Effects of increased atmospheric CO₂ on sea urchin early development

H. Kurihara*, Y. Shirayama

Seto Marine Biological Laboratory, Kyoto University, 459 Shirahama, Nishimuro, Wakayama, 649-2211, Japan

ABSTRACT: Increased carbon dioxide (CO_2) concentration in the atmosphere will change the balance of the components of carbonate chemistry and reduce the pH at the ocean surface. Here, we report the effects of increased CO_2 concentration on the early development of the sea urchins *Hemicentrotus pulcherrimus* and *Echinometra mathaei*. We examined the fertilization, early cleavage, and pluteus larval stage to evaluate the impact of elevated CO_2 concentration on fertilization rate, cleavage rate, developmental speed, and pluteus larval morphology. Furthermore, we compared the effects of CO_2 and HCl at the same pH in an attempt to elucidate any differences between the two. We found that fertilization rate, cleavage rate, developmental speed, and pluteus larval size all tended to decrease with increasing CO_2 concentration. Furthermore, CO_2 -seawater had a more severe effect than HCl-seawater on the fertilization rate. By contrast, the effects on cleavage rate, developmental speed, and pluteus larval morphology were similar for CO_2 - and HCl-seawater. Our results suggest that both decreased pH and altered carbonate chemistry affect the early development and life history of marine animals, implying that increased seawater CO_2 concentration will seriously alter marine ecosystems. The effects of CO_2 itself on marine organisms therefore requires further clarification.

KEY WORDS: $CO_2 \cdot pH \cdot Sea$ urchin · Early development · Biological impact · Global change

- Resale or republication not permitted without written consent of the publisher

INTRODUCTION

Increasing atmospheric carbon dioxide (CO₂) concentration (Houghton et al. 1992, Keeling & Whorf 1994) is predicted to lead to a rise in CO₂ concentrations at the ocean surface, due to continuous gas exchange between the air and seawater (Takahashi et al. 1997). By the end of this century, the concentration of CO₂ in the air is expected to rise to 750 ppm, which is about twice the present 365 ppm, and the CO₂ concentration in seawater will attain the same level (IPCC 2001). This will in turn increase the concentration of dissolved inorganic carbon (DIC) (i.e. $CO_{2(aq)}$, HCO_3 -, and CO₃²⁻), which will then lead to a decrease in pH at the ocean surface. Current models predict that the pH of surface seawater will drop from 8.0 to 7.8 by the year 2100 (Wolf-Gladrow et al. 1999). Between 2100 and 2200, the atmospheric CO₂ concentration is expected to increase to the range of 1500 to 2100 ppm resulting in a further decline in pH value (Sundquist 1993, Wigley et al. 1996).

To modulate the increasing atmospheric CO_2 , some have proposed the sequestration of CO_2 captured from power plants, which would be piped into the deep ocean (Nozaki 1991, Liro et al. 1992, Ohsumi 1995, Auerbach et al. 1996). The CO_2 concentration close to the injection point is, however, expected to be as high as 20 000 ppm and the pH there less than 5.8 (Herzog et al. 1996). Furthermore, models of CO_2 disposal suggest that the volume of acidic (pH <7.0) seawater could be hundreds of cubic kilometers (Caulfield et al. 1997a, Huesemann et al. 2002). These changes in seawater chemistry might have a serious, direct impact on marine organisms.

The effects of lowered pH have been studied experimentally in many species of marine organism using strong acids, such as hydrochloric acid (HCl) and sulfuric acid (H $_2$ SO $_4$) (Calabrese & Davis 1966, Kuwatani & Nishii 1969, Rose et al. 1977, Bamber 1987, 1990, Yamada & Ikeda 1999). These studies have shown that acidification below pH 7.0 has harmful consequences (such as decreased survival rate) in many organisms.

Otherwise, in most studies, pH fluctuations in the range of 0.5 to 1.0 unit have little influence on survival (Knutzen 1981). However, these studies are not sufficient to evaluate the environmental impact of raised atmospheric CO₂ on marine ecosystems for 3 reasons. First, studies have mostly examined survival rates at low pH; little attention has been paid to other biological processes, such as reproduction, growth rates, and morphological abnormalities, or to the effects of small changes in pH (<1.0 unit). Second, most studies have focused on the impact on adult organisms. The impact on early developmental stages should also be evaluated, since these stages are more sensitive to environmental stress. Third, increased CO₂ concentration may have other effects on marine organisms besides those caused by reducing pH. In this study, we therefore examined the effect of a small fluctuation in pH resulting from an increase in atmospheric CO₂ concentrations.

Watanabe et al. (2001) compared the survival rates of copepods raised in seawater acidified to the same degree with either CO₂ or HCl, and found that CO₂ had a greater impact. The impact of raised CO₂ concentrations in seawater has been examined mainly for coral reef organisms and phytoplankton. Elevated CO₂ concentrations and low pH are reported to reduce the calcite production rate of coccolithophores and corals (Riebesell et al. 1993, 2000, Gattuso et al. 1998). Thornton & Shirayama (2001) found that both sea urchins and snails raised in seawater containing elevated CO₂ concentrations had lower survival and growth rates after a 10 wk exposure.

