent of this analysis during June and July 2005–2007; 4 of these surveys were in 2006 (Table 3).

The estimated shore-based availability correction had a mean of 0.60 (SD = 0.038). The estimated availability correction for the systematic vessel surveys was 1.0, i.e. the slow vessel speed resulted in gray whales being available on the ocean surface for detection at least once during the time a given area of water was being searched.

The final detection function (N = 116) used a hazard rate model with no adjustments. The right truncation distance was 5.5 km. AICs indicated no improvements when covariates were added to the base model with distance (Table 4).

The mean probability of detection within the right truncation distance was 0.745 (SE = 0.044).

Values in the estimated WPKM² surface ranged from 0.00 to 0.99 whales km⁻² (Fig. 3).

Opportunistic WPUE surface estimation

One to 2 vessels were in the field during June–August 2005–2007, with most effort from mid-July through August (Table 5). Substantially more gray whales were seen during August than during June and July.

Table 4. Systematic vessel survey detection function candidate models with AICs. The base detection function that contained distance as a covariate was a hazard rate model with no adjustments and a right truncation distance of 5.5 km. Additional covariates were added singly to the base model. NumObs: number of observers

<table>
<thead>
<tr>
<th>Model</th>
<th>AIC</th>
</tr>
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<tr>
<td>Base model</td>
<td>1974.71</td>
</tr>
<tr>
<td>Group size</td>
<td>1976.73</td>
</tr>
<tr>
<td>Beaufort Scale</td>
<td>1976.73</td>
</tr>
<tr>
<td>Glare</td>
<td>1976.73</td>
</tr>
<tr>
<td>Visibility</td>
<td>1976.73</td>
</tr>
<tr>
<td>NumObs</td>
<td>1976.73</td>
</tr>
<tr>
<td>Year</td>
<td>1976.72</td>
</tr>
</tbody>
</table>
A total of 5051 on-effort segments, with a total length of 16,046.6 km, were designated for June−August 2005−2007 on the Sakhalin shelf. The mean segment length was 3.2 km (range 0.01 to 9.0 km; SD = 1.99) due to gaps in search effort along transects and rejection of segments having speed <5 km h\(^{-1}\). Mean segment speed was 15.6 km h\(^{-1}\) (range 5.0 to 39.6 km; SD = 5.06).

A total of 1170 sightings were associated with on-effort June−August 2005−2007 segments. The final detection function (N = 273) used a hazard rate model with covariates of visibility and distance. The right truncation distance was 4.2 km. The mean probability of detection within the right truncation distance was 0.603 (SE = 0.053).

June and July 2005−2007 vessel effort within the spatial extent of this analysis consisted of 1917 segments, with a total length of 5826.5 km. The mean segment length was 3.0 km (range 0.01 to 9.0 km; SD = 1.95). Mean segment speed was 15.3 km h\(^{-1}\) (range 5.1 to 39.4, SD = 5.3). A total of 479 sightings were associated with the 1917 effort segments. The estimated surface (Fig. 3) ranged in value from 0.00 to 2.00.

Calibration, merging and smoothing of the systematic and opportunistic surfaces

The spatial extents of the 2 surfaces overlapped, with the exception of north of shore-based Sta-
tion 3, where there was no opportunistic effort, and the coastal area south of shore-based Station 13 that lacked systematic effort. The opportunistic and systematic surfaces had 4533 and 2205 grid cells with effort, respectively, within the analysis spatial extent (Fig. 4). As expected, effort for a given cell differed between the 2 surfaces. Opportunistic surveys occurred mainly in the central part of the Piltun feeding area and within approximately 4 to 8 km offshore. The majority of shore-based systematic survey effort was concentrated within 5 km of shore, with most effort in the northern part of the feeding area where most observation stations were located.

The 2 surfaces were significantly correlated in cells with at least 2 units of effort in each surface \( (N = 1585, R = 0.35, p < 0.05 \text{ when zero values were included; } R = 0.26, p < 0.05 \text{ for only non-zero data}) \). The linear regression had \( R^2 = 0.47 (p < 0.05) \) and \( \beta = 0.92 (p < 0.05) \). The combined surface had values ranging from 0.00 to 0.46, with 4883 grid cells of effort (Figs. 3 & 4).

