

Monthly variation in the diet of harbour seals in inshore waters along the southeast Shetland (UK) coastline

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ABSTRACT: We studied the diets of harbour seals *Phoca vitulina* along the southeast Shetland (UK) coastline by analysing prey remains found in faeces (N = 733) at haul-out sites. A total of 44 325 fish otoliths were recovered. Sandeel (Ammodytidae) otoliths were the most numerous (38 704), followed by Gadidae (4707). Otoliths were measured and experimentally derived digestion coefficients were applied (correcting for digestion in the seal's gut) to estimate the size of ingested prey fishes. Gadids accounted for an estimated 53.4% of the annual diet by weight, sandeels 28.5% and pelagic fishes 13.8%. The dominant gadid fishes were whiting *Merlangius merlangus* (25.3%) and saithe *Pollachius virens* (11.1%). The range of species observed in the diet was similar to that recorded in other areas of the UK. One exception to this was garfish *Belone belone* accounting for 34.1% of the diet in September (1996), which is a species not previously reported for harbour seal diets in UK waters. There were strong seasonal patterns in the contribution of sandeels and gadids, with sandeels being important in spring and early summer, and gadids in winter. Pelagic species (mainly herring *Clupea harengus*, garfish and mackerel *Scomber scombrus*) were important in late summer and autumn. Observed seasonal patterns are similar to those previously recorded for harbour seal diets in the Moray Firth area of Scotland and appear to coincide with changes in prey availability. A comparison of the utility of using only otoliths to estimate seal diet with all identifiable structures showed that using otoliths alone underestimated the contribution of pelagic fish and overestimated the importance of gadids and sandeels.

KEY WORDS: Harbour seal · *Phoca vitulina* · Otoliths · Shetland · Diet · Seasonal variation

INTRODUCTION

Harbour seals or common seals *Phoca vitulina* have been extensively studied in UK waters. Many of these studies, particularly those on diet and feeding ecology, have focused on interactions with fisheries. Diets of harbour seals in UK waters have been described in Norfolk (Sergeant 1951), the northeast coast of Scotland (Rae 1960, 1968, 1973), the Moray Firth (Pierce et al. 1990, Pierce et al. 1991a, b, Tollit & Thompson 1996), Orkney (Pierce et al. 1990) and the west coast of Scotland (Boyle 1990). Only in the Moray Firth has seasonal and interannual variation in diet been described in detail. Changes in diet are generally assumed to

relate to changes in prey availability, as also suggested by Härkönen (1987) for this species in the Kattegat and Skagerrak.

The only previous work on harbour seal diets around the Shetland Islands, UK (60° N, 1 to 2° W; Fig. 1), documented the diet of harbour seals on the island of Mousa during the third quarter of 1994 (Brown & Pierce 1997). This work indicated that harbour seals preyed on a wide range of species, including most of the target species in local fisheries. There are around 6200 harbour seals present in the waters around the Shetland Islands (Hiby et al. 1996), representing over 20% of the UK population. The minimum population in UK waters is around 28 720 seals (Hiby et al. 1996). Numbers have increased substantially since protective legislation was introduced in the early 1970s.

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Shetland is thus an appropriate site for a new study of seasonal variation in diet of harbour seals. Such a study is also of interest to improve knowledge of interactions between harbour seals and commercial fisheries around Shetland. The southeast coast of Shetland was chosen as the focus of the present study because haul-out sites along this part of the coast are used by harbour seals all year round and are readily accessible for sampling. This stretch of coast accounts for approximately 20% of the total Shetland population of harbour seals (Duck et al. 1993).

Competition between seals and fisheries is particularly topical at a time when the total allowable catches for many commercial fish species are being reduced, and is a subject that has been extensively reviewed in the UK and elsewhere (e.g. Gulland 1987, Harwood 1987, 1992, des Clers & Prime 1996). The Shetland Islands are situated in the centre of some of the most productive fishing grounds in the North Sea. Fishing is one of the main traditional industries in the Islands, employing approximately 1870 persons in catching, processing and ancillary industries (Anon 1996a). Over 135 000 t of fish were landed into Shetland during 1995, worth £28.3 million at first sale (Anon 1996b).

In UK waters, grey seals *Halichoerus grypus* are generally perceived as a greater threat to fisheries, partly because of their greater numbers (an estimated all-age UK population of 108 500 seals in 1994; Hiby et al. 1996). However, in Shetland, harbour seals are more numerous than grey seals. It is estimated that there are around 3500 grey seals associated with breeding sites around the Shetland Islands, a figure which has remained relatively static since the 1970s (Brown 1995). Telemetry studies indicate that harbour seals generally forage in inshore waters (Thompson & Miller 1990, Thompson 1993). Given the value of inshore grounds to the local fishing fleet in Shetland, harbour seals may have greater direct impact on local fisheries than grey seals, which may travel considerable distances out to sea to forage (McConnell et al. 1992, SMRU 1994).

In recent years identifying and measuring fish otoliths recovered from faeces has been the main method of assessing the diets of seals from around the British Isles. Using this method to determine the diets of predators is subject to several important sources of error. Otoliths are eroded as they pass through the gut of seals (da Silva & Neilson 1985, Dellinger & Trillmich 1988, Harvey 1989), which can result in underestimation of the size of prey fish consumed (da Silva & Neilson 1985, Jobling 1987, Harvey & Antonelis 1994). Captive feeding experiments have been used to estimate 'digestion coefficients' to account for size reduction in otoliths (e.g. Tollit et al. 1997). Other problems with the methodology are that some species may be

completely absent or under-represented when the diet is assessed from otoliths alone, such as those with no otoliths (e.g. Rajidae), with small or fragile otoliths, or the heads of which are discarded by seals during feeding (Boyle et al. 1990, Pierce et al. 1991a). Nevertheless, analysis of faeces probably represents the single best method available to assess the diet of seals around the British Isles (Prime & Hammond 1990).

The aims of this study were: (1) to examine monthly variation in harbour seal diets along the southeast coastline of Shetland; (2) to relate changes in diet composition to known changes in prey availability; (3) to assess the utility of using other hard remains in addition to otoliths to detect prey species present in the samples and to improve quantitative estimates of diet composition; and (4) to identify potential competition between seals and local fisheries.

METHODS

Sample collection. Regular visits were made to 8 harbour seal haul-outs along the southeast Shetland coastline (Fig. 1) over the period 1 May 1995 to 30 April 1996. The numbers and species of seals at haul-out sites were noted before collection began. In some instances grey seals were present at sites but spatial separation between the species was the norm with grey seals preferring to lie close to the water's edge or on tidal rocks offshore. Areas which grey seals used were avoided when searching for faecal material and the numbers of grey seals that were actually intermixed with the harbour seals never exceeded 10% of the total number of all seals present. Each faecal sample was collected into a separate, light duty polythene bag and frozen at -20°C until further processing.

Processing and identification of prey remains. Faecal samples were washed through a series of sieves (2.00 and 0.355 mm); all hard parts (fish otoliths, cephalopod beaks and fish bones) were extracted from the sieves, and stored in alcohol until further processing. Samples were later sorted, otoliths were dried and stored in small glass vials, and cephalopod beaks and fish bones were stored in Industrial Methylated Spirits.

Otoliths were identified to the lowest possible taxon, usually to species, using a reference collection of local fish and a guidebook (Härkönen 1986). It was not always possible to identify them to species due to morphological similarities, particularly for small or badly digested otoliths. Thus haddock *Melanogrammus aeglefinus*, saithe *Pollachius virens* and pollack *Pollachius pollachius* could not always be distinguished from each other; similarly Norway pout *Trisopterus esmarki* and poor cod *Trisopterus minutus* were sometimes recorded as *Trisopterus* spp. It was not possible

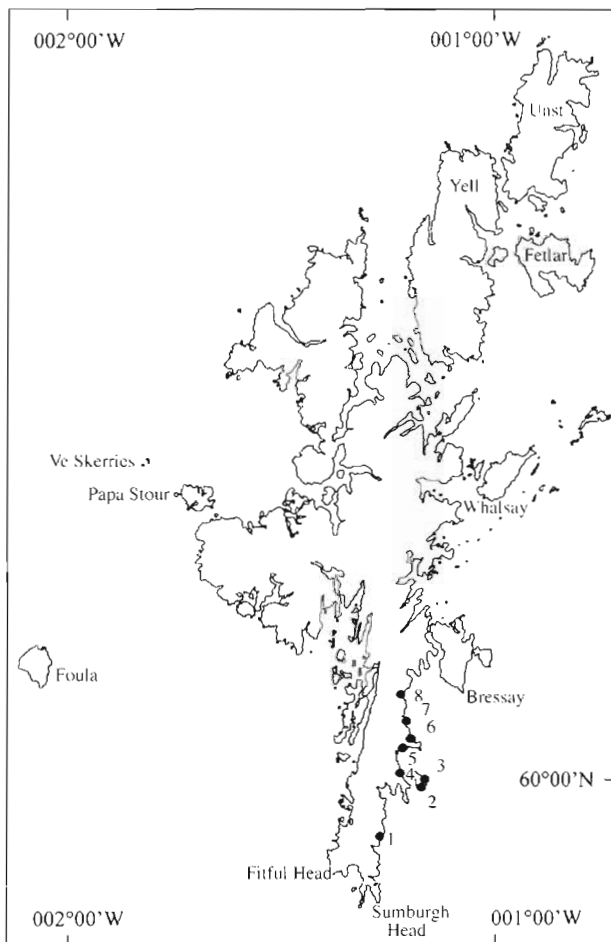


Fig. 1. Shetland Islands; locations of collection sites are indicated on map: (1) Trosswick Ness; (2) Mousa West Pool; (3) Mousa East Pool; (4) Leebiton; (5) Aiths Voe; (6) Aiths Wick; (7) Orkraquoy; (8) Quarff

to distinguish between the 5 species of sandeels found in Shetland waters and all sandeel otoliths were therefore recorded as sandeel (*Ammodytidae*). Cephalopod beaks were identified by M. B. Santos (University of Aberdeen).