We have focused here on the impacts on early development to evaluate sub-lethal effects of increased CO_2 concentrations. We studied 2 sea urchin species, Hemicentrotus pulcherrimus and Echinometra mathaei, as sea urchins are frequently indicator species of environmental alteration (Kobayashi 1971, 1995). The effects on the early development of the sea urchins were examined by studying (1) fertilization rate, (2) cleavage rate, (3) developmental speed, and (4) pluteus morphology. Furthermore, we compared the effects of CO_2 and HCl at the same pH to examine the influence of a change in DIC concentration on the early development of marine invertebrates.

MATERIALS AND METHODS

Test animals. The sea urchins *Hemicentrotus pulcherrimus* and *Echinometra mathaei* were collected from the subtidal rocky shore near the Seto Marine Biological Laboratory, Wakayama, Japan. Due to differences in breeding season, experiments using *H. pulcherrimus* were conducted from January to March at 14°C,

whereas those using *E. mathaei* were carried out from June to October at 24°C. To select the best quality gametes, only batches that had fertilization rates exceeding 95% in control seawater were used. *H. pulcherrimus* was used in all experiments, while *E. mathaei* was used only in the fertilization bioassay.

Preparing seawater for culture. Seawater filtered through a Millipore filter was placed in six 1 l beakers in each experiment, and the CO₂ concentration of the seawater was controlled by aerating the water with air containing different concentrations of CO₂ (CO₂-seawater). The flow rates of air and CO2 were adjusted using a flow meter (Kofloc). The control was aerated with 500 ml min⁻¹ air, which contained 365 ppm CO₂. To alter the CO₂ concentration, the water was also aerated with 0.25, 0.5, 1.0, 2.5 and 5.0 ml min^{-1} CO₂, respectively. The CO₂ concentrations of the gas were as follows: control, +500, +1000, +2000, +5000, and +10000 ppm. The seawater was aerated by experimental gases for more than 2 h, until the partial pressure of CO₂ concentrations (P_{CO2}) in the seawater became equilibrated with the gas (Table 1). The pH of the seawater was measured using a pH meter (Toledo) just before performing the experiment (Table 1). To prepare HCl-seawater, filtered seawater was placed in another six 1 l beakers and 1 N HCl was added to each to match the pH of the seawater in the respective CO₂ beakers.

Bioassay. To induce spawning, 0.1 M acetylcholine chloride was injected into the perivisceral cavity of individual sea urchins. Eggs were rinsed several times with filtered seawater and maintained in a 200 ml beaker at 4°C until use. Sperm were collected with a pipette and kept at 4°C as dry sperm. The dry sperm were diluted with filtered seawater just before the experiment. To determine the effect of CO_2 on the early development of sea urchins, eggs of *Hemicentrotus pulcherrimus* or *Echinometra mathaei* were placed in 6 Petri dishes $(10 \times 1.5 \text{ cm})$ filled with CO_2 -seawater with CO_2 concentrations of 365 (control), +500, +1000, +2000, +5000, and +10000 ppm. After 1 min, a drop of sperm suspension was added to each dish. Fifteen min

Table 1. The $P_{\rm CO2}$ and pH of seawater aerated using 6 different ${\rm CO_2}$ concentrations before the experiment and the pH of the seawater incubated after 3 d (ANOVA p < 0.01). Means \pm SD

CO ₂ condition	PCO ₂ (kPa)	H. pulcherrimus pH (before) pH (after)		E. mathaei pH (before)
Control	0.036	8.01 ± 0.10	7.97 ± 0.19	8.11 ± 0.03 7.82 ± 0.09 7.71 ± 0.09 7.33 ± 0.05 7.12 ± 0.02 6.79 ± 0.02
+ 500	0.086	7.77 ± 0.07	7.64 ± 0.15	
+ 1000	0.136	7.61 ± 0.05	7.57 ± 0.10	
+ 2000	0.236	7.38 ± 0.04	7.33 ± 0.05	
+ 5000	0.536	7.03 ± 0.07	7.10 ± 0.09	
+ 10000	1.036	6.83 ± 0.05	6.88 ± 0.09	

after insemination, about 500 embryos were transferred from each dish into three 20 ml vials and 100 embryos were transferred into a 50 ml vial.

To determine the fertilization rate, 10% buffered-formalin seawater solution was added to one of the 20 ml vials for each CO_2 concentration. The remaining 20 ml and 50 ml vials were filled with seawater at the respective CO_2 concentrations and any air spaces were removed to prevent exchange of CO_2 with the ambient air. One 20 ml vial for each CO_2 concentration was incubated for 105 min and the other for 210 min after insemination, to determine cleavage rate and developmental speed. The 50 ml vials were incubated for 3 d to determine pluteus larval morphology. After 3 d of incubation, the pH of the seawater was measured to verify that the concentration of CO_2 had not changed (Table 1), and one drop of concentrated formalin was then added to each bottle to halt development.

The same procedure as described above was also used for HCl-seawater instead of CO_2 -seawater.

Fertilization. To study the effect on the fertilization rate, 300 eggs were randomly sampled from each 20 ml vial, cultured for 15 min after insemination, and the numbers of fertilized eggs were counted under a microscope. Fertilization was defined as the presence of a fertilization membrane. For *H. pulcherrimus*, 6 and 4 batches were used for the CO₂- and HCl-seawater experiments, respectively. For *E. mathaei*, 3 and 5 batches were used.

Cleavage. To study the effect of CO_2 concentration on cleavage rate and developmental speed, 300 randomly sampled *Hemicentrotus pulcherrimus* eggs, that had been incubated for 105 or 210 min after insemination in the 20 ml vials, were observed under a microscope and the stage of each embryo was recorded. Six and 4 batches were used for the CO_2 -and HCl-seawater experiments, respectively.

