Annual sediment primary production and respiration in a large coral reef lagoon (SW New Caledonia)

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ABSTRACT: Sediment photosynthetic production and community respiration were investigated for 1 yr in the south-west lagoon of New Caledonia (surface area: 2000 km²; mean depth: 21 m). Metabolic fluxes were measured at the water-sediment interface using benthic enclosures, at 60 sampling stations randomly distributed in space and time. Mean gross primary production ($P_g$) was 12.06 mmol C m⁻² yr⁻¹. Mean respiration ($R$) was 13.68 mmol C yr⁻¹. $P_g/R$ was thus 0.88, indicating that the whole lagoon was net heterotrophic. There were, however, large differences along a land-ocean gradient. Muddy bottoms (35% of the lagoon surface), located along the shore and in the deeper part of the lagoon, were significantly heterotrophic ($P_g/R = 0.39$), whereas the ratio for sandy bottoms (65% of the area) towards the reef were close to 1 or slightly positive, indicating potentially autotrophic benthic communities. Both $P_g$ and $R$ followed a similar seasonal variation with higher values during the warm season (November to February), when maximum irradiance and the particulate organic carbon supply occurred. These results suggest that autotrophy, generally observed in coral reef flats, may also occur in lagoon sediments, indicating both a major importance of regeneration processes and a relative equilibrium of energy exchanges with adjacent ecosystems. Terrestrial influence results in a decrease of autochthonous production and a correlative increase of carbon sedimentation leading to benthic heterotrophy.

KEY WORDS: Coral reef lagoon · Benthic community metabolism · Sediments

INTRODUCTION

Most of our knowledge of coral reef ecosystems is about the reef compartment (Kinsey 1985a, Dubinsky 1990, Sorokin 1993). Community gross primary production of coral reefs is high, with a mean value of ca 0.6 mmol C m⁻² d⁻¹ (Kinsey 1985a); respiration is of the same order of magnitude and the budget of organic carbon is considered to be balanced (Crossland et al. 1991) or slightly positive (Gattuso et al. 1993, Smith 1995). Energy transfer from reef to the adjacent lagoon is thus necessarily low. Along islands or continental coasts, mangroves can supply organic carbon for surrounding communities, but their importance seems to be overrated because of the refractory nature of the exported organic matter and its low energetic value (Alongi & Christoffersen 1992). Thus, the functioning of the large sedimentary areas often observed in coral reef environments are still poorly understood (Gattuso et al. 1998). It is generally considered, however, that sediment respiration exceeds global production in terms of carbon (Yap et al. 1994). Sediment is thus regarded as a carbon sink for reef excess production or other energy sources.

The lagoons surrounding the mainland of New Caledonia (SW Pacific) cover a 24 000 km² area delimited by a 1600 km long barrier reef (Richer de Forges 1991). These coastal lagoons are under the influence of ter-
restriosal input unlike atoll lagoons. The south-west lagoon surrounding Nouméa peninsula was selected for an energy flux study. Previous studies on this site have investigated the biomass of functional compartments (Chardy & Clavier 1988a, Garrigue 1995, 1998), the vertical flux of particulate matter (Clavier et al. 1995) and sediment community metabolism (Boucher & Clavier 1990, Boucher et al. 1994a). A first attempt to estimate the carbon network was made by Chardy & Clavier (198813) with energy flows mainly derived from the literature. Two benthic energy sources were defined: the known vertical flux of organic particles (Clavier et al. 1995) and the unknown autochthonous photosynthetic production.

The aims of the present study were: (1) to evaluate, for the whole lagoon, the annual contribution of benthic primary production in the supply of carbon for benthic respiration, (2) to estimate the balance between benthic production and respiration in the different benthic communities in order to establish their relative importance in the metabolic status of the lagoon, and, (3) to determine seasonal changes in the metabolism of soft-bottom communities relative to physical factors.

