1. INTRODUCTION

Sympatric species, or populations within a species, that have similar prey preferences may compete for resources. Overlap in range does not always result in direct interaction due to the potential for spatiotemporal resource partitioning through variation in the timing and location of foraging behavior (Ratcliffe et al. 2014, Oviedo et al. 2018, Genov et al. 2019). This may lead to differences in habitat use among populations, resulting in different exposure to human activities (Genov et al. 2019, Teixeira et al. 2021).

In odontocetes, dietary differentiation is one of the main ecological processes that result in divergence in habitat use (Bearzi 2005), but social structure is also an important factor affecting long-term habitat use patterns in long-lived species such as killer whales Orcinus orca and sperm whales Physeter macrocephalus (Brent et al. 2015, Eguiguren et al. 2019). Understanding cetacean habitat use patterns is difficult due to their highly mobile nature and large home ranges. In many areas, visual monitoring is limited and cost prohibitive due to protracted periods of inclement weather, remote access, and short daylight hours. In these locations, remote acoustic monitoring...
is an effective alternative for vocally active species such as killer whales which produce discrete pulsed calls, whistles, and echolocations clicks (Schevill & Watkins 1966, Hanson et al. 2013, Rice et al. 2017).

Three sympatric yet distinct killer whale ecotypes occur in the coastal waters of Washington State, USA, that differ in diet, ecology, behavior, and morphology (Ford et al. 2000, Emmons et al. 2019a). Bigg’s killer whales forage on marine mammals, while offshore and resident killer whales are piscivorous (Ford et al. 1998, Dahlheim et al. 2008). Each ecotype can be identified acoustically by the group-specific pulsed calls they produce (Yurk et al. 2010, Riera et al. 2019). Within the resident ecotype, specific calls distinguish smaller nested social units, including populations, acoustic clans, or pods (Ford 1987).

Two populations of fish-eating or ‘resident’ killer whales, Northern Residents (NRKW) and Southern Residents (SRKW), have been well studied in their summer range of the inland waters of British Columbia and Washington State. NRKW are comprised of 3 acoustic clans (A, G, and R clans) of approximately 300 individuals (Towers et al. 2020), while SRKW are a single acoustic clan (J clan) of approximately 75 individuals (Center for Whale Research unpubl. data). NRKW and SRKW are both salmon specialists with a strong preference for Chinook salmon Oncorhyncus tshawytscha (Ford & Ellis 2006). The 2 populations are partially sympatric, but socially distinct from each other (Ford et al. 1998, 2000).

Population decline, small population size, and the existence of several anthropogenic threats led to SRKW being listed as ‘endangered’ under both the US Endangered Species Act (ESA) and Canada’s Species at Risk Act (SARA), while NRKW are listed as ‘threatened’ under SARA. Since these listings, NRKW have continued to grow at 2.9% yr⁻¹ while SRKW have failed to recover (Fisheries and Oceans Canada 2018, Towers et al. 2020). Critical habitat was designated for SRKW in the inland waters of Washington State in 2006 (NMFS 2006), but the whales typically only spend a fraction of the year in these waters (Olson et al. 2018). Passive acoustic monitoring (PAM) has been used at selected sites along the west coasts of Washington, Oregon, and California to better understand SRKW movements outside of the inland waters, especially during winter months (Hanson et al. 2013, Rice et al. 2017), and has led in part to the proposed expansion of critical habitat to include these waters (NMFS 2019).

In 2014, PAM efforts were increased to 15 monitoring sites along the Washington coast to better understand year-round SRKW movements and potential anthropogenic threats (Emmons et al. 2019b). Resident killer whale movements have been hypothesized to be driven by the abundance of Chinook salmon (Olson et al. 2018), and the continental shelf off the Washington coast is a part of the ocean distribution of a number of important Chinook salmon runs (Weitkamp 2010, Shelton et al. 2019). This part of the coast is bordered by the Strait of Juan de Fuca to the north and the Columbia River to the south, and the associated Juan de Fuca eddy and Columbia River plume play an important role in oceanographic and biological processes (Hickey & Banas 2003). The Strait of Juan de Fuca acts as a funnel for salmon returning to their natal rivers such as the Fraser River, the primary source of SRKW’s preferred summer prey (Hanson et al. 2010). The Columbia River supports large runs of spring, summer, and fall Chinook salmon, and previous analyses have suggested that spring runs are seasonally important to SRKW (Zamon et al. 2007, Ayres et al. 2012, Hanson et al. 2013).

