Ecology of a novel *Synechococcus* clade occurring in dense populations in saline Antarctic lakes

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ABSTRACT: The seasonal distribution and abundance of *Synechococcus*-like morphotypes was investigated in meromictic lakes and coastal areas of the Vestfold Hills, Antarctica. Populations were monitored by flow cytometry utilising phycoerythrin content and small cell size to distinguish the cells from other phytoplankton. In Ace Lake, the *Synechococcus* bloom commenced in September at the water temperature minimum and peaked in late November. Populations (up to $8 \times 10^6$ cells ml$^{-1}$) were maximally stratified at a depth of 11 m, corresponding to waters which were supersaturated with oxygen, high in phosphate and which received $>5 \mu$mol photons m$^{-2}$ s$^{-1}$ of light. At this depth, salinity (30 g kg$^{-1}$) was constant throughout the year and temperature ranged from 4.5°C in October to 10.5°C in February. In late November, high numbers of *Synechococcus* cells also occurred in the moderate salinity water bodies Lake Abraxas and Pendant Lake (salinity 16.5 to 31.0 g kg$^{-1}$), with populations highly stratified in Lake Abraxas (up to $1.5 \times 10^7$ cells ml$^{-1}$, temperature 8.0°C, salinity 20.3 g kg$^{-1}$) but less so in the cooler waters of Pendant Lake (max. $1.5 \times 10^7$ cells ml$^{-1}$, temperature 0.1 to 1.1°C, salinity 31.0 g kg$^{-1}$). *Synechococcus* populations did not occur in brackish, coastal marine or hypersaline water bodies in the Vestfold Hills. Populations appear to be controlled primarily by temperature and to a lesser extent by light availability and grazing. Characterization of non-axenic cultures indicated that the Antarctic lake *Synechococcus* populations were similar to other polar picocyanobacteria in terms of cardinal growth temperatures (minimum, optimum, maximum: $T_{\text{min}}$ -17.0°C, $T_{\text{opt}}$ 19.8°C, $T_{\text{max}}$ 29.5°C) and slow growth. Related only peripherally to *Synechococcus* sp. Cluster 5.2 (Marine Cluster B), the Antarctic strains represent a unique and highly adapted clade in the stable water columns of some saline Antarctic lakes.

KEY WORDS: *Synechococcus* · Antarctica · Meromictic lakes · Marine ecosystems

INTRODUCTION

The contribution of picocyanobacteria to biomass and primary production in marine waters is highly variable, but follows certain general patterns. Marine *Synechococcus* and *Prochlorococcus* spp. make relatively large contributions to primary production in oligotrophic open ocean waters. Populations occur down to the deep chlorophyll maximum (Waterbury 1992); however, vertical distribution patterns appear extremely variable. In the tropical oceans, picocyanobacteria contribute as much as 50% to the primary production, but this dramatically decreases in the polar regions, where small eukaryotic phytoplankters dominate (Gradinger & Lenz 1995). Populations vary from $10^2$ cells ml$^{-1}$ or less in polar seas (Marchant et al. 1987) to up to $10^9$ cells ml$^{-1}$ in low latitude, coastal zones (Agawin et al. 2003). *Synechococcus*-like picocyanobacteria also can occur in transitional and freshwater ecosystems including meromictic lakes, and populations can sometimes reach high levels, especially if eutrophic (Hawes 1990, Maeda et al. 1992). Clear
similarity in the distribution and ecological responses have been observed for picocyanobacteria in Arctic and Antarctic freshwater bodies (Vezina & Vincent 1997). Temperature appears to be a major limiting factor for polar ocean picocyanobacterial abundance (Wilmotte et al. 2002). Light, nutrients, salinity, grazing pressure and viral lysis also play a role to varying extents in controlling *Synechococcus* populations (Waterbury 1992, Bertrand & Vincent 1994, Mackey et al. 2002, Moore et al. 2002, Sullivan et al. 2003, Larsen et al. 2004).