MATERIAL AND METHODS

Study site. The south-west lagoon of New Caledonia covers 2066 km², 5 % of which is occupied by coral reef structures. The average depth is 21 m. Three main soft-bottom types have been identified (Fig. 1): the mud deposits near the coast and in the deeper parts, the grey sand bottoms in the middle, and the white sand bottoms close to the reef. They make up 35, 50 and 15 % of the lagoon surface, respectively. The classification was established for macrobenthic assemblages related to sediment granulometry (Chardy & Clavier 1988a, Chardy et al. 1988) and validated with other benthic parameters, especially aerobic and anaerobic metabolism (Boucher & Clavier 1990, Boucher et al. 1994a). This general distribution was refined using macrophyte association by Garrigue (1995).

Sampling strategy. Benthic fluxes at the water-sediment interface were estimated from photosynthetic production and respiration. Benthic metabolism, first linked to benthic biological activity, and hence community composition (Pichon 1997), is controlled through the physiology of plants and animals by environmental physico-chemical parameters (light, depth, substrate, turbidity, temperature), which can vary in space as well as in time. A spatio-temporal sampling was thus applied to the soft bottoms of the whole lagoon from July 1990 to June 1991. The following assumptions were used: (1) photosynthetic production only occurs between sunrise and sunset, (2) any fluctuation of photosynthetic production can be related to the time during the day, the period of the year and the location in the lagoon, (3) community respiration does not change during a 24 h cycle and only varies according to the season and the location in the lagoon.

The soft bottoms of the lagoon during 1 yr constitute the sampled statistic population. The day was divided into 1 h periods corresponding to the planned duration of production experiments. The months of the year were divided into 6 groups to take into account the seasonal fluctuations of light, temperature and day duration, and the technical limitations of the field work (availability of facilities at sea, diving and meteorological limitations, etc.). Geographical localisation of the

Fig. 1 Distribution of sampling stations in (A) space and (B) time. Dotted lines in B indicate seasonal cycle of sunrise and sunset.
sampling sites was defined in the lagoon with a 0.1 nautical mile precision both in latitude and longitude. The sampling unit consisted of three 0.2 m² areas over 1 h. It was defined by 4 co-ordinates: latitude, longitude, month and hour of the day. The total number of samples was set to 60 because of the time at our disposal. There was not enough information on benthic photosynthetic production in the lagoon to allow stratified sampling. Accordingly, the 4 co-ordinates of each station were randomly drawn (Fig. 1). This sampling was regarded as a simple random sampling (Cochran 1977). However, we can only obtain an approximation, because we studied three 0.2 m² and 1 h sub-samples by 0.1 X 0.1 nautical mile and 2 mo sampling unit. The number of stations by bottom type allocated by the random sampling over the whole lagoon was 21 in mud deposits, 27 in grey sand bottoms and 12 in white sand bottoms.

Incubation procedure. Incubations were carried out from the RV ‘Alis’. At each station, community primary production and respiration were obtained from changes in dissolved oxygen measured in three 60 l clear acrylic hemispheres fastened to 0.2 m² PVC bases (Boucher & Boucher-Rodoni 1985, 1988, Boucher et al. 1994a,b). Water motion was provided by adjustable submersible pumps connected to waterproof batteries. Oxygen concentrations were measured with the polarographic probes of 3 Yellow Spring Instrument (YSI 58) oxymeters refitted in a waterproof container and connected to a data logger. Oxygen concentrations and saturation rates could be checked underwater on liquid crystal displays. As metabolic response depends on hydrodynamics (Patterson et al. 1991, Forja & Gómez-Parra 1998), water flow in each enclosure was adjusted to the minimum value allowing stable measures from the probes. A quantum sensor (LI-192SA) was deployed inside one of the hemispheres to measure the photosynthetically active radiation (PAR, 400 to 700 nm) available for the enclosed plants. Oxygen concentrations and irradiance were logged every 5 s.

After a 30 min waiting period to lessen the disturbance created by experimental material setting, incubations were first carried out in light for 1 h to estimate the net oxygen photosynthetic production. Photosynthetic processes were then inhibited by injecting DCMU (final concentration of 5 x 10⁻⁵ mol L⁻¹), a photosynthesis inhibitor that is easy to handle in severe diving conditions and simulates dark incubation (Garrigue et al. 1992a). As oxygen tension could influence production (Dromgoole 1973), enclosed water was flushed prior to DCMU injection when the oxygen saturation rate exceeded 110%. Logging was resumed, for 1 h, after a period of 15 min to ensure that photosynthesis was fully inhibited.