Estimating prey size. To estimate original prey sizes, otoliths and beaks were measured using a binocular dissecting microscope fitted with an eyepiece graticule. Normally otoliths were measured lengthways. However, for species of fish such as herring *Clupea harengus* and whiting *Merlangius merlangus* (whose otoliths were almost always broken lengthways) otolith width was measured.

Experimentally derived digestion coefficients (Tollit et al. 1997) were applied to measurements on otoliths to account for digestive erosion. Tollit et al. (1997) distinguished a series of grades of digestion, based on changes in morphological features of the otoliths, as assessed by visual examination. Grade-specific correc-

tion factors are available for cod *Gadus morhua*, whiting and sandeels. We used the correction factor that best matched the degree to which our otoliths were eroded, e.g. for cod otoliths we selected the 'low' correction factor (1.07), for sandeels we selected the 'low' correction factor (1.16) and for whiting we selected the 'medium' correction factor (1.365). Tollit et al. (1997) also calculated species-specific coefficients for lemon sole *Microstomus kitt* and herring. For all other fish species, Tollit et al.'s average coefficients of 1.25 for otolith length and 1.24 for otolith width were used. All digestion coefficients used are given in Appendix 1.

Fish lengths and weights were estimated using regressions given in Appendix 1. Where possible, data for otolith size-fish size relationships are collected from fish caught in Shetland waters (E. G. Brown unpubl. data). Each otolith was assumed to represent 0.5 fish. Cephalopod weights were estimated using regressions from a published guide (Clarke 1986) and from unpublished data (G. J. Pierce & M. B. Santos). The total weight represented by each prey category was thus derived for each month and expressed as a percentage of the total for all categories. For each month, the mean number of otoliths per faeces (excluding the samples not containing any otoliths) was also calculated for the main prey species and groups.

Analysis of otolith size. For the purpose of carrying out a statistical comparison on the sizes of fish consumed we compared otolith size, since both the weight and length of the fish were estimated from the otolith. We feel that comparing the size of the otoliths directly is a more reliable method than comparing estimated fish lengths, as these are subject to additional error. Only for cod were fish lengths compared (several measurements on cod otoliths were made on broken otoliths and width had to be measured, in effect reducing the sample size available for comparison). For each of the main prey categories (sandeel, cod, haddock, herring, saithe, ling *Molva molva*, herring, Norway pout, poor cod, whiting), data were grouped by season (quarters of the year). Data for each quarter and the whole year were tested for normality using the Anderson-Darling test (Minitab software). For all species there were significant departures from normality, due to skewed or polymodal distributions. Comparisons between quarters were therefore made using the non-parametric Kruskal-Wallis 1-way analysis of variance. Post-hoc comparisons between each pair of seasons were made using the Mann-Whitney *U*-test. Since, for each species, there were 6 possible comparisons between seasons, a Bonferroni correction was applied to the probability level accepted for significance (we thus used $p < 0.0083$).

For sandeel, whiting and ling there were sufficient data to analyse monthly differences. The same proce-

dures were used, although with 66 possible inter-monthly comparisons, $p < 0.00076$ is required for significance.

Analysis of numbers of otoliths per faex. For each of the main prey categories (sandeel, cod, haddock, herring, garfish *Belone belone*, saithe, ling, *Trisopterus*, whiting, all gadid species and for all species combined), data on numbers of otoliths per faex (excluding faeces containing no otoliths of any species) were grouped by season (quarters of the year). Data for each quarter and the whole year were tested for normality using the Anderson-Darling test (Minitab software). For all species there were significant departures from normality, due to highly right-skewed distributions (zero otoliths was the most frequent class). Comparisons between months were therefore made using the non-parametric Kruskal-Wallis 1-way analysis of variance. Post-hoc comparisons between each pair of seasons were made using the Mann-Whitney *U*-test, with a Bonferroni correction as in the analysis of otolith size.

Errors associated with basing diet estimates on otoliths. To assess whether our use of otoliths to determine the diet of harbour seals was valid, we also identified other fish bones present in the faecal samples, using an extensive reference collection at the University of Aberdeen and a guide (Watt et al. 1997). Pierce et al. (1991a) demonstrated the usefulness of certain diagnostic fish bones in identifying fish prey consumed when otoliths were absent. Due to the similarity of the bones throughout the Gadidae, we could usually only classify to the family level. Fish bones were used in identification of herring (otic bullae, vertebrae and maxillae), Gadidae (pre-maxillae, maxillae, vertebrae, vomer and post temporals), Ammodytidae (atlas vertebrae and caudal vertebrae), flatfishes (vertebrae, pre-maxillae, maxillae and urohyals), garfish (vertebrae and pelvic girdle) and mackerel *Scomber scombrus* (vertebrae). For the terminology of fish bones see Watt et al. (1997). The utility of using otoliths alone was assessed by comparing the frequency of occurrence with which different prey categories were identified using (1) only otoliths and (2) all identifiable structures. Comparisons were made on raw frequencies using χ^2 tests.

Comparison of diet and fisheries. Data and information on fisheries around the Shetland Islands were obtained from the Fisheries Research Services Marine Laboratory at Aberdeen and the Shetland office of the Scottish Fisheries Protection Agency.

RESULTS

Collection of samples

A total of 733 faecal samples was recovered from 127 visits to harbour seal haul-outs at Troswick Ness, Leebitton, Mousa West Pool, Mousa East Pool, Aiths Voe, Aiths Wick, Okraquoy and Quarff (Fig. 1). The percentage of sampling trips that were successful (≥ 1 faeces collected), and the number of samples collected, were highest in August and September, coinciding with the annual moult. Fewest samples were collected in October, November and February (Table 1). Overall, 628 faecal samples (86%) were found to contain fish otoliths. The percentage of samples containing otoliths varied between months (Table 1).

Numbers of otoliths recovered

A total of 44 325 fish otoliths was recovered from the faecal samples. Table 2 gives the numbers of otoliths recovered by month for the main categories of fish prey. Sandeel otoliths were numerically the most abundant with 38 704 otoliths recovered (87.3% of the total number found), followed by Gadidae with 4707 otoliths recovered (10.6%), including *Trisopterus* spp. with 2793 otoliths (6.3%) (Table 2). Flatfish otoliths were only found occasionally with a total of 33 recovered, of which lemon sole accounted for 16.

The numbers of otoliths recovered per month ranged from 142 to 13 137. The number of sandeel otoliths recovered by month ranged from 19 to 10 985, while the number of Gadidae otoliths recovered ranged from 19 to 1703.