The genus *Synechococcus* currently includes a series of polyphyletic lineages within the phylum Cyanobacteria (Herdman et al. 2001, Scanlan & West 2002, Fuller et al. 2003). A substantial proportion of marine *Synechococcus* spp. cluster with *Prochlorococcus* spp. as a distinct clade containing various subgroups (Rocap et al. 2002, Fuller et al. 2003), e.g. *Synechococcus* Group 5.1 (Marine Cluster A) and Group 5.2 (Marine Cluster B). Characteristics that are used to categorize *Synechococcus* spp. such as photosynthetic pigmentation, salinity and nitrogen requirements and motility (Herdman et al. 2001), vary considerably within these and other clades (Rocap et al. 2002, Fuller et al. 2003), indicating that only closely related strains share the same ecotypic and physiological adaptations (Casamayor et al. 2002, Ernst et al. 2003). Much the same pattern of phenotypic variability occurs within various groups of the Bacteria in general, indicating that niche adaptation can often be sub-specific (Cohan 2002). This is certainly the case for picocyanobacteria, with comparisons of phylogenetically conserved genes revealing genetic variability and niche specialisation, as demonstrated by substantial divergences in genomes even amongst strains that show close relationships (Rocap et al. 2003).

In the Southern Ocean, genetic analysis of seawater down to 54°S revealed *Synechococcus* strains related to strains of Group 5.1, which are also common in temperate oceanic zones (Wilmotte et al. 2002). No specific data on *Synechococcus* in Antarctic zone waters is available. Sequences obtained from the bottom sediment of various Vestfold Hills water bodies grouped in the *Synechococcus/Prochlorococcus* clade (Bowman et al. 2000). Other evidence for the presence of *Synechococcus* in these lakes also comes indirectly from the occurrence of sediment glycolipid compounds specific only to cyanobacteria (Damsté et al. 2001). Various marine salinity lakes of the Vestfold Hills, Eastern Antarctica (Rankin et al. 1997, Vincent et al. 2000) were found to contain rich blooms of phycoerythrin-containing picocyanobacteria. Populations were 1 to 3 orders of magnitude higher than typically found in other marine water bodies (Waterbury 1992) or other polar lakes thought to be dominated by cold-adapted cyanobacteria (Tang et al. 1997, Vezina & Vincent 1997, Vincent et al. 2000). This prompted more detailed study of these cyanobacteria, with the primary study site located at Ace Lake, a well-studied meromictic lake in the Vestfold Hills (Gibson 1999).

Ace Lake has 3 stratified zones: (1) a top mixed zone; (2) below this a layer supersaturated with oxygen separated by a pycnocline (depth 7 to 9 m); (3) a bottom zone below another pycnocline (depth 12 to 15 m), which is anoxic and contains high methane levels but is sulphate-depleted. The ratio of major cations and anions are very similar to that of seawater (Franzmann 1996). The characteristics of these populations are investigated here in detail in order to determine the factors leading to their high density and to determine aspects of their ecology, physiology and phylogeny.

### MATERIALS AND METHODS

**Samples.** Water samples were collected from various lakes and coastal waters using a Kemmerer water sampler (Geneq) (see Table 1). For the seasonal study of Ace Lake, samples were collected monthly between February and August 1992 and then fortnightly between September 1992 and January 1993. Samples were kept at 4°C and flow cytometry analysis was conducted within 4 h of sampling. Subsamples were fixed in 1.5% (vol./vol.) formalin for epifluorescence microscopy. Photosynthetically available radiancy (PAR) within the lakes was determined using a digital scalar irradiance meter and measurements were taken as close to the solar noon as possible. The euphotic depth values were calculated from incident light readings (Kirk 1994).

**Water chemistry.** A submersible data logger was used to collect temperature and conductivity data. Dissolved organic carbon (DOC) was analysed from filter samples using a SK12 DOC analyser (Skalar). Soluble reactive phosphate (SRP) and nitrate were analysed using an Alkem Flow Solution Autoanalyser. Detection limits for SRP and nitrate were 0.04 and 0.02 µM, respectively.