Pluteus larvae. To determine the effects of CO_2 concentration on pluteus larval morphology, 10 *Hemicentrotus pulcherrimus* larvae were sampled randomly from each 50 ml vial and mounted on glass slides. The mounted larvae were photographed and the overall length (OL), postoral arm length (PL), and body length (BL) of each larva were measured under a microscope using an ocular micrometer (Fig. 1). Six and 3 batches were used for the CO_2 - and HCl-seawater experiments, respectively.

Data analysis. The means and standard deviations were computed for each experiment. Spearman's rank correlation coefficient (R_S) was calculated to compare the effects of lowering pH with CO_2 or HCl on fertilization rate, OL, PL, and BL. The Tukey-Kramer test was performed to evaluate effects of CO_2 and HCl on fertilization rate, cleavage rate and the lengths of OL, PL, and BL. Two-way repeated ANOVA was used to com-

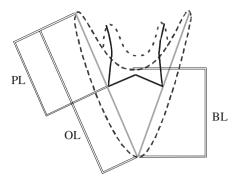


Fig. 1. Hemicentrotus pulcherrimus. Three morphometric measurements of the 4-armed pluteus larvae: (1) overall length (OL), (2) postoral arm length (PL), and (3) body length (BL)

pare the results of CO_2 -seawater and HCl-seawater experiments. One-way analysis of variance (ANOVA) was used to compare the effects of CO_2 and HCl on the percentage of embryos at each developmental stage in the cleavage analysis.

RESULTS

Relationship between CO₂ concentration and pH

The pH of the seawater decreased with increasing CO_2 concentration (Table 1; ANOVA, p < 0.001). A slight difference in pH was observed between the seawater used for *Hemicentrotus pulcherrimus* and *Echinometra mathaei*, but the difference was not significant. The pH of CO_2 -seawater incubated for 3 d changed within the range of 0.04 to 0.13 (Table 1). For HCl-seawater, the pH changed within the same range as CO_2 -seawater during the experiment (data not shown).

Effects on fertilization

The fertilization rate of *Hemicentrotus pulcherrimus* tended to decrease with pH in both CO_2 -seawater ($R_S = 0.74$, p < 0.01) and HCl-seawater ($R_S = 0.72$, p < 0.01; Fig. 2). However, the effect differed between the 2 acids. With CO_2 -seawater, the fertilization rates decreased slightly compared to the control when the concentration of CO_2 exceeded +500 ppm (pH 7.8), with increasingly severe effects as the pH decreased. At +10000 ppm (pH 6.8), the fertilization rate was almost half the control value, a significant difference (Tukey-Kramer; Fig. 2). By contrast, the fertilization rate in HCl-seawater only decreased to 77 % at pH 6.8. However, this was still significantly lower than in the control (Tukey-Kramer; Fig. 2). The fertilization rate was always lower in CO_2 -seawater than in HCl-

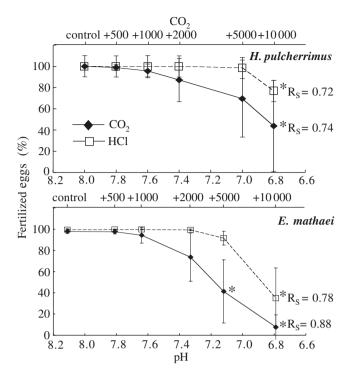


Fig. 2. Hemicentrotus pulcherrimus and Echinometra mathaei. Percentage of eggs fertilized in CO_2 -seawater and HClseawater at different pHs. For H. pulcherrimus, each value is the average of 6 experiments for CO_2 -seawater and 4 for HCl-seawater. For E. mathaei, each value is the average of 3 experiments for CO_2 -seawater and 5 for HCl-seawater. Error bars = SD. R_S = Spearman's rank correlation coefficient. *Means were significantly different from control (Tukey-Kramer)

seawater at the same pH, although the differences were not significant (ANOVA).

As in Hemicentrotus pulcherrimus, the fertilization rates of Echinometra mathaei tended to decrease with pH in both CO_2 -seawater ($R_S = 0.88$, p < 0.01, Fig. 2) and HCl-seawater ($R_S = 0.78$, p < 0.01; Fig. 2). The effects of pH on fertilization rate also differed between CO₂- and HCl-seawater (Fig.2). In CO₂-seawater, the fertilization rate of E. mathaei tended to drop linearly with increasing CO₂ concentration and was less than 10% at +10000 ppm (pH 6.8). The fertilization rates at +5000 (pH 7.1) and +10000 ppm were significantly different from that of the control (Tukey-Kramer; Fig. 2). Furthermore, E. mathaei seemed to be affected more severely than H. pulcherrimus, especially at higher CO₂ concentrations (Fig. 2). In HCl-seawater, the fertilization rates of E. mathaei decreased only after the pH dropped below 7.3. At pH 7.1, the fertilization rate decreased slightly to 94% and at pH 6.8, it decreased drastically to 34% (Fig. 2). The fertilization rate at pH 6.8 was significantly different from that of the control (Tukey-Kramer, Fig. 2). There was also a significant difference between the effects of CO_2 - and HCl-seawater on the fertilization rate in this species (ANOVA, p < 0.05).