Table 1. Numbers of site visits per month with % of visits that were successful in collecting 1 or more harbour seal *Phoca vitulina* scats and numbers of scats recovered per month, with % containing fish otoliths

	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Totals
No. of site visits	14	6	17	7	15	10	6	8	5	14	16	9	127
% successful	64 %	83 %	76 %	71 %	60 %	80 %	83 %	100 %	100 %	71 %	50 %	78 %	72 %
No. of scats found	49	9	75	26	29	55	47	208	145	38	26	26	733
No. with otoliths	47	9	64	17	28	49	44	188	111	30	17	24	628
% with otoliths	96 %	100 %	86 %	65 %	97 %	89 %	94 %	90 %	77 %	79 %	65 %	92 %	86 %

Number of otoliths per faex

The number of otoliths per faex for all species was generally lowest during autumn and winter and highest in spring (Table 2). Overall variation in the number of otoliths (all species combined) per faex in relation to season was significant (Kruskal-Wallis; $H = 33.77$, $p < 0.0005$). Otoliths were more numerous in quarter 2 (Apr–Jun) (79.9 to 142 per faex) than in quarters 3 (Jul–Sep) (46.1 to 69.9) ($p < 0.00005$) and 4 (Oct–Dec) (15.6 to 85.9) ($p = 0.0003$). Otoliths were also more numerous in quarter 1 (Jan–Mar) (15.8 to 133) than in quarter 3 ($p = 0.0081$). Gadid otoliths were most common during winter and least common during spring and early summer (Table 2). Overall variation in the number of gadid otoliths per faex in relation to season was significant ($H = 38.35$, $p < 0.0005$). Gadids were more numerous in quarters 1, 3 and 4 than in quarter 2 ($p < 0.00005$ in all cases). The number of whiting otoliths per faex was highest during the last quarter (1.82 to 11.8), and was lowest during the second quarter (0.18 to 0.76). Overall variation in the number of

whiting otoliths per faex relation to season was significant ($H = 44.91$, $p < 0.0005$). Whiting were more numerous in quarter 4 than in quarters 1 ($p = 0.0082$), 2 ($p < 0.00005$) and 3 ($p = 0.0037$). Whiting were also more numerous in quarters 1 and 3 than in quarter 2 ($p = 0.0027$ and $p = 0.0005$ respectively). Saithe was most common during November and January (2.83 and 2.77 otoliths per faex respectively) and infrequently recorded during the summer months (0.02 to 0.95). *Trisopterus* spp. was most common during December to January (15.8 to 5.06). Overall variation in the number of *Trisopterus* otoliths per faex in relation to season was significant ($H = 40.45$, $p < 0.0005$). *Trisopterus* were more numerous in quarter 1 than in quarters 2, 3 and 4 ($p < 0.00005$, $p = 0.0073$ and $p = 0.0005$ respectively), and more numerous in quarter 3 than in quarter 2 ($p < 0.00005$). Numbers of otoliths for pelagic fish species were highest during summer (1.32 to 2.39) and lowest during winter (0.11 to 0.71). Herring was most common from June to August (1.32 to 2.37) and lowest during winter (0.0 to 0.12). Overall variation in the number of herring otoliths per faex in relation to sea-

Table 2. Monthly variation in the numbers of fish otoliths recovered from harbour seal *Phoca vitulina* faeces, for the main prey species and group totals. Mean number of otoliths per faex for the main prey species and groups are given in parentheses

Species	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Species totals	% of total
Cod	1 (0.02)	0 (0)	8 (0.13)	1 (0.06)	0 (0)	1 (0.02)	2 (0.05)	33 (0.18)	3 (0.03)	2 (0.07)	3 (0.18)	1 (0.04)	55 (0.09)	0.12
Haddock	1 (0.02)	1 (11.1)	5 (0.08)	1 (0.06)	1 (0.04)	0 (0)	1 (0.02)	12 (0.06)	4 (0.04)	0 (0)	0 (0)	5 (0.21)	31 (0.05)	0.07
Saithe	130 (2.77)	0 (0)	10 (0.16)	0 (0)	2 (0.07)	1 (0.02)	42 (0.95)	29 (0.15)	28 (0.25)	9 (0.30)	9 (0.53)	68 (2.83)	328 (0.52)	0.74
Whiting	53 (1.13)	10 (1.11)	82 (1.28)	13 (0.76)	5 (0.18)	20 (0.41)	105 (2.39)	322 (1.71)	134 (1.12)	353 (11.8)	31 (1.82)	86 (3.58)	1214 (1.93)	2.74
Ling	10 (0.21)	1 (0.11)	7 (0.11)	1 (0.06)	4 (0.14)	8 (0.16)	1 (0.02)	12 (0.06)	9 (0.08)	4 (0.13)	1 (0.06)	1 (0.04)	59 (0.09)	0.13
Σ <i>Trisopterus</i>	238 (5.06)	109 (12.1)	247 (3.86)	54 (3.18)	7 (0.25)	75 (1.53)	90 (2.05)	1204 (6.40)	127 (1.14)	242 (8.07)	22 (1.29)	378 (15.8)	2793 (4.45)	6.30
Σ All Gadidae	442 (9.40)	121 (13.4)	361 (5.64)	70 (4.12)	19 (0.68)	105 (2.14)	261 (5.93)	1703 (9.06)	347 (3.13)	650 (5.85)	69 (4.06)	540 (22.5)	4707 (7.46)	10.6
Herring	2 (0.04)	0 (0)	3 (0.05)	7 (0.41)	1 (0.04)	116 (2.37)	58 (1.32)	382 (2.03)	51 (0.46)	4 (0.13)	2 (0.12)	0 (0)	626 (1.00)	1.41
Garfish	0 (0)	0 (0)	0 (0)	0 (0)	0 (0)	1 (0.02)	0 (0)	27 (0.14)	118 (1.06)	9 (0.30)	7 (0.41)	4 (0.17)	166 (0.26)	0.37
Σ All pelagic fish	5 (0.11)	2 (0.22)	10 (0.16)	8 (0.47)	1 (0.04)	117 (2.39)	58 (1.32)	426 (2.27)	179 (1.61)	14 (0.47)	12 (0.71)	4 (0.17)	792 (1.26)	1.79
Σ Sandeels	1484 (31.6)	19 (2.11)	8108 (127)	2328 (137)	2608 (93.1)	3691 (75.3)	1794 (40.8)	10985 (58.4)	4576 (41.2)	1411 (47.0)	184 (10.8)	1516 (63.2)	38704 (61.6)	87.3
Σ All species	1953 (41.6)	142 (15.8)	8483 (133)	2413 (142)	2644 (94.4)	3914 (79.9)	2116 (48.1)	13137 (69.9)	5120 (46.1)	2076 (69.2)	265 (15.6)	2062 (85.9)	44325 (70.6)	100

son was significant ($H = 81.85$, $p < 0.0005$). Herring were more numerous in quarter 3 than in quarters 1, 2 and 4 ($p < 0.00005$, $p = 0.0001$ and $p < 0.00005$ respectively), and more numerous in quarter 2 than in quarter 1 ($p = 0.0036$). Garfish was most common during September and November (1.06 and 0.41 respectively). Overall variation in the number of garfish otoliths per faex in relation to season was significant ($H = 35.04$, $p < 0.0005$). Garfish were more numerous in quarter 3 than in quarter 2 ($p = 0.002$). Numbers of sandeel otoliths were highest during March and April (127 and 137), after which they declined towards the end of the year and the lowest values were generally observed during winter (Table 2). Overall variation in the number of sandeel otoliths per faex in relation to season was significant ($H = 33.77$, $p < 0.0005$). Sandeel were more numerous in quarter 2 than in quarters 1 ($p = 0.0003$), 3 and 4 ($p < 0.00005$ in both cases). Sandeels were also more numerous in quarter 1 than in quarter 3 ($p < 0.0001$).

Diet composition by weight

Gadid fish contributed between 20.6 and 87.4 % to the diet by month throughout the year, and accounted for 53.4 % of the annual diet (Table 3). A clear temporal trend was evident: after February, gadids declined markedly in importance to reach their lowest value during the spring/early summer. Between July and December, gadids generally increased in importance

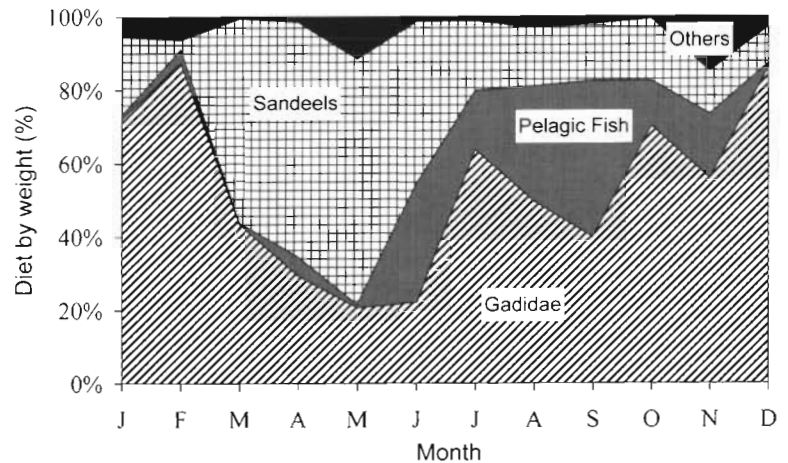


Fig. 2. Monthly variation in the percentage contribution of the main prey groups (by weight) to the diet of harbour seals *Phoca vitulina* along the southeast Shetland coast

(Fig. 2, Table 3). Whiting and saithe were the most important gadid species, contributing an estimated 36.4 % of the annual diet. Haddock was the least important, contributing only 0.9 % of the annual diet.

Pelagic fishes began to increase in importance after May to reach a peak during July to September and then sharply decreased towards the end of the year (Fig. 2). Herring was the most important pelagic species in the diet; during June and July, herring accounted for 32.3 and 16.1 % of the diet respectively. Garfish replaced herring as the main pelagic species after August, peaking during September when it formed 34.1 % of the diet, making it the top single species during September (Table 3). Overall, pelagic fish accounted for 13.8 % of the annual diet of harbour seals.