**Flow cytometry.** Cell counts for *Synechococcus* and phototrophic nanoplankton were conducted using a FACScan flow cytometer (Becton Dickinson), fitted with an argon laser with a 488 nm emission. Cell populations were detected using combinations of forward light scatter, side-angle light scatter, green, orange and red fluorescence emission using a 585/42 nm band-pass filter for phycoerythrin and propidium iodide-containing cells, and a 650 nm long-pass filter for chlorophyll a-containing cells. The instrument was set with a red or green fluorescence threshold and a
low acquisition rate. Microspheres of 1.98 µm diameter (Polysciences) were used for calibration for the determination of cell sizes and concentrations. Populations were gated on 2-parameter dot-plots using FACScan software. Non-photosynthetic bacterial populations were determined by staining with propidium iodide. Samples were centrifuged at 6000 × g at 4°C for 30 min, and the supernatant decanted. The sedimented cells were then washed (6000 × g, 4°C, 10 min) once with phosphate-buffered saline (PBS; 130 mM NaCl, 7 mM Na2HPO4, 3 mM NaH2PO4; pH 7.0) and resuspended in 90 µl of PBS. Propidium iodide was then added to a final concentration of 7 µg ml–1 and the suspension was incubated overnight at 4°C. Following incubation, the samples were washed twice with PBS and then resuspended in a known volume of PBS. Microspheres were used for size calibration as indicated above. Bacterial cells were then analysed by flow cytometry by comparing the cell-to-bead ratio.

Cell counts where required was subjected to ANOVA and pair-wise multiple comparison procedures (Student-Newman-Keuls tests) using the SAS program (SAS Institute).

Isolation of cultures. Lake samples were taken from depths at which concentrations of Synechococcus cells were maximal (Lake Abraxas: 19 m, Ace Lake: 11 m, Pendant: 6 m) and were enriched in SNAX and SOX media (Waterbury 1992) with incubation at 4 to 10°C on a 12 h dark/light cycle at 20 µmol photons m–2 s–1. The SNAX medium was prepared as described above. Bacterial cells were then analysed by flow cytometry to determine the cell-to-bead ratio. Cell count data where required was subjected to ANOVA and pair-wise multiple comparison procedures (Student-Newman-Keuls tests) using the SAS program (SAS Institute).

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**Pigment analysis.** Dense culture suspensions were examined over wavelengths of 400 to 750 nm using a GBC916 UV/VIS spectrophotometer (Dandenong) fitted with an integrating sphere. In vivo emission spectra were obtained between 500 and 700 nm with an excitation wavelength of 450 nm, while excitation spectra were collected over 400 to 665 nm with an emission wavelength of 680 nm. Lipid-soluble pigments in lake samples and cultures were extracted and analysed using an HPLC (Spectraphysics) fitted with a 68 nm emission wavelength of 680 nm. Lipid-soluble pigments were extracted by the modified 1-phase chloroform-methanol-water Bligh Dyer method (White et al. 1979). Lipid classes were quantified from the lipid extracts using an Iatroscan Kk. III TH-10 thin-layer chromatography-flame ionisation detector (Iatron, Tokyo). Following stabilisation of the absorbance, 20 × SSC (preheated to 90°C) was quickly added and mixed to increase the salinity of the solution to that of 2 × SSC. The temperature was adjusted to 71°C, the optimal temperature for renaturation for the Antarctic lake Synechococcus strains, assuming an average G + C of 56.5 mol%. The linear decline in absorbance at 260 nm was then monitored for 40 min. The percent DNA hybridisation was determined from the renaturation rate using the equation:

\[
\%\text{DNA hybridisation} = \frac{4AB - A - B\sqrt{2(AB)}}{A \times B} \times 100\%
\]

where \(AB\) is the renaturation rate for equal mixtures of DNA of the strains compared, and \(A\) and \(B\) are the rates for the unmixed controls of the respective strains. Values below 25% are considered insignificant due to random background renaturation. We performed 4 to 5 replicate analyses for each strain pair compared.

**DNA base composition and DNA:DNA hybridisation.** The DNA base composition was determined by the thermal denaturation procedure (Sly et al. 1986), using a GBC916 UV/VIS spectrophotometer fitted with a temperature controller. *Pseudoalteromonas haloplanktis* (mol% G + C 41.0) and *Escherichia coli* Strain K12 (mol% G + C 52.0) were used as reference standards. DNA:DNA hybridisation of genomic DNA between the 3 Antarctic lake strains was performed using the spectrophotometric renaturation kinetics technique as modified by Bowman et al. (1998). Briefly, genomic DNA at approximately 100 µg ml–1 was sheared to an average size of 1 kb using sonication and then dialysed at 4°C against SSC buffer (0.15 M NaCl and 0.015 M trisodium citrate, pH 7.0). The sheared DNA at 75 to 80 µg ml–1 from 2 strains were mixed in equal proportions and measured in comparison to separate unmixed controls of the same strains. The DNA was denatured at 95°C for 5 min in cuvettes in a GBC 916 UV/VIS spectrophotometer with a temperature controller. Following stabilisation of the absorbance, 20 × SSC (preheated to 90°C) was quickly added and mixed to increase the salinity of the solution to that of 2 × SSC. The temperature was adjusted to 71°C, the optimal temperature for renaturation for the Antarctic lake *Synechococcus* strains, assuming an average G + C of 56.5 mol%. The linear decline in absorbance at 260 nm was then monitored for 40 min. The percent DNA hybridisation was determined from the renaturation rate using the equation:

