Biological enhancement of estuarine benthic community structure

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ABSTRACT: Benthos in south Texas estuaries are normally concentrated in the top 3 to 4 cm where the sediment is well-oxygenated and less compact. Where larger infauna such as enteropneusts, ophiuroids, or echinoids occur in the sediments bioturbation by these infauna oxygenates and redistributes normally uninhabitated deeper sediments. A natural disturbance to Corpus Christi Bay benthos by these larger infauna could increase density and expansion of infaunal populations into deeper regions of the sediments, as well as enhance colonization by new infaunal species. During a 3.5 yr study of infaunal benthos there was a change in community structure associated with colonization of the soft-bottom habitat by the enteropneust Schizocardium n. sp. that resulted in a species composition atypical for a middle estuary habitat. After 2 yr the enteropneust population disappeared and the diverse and productive soft-bottom community regressed to pre-enteropneust characteristics. Increased aerobic sediments during enteropneust presence may have diminished predicted competition and encouraged development of a more diverse community than would have otherwise existed.

INTRODUCTION

Detection of patterns in community organization, the identification of dominant forces that underlie these patterns, and the more universal application of theories developed from these studies are often major goals in the study of an ecological community. Given the complexity of most ecological systems, however, the identification of causative mechanisms influencing community organization is extremely difficult. Observed community patterns may be the result of responses to spatial or temporal change in the environment (Woodin 1978, 1981), to competition (MacArthur & Levins 1967, Woodin 1976), to predation (e.g. Paine 1969, Connell 1975, Paine 1980), to chance events, or to the product of 2 or more interacting processes (Lewin 1983). Natural disturbances have been observed to be important structuring mechanisms that can alter both the habitat and associated organisms in soft-bottom communities (Woodin 1978, Vinnstein 1979, Thistle 1980, Zajac & Whitlatch 1982, Creed & Coull 1984). Although many of these disturbances are thought to have biological impact (Coull & Palmer 1984) the actual nature of the impact is often not identified.

Corpus Christi Bay, one of 7 major estuaries along the Texas coast of the Gulf of Mexico (Fig. 1), is located in a semi-arid, subtropical climate and usually receives less than 70 cm of rainfall annually. The low rainfall and small fluvial flow into this estuary often result in extended periods of higher salinities, more similar to oceanic conditions. The estuary is shallow, generally less than 5 m deep, and the presence of winds on almost a daily basis results in a great deal of wave scour and resuspension of muds at the sediment surface. Climatological and physical characteristics are thought to greatly influence the biological communities of this estuary.

Estuarine benthic infaunal communities in these south Texas coastal Gulf of Mexico waters have been investigated extensively (e.g. Flint & Younk 1983, Flint & Kalke 1985) in an attempt to elucidate factors responsible for their community organizations. Although long-term environmental fluctuations (Flint 1985) and spatial heterogeneity in sediment characteristics (Flint & Kalke 1985) have been shown to influence estuarine-wide community patterns, subtle biological factors are also thought to significantly affect community organization changes in these soft-bottom sediment habitats. We investigated for the occurrence of natural biological disturbances in the soft-bottom community of Corpus Christi Bay to determine how widespread these disturbances were and to
evaluate their impact on benthic community organization.

**METHODS**

Sampling sites in the Corpus Christi Bay estuary are illustrated in Fig. 1. Because of interesting long-term changes in benthic community structure and availability of historical benthic community characterization at Stn 7 (Holland et al. 1975), data from this sampling site are emphasized. Stn 7 has a water depth of 4.2 m and is characterized by sediments comprised of muddy clays. This collection site is representative of more than 60% of the estuary’s bottom habitat. The sampling intervals for which data from Stn 7 are reported included: July and October 1981; January, April, July, and October 1982; January, April, July, October, and November 1983; September and October 1984. Comparative benthic data from the other sites (Fig. 1) will be cited when necessary to illustrate specific points. More detailed descriptions of benthic communities at these other sites can be obtained from Flint & Younk (1983) and Flint & Kalke (1985).