Effects on early cleavage

The percentage of fertilized *Hemicentrotus pulcher*rimus eggs (fert) that had not yet cleaved (1 cell fert⁻¹) and of embryos that had already completed their first cleavage (2 cell fert⁻¹) 105 min after insemination differed significantly with pH in both CO₂-seawater (Fig. 3A; ANOVA, p < 0.01) and HCl-seawater (Fig. 3B; ANOVA, p < 0.01). In both CO₂- and HCl-seawater, the percentage of 1 cell fert⁻¹ increased as the pH of the seawater decreased, whereas that of 2 cell fert⁻¹ decreased with decreasing pH (Fig. 3). The percentage of 1 cell fert⁻¹ was significantly higher at +10000 ppm (pH 6.8) than in the control in CO₂-seawater (Fig. 3A; Tukey-Kramer) and at pH 7.0 and 6.8 in HCl-seawater (Fig. 3B; Tukey-Kramer). There was no significant difference in the percentage of either 1 or 2 cell fert⁻¹ between CO₂- and HCl-seawater at the same pH.

At 210 min after insemination, some *Hemicentrotus* pulcherrimus fertilized eggs reached the 8-cell stage while some remained at the 1-cell stage, and

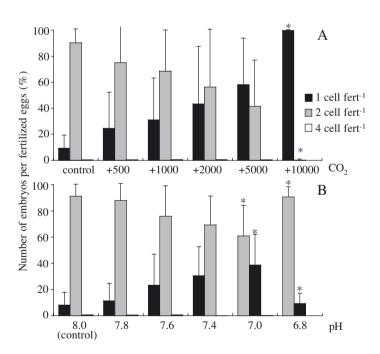


Fig. 3. Hemicentrotus pulcherrimus. Percentage of 1-, 2-, and 4-cell stage embryos in fertilized eggs cultured for 105 min after insemination in (A) CO₂-seawater and (B) HCl-seawater at different pHs. Each value is the average of 6 experiments for CO₂-seawater and 4 for HCl-seawater. Error bars = SD.*Means were significantly different from control (Tukey-Kramer).

abnormal embryos were observed in both CO_2 -seawater (Fig. 4A) and HCl-seawater (Fig. 4B). The percentage of earlier-stage embryos, such as 1-cell or 2-cell stage embryos, tended to increase as the pH decreased in both CO_2 -seawater (Fig. 4A; ANOVA, p < 0.01) and HCl-seawater (Fig. 4B; ANOVA, p < 0.01). The percentage of 1- and 2-cell stage embryos was significantly higher at pH 6.8 (CO_2 =+10000 ppm) than in the control in both CO_2 - and HCl-seawater (Fig. 4; Tukey-Kramer).

The percentage of later-stage embryos, such as 4-and 8-cell stage embryos, differed significantly with pH in both $\rm CO_2$ -seawater (Fig. 4A; ANOVA, p < 0.05) and HCl-seawater (Fig. 4B; ANOVA, p < 0.05). Although the average percentage of 8 cell fert⁻¹ tended to decrease with increasing pH, it did not differ significantly from the control at all pHs in both $\rm CO_2$ - and HCl-seawater (Fig. 4A,B; Tukey-Kramer). There were no significant differences between the percentages of embryos at each stage in $\rm CO_2$ -seawater and HCl-seawater for all pHs.

Effects on pluteus larvae

All 3 parameters measured (Fig. 1), overall length (OL), post-oral arm length (PL), and body length (BL),

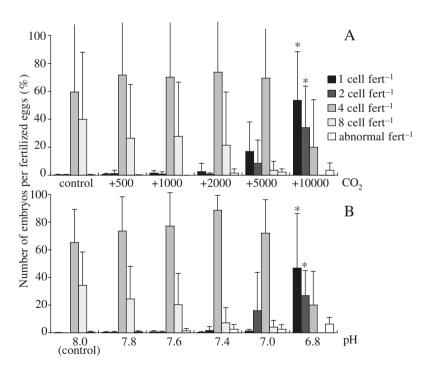


Fig. 4. Hemicentrotus pulcherrimus. Percentage of 1-, 2-, 4-, and 8-cell stage embryos in fertilized eggs cultured for 210 min after insemination in (A) $\rm CO_2$ -seawater and (B) HCl-seawater at different pHs. Each value is the average of 6 experiments for $\rm CO_2$ -seawater and 4 for HCl-seawater. Error bars = SD.* Means were significantly different from control (Tukey-Kramer)

tended to decrease with pH in both CO_2 -seawater (OL: $R_S=0.82$, p<0.01; PL: $R_S=0.75$, p<0.01; BL: $R_S=0.83$, p<0.01) and HCl-seawater (OL: $R_S=0.92$, p<0.01; PL: $R_S=0.85$, p<0.01; BL: $R_S=0.94$, p<0.01; Fig. 5). In CO_2 -seawater, all OL, PL, and BL were significantly different at all pH values from those in the control (Tukey-Kramer, Fig. 5). In HCl-seawater, OL and BL were significantly shorter at pH 7.6, 7.4 and 7.0 than in the control, whereas PL differed significantly only at pH 7.4 and 7.0 (Tukey-Kramer, Fig. 5). The data for pH 6.8 ($CO_2=+10\,000$ ppm) are not shown because in this treatment virtually no embryos developed to the pluteus stage. There were no significant differences between the effects of CO_2 - and HCl-seawater on OL, PL, and BL under these conditions (ANOVA, Fig. 5).