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**Physiological characterisation.** Duplicate cultures were grown at –1 to 30°C under 20 µmol photons m–2 s–1 on a 12 h dark/light cycle in a temperature gradient incubator (Toyo). Aseptically collected subsamples were taken periodically from the cultures and analysed by flow cytometry (as above) to determine cell numbers. Growth rates and cardinal temperatures (minimum, \(T_{\text{min}}\); optimum, \(T_{\text{opt}}\); maximum, \(T_{\text{max}}\)) were predicted using the 4-parameter square-root growth model.
kinetic model of Ratkowsky et al. (1983). The SAS statistical program (SAS Institute) was used to fit the model to the data and to determine the upper and lower root mean-square error (RMSE) limits for the model. The effect of light (tested between 5 and 300 µmol photon m⁻² s⁻¹) on growth was tested at 15°C using SNAX media with light monitored using a Li-Cor model LI-185 Quantum Ratiometer/Photometer (Li-Cor). The effect of salinity was tested in the SNAX medium prepared with dilutions of Ace Lake water in distilled water and with added sea salts (Sigma). Flow-cytometry was used to monitor growth over time as described above.

**In situ growth experiments.** The metabolic inhibitor procedure (Xiuren & Vaulot 1992) was used to assess environmental parameters on growth rates in Ace Lake during summer. Grazing pressure and effects of light, salinity, nutrients and temperature on growth rates were tested in combination using a factorial design (Campbell & Carpenter 1986).

For grazing pressure experiments, cycloheximide was added at 100 µg ml⁻¹ and was used to inhibit eukaryotic grazers, while ampicillin at 0.03 µg ml⁻¹ was used to inhibit growth of *Synechococcus*. At this concentration, ampicillin caused slight decreases in cell concentration, possibly due to cell lysis. Controls were also included which either contained no antibiotics or cycloheximide (100 µg ml⁻¹) and ampicillin (10 µg ml⁻¹).

Ace Lake water samples used in the experiments were collected from a depth of 6 m using a Kemmerer water sampler and added to plastic sample bags ('Whirl Pacs'). To assess the effects of temperature, the bags were lowered to 3 different depths (6, 8 or 10 m) in Ace Lake. Temperatures at each depth over the course of the experiment were quite stable: 3.0 to 3.4°C at 6 m, 6.3°C at 8 m and 7.0 to 7.1°C at 10 m. The effect of irradiance was assessed by using clear bags, bags that were a neutral density grey, and bags with shade cloth tied around them. This replicated light levels at 6 m, 6.3°C at 8 m and 7.0 to 7.1°C at 10 m. The pH in the lake was relatively constant, ranging from 8.1 to 8.5, with values slightly higher in summer than in winter. Light measurements indicated that nearly 100% of *Synechococcus* cells fixed in 1% (vol./vol.) glutaraldehyde (pH 7.0) at 5°C for 30 min and then frozen and stored in liquid N₂ were recoverable for flow cytometry as long as samples were rapidly thawed just before analysis (Lepesteur et al. 1993). Growth and grazing rates were determined using the equation: \[ r = k - g = \ln(N_t/N_0)/t \]
where \( N_0 \) is the cell abundance at Time zero, \( N_t \) is the cell abundance at Time \( t \), \( r \) is the net growth rate.
Specific growth rate (\( k \)) was measured from the rate of increased cell density in the cycloheximide treatments while grazing rate (\( g \)) was calculated from the rate of cell density reduction in the ampicillin treatment minus the control. Analysis of variance (ANOVA) was used to determine significance and interactions between treatments using the SAS programme (SAS Institute).