Table 3. Monthly variation in the diet, by percentage weight, of harbour seals *Phoca vitulina* along the southeast Shetland coast, as derived from the measurement of otoliths and cephalopod beaks found in faeces

Species	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Annual average
Cod	0.3	0.0	1.6	0.7	0.0	0.7	0.0	2.8	0.9	0.5	3.9	0.2	1.0
Haddock	0.2	4.3	0.7	1.7	0.2	0.0	0.8	1.1	0.3	0.0	0.0	1.8	0.9
Saithe	26.4	0.0	3.4	0.0	3.2	1.8	8.8	5.1	6.2	4.5	29.5	44.5	11.1
Whiting	17.3	35.3	23.4	18.9	5.0	7.2	49.6	27.9	21.9	48.9	17.2	31.2	25.3
Ling	10.5	18.0	4.1	1.8	11.6	10.4	0.4	3.0	7.7	10.1	1.0	0.7	6.6
Σ <i>Trisopterus</i>	14.5	29.6	8.9	6.3	0.5	1.8	3.5	9.3	2.5	6.1	4.0	8.2	7.9
Σ All Gadidae	71.9	87.4	43.3	29.3	20.6	21.9	63.9	49.7	40.0	70.6	55.9	86.6	53.4
Herring	0.4	0.0	0.3	2.6	0.9	32.3	16.1	23.9	5.9	1.8	3.3	0.0	7.3
Garfish	0.0	0.0	0.0	0.0	0.0	0.4	0.0	5.1	34.1	8.3	9.3	1.2	4.9
Σ All pelagic fish	1.0	3.7	0.4	5.2	0.9	32.7	16.1	31.4	42.5	12.4	17.6	1.2	13.8
Σ Sandeels	21.1	2.5	55.5	64.0	67.0	43.9	18.8	15.8	15.3	16.8	11.6	10.0	28.5
Σ Other fish	1.2	6.4	0.0	6.6	2.9	0.1	0.0	2.1	1.3	0.2	0.0	0.0	1.7
Σ Cephalopods	4.3	0.0	0.6	0.0	1.5	1.3	1.2	0.7	0.3	0.6	15.0	2.9	2.4

The importance of sandeels also showed clear evidence of seasonal trends, increasing after February and dominating the diet from March to June (Fig. 2). Sandeels contributed between 2.5 and 67.0% of the diet by month throughout the year and accounted for 28.5% of the annual harbour seal diet (Table 3). Generally, when sandeels were dominant in the diet (February to June), gadid fishes were least important and vice versa (Fig. 2).

Cephalopods were generally of highest importance during November to January; during November they accounted for 15% of the diet by weight (Table 3). However, overall they were of minor importance, accounting for 2.4% of the diet by weight.

Sizes of fish consumed

Length frequency distributions were constructed to allow examination of the estimated lengths of fish consumed. For sandeels (Fig. 3), distributions are on a monthly basis (except February) and for the other main species in the diet (cod, haddock, ling, whiting, Norway pout, poor cod, herring and garfish), data were pooled over the study period (Fig. 4). For the 4 species with largest sample sizes (whiting, ling, sandeel and *Trisopterus*), monthly trends in mean size eaten are also shown (Fig. 5).

The size distribution for sandeels eaten by seals was unimodal in every month except November (Fig. 3). Overall variation in sandeel size in relation to season was significant (Kruskal-Wallis; $H = 548.75$, $p < 0.0005$). Sandeels eaten in quarters 1 and 2 were larger than those eaten in quarters 3 and 4 ($p < 0.0005$) and those eaten in quarter 3 were larger than those eaten in quarter 4 ($p < 0.0005$). Overall variation in relation to month was also significant ($H = 1091.91$, $p < 0.0005$). Generally the size of sandeels declined with month (Fig. 5c) (Spearman's rank correlation $r = -0.435$) with the largest sandeels eaten in January to March and the smallest in August, September and December (see Table 4a).

All cod eaten (Fig. 4) were less than 50 cm in estimated length; the modal size was around 20 to 30 cm

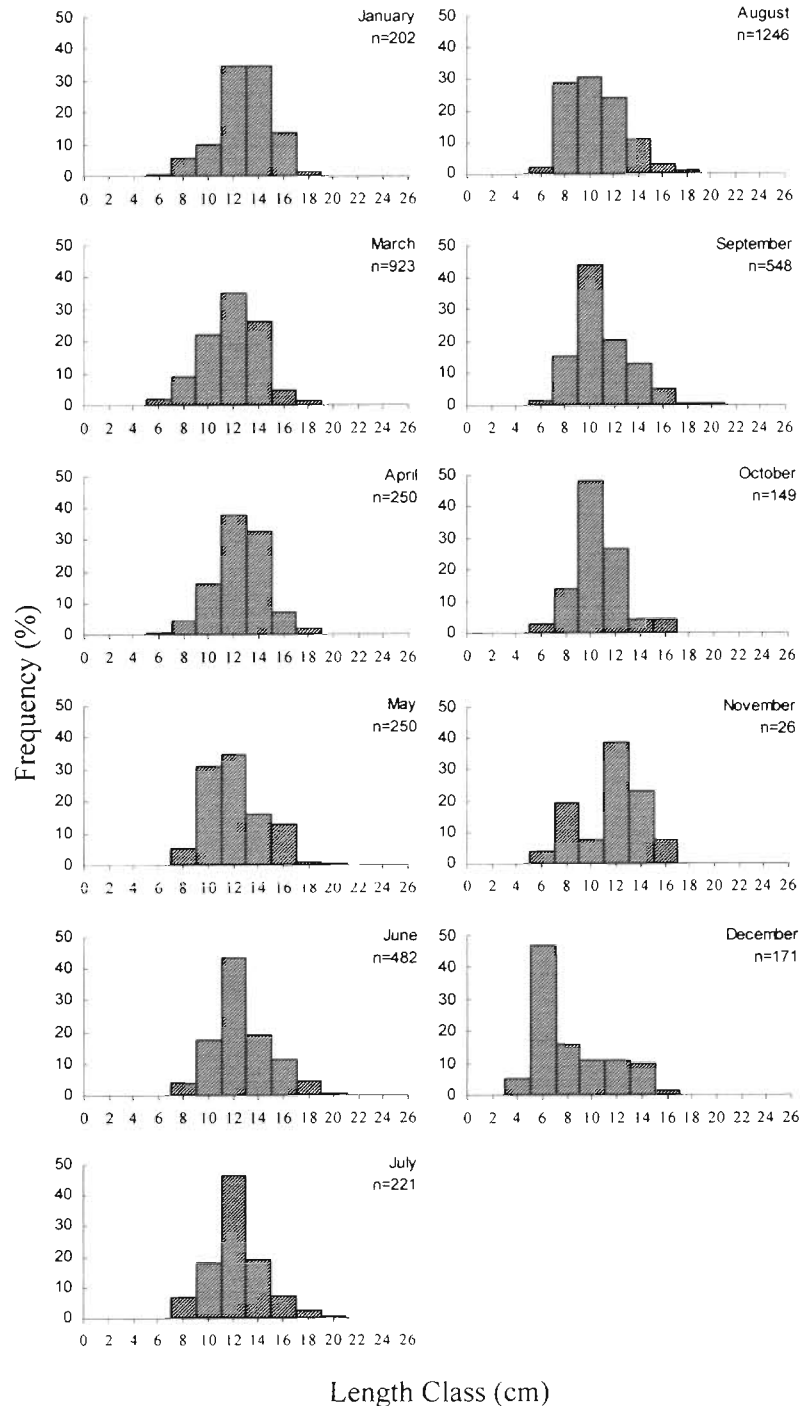


Fig. 3. Length-frequencies of sandeels consumed by *Phoca vitulina* as a percentage of the monthly total

with the distribution being slightly skewed towards larger fish. Slightly more than 20% of the cod eaten were in excess of the minimum legal landing size (MLS) of 35 cm.

Haddock (Fig. 4) had a modal size of 30 to 35 cm with the distribution being skewed towards smaller fish.

Approximately 40% of the haddock eaten were in excess of the MLS of 30 cm.

Whiting (Fig. 4) had 2 modes in the distribution, the first at 16 to 20 cm and a second, larger, mode at 32 to 36 cm. Approximately 60% of the whiting eaten were in excess of the MLS of 27 cm. Overall variation in whiting size in relation to season was significant ($H = 427.6$, $p < 0.0005$). Whiting eaten in quarters 2 and 3 were larger than those eaten in quarter 1 ($p = 0.0001$ and $p = 0.0005$ respectively), and whiting eaten in quarters 1, 2 and 3 were larger than those taken in quarter 4 ($p < 0.00005$ in all cases). When the data were

analysed on a month to month basis, whiting eaten in January, March and April were found to be generally larger than those eaten in other months (Table 4b, Fig. 5a).