The enclosed macroflora was collected by hand upon completion of the incubations. Three 5.31 cm² cores were taken in each enclosure to estimate photosynthetic pigments from the top cm of substrate. About 100 g of sediment was also collected for granulometric analysis.

Processing of the samples. The sediment samples were deep frozen pending analysis. Data were downloaded to a computer. Irradiance (I, mmol m⁻² h⁻¹) was integrated over the incubation hour and oxygen concentrations were averaged every minute. Net hourly photosynthetic production (Pₚ) was calculated as the difference between concentrations at the beginning and at the end of incubations carried out in light, after checking the regularity of the data. Total sediment oxygen consumption was computed by linear regression of the 60 mean values acquired after DCMU injection. To obtain only the benthic respiration of the enclosed system (R), total sediment oxygen consumption was adjusted for oxygen demand of the apparatus without sediment. The latter, assessed from 12 in situ isolated water dark incubations, was 0.8 x 10⁻³ mmol O₂ l⁻¹ h⁻¹ (SE = 0.1 x 10⁻³). Gross primary production (Pₚ) was estimated as Pₚ corrected from total sediment oxygen consumption. Net planktonic photosynthetic production of the enclosed water, very close to 0, was considered to be negligible compared to the benthic production. Oxygen fluxes were transformed to carbon fluxes using benthic community metabolic quotients calculated from the same area at different sites and seasons (Clavier et al. 1994); 1.03 for the community photosynthetic quotient (CPQ = gross O₂ production/gross CO₂ consumption) and 1.14 for the community respiratory quotient (CRQ = CO₂ production/O₂ consumption).

In the laboratory, macrophyte ash-free dry weight (AFDW) was measured after drying at 60°C for 48 h and oven heating at 550°C for 3 h. Photosynthetic pigments were estimated on lyophilised sediment samples after acetone extraction for 18 to 24 h at 5°C. After filtration of the extract, optical densities were measured on a spectrophotometer at 750 and 665 nm (Garrigue & Di Matteo 1991). Chlorophyll a (chl a) and phaeopigments were calculated as described by Lorenzen (1967). Sediment particle size fractions were determined according to Folk & Ward (1957).

Statistical treatment of data and modelling. ANOVAs were performed to test spatial and temporal homogeneity of the parameters under study. Normality of data (Shapiro and Wilk test) and homogeneity of variances (Bartlett test) were verified prior to analysis. The least significant difference (LSD) test was used to separate possible sets of homogeneous data (Sokal & Rohlf 1981). The Pₚ/R ratio and its confidence limits were calculated according to Cochran (1977) for a simple random sampling.
Gross primary production is essentially determined by the amount of plant material and irradiance. Gross primary production in carbon was plotted as a function of in situ irradiance. An exponential function (Gattuso et al. 1996, Boucher et al. 1998) was fitted to the community productivity:

\[ P_g = P_{g \text{ max}} \left[ 1 - \exp\left(-I/I_o\right) \right] + e \]

where \( P_g \): gross primary production; \( P_{g \text{ max}} \): rate of maximum gross productivity; \( I \): irradiance; \( I_o \): the optimum irradiance, where \( P_g = P_{g \text{ max}} \) and \( e \): a constant. Gross primary production was also related to macrophyte (AFDW) and microphyte (chl a) biomass using linear relationships.

The relation between gross primary production and the 3 pooled variables was then expressed as:

\[ P_g = P_{g \text{ max}} \left[ 1 - \exp\left(-I/I_o\right) \right] + a \text{ AFDW} + b \text{ chl a} + c \]

where \( c \): a constant; and \( a \) and \( b \): multiplicative factors. Non-linear estimation was computed using least-square estimates (Simplex procedure).

**RESULTS**

A detailed account of the raw data for each station is found in the report by Garrigue et al. (1992b).