**RESULTS**

**Characteristics of Ace Lake**

During the seasonal survey, the top 7 m of Ace Lake was mixed, with summer and winter temperatures of 2 to 5.5°C and −1°C, respectively, and salinity ranging between 13 to 16 g kg⁻¹ depending on the ice-cover thickness. Between 7 and 15 m, the oxygenated middle zone had a higher temperature, peaking at 10.5°C in February and dropping to 4.5°C in August to November (Fig. 1A). Salinity in this layer increases rapidly from 16 to 28 g kg⁻¹ with depth in the pycnocline from 7 to 9 m. The pH in the lake was relatively constant, ranging from 8.1 to 8.5, with values slightly higher in summer than in winter. Light measurements indicated that PAR reached a maximum of 35 µmol photons m⁻² s⁻¹ at 10 m during January, while at the same depth PAR was below the detection limit between April and October. The depth of the euphotic zone was <3 m during winter, but increased to 10–11 m over summer (Fig. 1B). DOC concentrations in the aerobic zone of Ace Lake were always less than 10 mg l⁻¹, with levels highest at the 10 m depth over winter (6 to 9 mg l⁻¹) and lowest at the end of summer (6 mg l⁻¹) (Fig. 1A). Nitrate was below detection limits in the aerobic zone of Ace Lake in late November. Phosphate concentration increased with increasing depth (see Fig. 4).

**Microbial populations in Ace Lake**

Flow cytometry analysis revealed that *Synechococcus* populations were below 10⁵ cells ml⁻¹ (Fig. 1C). During summer, numbers increased by 2 orders of magnitude up to the highest reading of 8 × 10⁶ cells ml⁻¹ at a
depth of 11 m in December (Fig. 1C). The number of *Synechococcus* cells was consistently greater at a depth of 11 m than at 8 and 10 m (Fig. 1C). Heterotrophic bacterial numbers on the other hand increased in summer by only about 4-fold at a depth of 10 m (Fig. 1A). The initiation of the picocyanobacterial bloom occurred in October, when PAR at 10 to 11 m increased to above 5 µmol photons m$^{-2}$ s$^{-1}$. By the time PAR had reached the maximum level in January, the picocyanobacterial bloom was in decline (Fig. 1B,C). The increase in numbers of picocyanobacteria also corresponded with supersaturation of dissolved oxygen below the pycnocline, with concentrations rising from 13 mg l$^{-1}$ in winter to about 17 mg l$^{-1}$ in summer (Fig. 1A).

**Distribution in relation to eukaryotic phytoplankton**

The contribution of picocyanobacteria to the primary production rates in Ace Lake was not directly measured; however, based on total phytoplankton cell surface area, picocyanobacteria dominated in Ace Lake below the pycnocline during late November when the bloom peaked (Fig. 2). Using flow cytometry and microscopy, 4 major eukaryotic phytoplankters were identified in Ace Lake in addition to *Synechococcus*: *Mesodinium rubrum*, *Pyramimonas gelidicola*, a *Cryptomonas* sp. and a phototropic nanoplankter; these algae dominated the upper layers of Ace Lake to varying degrees (Fig. 2). Flow cytometry indicated that the nanoplankter was 2 to 3 µm in diameter and produced a strong chlorophyll a fluorescent signal.

**Photoadaptation and diel periodicity**

Over the summer period (October to December), flow cytometry revealed that red (mostly chlorophyll a) and orange fluorescence (phycoerythrin) of individual picocyanobacterial cells declined with increased PAR at 10 m. The ratio of phycoerythrin to chlorophyll a decreased from 1.3 in October to 0.7 in December. The cyanobacterial population in Ace Lake was closely monitored in December to determine if diel periodicity in irradiance occurred. PAR over 12 h cycles at the 8 to 11 m depth varied from 0 to 2.5–4.5 µmol photons m$^{-2}$ s$^{-1}$ during the period, and only slight variations in cell distribution was observed.
over a series of 24 h experiments (data not shown). Diel periodicity, a common phenomenon in natural populations of Synechococcus spp. (Prézelin 1992), may not have been detected here due to low growth rates of the picocyanobacteria in Ace Lake (Fahnenstiel et al. 1991).

Controls on in situ growth rates of Ace Lake picocyanobacteria

The results from the metabolic inhibitor study did not indicate any significant correlation (p > 0.128) between picocyanobacterial growth rates and grazing pressure. In addition, different antibiotic treatments, which incorporated the effects of light levels as a further variable effect, showed no differences. No significant effects resulted from moderate variations in salinity ($F = 0.06$, p = 0.80) or nutrients ($F = 0.19$, p = 0.66). The only first-order interaction (ANOVA analysis) that corresponded to a significant change in growth rate was temperature, which was related directly to water depth and light (p < 0.035), with increased temperature and lower light resulting in greatest increases in growth rate (Fig. 3A). The reduction of light using shading also caused increases in growth rates, although this was only significant in samples suspended at 10 m (p = 0.008) (PAR 8: Fig. 3B).