The modal size of ling was 30 to 40 cm (Fig. 4) but the distribution was skewed towards larger fish, with approximately 50% of the ling eaten being in excess of the MLS of 46 cm. There was no clear seasonal pattern in size (Fig. 5b).

Saithe (Fig. 4) had a clumped distribution around 20 to 35 cm, which accounts for more than 70% of all the saithe eaten. Less than 20% of all the saithe eaten had lengths in excess of the MLS of 35 cm. Overall variation in saithe size in relation to season was significant ($H = 58.53$, $p < 0.0005$). Saithe eaten in quarters 3 and 4 were larger than those taken in quarter 1 ($p < 0.00005$ in both cases).

Poor cod (Fig. 4) had 2 modes in the distribution, the first around 9 to 15 cm (40% of all fish) and a second larger mode at 21 to 27 cm (30% of all fish). Overall variation in poor cod size in relation to season was significant ($H = 13.36$, $p = 0.004$). Poor cod eaten in quarters 1 and 3 were larger than those taken in quarter 4 ($p < 0.0007$ and $p < 0.002$ respectively). Norway pout (Fig. 4) had a mode around 15 to 18 cm. Overall variation in pout size in relation to season was significant ($H = 194.9$, $p < 0.0005$). Pout eaten in quarters 1, 2 and 3 were larger than those eaten in quarter 4 ($p < 0.00005$ in all cases). Taking the 2 *Trisopterus* species together, there was significant monthly variation in size ($H = 650.41$, $p < 0.0005$). Generally, fish taken in January, March and April were larger than those taken in the rest of the year (Table 4c, Fig. 5d).

Herring (Fig. 4) had a mode at 25 to 30 cm. Less than 5% of the herring eaten were smaller than the MLS of 20 cm for North Sea herring. Overall variation in herring size in relation to season was significant ($H = 13.36$, $p = 0.004$). Herring eaten in quarter 3 were larger than those taken in quarter 2 ($p = 0.001$). Very few herring were taken in quarters 1 or 4. Garfish (Fig. 4) had a clear mode at 70 to 80 cm and the distribution was skewed towards smaller fish.

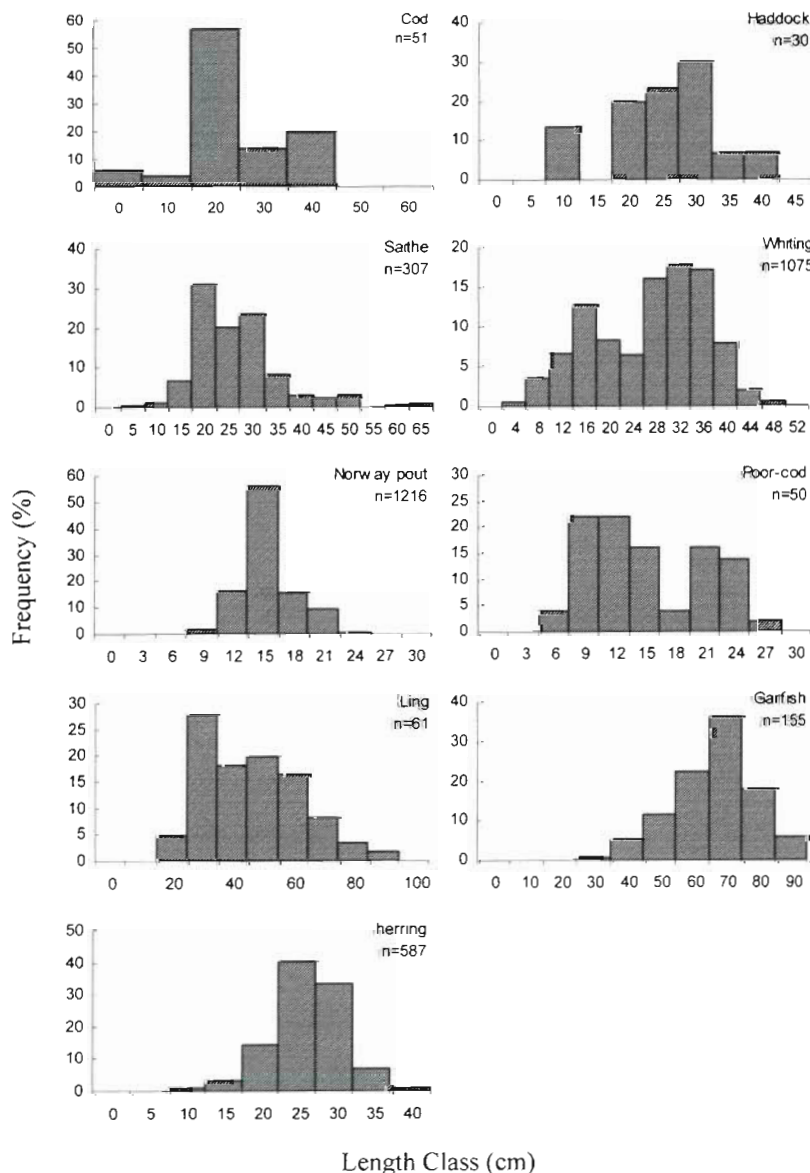


Fig. 4. Length-frequencies of fish consumed by *Phoca vitulina* as a percentage of the total, for cod, haddock, saithe, whiting, Norway pout, poor cod, ling, garfish and herring in spring and summer

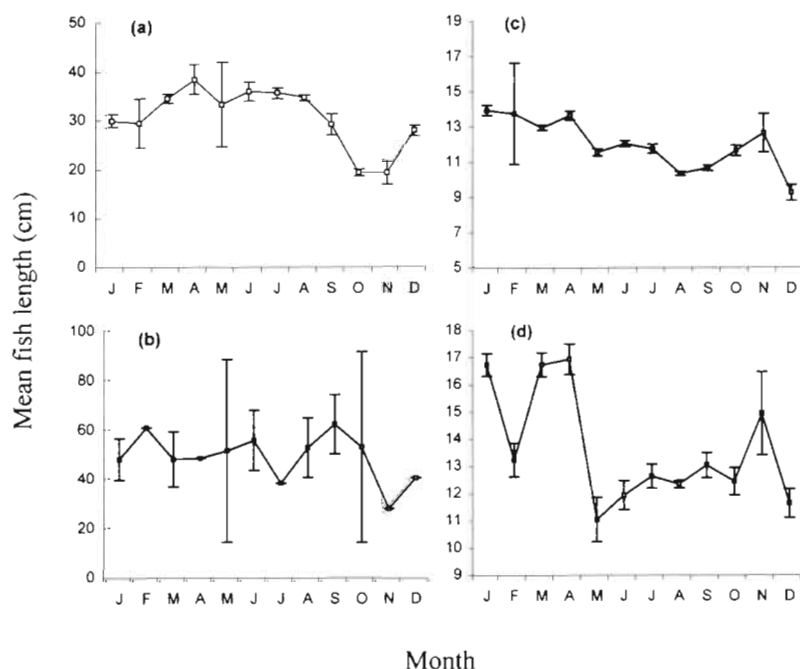


Fig. 5. Monthly variation in the mean length of fish consumed by *Phoca vitulina*, for species where sufficient data were available for all months (error bars show 95% CL): (a) whiting, $n = 1075$; (b) ling, $n = 58$; (c) sandeels, $n = 4326$; (d) *Trisopterus* spp., $n = 1616$

Assessment of error associated with using otoliths to determine seal diet

The frequencies with which different categories of prey were identified based on (1) otoliths alone and (2) all identifiable structures appear in Table 5. Comparisons using χ^2 tests indicate a significant increase in frequency of identification, when all hard structures were used, for gadids ($p < 0.02$; 15% increase), herring ($p < 0.01$; 41%), mackerel ($p < 0.01$; 323%) and garfish ($p < 0.01$; 102%). Although the detection of flatfish increased by 47% using all hard parts, this increase was not significant ($p > 0.10$), reflecting the low incidence of this group in the diet.

Correcting for samples containing no otoliths

We calculated a series of correction factors, to account for samples that contain bones of a fish species but not otoliths, based on the percentage increase in frequency of detection (as shown in the last column of Table 5). We assume that the number and size of fish (of a particular category) in samples containing only the bones was the same as for an average sample in which otoliths were present.

A modified estimate of diet composition was derived summing total prey weights for the main species or groups and applying the correction factors from Table 5. The contribution to the diet by weight was then re-scaled to sum to 100% (Table 6).

The apparent importance of gadids and sandeels is seen to decline slightly, whereas the importance of pelagic species increases, particularly for garfish (Table 6).