**Environmental parameters**

Water temperature varied from 22.3 to 28.1°C, with an average of 24.3°C. The mean irradiance inside the enclosures was 204 mmol m\(^{-2}\) h\(^{-1}\) and the maximal value was 791 mmol m\(^{-2}\) h\(^{-1}\). Monthly courses of both temperature and irradiance are shown in Fig. 2. They exhibit a seasonal pattern with extreme values of irradiance occurring before highest temperatures. The maximal irradiance observed in the September-October results, however, was the major contribution of 1 shallow station. The mean irradiance for the period decreases to 285 mmol m\(^{-2}\) h\(^{-1}\) (SE = 53) when this value is removed. The mean depth and sedimentological characteristics of the stations sampled in each bottom type are indicated in Table 1. Depth and mud percentage were significantly different for the 3 types of sediment (ANOVA, \( p < 0.05 \)), but only mud deposits were characterised by significantly higher mean size in phi scale corresponding to fine sand on the Wentworth (1922) scale. Other lagoon sediments corresponded to medium sand. Phaeopigment and chl a percentage (Table 2) differed significantly for the 3 bottom types (ANOVA, \( p < 0.05 \)). Mud deposits were characterised by a significantly lower chl a content, whereas grey sand bottoms exhibited the highest macroflora biomass (Table 2). Phaeopigments were lowest in July-August and highest in November-December. Macroflora biomass, chl a and the contribution of chl a to total pigments did not differ significantly over the year, except for the January-February highest values. For macroflora, this feature was due to the sampling during this period of a very dense *Halodule uninervis* seagrass bed at Stn 14. The January-February mean macroflora biomass would decrease to 13 g AFDW m\(^{-2}\) if this station was removed from the analysis.

**Gross photosynthetic production**

The mean benthic community gross photosynthetic production in oxygen estimated from the whole data set was 2.83 mmol m\(^{-2}\) h\(^{-1}\) (SE = 0.23), corresponding to a calculated carbon flux of 2.75 mmol m\(^{-2}\) h\(^{-1}\). With an annual mean of 12 h 02 min daylight per day in southern New Caledonia, the mean \( P_g \) was 33.03 mmol C m\(^{-2}\) d\(^{-1}\) (SE = 2.67) and the mean annual value was 12.06 mol C m\(^{-2}\) yr\(^{-1}\). For the whole lagoon, the total annual \( P_g \) in carbon was \( 2.38 \times 10^{10} \) mol C yr\(^{-1}\).
Table 2. Mean benthic biomasses (SE are in parentheses) in different bottom types over the year. chl a: chlorophyll a (mg m$^{-2}$); Phaeo: phaeopigments (mg m$^{-2}$); % chl a: percentage of chl a in total pigments; Macroflora: macroflora biomass as ash free dry weight (g m$^{-2}$). Chl a and phaeo refer to the top centimetre of sediment.

<table>
<thead>
<tr>
<th>Bottom type</th>
<th>Chl a</th>
<th>Phaeo</th>
<th>% chl a</th>
<th>Macroflora</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mud deposits</td>
<td>23 (1)</td>
<td>79 (2)</td>
<td>22 (1)</td>
<td>0.78 (0.21)</td>
</tr>
<tr>
<td>Grey sand bottoms</td>
<td>59 (4)</td>
<td>64 (2)</td>
<td>46 (1)</td>
<td>19.88 (3.33)</td>
</tr>
<tr>
<td>White sand bottoms</td>
<td>62 (5)</td>
<td>42 (3)</td>
<td>58 (1)</td>
<td>5.04 (1.20)</td>
</tr>
<tr>
<td>Whole lagoon</td>
<td>47 (2)</td>
<td>65 (2)</td>
<td>40 (1)</td>
<td>10.23 (1.65)</td>
</tr>
</tbody>
</table>

Table 3. Mean values of metabolic parameters in the 3 bottom types of the lagoon. $P_g$: gross photosynthetic production (mmol C m$^{-2}$ d$^{-1}$); R: respiration of the benthos (mmol C m$^{-2}$ d$^{-1}$); $P_g$/R: production/respiration ratio. Standard errors for $P_g$, R and 95% confidence limits for their ratios are in parentheses.