Simple kriging performed best of the 3 models to estimate missing values and smooth the combined surface (Fig. 5). The model created a stable extrapolation in data-void areas and estimated abundances that conformed well to raw data. Ordinary kriging produced reasonable estimates in areas with data but resulted in unreasonable extrapolations in data-void areas that were either too high in deep waters of the northeast area or negative in moderate water depths between 20 and 50 m. Co-kriging captured general trends in abundance but predicted poorly in both data-rich and data-poor areas. The cross-variogram indicated a weak negative correlation between abundance and depth, and the depth covariate produced a flat decrease of whale abundance with depth instead of capturing the narrow range of nearshore depths associated with the original abundance data.

An exponential covariance function was selected for fitting the variogram in all kriging models. The variogram values for the simple kriging model included the sill (0.0014), nugget (0.00086), and range (12 km). The low sill and nugget values were consistent with the low observed kriged values (Fig. 6). The kriging standard deviation across the prediction surface was relatively constant in areas with effort but increased in data-void regions located farther than the variogram range (12 km) from survey effort. We were unable to get a reasonable estimate of the anisotropy ellipse because the majority of whale observations were within 5 km of shore.

**Feeding area boundary estimation**

The estimated 95\% contour extended ~110 km along the Sakhalin Island coast and enclosed a continuous 607 km² nearshore area capturing the region of highest gray whale abundance in the Piltun feed-
ing ground (Fig. 6). The north-south portion of the boundary was an average of 5.4 km from shore (range 2.3 to 7.1 km) with a mean water depth of 24 m (range: 11 to 35 m). As expected, excluded cells were mainly located in deeper waters, or areas to the north or south of concentrations of gray whales observed to date in the feeding ground.

**DISCUSSION**

Identifying and delineating critical habitat for species of conservation concern is important when developing management plans and assessing and mitigating potential impacts of anthropogenic activities (Hauser et al. 2007, Williams et al. 2013). Critical habitat for baleen whales includes breeding areas, calving grounds, migration routes, and feeding areas (Gregr & Trites 2001, Wheeler et al. 2012). We used systematic and opportunistic survey effort to delineate the boundary of the Piltun nearshore feeding area. This boundary supported mitigation measures that aimed to prevent behavioural disturbance of feeding gray whales during a seismic survey conducted adjacent to their feeding ground. The 2 data sets provided complementary spatial coverage throughout the feeding area and its surroundings and improved estimation of the boundary. Systematic data provided little survey effort in deeper waters in the southern part of the feeding ground. Using only these data may have biased the estimated boundary shoreward, which could capture <95% of the feeding whales, thus exposing more feeding gray whales to sound levels sufficient to cause behavioural disturbance. Conversely, using only opportunistic surveys with most effort in deeper waters would place the feeding boundary too far offshore, possibly resulting in delays of the seismic survey to avoid disturbing gray whales in deeper waters that may have been transiting instead of feeding. Such delays could have conflicted with the primary mitigation of completing the survey as quickly as possible before many gray whales arrived on their feeding ground (Bröker et al. 2015).

Critical habitat demarcation can be difficult, particularly for mobile and cryptic species such as marine mammals that can also have large geographic ranges (Wheeler et al. 2012). It is crucial to incorporate spatial patterns of relative occurrence or density when identifying critical habitat (Wheeler et al. 2012, Williams et al. 2013). These patterns may be derived directly from survey data, predicted using spatiotemporal models, or based on expert opinion (Williams et al. 2013). Our study directly estimated a smoothed gray whale relative abundance surface from which we derived a 95% abundance contour as a proxy for the Piltun feeding area boundary. Other marine mammal studies have also used abundance contours estimated from a spatial surface to delineate core habitat and home ranges (e.g. Hobbs et al. 2005, Urian et al. 2009, Williams et al. 2013). Such studies
frequently use kernel density estimation (KDE) (Worton 1989) to create a smoothed surface from which the contour is derived. However, KDE results are considered sensitive to the smoothing parameter used (Worton 1989, Kie et al. 2010), and the smoothing is applied consistently in all directions (isotropy) (but see Amstrup et al. 2004). This approach assumes that an animal’s or species’ use of space extends uniformly in all directions (Amstrup et al. 2004), which is not supported for coastal marine mammal distributions. Rayment et al. (2009) addressed this issue by using univariate KDE on coastal Hector’s dolphin *Cephalorhynchus hectori* sightings, which were projected onto a line parallel to shore, whereas Amstrup et al. (2004) modelled a 2D kernel that estimated smoothing distances separately for the X and Y dimensions.