NRKW also occur along the Washington coast (Rice et al. 2017), but to what extent is uncertain. NRKW and SRKW are ecologically similar, especially in regard to their diet (Ward et al. 2013). Thus, it is important to understand the role that intraspecies competition may play in limiting the recovery of SRKW (Chasco et al. 2017). Here, we used a large network of passive acoustic recorders to assess the habitat use patterns of both NRKW and SRKW along the Washington coast. Specifically, we compared monthly occurrence of both populations at recorder locations grouped by their proximity to the Strait of Juan de Fuca to the north and the Columbia River to the south in one analysis and by their distance from shore in a second analysis. Based on previous studies (Yurk et al. 2010, Rice et al. 2017, Riera et al. 2019), we hypothesized that we would observe spatial and/or temporal habitat partitioning by SRKW and NRKW in order to minimize competitive interactions for Chinook salmon.

2. MATERIALS AND METHODS

2.1. Field efforts

PAM was conducted at 15 sites off the Washington coast using ecological acoustic recorders (EARs; Lammers et al. 2008). These sites spanned the width of the continental shelf from the Canadian border to the north and the Columbia River to the south.
Sites were selected based on previous sightings and satellite tag locations of SRKW, bathymetric features, accessibility for mooring deployment and recovery, and minimizing the likelihood of fishery interactions.

An EAR can be programmed either as an event recorder triggered by defined acoustic parameters or to record on a dedicated duty cycle. In this study, EARs were programmed to record on a 15% duty cycle, resulting in 90 s of continuous recording every 600 s. The duty cycle was chosen based on several factors, including the likelihood of capturing group-specific killer whale pulsed calls, expected power consumption, and drive space that would allow for data collection for up to 1 yr. The sampling rate used on all deployments was 25 kHz, which provided approximately 12.5 kHz of bandwidth. This sampling rate was chosen as a trade-off for preserving disk drive space and battery life while still allowing for identification of killer whale pulsed calls. These calls typically range in frequency between 1 and 10 kHz, but some calls contain a high frequency component and contain harmonics up to 30 kHz (Ford 1989, Miller 2002).

2.2. Data analysis

All 90 s recordings were sorted by day and then concatenated and converted into .wav files using custom Matlab v.2014b (The MathWorks) scripts. Daily files were inspected visually and aurally for cetacean sounds by visualizing the entire file using TRITON (Wiggins & Hildebrand 2007). The daily occurrence of all species detected was summarized. Those files containing killer whale sounds were further reviewed, and discrete calls were compared to a catalog of pod- and population-specific pulsed calls to determine the killer whale ecotype, population, clan (in the case of NRKW), and pod (if possible, in the case of SRKW) (Ford 1987). Killer whale detections of insufficient quality to determine ecotype and population were not included in this analysis.

For all resident killer whale detections, the types of sounds (clicks, calls, buzzes) in each 90 s recording were identified, and these recordings were organized into acoustic encounters using methods described by Riera et al. (2013). Encounters were defined as periods in which sounds produced by the same group of whales could be detected over multiple sequential recordings separated by less than 3 h, and the duration of these encounters was also determined in decimal hours. Monthly median encounter length was calculated for each population. We compared the effect of resident killer whale population on encounter length using a Mann Whitney U-test.

Rates of monthly occurrence were calculated by dividing the number of days in a given month by the total number of days of effort in that month. To investigate spatial differences in monthly occurrence,
recorder sites were first grouped into north and south groups. These groups were determined by proximity to the Straits of Juan de Fuca to the north and the Columbia River to the south. The northern sites were Juan de Fuca, Cape Flattery mid-shelf, Cape Flattery offshore, Cape Flattery deep, Sandpoint, and La Push. The southern sites were Quinault deep, Quinault mid-shelf, Quinault inshore, Westport, Westport mid-shelf, Westport deep, Willapa, Columbia River, and Columbia River south. Secondly, sites were grouped by distance from shore: inshore, mid-shelf, and offshore. The inshore sites were Juan de Fuca, Sandpoint, La Push, Quinault inshore, Westport, Willapa, and Columbia River. The mid-shelf sites were Cape Flattery mid-shelf, Quinault mid-shelf, Westport mid-shelf, and Columbia River south. The offshore sites were Cape Flattery offshore, Cape Flattery deep, Quinault deep, and Westport deep.