<table>
<thead>
<tr>
<th>Bottom type</th>
<th>$P_g$</th>
<th>R</th>
<th>$P_g$/R</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mud deposits</td>
<td>7.72 (1.66)</td>
<td>19.55 (0.96)</td>
<td>0.39 (0.07-0.72)</td>
</tr>
<tr>
<td>Grey sand</td>
<td>44.96 (4.47)</td>
<td>47.67 (3.48)</td>
<td>0.94 (0.60-1.28)</td>
</tr>
<tr>
<td>White sand</td>
<td>49.85 (4.45)</td>
<td>45.47 (3.86)</td>
<td>1.10 (0.63-1.54)</td>
</tr>
</tbody>
</table>

Benthic respiration

The mean benthic community respiration in oxygen was 1.37 mmol m$^{-2}$ h$^{-1}$ (SE = 0.10) and the carbon equivalent was 1.56 mmol m$^{-2}$ h$^{-1}$. The daily respiration was therefore 37.52 mmol C m$^{-2}$ d$^{-1}$ (SE = 2.02), corresponding to an annual rate of 13.69 mol C m$^{-2}$ yr$^{-1}$. The annual benthic respiration was $2.74 \times 10^{10}$ mol C yr$^{-1}$ (SE = $0.11 \times 10^{10}$) for all sediments of the lagoon. The 3 replicate samples were not significantly different (p = 0.51) but the hypothesis of respiration homogeneity between stations can be rejected (p < 0.001). The distribution was similar to $P_g$ with similar values for the grey and the white sand bottoms and a significantly lower respiration for the muddy bottoms (Table 3). Respiration varied significantly (ANOVA, p < 0.01) over the day (Fig. 3), with peaks at 8 to 9 h and 12 to 13 h, but no clear pattern emerged. $P_g$ exceeded respiration, and $P_g$ was >0, from 8 to 16 h. In contrast, a clear seasonal respiration cycle, similar to that of $P_g$ can be observed (Fig. 4). In fact, the similarity of trends in $P_g$ and respiration during the day and the year suggests a close coupling between these 2 parameters. This hypothesis is
Fig. 5. Relationship between gross primary production and benthic respiration. The number of observations is 170 supported by the significant correlation ($R^2 = 0.66$) between the 2 variables on the whole data set (Fig. 5).

Production/respiration ratio and net photosynthetic production

The $P_g/R$ ratio was 0.88 for the whole lagoon, with 95% confidence limits of 0.63 and 1.12. Mean $P_g/R$ ratios varied monthly from 0.66 to 0.99 (Fig. 6). Lower values were observed from May to September, during the cool season, while the ratios were higher from October to February, indicating the importance of $P_g$ during the warm season. This general feature was the result of different contributions from the 3 bottom types (Table 3). Highest $P_g/R$ were observed for the grey and the white sand bottoms, whereas, respiration greatly exceeded $P_g$ in muddy bottoms. The $P_g/R$ ratio of muddy bottoms was the only one that was significantly different from 1 ($p < 0.05$). If one considers the area covered by each bottom type, 65% of the southwest lagoon of New Caledonia can be regarded as autotrophic, whereas 35% is markedly heterotrophic, leading to an annual carbon deficit greater than $3.5 \times 10^6$ mol C yr$^{-1}$ (Table 4).

Modelling of primary production

Factors affecting primary production, irradiance ($I$) and plant biomass (Macroflora AFDW and chl a) were investigated. Values of regressions between these parameters are given in Table 5. The bulk of gross primary production variance was explained by light (Fig. 7). Contributions of chl a in the first cm of sediment and macrophyte biomass were less important. A general model including the 3 parameters has been adjusted to data. The equation for gross primary production is expressed as:

$$P_g = 5.84 [1 - \exp(-I/319.73)] + 0.04 \text{Macroflora AFD} + 0.01 \text{chl a} - 0.44$$

This model, based upon 150 observations, explains 81% of gross primary production variance (Fig. 8). The relatively small increase in precision, compared to the $P_g$-light relationship ($R^2 = 0.72$), is the result of significant correlation between light and both macrophyte biomass ($r = 0.23, p < 0.01$) and chl a ($r = 0.43, p < 0.01$).

Temperature is often an important causal factor in the variation of primary production in sediment areas (Moriarty et al. 1990). However, in this study, temperature only explains 4% of $P_g$ and less than 1% of $R$ variance and we do not consider this parameter for modelling.