DISCUSSION

Seasonal variation in diet

Our results show strong seasonal trends in diet, with sandeels being the most important prey in March to June and gadids dominating the diet in much of the rest of the year. The importance of the third major category of prey, pelagic fish species, is probably significantly underestimated by basing conclusions on use of otoliths alone. Pelagic fish species were important in the diet during the summer months. Garfish in particular appear to be seasonally important, especially in September.

The inshore commercial sandeel fishery takes place around Shetland during spring and early summer and it may therefore be assumed that their availability to seals in inshore waters is highest at this time. It is also generally assumed that sandeels overwinter buried in the substrate, although Tollit & Thompson (1996) showed that sandeels may be available to seals in winter (see below).

Seasonal variation in the contribution of herring to the diet appears to broadly follow known migratory patterns. Herring migrate southwards past Shetland during late summer (J. Morrison, SOAEFD, Marine Lab., pers. comm.) and are likely to be at peak availability at some time during July to September, after which their continued migration would take them out of the range of the Shetland population of harbour seals.

The importance of garfish in the diet is worthy of comment, since this species has not been reported in seal diets in other areas of the North Sea. Garfish are occasionally by-caught with herring and mackerel by pelagic fishing vessels and have been observed inshore around Shetland in recent years (Brown pers. obs.).

Table 5. Number of *Phoca vitulina* faecal samples containing prey belonging to that group or species, for (A) using only otoliths to identify prey and (B) using all diagnostic bone structures (including otoliths) to identify prey. Difference between the identifications is expressed as a percentage change

Species/group	No. of identifications based on		Increase $100 \times (B/A) - 100$ (%)
	A	B	
Gadidae	329	377	+15
Sandeels	316	318	+1
Flatfish	19	28	+47
Herring	153	215	+41
Mackerel	13	55	+323
Garfish	59	119	+102

are likely to be available in high densities in winter and this may be a feasible explanation for the observed increase in consumption of gadids.

Comparisons with other studies

The range of prey species seen in the present study was broadly similar to that reported for other studies of harbour seal feeding in Scottish waters, with sandeels and Gadidae figuring prominently in the diet. However, feeding on garfish has not been reported in other studies.

In the Moray Firth area of Scotland, Pierce et al. (1991b) reported that clupeids predominated in winter diets and sandeels in summer diets during 1988. The decline in importance of sandeels in the diet towards the end of the year is also recorded for grey seals (Prime & Hammond 1990). In the Kattegat and Skagerrak, sandeels reached their highest contribution to the diet of harbour seals during April to June (18%; Härkönen 1987). These observations are consistent with the temporal pattern observed in the contribution of sandeels to the diet of harbour seals in the present study. The decline in importance of sandeels in winter is not always observed, however. Tollit & Thompson (1996) found that consumption of sandeels by harbour seals in the Moray Firth remained high in the winters of 1989-90, 1990-91 and 1991-92, years in which clupeid abundance in the Firth was low.

Härkönen (1987) reported cod to contribute 12 to 22% to the diet by weight, depending on the season. Our results show that cod only con-

tributes 1% to the annual diet of harbour seals inshore along the southeast Shetland coast, and cod was generally unimportant in the Moray Firth diet (1988 to 1992), except in the winter of 1991-92 (Pierce et al. 1991b, Tollit & Thompson 1996).

The most evident difference between the diets of harbour seals in Danish and Shetland waters is the contribution of flatfishes to the diet. Härkönen (1987) reported the maximum contribution to be 34% for any one season. Along the southeast Shetland coastline, flatfishes contributed less than 1% of the annual diet. In the Moray Firth, flatfish contributed less than 20% to summer diets in 1988 and 1990-1992, being absent altogether in the 1989 diet (Pierce et al. 1991b, Tollit & Thompson 1996).

There is little quantitative information available from other studies on the diet of harbour seals to allow comparison of the sizes of fish consumed. Olsen & Bjørge (1995) reported that the size of fish consumed by harbour seals in Norway ranged from 5 to 92 cm. We found a similar range, 3 to 99 cm. Tollit & Thompson (1996) reported that the mean size of whiting eaten by harbour seals, in the Moray Firth, ranged from 8.7 to 11.9 g. Our data showed that whiting eaten by harbour seals along the southeast Shetland coast had a mean weight of 245 g, markedly larger than those in the Moray Firth. The Moray Firth is a nursery area for herring, sprat and small gadids (Hopkins, 1986) and it is, therefore, not surprising to find that harbour seals exploit this abundant small prey in areas such as the Moray Firth.

The question remains as to whether harbour seals around Shetland are deliberately selecting larger prey in Shetland waters or if the fish available are generally larger than elsewhere. It is possible that some of the fish eaten include discarded fish. Seals are sometimes by-caught by whitefish boats working inshore (Brown pers. obs.). Discards form an important part of the diet of some seabirds (Evans et al. 1994, Garthe & Huppopp

Table 6. Percentage diet by weight for the main *Phoca vitulina* prey species and groups, before and after correction factors to account for samples containing no otoliths were applied

Species or group	Initial weight (g)	% weight (A)	Correction factor	Corrected weight (g)	% weight (B)	Difference (B - A)
Gadidae	265 860	49.2	1.15	305 739	45.7	-3.5
Herring	72 068	13.3	1.41	101 616	15.1	+1.8
Mackerel	5 208	1.0	3.23	16 822	2.5	+1.5
Garfish	44 252	8.2	2.02	89 389	13.3	+5.1
All pelagic	124 805	23.1	1.69	211 104	31.4	+8.3
Flatfish	2 689	0.5	1.47	3 953	0.6	+0.1
Sandeels	134 226	24.8	1.01	135 568	20.2	-4.6
Others	12 720	2.4	1.00	12 720	1.9	-0.5

1994, Garthe et al. 1996) but their importance for other piscivores is not well known. Herring, mackerel and garfish are all fast swimming species and their presence in discards might explain why seals were able to take them in relatively large numbers. However, this would not explain why harbour seals feed on these species when there is no pelagic fishing vessel activity in the area, e.g. peak predation on garfish occurs after the bulk of the herring would have migrated past Shetland, and fishing effort would have moved away to follow the herring shoals.

Methodology; use of bones to identify fish prey and implications for diet estimates

The methodology used in this study, based on faeces collected at haul-out sites, is expected to have provided representative information on the diet of seals feeding locally. Telemetry studies have shown that harbour seals are essentially inshore feeders (Thompson & Miller 1990, Thompson 1993) and it can therefore be argued that most feeding will be represented in faeces found on the haul-out. The number of samples collected varied from month to month, reflecting changing availability, as also reported by Pierce et al. (1991b) and Hammond et al. (1994a, b). As is usual in such studies, the 'weighting' given to each faex was determined solely by the weight of prey represented by the hard remains contained therein. Alternative formulations (e.g. equal weighting for each faex, equal total weighting for faeces for each month) are possible but were not explored here.

Our results clearly show that some species or groups of fish are likely to be underestimated by using only the otoliths found in faecal samples, particularly those with small or fragile otoliths (e.g. herring, mackerel and salmon). There was a 323% increase in the apparent incidence of mackerel when all identifiable mackerel remains were used as opposed to only otoliths. Boyle et al. (1990) reported that during a captive feeding experiment they recovered only 1 salmon otolith out of a total of 38 ingested (2.6% recovery rate). Cottrell et al. (1996) reported an average recovery rate of 54% for otoliths of several species of fish fed to captive harbour seals. They also found that using bone structures other than otoliths significantly improved the detection of prey species and concluded, 'identifying several different prey structures increases the likelihood of identifying a prey type'.

Prime & Hammond (1990) argued that traditional faecal analysis, based on identification and measurement of otoliths, does not miss any significant part of the diet. SMRU (1984) compared the energy content of several grey seal faeces from the southwestern North

Sea with the energy content of the fish represented by the otoliths in the faeces, and found that the ratio of the 2 values was consistent with reported values for assimilation efficiency. Thus no major component of the diet was missed. However, their samples were collected in an area of the North Sea where the seals were unlikely to encounter salmonid or pelagic fishes.

Recent work has been directed towards refining digestion coefficients to account for partial digestion of the otolith while in the seal's gut (e.g. Tollit et al. 1997), rather than dealing with loss or complete digestion of otoliths. Brown & Pierce (1997) reported that the percentage composition of the diet by weight was relatively insensitive to the inclusion of experimentally derived correction factors. Our present work shows that the apparent diet composition is sensitive to the inclusion of fish represented by remains other than otoliths found in the faeces.

Assessment of seal fisheries interactions

The main species exploited by commercial fisheries around Shetland are haddock, whiting, ling, saithe and cod (Anon 1996b). One species that is important in local landings but was not detected in the harbour seal diet is the anglerfish *Lophius piscatorius*. This species has very small otoliths and a soft skeleton so it is possible that its remains would not be detected.