We used kriging (Cressie 1993) to smooth our coastal gray whale density surface because the spatial smoothing was estimated from the observations through a variogram that can incorporate directionality (anisotropy). Kriging models allow additional information such as covariates and assumed spatial means to be used for predictions. We were unsuccessful in modelling an anisotropic variogram to capture the long, narrow shape of the gray whale distribution because the empirical variogram in the east-west direction had no asymptote to indicate a km range at which data points became independent, and we were unable to estimate a reasonable range parameter in this direction. In addition, our data contained many observed zeros in offshore areas that confounded variogram estimation in the east-west direction. We therefore used an isotropic variogram model that smoothed equally in all directions. The estimated unidirectional smoothing amount was a compromise that likely was too great in the east-west direction and undersmoothed the north-south direction. The estimated contour was thus likely located farther offshore and was more protective of whale habitat compared to a contour from an estimated surface using anisotropic smoothing. Improvements to the kriging could be made by blocking the study area and fitting variograms separately to cohesive regions of similar stationarity and by including, if available, additional covariates associated with gray whale habitat use (e.g. locations with high prey biomass).

Separate detection functions and values to correct for availability (probability of detecting a gray whale at the surface) were estimated for each platform (vessel and shore-based) used in the systematic survey relative density surface. Borchers et al. (2013) suggested that the McLaren (1961) availability correction used in our estimates can be biased, with differing amounts of bias for the 2 platforms. Given the low number of systematic vessel surveys, it is unlikely that this source introduced substantial bias into the density surface. The availability correction for shore-based density estimates was a constant; thus, its bias would equally affect all density estimates in the surface. The shore-based detection function had limitations in that effects of environmental covariates were not tested due to low sample sizes. It is possible that detection at farther distances decreased as environmental conditions (e.g. visibility or Beaufort Scale) deteriorated, resulting in underestimation of densities. Inclusion of the greater opportunistic surface effort in deeper waters of the feeding area may have helped reduce this potential bias.

The rules for delineating effort segments for the opportunistic vessel survey tracks provided a nominal but reasonable method of identifying on-effort segments in the absence of formal observer declarations of going ‘on’ and ‘off’ effort. The use of daylight hours and environmental conditions conducive to observing gray whales likely minimized the number of ‘off’ effort segments with zero sightings that were mistakenly taken as ‘on’ effort and would negatively bias WPUE estimates. The opportunistic survey effort coverage was based on the modelled detection function right truncation distance. Gray whale detection during opportunistic surveys may have been affected by use of different vessels and different observers. However, these vessels had comparable observer platform height, and potential covariates affecting detection such as observer and environmental conditions were tested. As absolute abundance was not required, no attempt was made to estimate an availability correction value. This probability was assumed constant across the study area. This may not be the case if, for example, whales surface at different rates at different depths.

The estimated boundary corresponded well to historically observed whale use in the feeding ground and captured regions of highest abundance. The boundary was determined specifically for the June to July time period when gray whales are migrating into the feeding area and abundance is relatively low compared to later in the season. Use of the boundary should therefore be limited to the same June to July time frame for which the boundary was developed. However, our methods can be applied to re-estimate the boundary for a different time period or used as a framework to delineate important habitat for other species.
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LITERATURE CITED


McLaren IA (1961) Methods for determining the numbers and availability of ringed seals in the eastern Canadian Arctic. Arctic 14:162–175


Wheeler B, Gilbert M, Rowe S (2012) Definition of critical summer and fall habitat for bowhead whales in the eastern Canadian Arctic. Endang Species Res 17:1–16


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