

Effect of a Free-Living Marine Nematode (*Diplolaimella chitwoodi*) on Detrital Carbon Mineralization*

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ABSTRACT: Natural densities (10^6 nematodes m^{-2}) of the free-living marine nematode *Diplolaimella chitwoodi* increased maximum rates of carbon mineralization of *Gracilaria* detritus by 300%; lower densities ($10^4 m^{-2}$), by 100%. In the presence of nematodes increased mineralization was observed for 2 different particle sizes of *Gracilaria* detritus, 63 to 100 and 250 to 308 μm . In the absence of nematodes, there were only slight differences in mineralization rates between the 2 particle sizes. Nematodes also increased the rate of mineralization of *Spartina* detritus but to a lesser degree (50%). Meiofaunal nematodes can significantly affect benthic carbon flow by such an effect on the rate of detritus mineralization.

INTRODUCTION

Detritus is a major energy source for most marine benthic systems. Salt marshes and submerged angiosperms are very productive but little of this organic material is consumed while the plants are alive (e.g. Teal, 1962; Fenchel, 1977). Macroalgae, although more susceptible to direct grazing, may also be a significant source of dead organic matter. While sedimentation inputs to the benthos are becoming known, the fate of organic matter once it enters benthic systems is poorly understood. Rates of oxygen uptake by the benthos have been measured (Pamatmat, 1977) but these shed little light on what mechanisms are important in the transfer of detrital carbon to benthos.

Microbes, meio- and macrofauna assimilate detrital carbon and subsequently mineralize organic C to CO_2 . The proportion of carbon incorporated into tissue relative to that mineralized to CO_2 will affect benthic production because, once mineralized, detrital carbon is essentially lost as an energy source for the benthos (Christian and Wetzel, 1978). This represents a com-

petition between 'wasteful' (with respect to macrofaunal production) mineralization and 'useful' assimilation/production. The ratio of production to organic carbon mineralization (P:O; Tenore and Hanson, 1980) is a benefit:cost ratio for the system. Increasing the rate of mineralization of detritus relative to the assimilation will tend to decrease benthic production.

Microbes probably mineralize the major portion of detrital carbon sedimenting to the benthos but direct measurements have not been made. Partitioning of benthic oxygen uptake implies that bacteria are responsible for the majority of oxygen consumption (\approx mineralization) (e.g. Smith et al., 1972). Therefore, changes in microbial metabolic rates affect mineralization rates. In particular, increasing microbial activity will increase losses of organic carbon thereby decreasing the amount of food available to other consumers. Conversely, it can be argued that microbial activity converts refractory detritus into nutritionally available food (e.g. Ward and Cummins, 1979) but the general importance of microbial biomass in detritivore nutrition is questionable (Cammen, 1980).

Microbial activity, and therefore detritus mineralization, is generally increased by warmer temperatures and greater availability of oxygen and nitrogen (Haines and Hanson, 1979). Overall, detritus derived from vascular plants decomposes more slowly than detritus derived from seaweeds due to lower N content

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and a larger proportion of 'recalcitrant' material such as lignin and humic acids in vascular plants (Godshalk and Wetzel, 1978). For any particular detritus type and conditions, microbial activity is greater for detritus of smaller particle size (Gosselink and Kriby, 1974) probably due to the greater surface/volume ratio for smaller particles. Temporal aspects of detritus dynamics may confuse these generalities because microbial activity reaches its maximum rate sooner for 'available' detritus such as seaweeds and relatively later for vascular plant detritus (Tenore, 1977; Tenore and Hanson, 1980). While detrital type, N-content and particle size directly affect microbial activity, other organisms may directly and/or indirectly influence rates of decomposition.