Our results show that the 5 main commercial species account for 45% of the annual diet of harbour seals in this area. However, many of the fish eaten were under the current legal minimum landing sizes. It remains difficult to quantify the degree of direct competition since:

(1) It is difficult to establish whether harbour seals are taking fish in the same area and at the same period of time as commercial fishing vessels because at present there are no data on where harbour seals are feeding in relation to the main centres of commercial fishing.

(2) The proportion of the small fish eaten which would have become available to fisheries in the future is unknown. Indeed, as discussed above, the seals could have been directly feeding on discarded fish.

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Appendix 1. Regression equations used to predict fish lengths and weights from otolith measurements; sources (S) and digestion coefficients (DC) are also given. B = Bedford et al. (1986); C = Coull et al. (1989); Cl = Clarke (1986); E = E. G. Brown (unpubl. data); H = Härkönen (1986); J = J. R. G. Hislop (unpubl. data); P = G. J. Pierce (unpubl. data); * combined data. OL: otolith length; OW: otolith width; LHL: lower hood length; UHL: upper hood length; URL: upper rostral length; LRL: lower rostral length

Species of fish with scientific name	Estimated fish length (mm)	S	Estimated fish weight (g)	S	DC
Cod <i>Gadus morhua</i>	FL = 9.535 (OL ^{1.475})	E	FW = 0.007823 (OL ^{4.452569})	E	1.070
	FL = 42.683 (OW ^{1.300})	E	FW = 0.651869 (OW ^{3.986982})	E	1.070
Haddock <i>Melanogrammus aeglefinus</i>	FL = 14.096 (OL ^{1.198})	E	FW = 0.019742 (OL ^{3.709555})	E	1.250
	FL = 45.897 (OW ^{1.215})	E	FW = 0.723869 (OW ^{3.793770})	E	1.240
Saithe <i>Pollachius virens</i>	FL = 13.829 (OL ^{1.332})	E	FW = 0.028654 (OL ^{3.931842})	E	1.250
	FL = 42.391 (OW ^{1.525})	E	FW = 0.762794 (OW ^{4.513928})	E	1.240
Haddock/saithe	FL = 16.274 (OL ^{1.197})*	E	FW = 0.039122 (OL ^{3.600289})*	E	1.250
	FL = 49.497 (OW ^{1.269})	E	FW = 1.066829 (OW ^{3.844856})	E	1.240
Ling <i>Molva molva</i>	FL = -128.038 + (OL×67.634)	E	FW = 0.198199 (OL ^{3.620808})	E	1.250
	FL = -130.941 + (OW×186.906)	E	FW = 6.559731 (OW ^{3.738733})	E	1.240
Whiting <i>Merlangius merlangus</i>	FL = -54.114 + (OW×79.671)	E	FW = 0.790806 (OW ^{3.705954})	E	1.365
Norway pout <i>Trisopterus esmarki</i>	FL = -4.794 + (OL×23.624)	E	FW = 0.046490 (OL ^{3.324836})	E	1.250
	FL = -1.640 + (OW×55.204)	E	FW = 0.886564 (OW ^{3.279030})	E	1.240
Poor cod <i>Trisopterus minutus</i>	FL = -66.617 + (OL×30.734)	E	FW = 0.002926 (OL ^{4.761464})	E	1.250
	FL = -34.836 + (OW×56.691)	E	FW = 0.218384 (OW ^{4.211111})	E	1.240
<i>Trisopterus</i> spp.	FL = -5.886 + (OL×23.443)*	E	FW = 0.033918 (OL ^{3.531258})*	E	1.250
	FL = 15.515 + (OW×45.404)*	E	FW = 0.916531 (OW ^{3.157323})*	E	1.240
Tusk/torsk <i>Brosme brosme</i>	FL = 24.100 + (OL×44.424)	E	FW = 0.005140 (FLcm ^{3.189000})	B	1.250
Unidentified gadids	FL = -61.590 + (OL×33.304)*	H	FW = 0.016042 (FLcm ^{3.035950})*	B/C	1.250
	FL = -54.350 + (OW×76.582)*	P	FW = 0.016042 (FLcm ^{3.035950})*	P	1.240
Ling/rockling	FL = 36.900 + (OL×51.700)*	H/P	FW = 2.485710 (OL ^{2.592760})*	H/P	1.250
Herring <i>Clupea harengus</i>	FL = -87.490 + (OW×184.390)	H	FW = 0.006030 (FLcm ^{3.090400})	C	1.170
Mackerel <i>Scomber scomberus</i>	FL = 41.363 + (OL×74.075)	E	FW = 4.745687 (OL ^{3.129485})	E	1.250
	FL = -33.539 + (OW×255.071)	E	FW = 78.642882 (OW ^{4.046553})	E	1.240
Garfish <i>Belone belone</i>	FL = 0.000 + (OL×140.000)	H	FW = 0.000200 (FLcm ^{3.442000})	C	1.250
Scad <i>Trachurus trachurus</i>	FL = -27.020 + (OL×34.939)	P	FW = 0.003400 (FLcm ^{3.294300})	C	1.250
	FL = -26.110 + (OW×79.010)	P	FW = 0.003400 (FLcm ^{3.294300})	C	1.240
Sprat <i>Sprattus sprattus</i>	FL = -50.520 + (OL×151.950)	J	FW = 0.002168 (FLcm ^{3.474600})	C	1.090
Lemon sole <i>Microstomus kitt</i>	FL = 10.930 + (OL×88.460)	H	FW = 0.026520 (FLcm ^{2.764300})	C	1.230
	FL = -71.440 + (OW×176.450)	P	FW = 0.026520 (FLcm ^{2.764300})	C	1.230
Dab <i>Limanda limanda</i>	FL = -50.960 + (OL×58.470)	H	FW = 0.005450 (FLcm ^{3.195000})	B	1.250
	FL = -97.830 + (OW×108.584)	P	FW = 0.007400 (FLcm ^{3.112800})	C	1.240
Bothidae	FL = -11.420 + (OL×54.770)*	H	FW = 0.024920 (FLcm ^{2.857000})*	B	1.250
Long rough dab	FL = -24.520 + (OL×48.350)	H	FW = 0.00400 (FLcm ^{3.203900})	C	1.250
<i>Hippoglossoides platessoides</i>	FL = -20.830 + (OW×59.433)	P	FW = 0.00400 (FLcm ^{3.203900})	C	1.240
Witch <i>Glyptocephalus cynoglossus</i>	FL = -100.650 + (OL×78.290)	H	FW = 0.001700 (FLcm ^{3.389800})	C	1.250
	FL = -272.000 + (OW×116.000)	P	FW = 0.001700 (FLcm ^{3.389800})	C	1.240
LRD/witch	FL = -52.600 + (OL×61.102)*	H	FW = 0.125180 (OL ^{4.144100})*	H	1.250
Unidentified flatfish	FL = -25.950 + (OL×53.274)*	H	FW = 0.009923 (FLcm ^{3.035950})*	B/C	1.250
	FL = -38.100 + (OL×76.600)*	P	FW = 0.009923 (FLcm ^{3.035950})*	B/C	1.240
Sandeels <i>Ammodytes</i> spp.	FL = 18.376 + (OL×51.441)	E	FW = 1.083343 (OL ^{2.446703})	E	1.160
and <i>Gymnodytes</i> spp.	FL = 10.589 + (OW×110.199)	E	FW = 5.731932 (OW ^{2.679693})	E	1.160
Greater sandeels <i>Hyperoplus</i> spp.	FL = -4.024 + (OL×56.840)	H	FW = 1.083343 (OL ^{2.446703})	E	1.160
Argentines <i>Argentina</i> spp.	FL = -35.049 + (OL×46.370)	E	FW = 0.082530 (OL ^{3.830400})	E	1.250
	FL = -50.464 + (OW×73.583)	E	FW = 0.140400 (OW ^{4.654200})	E	1.240
Wrasses Labridae	FL = 3.320 (OL ^{53.440})*	H	FW = 2.330310 (OL ^{2.934000})*	H	1.250
Catfish <i>Anarhichas lupus</i>	FL = -242.270 + (OL×216.510)	H	FW = 0.003300 (FLcm ^{3.249100})	C	1.250
Dragonet <i>Callionymus</i> spp.	FL = -51.390 + (OL×84.120)	P	FW = 0.022000 (FLcm ^{2.590700})	C	1.250
	FL = -68.660 + (OW×167.300)	P	FW = 0.022000 (FLcm ^{2.590700})	C	1.240
Norway haddock <i>Sebastes viviparus</i>	FL = -14.460 + (OL×38.810)	H	FW = 0.000750 (OL ^{5.410000})	H	1.250
Curled octopus <i>Eledone</i>			FW = 5.365600 (LHL ^{2.850000})	Cl	
			FW = 8.250720 (UHL ^{2.337400})	P	
Rossia <i>Rossia macrosoma</i>			FW = 8.846306 (LHL ^{1.650000})	Cl	
Market squid <i>Loligo forbesi</i>			FW = 19.105950 (LRL ^{84.274})	Cl	
			FW = 31.092000 (URL ^{2.497120})	P	
Sepiolidae			FW = 0.642454 (UHL ^{0.350000})	Cl	
			FW = 1.491820 (LHL ^{0.350000})	Cl	