All sediment sampling at Stn 7 was accomplished using SCUBA. During each sampling interval triplicate 7.5 cm diameter, 30 cm long sediment cores were taken. The cores were capped and brought on board ship. After a minimum of 1 h for sediment stabilization in the cores, vertical Eh profiles were measured every 10 mm through each core using a Radiometer pH/Eh Meter with a platinum/calomel electrode system. The Eh measurements were made to determine the depth of the sediment redox potential discontinuity (RPD) layer (millivolt reading equals zero). This depth was determined through linear interpolation. As described by Rhoads & Boyer (1982), it is believed that the RPD indicates the deepest extent of oxygenated sediments and is thought to be influenced by infaunal bioturbation (McCall 1977, Rhoads et al. 1978).

The walls of each sediment core tube were split to aid in sectioning the sediments. Since most infauna are usually observed in the top 3 cm of Corpus Christi Bay.
sediments and the sediments of this estuary usually turn anaerobic at depths greater than 3 cm (Flint et al. 1982), the cores were sectioned into the following depth strata: 0 to 3, 3 to 10, and 10 to 20 cm sediment depth. Each stratum was placed in a 1 l jar and preserved with 10% seawater formalin containing Rose Bengal (Flint & Holland 1980). In the laboratory, each sediment section was sieved through 0.5 mm mesh and the retained benthic macroinfauna identified and counted. Wet weight biomass was measured individually on the dominant macroinfaunal species and on the entire sample.

During each sampling interval additional sediment cores were obtained for water content and grain size distribution and for determination of sediment organic carbon. Sediment texture parameters were determined according to methods described by Flint & Younk (1983). Percent total organic carbon was measured using a high temperature combustion technique, after acidification, according to methods of Hedges & Parker (1976). During each sampling interval temperature, salinity, and pH (Hydrolab Surveyor 6) were measured along with bottom water dissolved oxygen (YSI Model 57 Oxygen Meter) to characterize temporal variability for the habitat. Statistical analysis methods including analysis of variance and Pearson correlation, as described in Sokal & Rohlf (1969), were used to detect differences in temporal and spatial trends and to investigate for interrelations between species and abiotic factors. Transformation of benthos data did not significantly improve the results of these analyses so raw data were used in all cases for statistical analysis.

RESULTS

The hydrology of the middle estuary area during the study interval exhibited seasonal temperature fluctuations between 12 and 30 °C. Salinities ranged from 23 to 37 ppt but were above 27 ppt for much of the study interval. Dissolved oxygen of the bottom waters was never observed below 4.2 mg l⁻¹ and usually occurred above 6.0 mg l⁻¹. Water column stratification was never observed.

Average sediments in the middle estuary (Stn 7) were characterized by 76% clay, 21% silt, and 2% sand. Sediment water content (to depth of 6 cm) was extremely high with values ranging from 57 to 73% (Fig. 2), and during most observations was greater than 65%. From April through July 1982, sediment water dropped below 60%, which was significantly lower (P < 0.05) than other periods according to analysis of variance. Sediment organic carbon was often greater than 1% by weight (Fig. 2). Initially, in 1982, surface sediments contained greater amounts of organic carbon than deeper sediments. This pattern was reversed later in the study period, and by January 1983 the deeper sediments contained more organic carbon than the surface sediments. In contrast, other benthic sites in the estuary almost always illustrated the vertical distribution for sediment organic carbon shown by the early 1982 samples at Stn 7 (Flint & Kalke 1985).

The most variable characteristic was the depth of the RPD (Fig. 2). From July 1981 through January 1982 an RPD depth of 1.5 to 3.2 cm was similar to other benthic study sites in the estuary (Flint & Kalke 1985). After January 1982 the RPD migrated deeper in the sediments and by April 1983 had reached a maximum of 10 cm depth. With the shallower RPD early in the study, a blackened anaerobic (sulfur) zone was associated with the measured zero mv Eh readings. As the RPD deepened this blackened layer was no longer observed at 2 to 3 cm depth but occurred much deeper in the sediments. After maximum migration of the RPD to 10 cm sediment depth in April 1983, it returned to shallower sediments in July 1983 and remained
between 0.5 and 2.8 cm deep for the rest of the study period (Fig. 2). Although a shallower RPD was measured late in the study, the characteristic black layer was not observed, as had been the case earlier, until the final 2 sample intervals, September and October 1984. In addition, deep burrows in the cores between July and November 1983 always had an extensive area of light-brown sediments surrounding their external walls, indicative of oxygenated sediments in the immediate vicinity of the burrows.