3. RESULTS

Between 2014 and 2017, the number of days monitored by location ranged from 69 d at Quinault mid-shelf to 1149 d at Columbia River North (Fig. 2). Resident killer whales were detected acoustically 429 times, with detections at all sites except for Quinault Deep. Total duration was 490.9 h for the 263 SRKW encounters, and 441.3 h for the 166 NRKW encounters. SRKW were detected at 10 of the 15 sites, while NRKW were detected at 11 of the 15 sites. SRKW were detected in every month of the year but August (Fig. 3B), and 66.2% of these encounters were between February and May (n = 174). NRKW were detected in every month of the year with 2 peaks in occurrence (Fig. 3B): one peak was between February and April with 42.8% of all NRKW encounters (n = 71) and the second between August and October with 37.3% of all NRKW encounters (n = 62). All 3 SRKW pods were detected at the northern sites, while only 2 (K and L pods) were detected at the southern sites. The majority of NRKW encounters were of G clan, but all 3 clans were detected. The median SRKW encounter length was 1.33 h and ranged from

Fig. 2. Monthly and yearly effort for each acoustic recorder site.
The median NRKW encounter length was 1.92 h and ranged from 0.17–10.33 h. The proportion of encounters longer than 2 h was lower for NRKW (33.5%) than for SRKW (50%) encounters. Encounter lengths for NRKW (median = 1.92) were longer than those for SRKW (median = 1.33). A Mann-Whitney test indicated that this difference was statistically significant ($U = 26027.5$, $p < 0.001$).

The monthly median encounter duration for SRKW ranged from 0.7–2.2 h (Fig. 3). It was between 1 and 2 h for all months but June (median = 2.2 h) and September (median = 0.7 h). For NRKW, monthly median encounter duration ranged from 1.2–3.5 h (Fig. 3A) and was greater than 2 h in May and July–December. Median encounter duration was longer for NRKW than SRKW in all months but February and June.

The majority of resident killer whale (NRKW and SRKW) encounters (63%) were at the northern sites. Less than half of the SRKW encounters (45%) were at the northern sites and 55% were at the southern sites (Fig. 4B). In contrast, 93% of all NRKW encounters were at the northern sites (Fig. 4B). The occurrence of encounters peaked in the fall at northern sites and in the spring at southern sites (Fig. 4A).

The inshore sites accounted for a little over half of all resident killer whale (NRKW and SRKW) encounters (53%). The majority (83%) of SRKW encounters were at these inshore sites, while only 5% of NRKW were at these sites (Fig. 5B). The mid-shelf sites accounted for 16% of resident encounters: 14% of SRKW and 18% of NRKW encounters (Fig. 5B). The offshore sites accounted for 31% of all resident encounters. Only 3% of SRKW encounters were at these sites, but 76% of NRKW encounters were at these offshore sites (Fig. 5B). The few SRKW encounters at the offshore sites were in the spring months. The spring and fall peaks in NRKW encounters occurred at the mid-shelf and offshore sites.

4. DISCUSSION

Passive acoustic monitoring was used to describe the temporal and spatial occurrence of resident killer whales off the Washington coast. SRKW and NRKW were both detected throughout the year with detection peaks in the spring and fall, seasons in which there are very few vessel surveys due to challenging weather conditions, further illustrating the importance of PAM for year-round monitoring of these populations.
Interpretation of PAM results is enhanced by an understanding of factors that may impact the detectability of resident killer whales, such as detection range and vocal behavior. The detection range can be influenced by both natural and anthropogenic sounds that may mask killer whale vocalizations. Our ability to detect killer whales may be impacted by seasonal changes in wind speed and precipitation (Lemon et al. 1984, Robinson et al. 2019). Riera et al. 2019 modeled the detection range for resident killer whales at Swiftsure Bank, to the north of the study area, for 4 months of the year: January, June, July, and September. The monthly mean detection distance varied seasonally from 5 km in January to 6.5 km in September, with intermediate values in June and July. These results indicate that the detection range did not vary greatly throughout the year, but decreased slightly in the winter.