Macrofauna generally increase rates of detritus degradation (Fenchel, 1970; Hargrave, 1970; Briggs et al., 1979). Macrofauna may indirectly influence microbial activity via bioturbation, thus increasing oxygen availability and bringing up nutrients from deeper sediment layers. Aside from bioturbation, some macrofauna physically shred large fragments of detritus thereby decreasing particle size and increasing rates of decomposition (Robertson and Mann, 1980). Apart from these physical changes associated with macrofaunal grazing, it has been postulated that grazing, by removing senescent cells, may maintain the microbial community in log phase of growth (Cooper, 1973; Fenchel, 1977). Direct confirmation of this mechanism for increasing microbial activity is lacking but it is a reasonable explanation for increased decomposition in the presence of macrofauna or grazers. Conversely, direct macrofaunal respiration of detrital carbon is probably a minor component of total C flow in benthic systems. For example, the presence or absence of *Capitella*, a non-bioturbator, did not significantly affect oxidation rates in flow-through microcosms (Tenore, 1977).

While several macrofaunal species are known to affect rates of decomposition, there are very few similar studies with meiofauna. Grazing by protozoans increased bacterial activity and detritus mineralization (Barsdate et al., 1974). Conversely, ciliates did not affect mineralization of *Gracilaria* detritus but did increase incorporation of detritus by *Nereis succinea* (Briggs et al., 1979). Nematodes increase orthophosphate regeneration (Tietjen, 1980), which implies that nematodes might enhance mineralization of detritus.

Meiofauna probably are an important component of benthic energy flow because of their high abundance and metabolic rates (Coull and Bell, 1979). Moreover, meiofauna function at temporal and spatial scales similar to the microbes, allowing close coupling between microbes and meiofauna. Their rapid turnover rates would enable meiofauna to respond quickly

to changes in microbial biomass or activity. By the same token, this coupling implies that meiofauna grazing or physical activity may exert a significant positive or negative effect on microbial biomass and turnover. Also, meiofauna are small enough to allow them to directly and perhaps selectively feed on specific microbes, thus changing the composition of the microbial community. For these reasons, one might expect meiofauna in general, and particularly nematodes, the dominant taxon, to affect microbial activity and thereby affect detrital decomposition.

Therefore, ^{14}C tracer techniques were used to explore the possible effect of the meiobenthic nematode *Diplolaimella chitwoodi* on rates of detrital carbon mineralization. We examined the effect of (1) nematode density on mineralization rate of *Gracilaria* and *Spartina* detritus; (2) nematodes on mineralization rates of 2 particle sizes.

METHODS

Detritus was homogeneously labelled with ^{14}C by growing *Gracilaria* or *Spartina* in the presence of $^{14}\text{CO}_2$ (Tenore et al., 1979). Dried plant material was ground in a Wiley mill to pass a 250 μm screen. Immediately before use, detritus was aged for 3 d in 100 ml of autoclaved seawater g^{-1} dry wt detritus. Two ml of filtered (3 μm) seawater from nematode cultures were added to provide a microbial inoculum. Detritus for the particle size experiment was aged, freeze-dried, and ground with mortar and pestle. The size classes, 63 to 100 μm and 250 to 308 μm , were separated by sieving.

The nematode *Diplolaimella chitwoodi* (Monhysteridae) was maintained on Gerbers Mixed Cereal (Pabulum) in 75 cm^2 tissue culture flasks. Only cultures in late log phase of growth were used in experiments. Nematodes were separated from the Pabulum by a modified Baermann extraction (Sikora et al., 1977) resulting in a clean suspension of worms. Various volumes of this worm suspension were added to experimental chambers to give the required density of nematodes.

Flow-through chambers (Tenore et al., 1977) with an area of 32 cm^2 and a volume of 203 cm^3 were layered with fine sand (2 cm) and 300 mg dry wt (50 mg dry wt for particle size experiment) ^{14}C labelled detritus. Worms were added and the chambers sealed. Autoclaved, aerated seawater was pumped through at ~ 4 mls min^{-1} (~ 28 turnovers d^{-1}). All experiments were conducted at $20^\circ \pm 1^\circ$ and in the dark for 6 d. Two replicates were run for each treatment.