The larvae grown at pH 7.8 ($\rm CO_2 = +500~ppm$) were smaller (Fig. 6B,H) than the controls (Fig. 6A,G), although of similar shape. In contrast, larvae grown in seawater more acidic than pH 7.6 ($\rm CO_2 > +1000~ppm$), were smaller and tended to be trapeziform rather than of the normal triangular form (Fig. 6C,D,E,I,J & K). Larvae grown at pH 6.8 ($\rm CO_2 = +10\,000~ppm$) were extremely abnormal, lacking arms and we were unable to find any spicules (Fig. 6F,L).

DISCUSSION

Effects of elevated CO₂ on the early development of sea urchin embryos

In this study, we have assessed the effects of high CO₂ concentrations on the early development of sea urchin embryos. All of the parameters studied, i.e. fertilization rates, cleavage rates, and size of pluteus larvae, tended to decrease with increased CO₂ concentration. We also compared the effects of CO₂ and HCl at the same pH to evaluate any differences between the two. The 2 compounds had different effects on fertilization rate, whereas their effects on cleavage rate, developmental speed, and pluteus larval morphology were the same. It was especially noteworthy that, in contrast to CO₂-seawater, there seemed to be a threshold in HCl-seawater where the fertilization rate only began to decrease when the pH was <7.1-7.0 (Fig. 2). Christen et al. (1983) pointed out that the low pH (7.0) caused by strong acids prevents sperm motility. In our experiment, however, we observed many sperm swimming around eggs that failed to be fertilized in both CO2- and HCl-seawater (data not shown). In another species of sea urchin, Anthocidaris crassipina, polyspermy

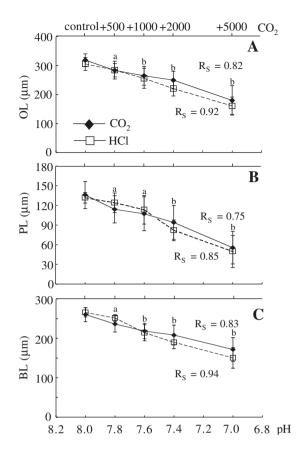


Fig. 5. Hemicentrotus pulcherrimus. Three morphometric measurements for 4-armed pluteus larvae cultured for 3 d at different pHs in CO_2 -seawater and HCl-seawater. (A) OL; (B) PL; (C) BL. Each value is the average of 5 experiments for CO_2 -seawater and 3 for HCl-seawater. Error bars = SD. R_S = Spearmen's rank correlation coefficient. Means were significantly different from the control (Tukey-Kramer) (a) only CO_2 -seawater and (b) both CO_2 - and HCl-seawater

was induced when eggs were fertilized at pH 7.0 (Kobayashi 1973), although we could not find any such eggs. Gregg & Metz (1976) indicated that low pH (6.5) caused by strong acids inhibited the acrosomal reaction of sea urchin sperm, which could explain the lower fertilization rates observed in our study. Since the fertilization rate decreased at a much higher pH in CO₂-seawater, however, the increased CO₂ concentration itself may have an inhibitory effect. Elevated CO₂ concentration leads to an increase in total inorganic carbon (TC = CO_{2(aq)} + HCO₃⁻ + CO₃²⁻) in seawater, and such changes in CO₂ chemistry may affect the fertilization mechanism, leading to reduced fertilization.

When fertilized *Hemicentrotus pulcherrimus* eggs were kept in CO_2 -seawater for 105 and 210 min, the percentage of embryos at later stages tended to decrease with increasing CO_2 concentration (Figs. 4 & 5). Furthermore, some of the fertilized eggs kept at CO_2 concentrations higher than +5000 ppm for

210 min did not cleave at all (Fig. 5). These results reveal 2 effects of CO2 on the development of sea urchin embryos: a dose-dependent developmental delay and a decline in the cleavage rate at CO₂ levels above +5000 ppm (i.e. a decrease in the number of zygotes capable of cleaving at concentrations higher than +5000 ppm CO₂). In chick embryos, the rates of thymidine-³H incorporation into DNA and cell multiplication both show a positive correlation with pH (Rubin 1971). If the same relationships exist for sea urchins, development could be delayed in acidic CO₂and HCl-seawater. The reduction in intracellular pH after fertilization, caused by incubating zygotes in strong acids, leads to a reduction of protein synthesis (Grainger et al. 1979). When Sphaerechinus granularis sea urchin embryos were reared in seawater acidified with HCl or H₂SO₄, mitotic abnormalities were induced at pH <6.5 (Pagano et al. 1985 a,b, Cipollaro et al. 1986). Such impacts on protein synthesis and on mitotic activity are thought to lead to decreased growth and cleavage rates.

When embryos were cultured for 3 d in CO₂- or HClseawater, the morphology and size of the pluteus larvae were both affected in a pH-dependent manner (Figs. 5 & 6). Embryos cultured at +10 000 ppm (pH 6.8) completely lacked a skeleton (Fig. 6F,L). The larval skeleton of the sea urchin is composed of calcium carbonate (Okazaki & Inoue 1976); therefore, we postulate that these decreases in the length of pluteus larvae result from decreased calcium carbonate assimilation efficiency due to the low pH, as has been observed in phytoplankton (Riebesell et al. 1993, 2000) and coral (Gattuso et al. 1998). The size decrease may also be the result of retarded development. Furthermore, low pH is also thought to induce morphological abnormalities. Pagano et al. (1985a) used HCl to study the effects of low pH on embryos of the sea urchin Paracentrotus lividus, and demonstrated that larval morphology is critically affected at pH 7.5.