The dominant macroinfauna observed at Stn 7 consisted of polychaetes, with the exception of the enteropneust (acorn worm), *Schizocardium* new species (n. sp.). During the investigation the polychaetes *Polydora caulleryi* and *Mediomastus Californiensis* dominated in total abundance (45 and 31 %), respectively while *Schizocardium* n. sp. dominated in total biomass (91 %). Dominant species represented several different functional groups (feeding strategies) used to characterize benthic populations (Fauchald & Jumars 1979). These groups ranged from tube builders to burrowers and from surface suspension-feeders to subsurface deposit-feeders.

Total macroinfaunal density and biomass were low and mostly confined to the shallow sediments during the first 3 observation periods at the middle estuary site (Fig. 3). Total infauna species richness did not exceed 6 taxa during this interval. Up until January 1982 the RPD depth occurred between 2 and 3 cm sediment depth. By April 1982, the extent of infaunal habitation had increased to at least 20 cm depth with an associated increase in depth of the RPD. Corresponding with these changes was the colonization of the enteropneust *Schizocardium* n. sp. at the study site. After April 1982 both macroinfaunal density and biomass continued to extend deeper into the sediments and species richness exhibited corresponding increases (Fig. 3). Peak biomass was usually observed at depth in the sediment during 1983 and early 1984, and on most occasions peak densities were usually noted at greater than 3 cm sediment depth. Between April 1982 and April 1983, when infaunal peak densities and biomass were at depths greater than 3 cm, the RPD continued to migrate deeper into the sediments (Fig. 3). The collection periods of September and October 1984 revealed that both density and biomass of the community had decreased at the middle estuary site and species richness had also shown a decline. These changes corresponded to the disappearance of the enteropneust population, which occurred during the first half of 1984. Species richness, community biomass, and community density all exhibited significant increases (P < 0.05) after the enteropneust was first observed at the middle estuary site, and these variables all showed significant decreases after the enteropneust's disappearance in 1984.

The macroinfaunal community at Stn 7 was comprised exclusively of spionid polychaetes during the 1981 sampling intervals (Fig. 4) and the most abundant were *Parapinnospio pinna* and *Streblospio benedicti*. Following a decline in these populations after October 1981, the capitellid polychaete *Mediomastus Californiensis* appeared in samples and steadily increased in density. In April 1982 the enteropneust *Schizocardium* n. sp. occurred in extremely high densities. These initial colonizers were small-sized, with low population biomass (Fig. 4), and were likely the result of a massive settlement of this species in the area, since none had been observed previously (e.g. Holland et al. 1975). The *Schizocardium* population

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**Fig. 3.** Vertical distribution of benthic macroinfaunal community density and biomass at Stn 7 in Corpus Christi Bay for 14 sampling intervals during 1981 to 84. Total taxa richness is listed above each plot and the depth of the RPD layer is indicated by the lower border of stippling.
The large increase in biomass of the *Schizocardium* n. sp. population in July 1982 paralleled a drop in *Mediomastus californiensis* and spionid densities (Fig. 4). With a decrease in *Schizocardium* standing crop after this sampling interval, *M. californiensis* densities again increased whereas the spionid densities (*Parapnonospio pinna* and *Streblospio benedicti*) did not. In January 1983 *M. californiensis* reached its maximum population densities. Sediment colonization by *Schizocardium* n. sp. paralleled an increased taxa richness at Stn 7 (Fig. 3) which included the polychaetes *Gypis vittata* and *Polydora caulleryi* in July 1982. In July 1983 *P. caulleryi* became the dominant infaunal species at the site (Fig. 4). As *P. caulleryi* populations became more dense, *M. californiensis* as well as other polychaete populations began to decline (Fig. 5).