Mineralization rate (organic ^{14}C mineralized to $^{14}\text{CO}_2$) was calculated from the concentration of $^{14}\text{CO}_2$

in effluent water, specific activity of the detritus and flow rate through the chamber (Tenore et al., 1979). The amount of detritus mineralized over the course of an experiment was determined by plotting mineralization as a function of time and integrating the area under the curve.

The Wilcoxon test was used to examine differences between treatments. With 2 replicates per treatment, the lowest P-value tabulated is 0.13 for the case where both replicates of Treatment A are greater than both replicates of Treatment B.

RESULTS

Nematode Density

Mineralization rates in control chambers (no nematodes present) were $\sim 25 \mu\text{g C detritus mineralized h}^{-1}$; $< 6\%$ of the initial dry wt was mineralized over 1 wk (Fig. 1). With 0.01×10^6 nematodes m^{-2} , maximum mineralization rates were

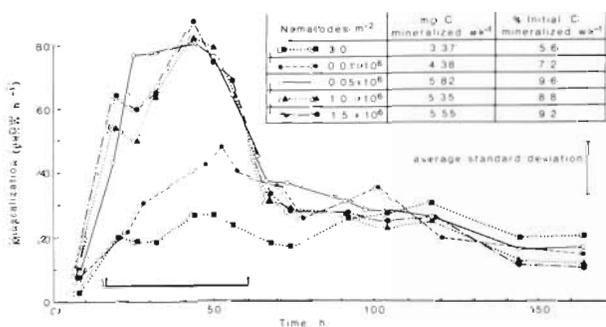


Fig. 1. Carbon mineralization in chambers containing 300 mg dry wt *Gracilaria* detritus and various densities of *Diplolaimella chitwoodi*. Each line is the mean for 2 chambers. Brackets enclose time-span where treatments are significantly ($P = 0.13$) different

increased by 100% relative to controls and the initial dry wt increased to 7.2%. With natural densities (0.05 to 1.5×10^6 nematodes m^{-2}) maximum mineralization was $> 300\%$ greater than controls and the amount mineralized w^{-1} was 8.8–9.6%.

Particle Size

Presence of nematodes ($0.5 \times 10^6 \text{ m}^{-2}$) increased the maximum mineralization rate and the amount mineralized w^{-1} relative to controls for both large (250 to 308 μm) and small (63 to 100 μm) particle sizes of *Gracilaria* detritus (Fig. 2). The presence of nematodes caused greater differences in mineralization rate than

Effect of Nematodes on Mineralization of Different Particle Sizes

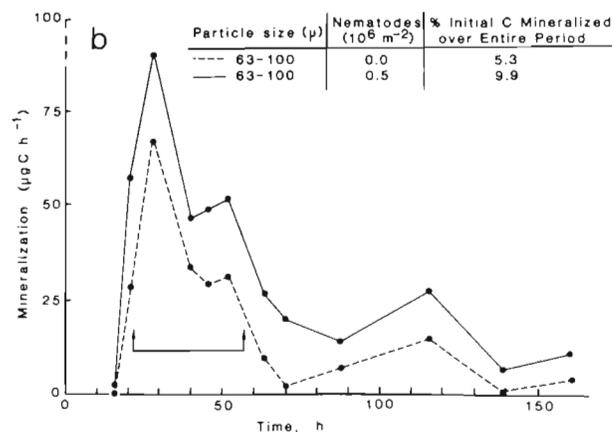
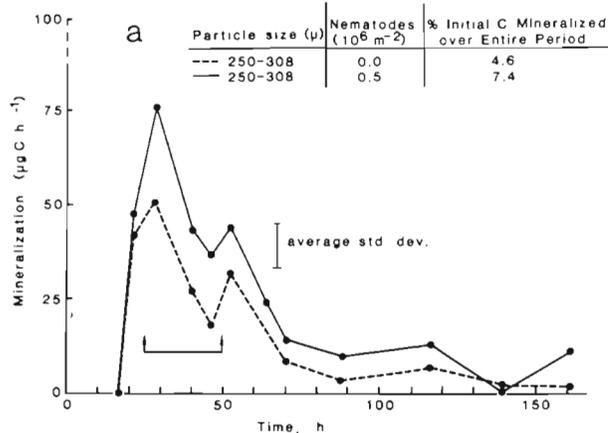