Ecological impacts of increased CO₂ concentration in the seawater

Our results reveal that increased CO_2 concentrations in seawater affect various stages in the early life history of sea urchins. A delay in the developmental rate increases mortality of embryos during the planktonic period due to predation (Vance 1973a,b, Christiansen & Fenchel 1979, Emlet et al. 1986). The typical triangular form and long arms of the pluteus larvae are important for swimming and feeding (Strathmann 1971, Emlet 1983). The clearance rate of food in suspension increases with increasing length of the ciliated band (Strathmann 1971, Hart 1991), which is known to

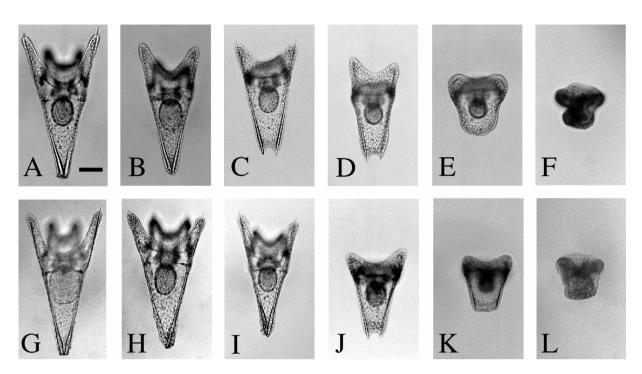


Fig. 6. Hemicentrotus pulcherrimus. Morphology of 4-armed pluteus larvae cultured for 3 d in (A–F) CO_2 -seawater and (G–L) HCl-seawater. (A & G) control (pH 8.0); (B & H) +500 ppm (pH 7.8); (C & I) +1000 ppm (pH 7.6); (D & J) +2000 ppm (pH 7.4); (E & K) +5000 ppm (pH 7.0); (F & L) +10000 ppm (pH 6.8). Scale bar = 50 μ m

increase allometrically with respect to overall length, arm length, and body length (Fenaux et al. 1988, McEdward 1984). Therefore, abnormal morphology and shortening of OL, PL and BL may reduce swimming and feeding capability and, in turn, survival. The skeleton is presumed to serve supportive or protective functions (Emlet 1982, Pennington & Emlet 1986); therefore, changes in skeletogenesis may also reduce the survival rate. All the effects of raised CO₂ concentration observed in this study would have a negative impact on the survival of sea urchin embryos in their early life history.

Previous studies of the impact of acidification on marine organisms focused on the survival rate under conditions prepared using strong acids, such as HCl or H₂SO₄ (Calabrese & Davis 1966, Kuwatani & Nishii 1969, Rose et al. 1977, Bamber 1987, 1990, Yamada & Ikeda 1999). As a result, a fluctuation in pH of less than 1.0 was thought to have only a minor impact (Knutzen 1981). However, we have shown that the effect of CO₂seawater is stronger than that of HCl-seawater at the same pH. Watanabe et al. (2001) compared the survival rates of copepods raised in seawater acidified to the same degree using CO₂ or HCl, and also found that CO₂ had a greater impact. Furthermore, even a small change in pH affected the early development of sea urchins, and the effect tended to become more severe as the CO2 concentration increased. Some aspects of

the mechanisms affecting early development are common among marine invertebrates. Therefore, increasing atmospheric CO_2 concentration may affect the early stages of many marine invertebrates in the same way that it affects sea urchins. This is especially serious for species with a high mortality rate in the early stages of development (Legendre & Rivkin 2002). It is also speculated that animals that use calcium carbonate in their bodies such as corals, mollusks, echinoderms, and foraminiferans will be highly affected. We have evaluated only the direct effects of CO_2 and low pH in this study. Low pH is, however, known to increase the impacts of certain inorganic pollutants (Pagano et al. 1986) which may further exacerbate these effects.

Recently, a large number of experimental studies evaluating the environmental impact of the oceanic disposal of $\rm CO_2$ have been carried out (Auerbach et al. 1997, Haugan 1997, Omori et al. 1998). Most of these studies considered how lowered pH would affect the percentage mortality of adult organisms (Herzog et al. 1996, Adams et al. 1997, Caulfield et al. 1997b, Takeuchi et al. 1997, Yamada & Ikeda 1999). We have shown that acidic (pH <7.0) seawater drastically affects the early development of sea urchins, resulting in a reduced fertilization rate, blocked cleavage capacity and a complete absence of skeleton in pluteus larvae.

In conclusion, to anticipate the impact of increased CO_2 concentrations on marine ecosystems accurately, comprehensive information on the effects of CO_2 fluctuations on the entire life cycles of various marine organisms must be gathered.

Acknowledgements. We gratefully acknowledge S. Amemiya and S. Kinjo for their critical and helpful comments on the manuscript. We thank G. Itani and R. Shimamura for their advice on statistical analysis. We are grateful for the assistance of all other members of Seto Marine Biological Laboratory. We also thank 3 anonymous reviewers for suggestions that improve the quality of this paper.