DISCUSSION

Benthic metabolism is seldom assessed on large spatial and temporal scales. Fluxes at the water-sediment interface are important in the carbon budget of coastal environments. This study focused on primary production and respiration, which are the most important processes regulating carbon flow in sediment ecosystems.
Table 5. Values of parameters (SE are in parentheses) calculated by regressions of gross primary production \( (P_g, \text{ in mmol } \text{C m}^{-2} \text{ h}^{-1}) \) on light \( (I, \text{ mmol } \text{m}^{-2} \text{ h}^{-1}) \), Macroflora ashfree dry weight (AFDW) \( \text{g m}^{-2} \) and chl a \( \text{mg m}^{-2} \). N: number of observations; \( R^2 \): proportion of variance explained by the regression

<table>
<thead>
<tr>
<th>Adjusted model</th>
<th>Estimated parameters</th>
<th>N</th>
<th>( R^2 )</th>
</tr>
</thead>
<tbody>
<tr>
<td>( P_g = P_{\text{max}} \left[ 1 - \exp \left(-\frac{I}{I_h} \right) \right] + e )</td>
<td>( P_{\text{max}} = 6.25 \ (0.46), \ I_h = 226.87 \ (45.7), \ e = -0.09 \ (0.21) )</td>
<td>153</td>
<td>0.72</td>
</tr>
<tr>
<td>( P_g = a_1 \text{ chl a} + b_1 )</td>
<td>( a_1 = 0.05 \ (0.01), \ b_1 = 0.36 \ (0.30) )</td>
<td>170</td>
<td>0.30</td>
</tr>
<tr>
<td>( P_g = a_2 \text{ Macroflora AFDW} + b_2 )</td>
<td>( a_2 = 0.05 \ (0.01), \ b_2 = 2.01 \ (0.17) )</td>
<td>167</td>
<td>0.24</td>
</tr>
</tbody>
</table>

interface are mainly derived from small-scale experiments in time or space, and, as already established for coral communities (Pichon 1997), generalisation of such data may hinder a realistic view of the complexity of nature. Benthic biomass and activity display large spatial and temporal heterogeneity; careful sampling is therefore required to derive reliable estimates of global metabolic parameters. Several assumptions are often made. Benthic respiration is generally considered as constant during a 24 h period and only varies with space and season. Primary production, in contrast, depends on irradiance and also varies according to the time of day. Benthic metabolism variation is sometimes considered in space only (Boucher et al. 1994a), or extrapolated using relationships established from partial sampling (Charpy-Roubaud 1988) of the survey area. In this study, we tried to enhance the accuracy of metabolic estimates by using a sampling design which takes into account the main factors influencing benthic metabolism (Rizzo & Wetzel 1985).

The use of metabolic coefficients to calculate carbon fluxes from differences in oxygen concentration is a limitation of our method. The validity of community metabolic coefficients has been discussed by Clavier et al. (1994), and they are assumed to be representative of the south-west lagoon of New Caledonia. CRQ and CPQ are, however, only estimates of the lagoon characteristics and they explain 64 and 91 % of the variance respectively. Variability of metabolic fluxes are thus underestimated in the present study. Other limitations of metabolic estimates are the necessity to extrapolate daytime values over the 24 h period and the fact that it is impossible to directly measure daytime respiration (Sournia 1977, Sorokin 1993). We postulated that day and night community respiration are equivalent at the same location and only vary seasonally and geographically in the lagoon. This hypothesis is obviously a simplification of natural processes (Revsbech et al. 1981, Lindeboom et al. 1985, Epping & Jørgensen 1996) enhanced by the use of DCMU in a clear environment (Boucher et al. 1998). The spatio-temporal random sampling we used allows respiration fluctuation during daytime to be smoothed, and global estimations of \( P_\delta \) are not biased by \( R \) fluctuations. In contrast, the extrapolation of \( R \) over 24 h periods remains uncertain and further information on dark respiration is required. In the lagoon under study, evolution of

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**Fig. 7.** Benthic gross primary production as a function of irradiance measured at the sediment surface

**Fig. 8.** Relationship between observed and predicted values of gross primary production in mmol C m\(^{-2}\) h\(^{-1}\)
temperatures exceed 25°C. In contrast, metabolism is global metabolic estimates result from the specific con-

sequently, as already observed by Shaf-

The estimated $P_g/R$ ratio suggests that the whole south west lagoon of New Caledonia is heterotrophic. This lagoon harbours distinct environments and communities. Consequently, as already observed by Shaffer & Onuf (1983) in a Southern Californian lagoon, global metabolic estimates result from the specific con-