Fig. 2. Carbon mineralization of 50 mg dry wt of 2 particle sizes of *Gracilaria* detritus with and without *Diplolaimella chitwoodi*. See also legend of Fig. 1

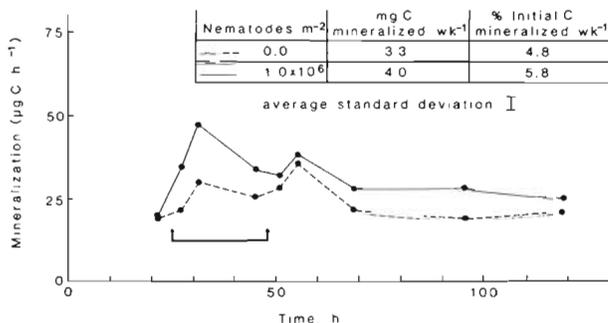


Fig. 3. Carbon mineralization of 300 mg dry wt of *Spartina* detritus with and without *Diplolaimella chitwoodi*. See also legend of Fig. 1

the differences in particle size. Presence of nematodes increased maximum mineralization roughly 25% while mineralization of large vs. small particles was only slightly different. Amount of detritus mineralized

w^{-1} was 25% higher with nematodes compared to no nematodes for both particle sizes. The amount mineralized w^{-1} was roughly the same for large vs. small particles in control chambers.

Spartina detritus

The presence of 1.0×10^6 nematodes m^{-2} increased maximum rate of mineralization of *Spartina* detritus by 50% relative to no nematodes and amount mineralized w^{-1} by 20% (Fig. 3). The initial dry wt of *Spartina* detritus was 300 mg, and so is comparable to the results of 'nematode density' experiments. Rates of mineralization in control chambers were roughly the same for the 2 types of detritus (cf. Tenore et al., 1979; 0% N-supplement). Mineralization of *Spartina* in the presence of nematodes was slower than mineralization of *Gracilaria* in the presence of nematodes.

DISCUSSION

Nematode densities in detritus-based systems average 10^6 individuals m^{-2} (Coull and Bell, 1979) with 'deposit-feeders' such as *Diplolaimella* being the numerically dominant trophic type (Levy and Coull, 1977). The results show that these densities potentially double the rate of detrital carbon mineralization. Caution must be used when extrapolating from laboratory results to field conditions. For example, mixed communities of meiofauna might not exhibit the same results or mineralization may be limited by lack of oxygen. Even so, the potential magnitude of this nematode effect is a significant fraction of benthic carbon flow.

Presence of nematodes increased mineralization rate of 2 particle sizes, indicating that the observed increase was not specific to 1 particle size. Moreover, nematodes had a greater effect (25% increase) on mineralization rate than the difference in particle size. Although it is dogmatic that small particles decompose faster than large particles (Godshalk and Wetzel, 1978), there are no direct unequivocal measurements of weight loss from small vs. large particles in this size range. Therefore, nematodes must be considered at least as important as particle size differences when examining factors that affect rates of detritus mineralization.