LITERATURE CITED

- Adams EE, Caulfield JA, Herzog HJ, Auerbach DI (1997) Impacts of reduced pH from ocean disposal: sensitivity of zooplankton mortality to model parameters. Waste Mgmt 17:375–380
- Auerbach D, Caulfield J, Herzog H, Adams E (1996) Environmental impacts of ocean disposal of CO₂. In: Battaglia B, Beardmore JA (eds) Ocean storage of carbon dioxide. Workshop 2-Environmental impact. Plenum Press, New York, p 41–55
- Auerbach DI, Caulfield JA, Adams EE, Herzog HJ (1997) Impacts of ocean disposal on marine life: I. A toxicological assessment integrating constant-concentration laboratory assay data with variable-concentration field exposure. Environ Model Assess 2:333–343
- Bamber RN (1987) The effects of acids seawater on young carpet-shell clams *Venerupis decussata* (L.) (Mollusca: Veneracea). J Exp Mar Biol Ecol 108:241–260
- Bamber RN (1990) The effects of acidic seawater on three species of lamellibranch mollusc. J Exp Mar Biol Ecol 143: 181–191
- Calabrese A, Davis HC (1966) The pH tolerance of embryos and larvae of Mercenaria mercenaria and Crassostrea virginica. Biol Bull (Woods Hole)131:427–436
- Caulfield JA, Adams EE, Auerbach DI, Herzog HJ (1997a) Impacts of ocean disposal on marine life: II. Probabilistic plume exposure model used with a time-varying doseresponse analysis. Environ Model Assess 2:345–353
- Caulfield JA, Auerbach DI, Adams EE, Herzog HJ (1997b) Near field impacts of reduced pH from ocean CO₂ disposal. Energy Convers Mgmt 38 Suppl: 343–348
- Christen R, Schackmann RW, Shapiro BM (1983) Metabolism of sea urchin sperm. Interrelationships between intracellular pH, ATPase activity, and mitochondrial respiration. J Biol Chem 258:5392–5399
- Christiansen FB, Fenchel TM (1979) Evolution of marine invertebrate reproductive patterns. Theor Popul Biol 16: 267–282
- Cipollaro M, Corsale G, Esposito A, Ragucci E, Staiano N, Giordano GG, Pagano G (1986) Sublethal pH decrease may cause genetic damage to eukaryotic cell: A study on sea urchins and *Salmonella typhimurium*. Teratog Carcinog Mutagen 6:275–287
- Emlet RB (1982) Echinoderm calcite: a mechanical analysis from larval spicules. Biol Bull (Woods Hole) 163:264–275
- Emlet RB (1983) Locomotion, drag, and the rigid skeleton of larvae echinoderms. Biol Bull (Woods Hole) 164:433–445
- Emlet RB, McEdward LR, Strathmann RR (1986) Echinoderm larval ecology viewed from the egg. In: Jangoux M,

- Lawrence LM (eds) Echinoderm studies, Vol 2. Balkema AA, Rotterdam, p 55-136
- Fenaux L, Cellario C, Rassoulzadegan F (1988) Sensitivity of different morphological stages of the larva of *Paracentrotus lividus* (Lamarck) to quantity and quality of food. In: Burke RD, Mladenov PV, Lambert P, Parsley RL (eds) Echinoderm biology. AA Balkema, Rotterdam, p 259–266
- Gattuso J-P, Frankignoulle M, Bourge I, Romaine S, Buddemeier RW (1998) Effect of calcium carbonate saturation of seawater in coral calcification. Global Planet Change 18: 37–46
- Grainger JL, Winkler MM, Steinhardt RA (1979) Intracellular pH controls protein synthesis rate in the sea urchin egg and early embryo. Dev Biol 68:396–406
- Gregg KW, Metz CB (1976) Physiological parameters of the sea urchin acrosomal reaction. Biol Reprod 14:405–411
- Haugan PM (1997) Impacts on the marine environment from direct and indirect ocean storage of CO_2 . Waste Mgmt 17: 323-327
- Hart MW (1991) Particle captures and the method of suspension feeding by echinoderm larvae. Biol Bull (Woods Hole) 180:12–27
- Herzog HJ, Adams EE, Auerbach D, Caulfield J (1996) Environmental impacts of ocean disposal of CO₂. Energy Convers Mgnt 37 (6–8):999–1005
- Houghton JT, Callander BA, Varney SK (1992) Climate Change: the supplementary report to the IPCC Scientific. Cambridge University Press, Cambridge
- Huesemann MH, Skillman AD, Crecelius EA (2002) The inhibition of marine nitrification by ocean disposal of carbon dioxide. Mar Poll Bull 44:142–148
- Keeling CD, Whorf TP (1994) Atmospheric CO_2 records from sites in the SIO air sampling network. In: Boden TA, Kaiser DP, Sepanki RJ, Stoss FW (eds) Trends '93: a compendium of data on global change. Carbon Dioxide Information Analysis Center, Oak Ridge National Laboratory, Oak Ridge, TN, p 16–26
- Knutzen J (1981) Effects of decreased pH on marine organisms. Mar Pollut Bull 12:25–29
- Kobayashi N (1971) Fertilized sea urchin eggs as an indicatory material for marine pollution bioassay, preliminary experiments. Publ Seto Mar Biol Lab 18(6):379–406
- Kobayashi N (1973) Studies on the effects of some agents on fertilized sea urchin eggs, as part of the bases for marine pollution bioassay. Publ Seto Mar Biol Lab 21 (2):109–114
- Kobayashi N (1995) Bioassay data for marine pollution using echinoderms. In: Cheremisinoff PN (ed) Environmental control technology, Vol 9. Gulf Publ. Houston, p 536–609
- Kuwatani Y, Nishii T (1969) Effects of pH of culture water on the growth of the Japanese pearl oyster. Bull Jpn Soc Sci Fish 35 no4:342–350
- Legendre L, Rivkin RB (2002) Pelagic food webs: Responses to environmental processes and effects on the environment. Ecol Res 17:143–149
- Liro CR, Adams EE, Herzog HJ (1992) Modeling the release of ${\rm CO_2}$ in the deep ocean. Energy Convers Mgnt 33: 667-674
- McEdward LR (1984) Morphometric and metabolic analysis of the growth and form of an echinopluteus. J Exp Mar Biol Ecol 82:259–287
- Nozaki Y (1991) Feasibility of dumping fossil-fuel CO_2 into the deep ocean. J Oceanogr Soc Jpn 47:221–225
- Ohsumi T (1995) CO_2 storage options in the deep sea. Mar Tech Soc J 20:58–66
- Okazaki K, Inoue S (1976) Crystal property of the larval sea urchin spicule. Dev Growth Differ 18:413–434
- Omori M, Norman CP, Ikeda T (1998) Oceanic disposal of