The vertical pattern of dominant populations illustrated specific faunal distributions in the Stn 7 soft-bottom community (Fig. 5). Tube building, surface deposit-feeding spionids, such as *Parapnonospio pinna*, inhabited the sediments early in the investigation and confined their populations to the upper 3 cm.
*Mediomastus californiensis*, a burrowing deposit-feeder, was observed in January 1982 samples and occurred to 10 cm sediment depth. With the occurrence of *Schizocardium* n. sp. throughout the sediments in April 1982, both the spionid polychaete populations and *M. californiensis* extended their ranges into deeper sediments and the spionids exhibited increased densities over the previous 2 sampling intervals (Fig. 5). Following establishment of enteropneust biomass concentrations in deeper sediments, the burrowing, subsurface deposit-feeding polychaete, *Gyptis vitata* and the tubiculous polychaete *Polydora caulleryi* colonized the study site sediments and distributed themselves throughout the first 20 cm. *P. caulleryi* was not observed until 1983 and when peak densities for this population occurred, they were always observed in the deeper sediment strata, which was unusual for a tubiculous, surface deposit-feeder. The taxa discussed above continued to inhabit the site and maintained their vertical range over a considerable sediment depth through January 1984. In several 1 m long cores macroinfauna were never observed deeper than 20 cm in the sediments, although *Schizocardium* n. sp. tubes were noted the full extent of the cores. Between January and September 1984 spionids increased their densities to 10 cm sediment depth. *Schizocardium* and *P. caulleryi* populations disappeared from the site and *G. vitata* populations declined and were confined to depths less than 10 cm (Fig. 5).

**DISCUSSION**

We believe that the changes in soft-bottom benthic community structure observed during this investigation were directly attributed to the colonization of an enteropneust population that was first observed in April 1982. In the spring of 1984 a synoptic sampling cruise was conducted in the Corpus Christi Bay estuary to determine if the occurrence of the enteropneust at Stn 7 was an isolated event. All of the areas represented by a circle in Fig. 1 were sampled for benthic invertebrates. As shown by the circles containing an X, a good part of the estuary supported *Schizocardium* populations, indicating that Stn 7 was not unique in this respect.

Total macroinfaunal species richness increased from an average of 4 prior to April 1982 to an average of 15 after April 1982 at Stn 7. Number of taxa declined again in September 1984, after the enteropneust population had disappeared from the site. In contrast, benthic collections taken monthly between October 1972 and May 1975 at the same site (Holland et al. 1975) showed a 3 yr average of 10 taxa inhabiting these sediments and total infaunal density during this study averaged 1038 ind m\(^{-2}\). In the present investigation, total macroinfaunal density prior to April 1982 averaged 2974 ind m\(^{-2}\), whereas after April 1982 average total density was 19 039 ind m\(^{-2}\). After disappearance of the enteropneust population, between January and September 1984, average density declined to 4089 ind m\(^{-2}\). In addition, historical records on this site (1972–75) indicated that *Paraprionospio pinnata, Mediomastus californiensis* and the small deposit-feeding polychaete *Cossura delta* were the dominant populations during earlier years. *Schizocardium* n. sp. and *Polydora caulleryi* were never observed during the earlier 3 yr study. We concluded therefore, that the community structure changes between April 1982 and January 1984 deviated from long-term patterns at this middle estuary site and probably other sites in the estuary supporting enteropneust populations.

In contrast to Stn 7, vertical distributions of the macroinfaunal community at other benthic sites in the estuary between 1981 and 1984 not supporting enteropneusts did not show similar patterns. For example, the vertical and temporal distributions of total density at Stns 2 and 10 (Fig. 1) showed peak densities in the surface sediments, to approximately 3 cm depth (Fig. 6). On numerous occasions fauna were not even present in sediments deeper than 10 cm. At both of these sites the RPD never exceeded 4 cm depth and usually occurred at less than 3 cm. Patterns of total community biomass at these other benthic sites followed similar trends as those observed for density (Flint & Kalke 1985).

We suggest that benthic community organization at Stn 7 was the result of interactions between dominant taxa of the community mediated by the natural disturbance effects of the enteropneust on the community.
Table 1. Two-way analysis of variance results for the investigation of differences in the factors of time and sediment depth strata on the dominant taxa distributions in the benthic community of Corpus Christi Bay. Probability levels for each comparison are shown along with the interaction effects between factors. Degrees of freedom for each factor are also given.