Picocyanobacteria in other Vestfold Hills lakes and coastal areas

Several Vestfold Hills lakes and coastal waters (Table 1) were surveyed during November for picocyanobacteria. Besides Ace Lake (Fig. 4), populations were only de-

Table 1. Synechococcus and bacterial populations in different Vestfold Hills lakes and coastal areas sampled in mid-summer. The level of PAR was measured at noon at the time of sampling and is shown in relation to the depth range sampled. bld: below limits of detection; nd: no cells detected; na: not available

<table>
<thead>
<tr>
<th>Water body</th>
<th>Depth range sampled (m)</th>
<th>% of PAR</th>
<th>Synechococcus ($\times 10^4$ cells ml$^{-1}$)</th>
<th>Bacteria ($\times 10^4$ cells ml$^{-1}$)</th>
<th>Salinity (°C)</th>
<th>Phosphate (µM)</th>
<th>Nitrate (µM)</th>
<th>DOC (µg l$^{-1}$)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Ace Lake</td>
<td>4–11</td>
<td>0.13–80.7</td>
<td>6.4–2.4</td>
<td>63.1–532</td>
<td>1.4–6.6</td>
<td>18.1–29.9</td>
<td>0.26–8.8</td>
<td>bld</td>
</tr>
<tr>
<td>Lake Abraxas</td>
<td>3–19</td>
<td>0.005–1480</td>
<td>7.2–1.0</td>
<td>10.3–24</td>
<td>0.5–8.0</td>
<td>16.8–20.3</td>
<td>0.06–0.2</td>
<td>bld–0.21</td>
</tr>
<tr>
<td>Pendant Lake</td>
<td>5–13</td>
<td>871–1500</td>
<td>1.1–0.7</td>
<td>480–667</td>
<td>0.2–0.8</td>
<td>16.5–31.0</td>
<td>0.8–1.3</td>
<td>bld</td>
</tr>
<tr>
<td>Taynaya Bay</td>
<td>3–16</td>
<td>3.6–0.1</td>
<td>50.3–80.6</td>
<td>–1.9–1.6</td>
<td>35.2–44.2</td>
<td>0.9–14.4</td>
<td>5.1–21.9</td>
<td>bld</td>
</tr>
<tr>
<td>Coastal marine</td>
<td>5</td>
<td>nd</td>
<td>0.09</td>
<td>8.6</td>
<td>–1.9</td>
<td>34.6</td>
<td>2.0</td>
<td>14.9</td>
</tr>
<tr>
<td>Burton Lake</td>
<td>4–12</td>
<td>nd</td>
<td>0.01–0.04</td>
<td>115–158</td>
<td>–2.3</td>
<td>43.3–43.8</td>
<td>10.5–27.4</td>
<td>0.6–1.1</td>
</tr>
<tr>
<td>Lake McCallum</td>
<td>5–20</td>
<td>nd</td>
<td>3.6–0.1</td>
<td>4.2–8.6</td>
<td>1.4–6.6</td>
<td>14.8–18.2</td>
<td>0.09–0.13</td>
<td>bld</td>
</tr>
<tr>
<td>Clear Lake</td>
<td>5–30</td>
<td>nd</td>
<td>17.0–0.35</td>
<td>9.4–11.7</td>
<td>1.4–7.0</td>
<td>10.3–12.7</td>
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<td>na</td>
</tr>
<tr>
<td>Eko Lake</td>
<td>4–20</td>
<td>nd</td>
<td>44–1.7</td>
<td>7.1–585</td>
<td>8.2–17.6</td>
<td>59.8–124.6</td>
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<tr>
<td>Organic Lake</td>
<td>2–6</td>
<td>nd</td>
<td>2.3–0.02</td>
<td>530–1360</td>
<td>–10.7–7.8</td>
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<td>Fletcher Lake</td>
<td>2–6.5</td>
<td>nd</td>
<td>6–1.4</td>
<td>19.7–233</td>
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<td>56.4–81.8</td>
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</table>