Sediments have received less attention than coral reef/flats (Gattuso et al. 1998). The reef environment is among the most productive ecosystems of the planet (Sorokin 1992); in contrast, adjacent lagoon sediments are regarded as of minor importance. In fact, besides the production per unit of surface, we have to consider the relative size of both systems. The highly productive reef flat corals are generally restricted to narrow reefs lining sometimes vast sediment areas. The south-west lagoon of New Caledonia is a good model of such a distribution, with a ratio of sediment to coral reef surface around 20 (Chardy et al. 1988). The actual reef production is still unknown in New Caledonia. However, if we accept that the order of magnitude matches the ‘standard’ reef flats, average value of ca 210 mol C m$^{-2}$ yr$^{-1}$ (Kinsey 1985a), the sediment $P_g$ value is about 20 times lower. This ratio is in agreement with the review of coastal ecosystem metabolism by Gattuso et al. (1998), with a $P_g$ ranging from 30 to 1369 mol C m$^{-2}$ yr$^{-1}$ and from 8 to 82 mol C m$^{-2}$ yr$^{-1}$ for algal-dominated reef communities and sediments, respectively. At the whole reef ecosystem scale, coral reef and sediment production are therefore equivalent in the studied area. Comparison of $P_g$ with other data is difficult because of different study scales and water depth. The importance of sedimentary processes in whole reef metabolism decreases as a function of depth. For example, sedimentary areas cover 40 to 60% of the Tuahura reef flat surface (French Polynesia) and accounted for 3 to 30% of organic carbon excess production (Boucher et al. 1998). Our estimate is 1.6 times higher than that obtained by Charpy-Roubaud (1988) in the Tikehau lagoon with a similar mean depth (25.2 m). It is interesting to note that estimated production in the New Caledonia lagoon (12 ± 2 mol C m$^{-2}$ yr$^{-1}$, 95% confidence level) is similar to the average annual production (11 ± 5 mol C m$^{-2}$ yr$^{-1}$) calculated from a large set of studies considering various habitat, substrate type and latitude by Rizzo & Wetzel (1985).

The estimated $P_g/R$ ratio suggests that the whole south west lagoon of New Caledonia is heterotrophic. This lagoon harbours distinct environments and communities. Consequently, as already observed by Shaffer & Onuf (1983) in a Southern Californian lagoon, global metabolic estimates result from the specific con-

bolic metabolism is different in the 3 bottom types described in the lagoon (Chardy et al. 1988), with maximum production near the reef and minimum production on muddy sediments. In fact, 2 major bottom groups may be distinguished. Significantly heterotrophic sediments with low metabolic fluxes correspond to mud deposits (see Alongi 1990) near the coast and in the deeper parts of the lagoon. These sediments are located in zones with low residual velocity of tidal currents (Douillet 1998) and are the main site in which particulate organic carbon deposition occurs (Clavier et al. 1995). They also exhibit the lowest macro- and microphyte biomasses. In contrast, grey and white sand bottoms are autotrophic, with $P_g/R$ ratios equal to or greater than 1. They are, therefore, similar to the lagoon of French Frigate Shoals in Hawaii, where the mean $P_g/R$ ratio was 1.09 (Atkinson & Grigg 1984). Grey and white sand bottoms develop where current velocities are relatively high, carrying away small sediment particles though the passes (Douillet 1998). In these communities with high plant biomass, the bulk of primary production is used for local metabolism (Lindeboom & Sandee 1989, Moriarty et al. 1990). Macrophyte biomass is maximal on grey sand bottoms, while chl a concentrations are similar in both communities. We suppose that irradiance in shallower white sand bottoms balances the macrophyte deficit to give similar primary production for the 2 bottom types. Biomass of Sargassaceae may be abundant on a seasonal basis on back reef white sand bottom (Garrigue 1995). These algae are too large to be enclosed in the chambers that we used. Therefore, both biomass and production of white sand communities were underestimated. In brief, we observed that production increased from the relatively deep and turbid coastal waters to the shallow back reef sediments overlaid with clear oceanic waters flowing above the reef. This feature corresponds to a gradient in the ratio of terrestrial to oceanic influence, with allochthonous particulate carbon replacing autochthonous primary production near the shore (Clavier et al. 1995). The gradient is generated by water movement that is the driving mechanism for transporting terrestrial inputs. Hydrodynamics is, therefore, the major factor influencing spatial distribution of benthic metabolism in the lagoon (see also Lewis et al. 1985). It determines, though turbidity, the irradiance on the sea floor, and influences the quantity of fine particles in the sediment, thereby controlling benthic community composition.