<table>
<thead>
<tr>
<th>Taxon</th>
<th>Time (df: 10)</th>
<th>Depth strata (df: 2)</th>
<th>Interactions (df: 20)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Spionids</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
</tr>
<tr>
<td>Mediomastus californiensis</td>
<td>0.001</td>
<td>0.001</td>
<td>0.001</td>
</tr>
<tr>
<td>Schizocardium n. sp.</td>
<td>0.001</td>
<td>0.216</td>
<td>0.005</td>
</tr>
<tr>
<td>Gyptis vittata</td>
<td>0.009</td>
<td>0.001</td>
<td>0.005</td>
</tr>
<tr>
<td>Polydora caulleryi</td>
<td>0.001</td>
<td>0.001</td>
<td>0.010</td>
</tr>
</tbody>
</table>

Differences in distribution of the dominant fauna occurred from October 1981 to October 1984 (Fig. 4 & 5). Analysis of variance (2-way on factors of time and sediment depth) indicated that temporal differences in taxa distribution were significant (Table 1). Either (1) seasonal environmental characteristics exerted an effect on densities (e.g. reproductive cycles) or (2) species interactions occurred that caused temporal changes in distribution. Significant differences in sediment vertical distribution for all taxa, except Schizocardium n. sp., also were noted (Table 1). Peak densities of each taxa occurred at specific depths in the sediment column (also seen in Fig. 5). For example, spionids, with the exception of Polydora caulleryi, occurred in the 0 to 3 cm depth range while peak Mediomastus californiensis densities occurred in sediments to 10 cm depth. P. caulleryi and Gyptis vittata peak densities were usually observed below 3 cm depth. Interactions between the factors of time and sediment depth were also noted (Table 1) and were the result of vertical changes in peak distributions of the different taxa over the study period (Fig. 5).

Species interactions were suggested above as one potential factor influencing the significant differences observed in temporal distributions of dominant taxa. Changes in the RPD layer (depth of oxygenated sediments) during this investigation (Fig. 2) also appeared to parallel patterns in some of these dominant infauna. Correlations found to be significant between specific taxa distributions and between taxa and the depth of the RPD included the following. There was a positive correlation between depth of the RPD and both Mediomastus californiensis density (0.70) and Schizocardium n. sp. biomass (0.68). There was a negative correlation between depth of the RPD and Polydora caulleryi density (−0.65). Schizocardium n. sp. standing crop was negatively correlated with Parapnonospio pinnata density (−0.81) and positively correlated with M. californiensis density (0.64). P. caulleryi distributions were negatively correlated with both P. pinnata (−0.45) and M. californiensis densities (−0.66). M. californiensis and Gyptis vittata, both burrowing, deposit-feeders, were strongly correlated with one another in their distributions (0.87).

Other benthic study sites in Corpus Christi Bay supported several of the dominant fauna observed at Stn 7 after April 1982 (Flint & Kalke 1985). Where Mediomastus californiensis and Polydora caulleryi occurred together at other sites, a correlation of \( r = -0.56 \) (\( n = 25 \)) on densities between the 2 species was measured, indicating that the 2 taxa did not normally occur in high densities together. Stn 7 dominant infaunal species collected from other benthic study sites not supporting populations of Schizocardium usually occurred in greatest density in shallower sediments. For example, Table 2 illustrates 3 of the Stn 7 dominant polychaetes and their distribution at other benthic sites. Average depth of the RPD layer was shallower and all 3 species limited their peak distributions to less than 10 cm sediment depth. It would appear that these populations were distributed differently in sediments devoid of Schizocardium n. sp. compared to when it was present in the community (e.g. Fig. 5).

A similar pattern was observed when burrowing ophiuroids occurred in sediments. For example, Fig. 7 shows benthic infaunal distribution for 3 similar sediment cores taken in close proximity to one another in Corpus Christi Bay. Two cores exhibited vertical distributions similar to that of the dominant fauna at Stn 7. The other core had a similar distribution, but lacked Mediomastus californiensis and Schizocardium n. sp.