- CO₂: Potential effects on deep-sea plankton and micronekton. A review. Plank Biol Ecol 45 (2):87–99
- Pagano G, Cipollaro M, Corsale G, Esposito A, Ragucci E, Giordano GG (1985a) pH-Induced changes in mitotic and developmental patterns in sea urchin embryogenesis. I. Exposure of embryos. Teratog Carcinog Mutagen 5: 101–112
- Pagano G, Cipollaro M, Corsale G, Esposito A, Ragucci E, Giordano GG (1985b) pH-Induced changes in mitotic and developmental patterns in sea urchin embryogenesis. II. Exposure of sperm. Teratog Carcinog Mutagen 5:113–121
- Pagano G, Cipollaro M, Corsale G, Esposito A, Ragucci E, Giordano GG, Trieff NM (1986) The sea urchin: Bioassay for the assessment of damage from environmental contaminants. In: Cairns J Jr (ed) Community toxicity testing. Philadelphia: ASTM STP-920, American Society for Testing and Materials, p 66–92
- Pennington JT, Emlet RB (1986) Ontogenatic and diel vertical migration of a planktonic echinoid larva, *Dendraster excentricus* (Eschscholtz): occurrence, causes, and probable consequences. J Exp Mar Biol Ecol 104:69–95
- Riebesell U, Wolf-Gladrow DA, Smetacek V (1993) Carbon dioxide limitation of marine phytoplankton growth rates. Nature 361:249–251
- Riebesell U, Zondervan I, Rost B, Tortell PD, Richard EZ, Morel FMM (2000) Reduced calcification of marine plankton in response to increased atmospheric ${\rm CO_2}$. Nature 407: 364-367
- Rose CD, Williams WG, Hollister TA, Parrish PR (1977) Method for determining acute toxicity of an acid waste and limiting permissible concentrations at boundaries of an oceanic mixing zone. Envir Sci Technol 11(4):367–371 Rubin H (1971) pH and population density in the regulation of

Editorial responsibility: Otto Kinne (Editor), Oldendorf/Luhe, Germany

- animal cell multiplication. J Cell Biol 51:686-702
- Strathmann RR (1971) The feeding behavior of planktotrophic echinoderm larvae: mechanisms, regulation, and rates of suspension-feeding. J Exp Mar Biol Ecol 6:109–160
- Sundquist ET (1993) The global carbon dioxide budget. Science 259:934–941
- Takahashi T, Feely RA, Weiss RF, Wanninkhof RH, Chipman DW, Sutherland SC, Timothy TT (1997) Global air-sea flux of CO₂ difference. PNAS 94:8292–8299
- Takeuchi K, Fujioka Y, Kawasaki Y, Shirayama Y (1997) Impacts of high ${\rm CO}_2$ concentration of on marine organisms: a modification of ${\rm CO}_2$ ocean sequestration. Energy Convers Mgmt 38 Suppl: 337–341
- Thornton H, Shirayama Y (2001) III-1 Effects on benthic organisms. In: CO_2 ocean sequestration and its biological impacts. Bull Jpn Soc Sci Fish 67(4):756–757
- Vance RR (1973a) On reproductive strategies in marine benthic invertebrates. Am Nat 107:339–352
- Vance RR (1973b) More on reproductive strategies in marine benthic invertebrates. Am Nat 107:353–361
- Watanabe Y, Ishida H, Yamaguchi A, Ishizaka J (2001) III-5 Effects of high concentration of CO₂ on deep-sea plankton. In: CO₂ ocean sequestration and its biological impacts. Bull Jpn Soc Sci Fish 67(4):764–765
- Wigley TML, Richels R, Edmonds JA (1996) Economic and environmental choices in the stabilization of atmospheric CO_2 concentrations. Nature 379:240–243
- Wolf-Gladrow DA, Riebesell U, Khardt SB, Bijma J (1999) Direct effects of CO_2 concentration on growth and isotopic composition of marine plankton. Tellus 51B: 461–476
- Yamada Y, Ikeda T (1999) Acute toxicity of lowered pH to some oceanic zooplankton. Plankton Biol Ecol 46 (1): 62-67

Submitted: July 17, 2003; Accepted: February 24, 2004 Proofs received from author(s): June 2, 2004