LITERATURE CITED

- Anon (1996a) Shetland in statistics. Shetland Islands Council's Development Department, Lerwick, Shetland
- Anon (1996b) Scottish sea fisheries statistical tables 1995. Scottish Office Agriculture Environment and Fisheries Department, Edinburgh
- Bedford BC, Woolner LE, Jones BW (1986) Length-weight relationships for commercial fish species and conversion factors for various presentations. Fisheries Data Report. MAFF, Lowestoft
- Boyle GJ (1990) The feeding ecology of common seals (*Phoca vitulina* L.) in Loch Linnhe, West Scotland. MSc thesis, Imperial College of Science, Technology and Medicine, University of London
- Boyle PR, Pierce GJ, Diack JSW (1990) Sources of evidence for salmon in the diets of seals. Fish Res 10:137–150
- Brown EG (1995) Survey and review of data on grey seal populations around Shetland. Unpublished report to Scottish Natural Heritage, Contract No. NE/S/32/94, Lerwick, Shetland
- Brown EG, Pierce GJ (1997) The diet of harbour seals at Mousa, Shetland, during the third quarter of 1994. J Mar Biol Assoc UK 77:539–555
- Clarke MR (ed) (1986) A handbook for the identification of cephalopod beaks. Clarendon Press, Oxford
- Cottrell PE, Trites AW, Miller EH (1996) Assessing the use of hard parts in faeces to identify harbour seal prey: results of captive feeding trials. Can J Zool 74:875–880
- Coull KA, Jermyn AS, Newton AW, Henderson GI, Hall WB (1989) Length-weight relationships for 88 species of fish encountered in the North East Atlantic. Scottish Fisheries Research Report 43, Aberdeen
- da Silva J, Neilson JD (1985) Limitations of using otoliths recovered in scats to estimate prey consumption in seals. Can J Fish Aquat Sci 42:1439–1442
- Dellinger T, Trillmich F (1988) Estimating diet composition from scat analysis of otariid seals (Otariidae): is it reliable? Can J Fish Aquat Sci 42:1865–1870
- des Clers S, Prime J (1996) Seals and fishery interactions: observations and models in the Firth of Clyde, Scotland. In: Greenstreet SPR, Tasker ML (eds) Aquatic predators and their prey. Blackwell Science, Oxford, p 124–132
- Dorman JA (1989) Some aspects of the biology of the garfish *Belone belone* (L.) from southern Ireland. J Fish Biol 35:621–629
- Dorman JA (1991) Investigations into the biology of the garfish, *Belone belone* (L.) in Swedish waters. J Fish Biol 39:59–69
- Duck CD, Hall AJ, Ingram SS (1993) Monitoring the impact of the M.V. Braer oil spill in Shetland: common seal aerial survey, August 1993. Unpublished report to the Ecological Steering Group on the Oil Spill in Shetland, Sea Mammal Research Unit, NERC, Cambridge
- Evans SM, Hunter JE, Elizal, Wahju RI (1994) Composition and fate of the catch and by-catch in the Farne Deep (North Sea) *Nephrops* fishery. ICES J Mar Sci 51:155–168
- Garthe S, Camphuysen CJ, Furness RW (1996) Amounts of discards by commercial fisheries and their significance as food for seabirds in the North Sea. Mar Ecol Prog Ser 136: 1–11
- Garthe S, Huppopp O (1994) Distribution of ship-following seabirds and their utilization of discards in the North Sea in summer. Mar Ecol Prog Ser 106:1–9
- Gulland JA (1987) The impact of seals on fisheries. Mar Pol (July 1987), p 196–204
- Hammond PS, Hall AJ, Prime J (1994a) The diet of grey seals around Orkney and other island and mainland sites in north-eastern Scotland. J Appl Ecol 31:340–350
- Hammond PS, Hall AJ, Prime J (1994b) The diet of grey seals in the Inner and Outer Hebrides. J Appl Ecol 31:737–746
- Härkönen T (1986) The guide to the otoliths of the bony fishes of the north-east Atlantic. Danbiu ApS, Hellerup, Denmark
- Härkönen T (1987) Seasonal and regional variations in the feeding habits of the harbour seal, *Phoca vitulina*, in the Skagerrak and the Kattegat. J Zool 213:535–543
- Harvey JT (1989) Assessment of the errors associated with harbour seal (*Phoca vitulina*) faecal sampling. J Zool 219:101–111
- Harvey JT, Antonelis GA (1994) Biases associated with non-lethal methods of determining the diet of northern elephant seals. Mar Mamm Sci 10:178–187
- Harwood JA (1987) Competition between seals and fisheries. Sci Prog 71:429–437
- Harwood JA (1992) Assessing the competitive effect of marine mammal predation on commercial fisheries. In: Payne AIL, Brink KH, Mann KH, Hilborn R (eds) Benguela trophic functioning. S Afr J Mar Sci 12:689–693
- Hiby L, Duck D, Thompson D, Hall A, Harwood J (1996) Seals stocks in Great Britain. NERC News, January 1996, Swindon
- Hopkins PJ (1986) Exploited fish and shellfish populations in the Moray Firth. Proc R Soc Edinb 91B:57–72
- Jobling M (1987) Marine mammal faeces samples as indicators of prey importance: a source of error in bioenergetic studies. Sarsia 72:255–260
- McConnell BJ, Chambers C, Nicholas KS, Fedak MA (1992) Satellite tracking of grey seals (*Halichoerus grypus*). J Zool 226:271–282
- Muus BJ, Dahlström P (1977) Guide to the sea fishes of Britain and north western Europe. Wm Collins and Sons Ltd, London
- Olsen M, Bjørge A (1995) Seasonal and regional variations in the diet of harbour seals in Norwegian waters. In: Blix AS, Walløe L, Ultang Ø (eds) Whales, seals, fish and man. Elsevier Science BV, Amsterdam, p 271–285
- Pierce GJ, Boyle PR, Diack JSW (1991a) Identification of fish otoliths and bones in faeces and digestive tracts of seals. J Zool 223:320–328
- Pierce GJ, Boyle PR, Thompson PM (1990) Diet selection by seals. In: Barnes M, Gibson RN (eds) Trophic relationships in the marine environment. Proc 24th Eur Mar Biol Symp, Oban 1989. Aberdeen University Press, Aberdeen, p 222–238
- Pierce GJ, Thompson PM, Miller A, Diack JSW, Miller D, Boyle PR (1991b) Seasonal variation in the diet of common seals (*Phoca vitulina*) in the Moray Firth area of Scotland. J Zool 223:641–652
- Prime JH, Hammond PS (1990) The diet of grey seals from the south-western North Sea assessed from the analyses of hard parts found in faeces. J Appl Ecol 27:435–447
- Rae BB (1960) Seals and Scottish Fisheries. Department of Agriculture and Fisheries for Scotland Marine Research, 1960, No. 2, Edinburgh
- Rae BB (1968) The food of seals in Scottish waters. Department of Agriculture and Fisheries for Scotland Marine Research, 1968, No. 2, Edinburgh
- Rae BB (1973) Further observations on the food of seals. J Zool 169:287–297
- Sergeant DE (1951) The status of the common seal (*Phoca vitulina* L.) on the East Anglian coast. J Mar Biol Assoc UK 29:707–717
- SMRU (1984) Interactions between grey seals and UK fish-

- eries. Report on research conducted for the Department of Agriculture and Fisheries for Scotland by the Sea Mammal Research Unit. NERC, Cambridge
- SMRU (1994) Grey seals in the North Sea and their interactions with fisheries. Final report to the Ministry of Agriculture, Fisheries and Food on contract MF 0503. NERC, Cambridge
- Thompson PM (1993) Harbour seal movement patterns. Symp Zool Soc Lond 66:225–239
- Thompson PM, Miller D (1990) Summer foraging activity and movements of radio-tagged common seals (*Phoca vitulina* L.) in the Moray Firth, Scotland. J Appl Ecol 27:492–501
- Tollit DJ, Thompson PM (1996) Seasonal and between year variations in the diet of harbour seals in the Moray Firth, Scotland. Can J Zool 74:1110–1121
- Tollit DJ, Steward M, Thompson PM, Pierce GJ, Santos MB, Hughes S (1997) Species and size differences in the digestion of otoliths and beaks; implications for estimates of pinniped diet composition. Can J Fish Aquat Sci 54: 105–119
- Watt J, Pierce GJ, Boyle PR (1997) A guide to the identification of North Sea fish using premaxillae and vertebrae. Co-operative Research Report No. 220, International Council for the Exploration of the Sea, Copenhagen
- Wheeler A (1978) Key to the fishes of Northern Europe. Frederick Warne Ltd, London

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