Table 2. Mean and standard error of benthic taxa density distributions at other sampling sites in Corpus Christi Bay where the enteropneust Schizocardium n. sp. did not occur. Species are also species characteristic of the Stn 7 benthic community. RPD depth range of these other sampling sites is also shown.

<table>
<thead>
<tr>
<th>Species</th>
<th>Sediment depth (cm)</th>
<th>Mean density (organisms (0.01) m(^{-2}))</th>
<th>RPD range (cm)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Mediomastus californiensis</td>
<td>0 - 3</td>
<td>26.1 ± 9.2</td>
<td>2 - 3</td>
</tr>
<tr>
<td>(n = 20)</td>
<td>3 - 10</td>
<td>33.2 ± 8.9</td>
<td></td>
</tr>
<tr>
<td>Gyptis vittata</td>
<td>0 - 3</td>
<td>2.1 ± 1.1</td>
<td>2 - 4</td>
</tr>
<tr>
<td>(n = 8)</td>
<td>3 - 10</td>
<td>0.4 ± 0.2</td>
<td></td>
</tr>
<tr>
<td>Polydora caulleryi</td>
<td>0 - 3</td>
<td>128.0 ± 26.2</td>
<td>2 - 4</td>
</tr>
<tr>
<td>(n = 6)</td>
<td>3 - 10</td>
<td>16.5 ± 5.8</td>
<td></td>
</tr>
<tr>
<td></td>
<td>10 - 20</td>
<td>1.3 ± 0.6</td>
<td></td>
</tr>
</tbody>
</table>

\(^1\) Indicates the number of samples that contained the species of interest without populations of Schizocardium n. sp. present.
The distribution of fauna with maximum abundances and diversity in the top 1 to 2 cm of sediment. The range of fauna and greater abundances throughout the top 6 cm in the third core were correlated with the presence of an ophiuroid in the deeper sediments. We believed that natural disturbance caused by sediment reworking activities (bioturbation) of the ophiuroid (Woodley 1975) created a habitat for other infaunal populations to expand deeper into the sediments, and for the community to become more diverse.

We propose the following to explain the results observed during this study. Prior to enteropneust colonization at Stn 7 sediments were characterized by a homogeneous mud to 20 cm, very little faunal burrowing, and a black anoxic layer at approximately 3 cm depth. After April 1982 the sediments were intermeshed with tubes and pockets and enteropneusts often were observed in these pockets or moving through the large burrow network in split cores. The sediments were usually light brown in color and the characteristic blackened layer was absent, suggesting that the sediments were aerobic to considerable depths (Yingst & Rhoads 1980), as verified by the extent of the measured RPD layer (Fig. 2). Corresponding to extensive burrowing, a shift in organic carbon content of various sediment strata occurred while water content decreased initially and the RPD of these sediments shifted deeper than was observed previously (Fig. 2). The depth of the RPD layer was correlated with both increasing density of *Mediomastus californiensis* and increasing biomass of *Schizocardium* n. sp. Although this deeper movement paralleled the colonization by *M. californiensis* (Fig. 4), the RPD depth of 3.2 cm was not uncharacteristic of other benthic sites in Corpus Christi Bay where this species dominated the community (e.g. Table 2). The more dramatic shifts in RPD were associated with the greater biomasses of *Schizocardium* n. sp. that developed at this site in April 1982 (Fig. 5), and implicated this species' impact on the environment.

*Polydora caulleryi* populations usually recruited into the benthic communities of this estuary in spring (Flint et al. 1982). In April 1982 large densities of *Paraprionospio pinnata* (3230 m⁻²) inhabited the surface sediments of Stn 7 along with large densities of *Schizocardium* n. sp. (Fig. 5). This spionid feeds on the surface sediments with wide ranging palps and has been observed to take small animals (authors' unpubl. lab. obs., Dauer 1985). Although *P. caulleryi* colonized the sediments between April and July 1982 (Fig. 4) its density remained small and we believe that the high densities of *P. pinnata* as well as extensive bioturbation activities by *Schizocardium* during this period eliminated many of the recruiting *P. caulleryi*, as predicted by adult-larval interaction theory (Woodin 1976).