In the study area, benthic metabolism shows a clear seasonal trend with about a 3-fold variation throughout the year. Maximum values are observed from November to February, during the warm season, when water temperatures exceed 25°C. In contrast, metabolism is
markedly lower from April to September, during the cool season, with relatively lower temperatures (<25°C). A similar temporal course of metabolism has already been described by Kinsey (1985b), Klumpp & McKinnon (1989), Hansen et al. (1992) and Uthicke & Klumpp (1998) in Australia. Unlike Grant’s (1986) findings in a temperate ecosystem, temperature is not the main cause of seasonal cycle in the lagoon benthic metabolism; it depends, rather, on light availability (Yap et al. 1994) and plant biomass. In New Caledonia, irradiance is maximum between October and December and minimum in June (Anonymous 1981). Similar seasonal variations are observed for micro- and macroflora biomasses and macroflora production (Garrigue 1991) with significantly higher values in January-February. Fluctuation of the New Caledonia climate between a warm and a cool season has, therefore, some relevance to benthic metabolism. The warm season is also the period of maximum runoff associated with the main rainfall season, whereas river water levels are lowest during the cool season. Material carried by rivers towards the lagoon causes a rise in suspended material concentration and a related increase in particulate matter sedimentation (Clavier et al. 1995). The terrestrial influence gradient is most extreme during the warm season when benthic primary production is limited in mud deposits, but minor influence is exerted on back-reef communities due to the large distance (>10 km) across the lagoon. Seasonal fluctuation of light related to solar illumination, or to variations in water turbidity, is likely to reinforce the spatial gradient observed in the lagoon.

If $P_g/R$ varies on a spatial basis, a close relationship is observed between benthic production and respiration at different periods of the year. $P_g$ is essentially related to irradiance and plant biomass, but $R$ is a more complex indication of benthic energy demand. The latter is detritus based (Chardy & Clavier 1988b, Alongi 1989) and fuelled by autochthonous primary production and by allochthonous carbon sedimented from the water column (Hansen et al. 1992). In the south-west lagoon of New Caledonia, organic matter originating from seston has a high C/N ratio indicating its refractory nature (Clavier et al. 1995), in contrast to benthic plant material which is more labile. The total lagoon is, however, heterotrophic and depends on the seasonal supply of particulate organic matter (Clavier et al. 1995), with maximum organic carbon sedimentation from February to June, medium sedimentation in November, and minimum from July to October. The seasonal cycle of $P_g/R$ reflects this input, with decreasing values during the high sedimentation period and a rise after June, with enhancement of the relative amount of energetic autochthonous organic matter. Our data support and generalise the findings of Hansen et al. (1992) in Davies Reef lagoon. A similar variation in $P_g/R$ ratio was also found by Hopkinson (1985) in an heterotrophic temperate environment where organic matter that sustain the metabolic activity varied in quality and quantity.

The present study suggests that soft bottoms of the New Caledonia south-west lagoon are important sites of organic matter production and decomposition. Overall, the lagoon benthos is heterotrophic because of the carbon deficit of coastal and deep muddy bottoms. Carbon production balances respiration demand, however, in about two-thirds of the area of the bottom. This feature suggests that the autotrophic character of coral reef ecosystems (Gattuso et al. 1998) is also characteristic of nearby sediments, indicating both a major importance of regeneration processes and a relative equilibrium of energy exchanges with adjacent ecosystems. Terrestrial influence on such a reef ecosystem leads to a decrease of autochthonous production and a correlated increase of carbon sedimentation leading to benthic heterotrophy.

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