Detrital source (derived from vascular vs. non-vascular plants) is also considered significant in controlling rates of decomposition (Tenore, 1977). There was no difference in mineralization rate of *Gracilaria* vs. *Spartina* detritus in the absence of nematodes (Figs. 1 and 3; see also Tenore et al., 1979). Nematodes increased

mineralization of *Spartina* detritus but to a lesser degree than they increased mineralization of *Gracilaria*. *Gracilaria* detritus is of higher nutritional quality (Tenore, 1981; Hanson, submitted; Findlay, submitted) and therefore it is predictable that nematodes would exert a greater influence (whatever the mechanism) on *Gracilaria* mineralization relative to *Spartina*. Apparently, a larger proportion of *Gracilaria* detritus is immediately 'available' (i.e. is susceptible to rapid mineralization) whereas *Spartina* detritus must be aged for months before it reaches maximum 'availability' (Tenore and Hanson, 1980). Rates of mineralization in the absence of nematodes were roughly the same for *Spartina* and *Gracilaria* implying that this labile fraction in *Gracilaria* is not necessarily mineralized faster than the labile fraction in *Spartina*. These experiments used 3 d aged detritus; over longer time spans peak mineralization of *Gracilaria* occurs after 40 d of aging whereas peak mineralization of *Spartina* occurs after 150 d of aging (Tenore and Hanson, 1980). However, changes in extrinsic regulatory factors such as presence of nematodes or added nitrogen (Tenore et al., 1979) do lead to immediate differences in mineralization of *Gracilaria* vs. *Spartina*. The important point is that immediately 'available' detritus (*Gracilaria*) is much more susceptible to stimulation of mineralization by, for example, nematodes or nitrogen than is 'refractory' detritus (*Spartina*).

Temporal aspects of detrital decomposition may be complex and these experimental systems are not intended to represent an equilibrium or steady-state condition. It appears that nematodes change the time-course of carbon mineralization. Apparently, there is a labile pool of carbon which is rapidly mineralized and depleted in the presence of nematodes (Fig. 1). In the absence of nematodes, this same pool is mineralized at a slower rate.

The actual mechanism by which nematodes enhance mineralization is unknown. Nematode-respired CO_2 represents less than 4% of the carbon mineralization of *Gracilaria* in chambers with 10^6 nematodes m^{-2} (assuming *Diplolaimella chitwoodi* respire at twice the maximum rate reported in Gerlach, 1971). Nematode respiration is a larger proportion of total mineralization of *Spartina* but still does not account for the increase in CO_2 production in the presence of nematodes. Apparently, nematodes increase microbial activity, thereby increasing carbon mineralization.

The observed increase in carbon mineralization due to the presence of nematodes will decrease the standing stock of detrital carbon available to other detritivores. Although this in itself should decrease production, concurrent changes in microbial biomass and/or turnover may in fact increase the nutritional quality of

the detritus (Ward and Cummins, 1979). For detritus such as seaweeds, which are directly available to consumers, it is quite likely that increased mineralization would decrease overall production. For instance, the population growth rate of *Diplolaimella chitwoodi* was slower at low vs. high ratios of seaweed detritus but population growth was unaffected by low vs. high ratios of *Spartina* detritus (Findlay, submitted).

Microbes, in the presence of nematodes, mineralize seaweed detritus faster than *Spartina* detritus (Figs. 1 and 3). Therefore, competition may occur between microbes and nematodes for *Gracilaria* detritus. Conversely, for more refractory detrituses such as *Spartina* or *Zostera*, the effect of nematodes on the microbial community may lead to an enhancement of microbial and/or consumer production. In either case, nematodes probably play a significant role in regulating the direction and magnitude of detrital carbon flow.

There is a long-standing discussion over the significance of nematodes and other meiofauna as food for higher trophic levels (e.g. Sikora, 1977) vs. their importance as regenerators of otherwise unavailable nutrients (Marshall, 1970; Lasserre et al., 1976; Tietjen, 1980). The present results demonstrate: *Diplolaimella chitwoodi* can play a central role in regulating carbon mineralization (microbial activity) that will indirectly affect other aspects of benthic carbon flow. The significance of nematodes in natural systems probably lies in their coupling and interactions with the microbial community. Consequently, because microbes are the most important component of marine benthos, factors such as nematodes, that affect microbial abundance and/or activity will play a larger role in benthic carbon flow than would be expected on the basis of nematode biomass or production.

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