There was a strong negative correlation between *Paraprionospio pinnata* and *Schizocardium* n. sp. and by April 1983 *P. pinnata* was less dense (230 m⁻²) than in 1982. *Polydora caulleryi* thus encountered less interference in its recruitment, and was able to establish a significantly denser population in 1983 than it did in the spring of 1982 (Fig. 5). The fact that *P. pinnata* was historically a dense population at this study site (e.g. Holland et al. 1975) may further serve to explain why *P. caulleryi* populations were never abundant until
more recently, coinciding with the colonization of the enteropneust. The *P. caulleryi* population was also able to maintain high densities in the same sediment column that supported high densities of the burrowing polychaete *Mediomastus californiensis* (Fig. 4), contrary to observations at other sites. This co-dominance occurred because a greater extent of oxygenated sediment was able to be inhabited.

With the establishment of dense populations of the tube-building spionid *Polydora caulleryi* at the study site the RPD depth of the sediments began to migrate toward the surface again (Fig. 2). There was a negative correlation between RPD depth and *P. caulleryi* densities. The stabilization of sediments by the high density of *P. caulleryi* tubes, as described for other tube building polychaetes (Woodin & Jackson 1979), could have inhibited the sediment reworking abilities and ventilating activities of dominant subsurface burrowers, causing the RPD depth to decrease. Although both populations were large, there was a negative correlation between *P. caulleryi* and *Mediomastus californiensis* densities, the dominant subsurface, burrowing polychaete population at this study site.

Between January and September 1984 *Schizocardium* populations disappeared from the benthic community at Stn 7 in Corpus Christi Bay. Possible explanations for this disappearance could include changes in sediment habitat or changes in the more general estuarine environment. As discussed above, the sediment habitat showed changes in RPD layer depth (July 1983) which were thought to be affected by a tubicolous polychaete. Sulfate reduction might have increased in the sediments at this site because of greater summer water temperatures, which would serve as an alternative explanation for a shallower anoxic layer. Migration of the RPD to 1 to 2 cm depth, however, was not observed in 1982. A much deeper RPD layer was evident during this period than in July 1983 (Fig. 2) suggesting dynamics were more complex than simply sulfate reduction. Sediment organic content also decreased later in the study (Fig. 2), which could have reflected a decrease in food availability for the enteropneust population. On an estuarine-wide scale, salinities in excess of 36 ppt were observed in the summer of 1984 (Fig. 8). This abrupt change in salinity may have been sufficient to eliminate *Schizocardium* populations. As Fig. 8 illustrates, other periods of *Schizocardium* occurrence in the estuary paralleled intervals of lower salinity. On all occasions, when salinity increased the enteropneust disappeared.

Whatever the cause, the disappearance of the large enteropneust correlated with some major changes in benthic community organization. Most obvious was the re-establishment of *Paraprionospio pinnata* and *Streblospio benedicti*, and overall increase in other spionid populations (Fig. 5). *Polydora caulleryi* disappeared from the habitat and the diversity of the community was much less. These results identified several of the mechanisms at work in structuring benthic communities when a mobile, large-sized organism colonizes a soft-bottom habitat, such as the echiuran described by Stull & Haydock (1982), the ophiuroid illustrated in Fig. 7, the enteropneust described by Thistle (1980), or the enteropneust discussed here. We believe that the presence of the *Schizocardium* n. sp. population in the soft-bottom community, with its sediment reworking abilities, limited population density of one dominant spionid (*P. pinnata*) and created a larger vertical extent to the otherwise confined aerobic conditions in these sediments, allowing other populations to extend into deeper than normal sediments. Numerous studies of soft-bottom communities have implicated competition for space as one organizing force acting on community structure (e.g. Levinton 1977, Woodin 1978, Heck & Orth 1980, Peterson & Andre 1980, Woodin 1981). Differences in vertical space utilization apparently are an effective mechanism for reducing interspecific competition in soft-bottom communities.

![Figure 8](image-url)

*Fig. 8.* Eleven yr record of surface salinity in the Corpus Christi Bay estuary. Periods of occurrence for the enteropneust *Schizocardium* n. sp. are indicated along with total number of individuals collected.
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