Sound associated with large commercial vessels, fishing vessels, and small personal craft can also reduce the detection range, and the intensity of these activities varies across the continental shelf. In this study, the detection rates of resident killer whales did not vary greatly between inshore sites (4.8% of days) and offshore sites (5.9% of days), but we acknowledge the uncertainty that seasonal and temporal changes in detection range may play in interpreting our results.

Resident killer whales have been shown to vocalize more while foraging and socializing than when traveling and resting (Ford 1989, Holt et al. 2013). However, SRKW have been documented to vocalize less than 44% of the time in these coastal waters in the winter (Hanson et al. 2018). This means that the results of this study represent the minimum occurrence of these 2 resident killer whale populations. Encounter duration can provide further information on habitat use; longer durations indicate that the whales are using the area for a range of behaviors, including foraging and socializing, rather than just transiting through it (Riera et al. 2019).

Previous studies using PAM and visual surveys have described limited use of the Washington coast by NRKW (Ford et al. 2017, Rice et al. 2017), although in the present study they accounted for over one-third of all resident killer whale encounters. The majority of these encounters were at northern sites and were longer than those of SRKW, indicating more extensive use of the Washington coast than previously reported.

With the overlapping ranges and similar prey preferences of these 2 resident killer whale populations, it is important to understand the role that competition may play in limiting the recovery of SRKW (Chasco et al. 2017). Competition for prey resources may result in spatial or temporal habitat segregation (Johnson & Crowley 1980, Martin & de Silva 2004). In
addition to temporal habitat partitioning reported in previous studies (Yurk et al. 2010, Rice et al. 2017, Riera et al. 2019), the data presented here indicate that spatial separation of the 2 resident killer whale populations is also occurring along the Washington coast. Both NRKW and SRKW exhibited spring and fall peaks in occurrence, but the spring peak was driven by an increase in SRKW encounters at the southern sites, while the fall peak was driven by an increase in NRKW encounters at northern sites. The 2 populations were also segregated across the shelf, with SRKW occurring primarily at inshore sites and NRKW at offshore sites. These findings were made possible by monitoring an expanded area using a large network of recorders across diverse sites as opposed to one or 2 sites as in previous studies.

The spring and fall peaks in occurrence for both populations suggest that both may be targeting similar prey. Previous studies in the inland waters have shown that both NRKW and SRKW exhibit a strong preference for Chinook salmon (Ford et al. 1998, Ford & Ellis 2006, Hanson et al. 2010, Ford et al. 2016), but there is limited diet data available for the coastal waters. The spring peak in occurrence aligns with previous sightings of SRKW near the Columbia River (Zamon et al. 2007, Hanson et al. 2018) and the return of spring Chinook salmon to the Columbia River, the largest run in the region at this time of year (Hanson et al. 2013). Diet data collected along the outer coast have been limited to spring months and indicate that Chinook salmon are an important component of SRKW throughout the year (Hanson et al. 2021). While fall coastal diet data are lacking, Chinook salmon typically return to the Fraser River from July–October (Groot & Margolis 1991). However, care must be taken when correlating killer whale and salmon movements. For example, observations of NRKW predation on sablefish near the shelf break mainly in the month of September (Ford et al. 2017) suggest that killer whale movements may not be entirely dependent upon the distribution of Chinook salmon.

In conclusion, we observed both spatial and temporal habitat partitioning by NRKW and SRKW along the coast of Washington State. Our data show that populations with similar prey preferences can parti-
tation to reduce competitive interactions. Recognizing these differences in habitat use is critical, since the whales’ exposure to anthropogenic activities may differ as a result. For example, sounds associated with military activities were most frequently detected at the offshore sites, while sounds associated with fishing were most frequently detected at the mid-shelf sites (Emmons et al. 2019b). This study provides further information about the habitat use of these resident killer whale populations with implications for their management and conservation in proposed critical areas.

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