Fig. 3. Synechococcus. In situ growth of cells in Ace Lake showing interactions of temperature and light. Temperature and photosynthetically available radiance (PAR) (µmol photons m$^{-2}$ s$^{-1}$) to which sample bags were exposed are indicated. Temperature remained nearly constant during course of experiment; PAR levels indicated were maximum levels reached at noon. (A) Mean (+SE) growth rates at 3 depths and light intensities. (B) Mean (+SE) growth rates in sample bags suspended and incubated at 2 depths, with some bags shaded from light such that maximum PAR was reduced to approximately 2 µmol photons m$^{-2}$ s$^{-1}$ (see ‘Materials and methods’ for further details)
ected in Lake Abraxas and Pendant Lake on the basis of flow cytometry (Table 1). In Lake Abraxas Synechococcus populations were stratified, with greatest cell numbers at depths of 19 to 20 m, corresponding to a temperature of 7 to 8°C (Fig. 4). PAR at this depth was about 1% of the total incident light recorded at the time (Fig. 4). In comparison, Synechococcus cells were much less stratified in Pendant Lake, which had a more even temperature distribution (0.1 to 1.1°C) (Fig. 4). Only slightly greater populations occurred below the pycnocline at 11 m. PAR throughout much of Pendant Lake was between 0.7 and 1.1% of the total incident light (Fig. 4).

Isolation and morphological data

Non-axenic cultures (designated strains Abraxas, Ace and Pendant) were successfully obtained from 3 Antarctic lakes in the SNAX medium. Growth also occurred to a lesser extent in f/2 and FE media, with the Synechococcus representing a minority component of enrichments, dominated by diatoms. No growth occurred in SOX medium (Waterbury 1992), suggesting that the lake Synechococcus were unable to fix nitrogen. No growth was found in media prepared with coastal seawater. After the primary inoculation,
growth did not occur on SNAX media solidified with agar or agarose. Antibiotic selection reduced heterotrophic bacterial contaminants to a level of <1% of the total cell population in SNAX media, and only 1 homogenous strain was observed by microscopy (confirmed subsequently by 16S rRNA gene sequence analysis). Under epifluorescent microscopy, all 3 strains fluoresced orange-red under green light and were fairly uniform in size (0.9 to 1.1 µm x 1.5 to 1.8 µm) and appeared as coccibacilli or short rods. Transmission electron microscopy revealed that the cells contained thylakoid membranes, phycobilisomes (20.3 ± 5.7 diameter) and carboxysomes (85 nm diameter) (Fig. 5). The cultures were confirmed as representatives of the dominant population in Ace Lake, Lake Abraxas and Pendant Lake on the basis of lipid-soluble pigment profiles from the cultures and lake water particulates, which have the same relative proportions of photosynthetic and carotenoid pigments (Fig. 6A). Pigments in extracts from Lake Abraxas and Pendant Lake were very similar to those of Ace Lake. Water bodies that did not contain Synechococcus cells (as determined by flow cytometry) had different pigment profiles.

**Chemical aspects of Synechococcus isolates**

*In vivo* spectra were similar for all strains, with wavelength peaks indicating that the cells contained mostly chlorophyll a, phycoerythrin and lesser amounts of phycocyanin and allophycocyanin. An absorbance peak at 495 nm indicated the presence of β,β-carotene. HPLC analysis of the Ace strain (Fig. 6B,C), confirmed the presence of chlorophyll a, β,β-carotene, zeaxanthin and 2 other unknown zeaxanthin-like carotenoids, which had absorption spectra at 454/484 nm and 455/485 nm, respectively. When grown under high light conditions (100 µmol photons m⁻² s⁻¹) the ratios of lipid-soluble pigments shifted, with a 7-fold decline in chlorophyll a, a 3-fold decline in β,β-carotene, and 2 to 3-fold increases in the proportions of the zeaxanthin-type carotenoids (Fig. 6B). The fatty acids of Strain Ace consisted mainly of 16:1ω7c (43 to 50%), 14:0 (18 to 28%), 16:0 (10 to 12%), 18:1ω7c (1 to 3%) and 18:0 (2 to 6%), with profiles of cells grown at early and late logarithmic growth phase only varying slightly.

**Phylogenetic and genotypic characteristics**

All 3 lake strains were very closely related (16S rRNA gene sequence similarity >98%) and formed a distinct, peripheral lineage within the *Prochlorococcus**/Synechococcus* clade (Fig. 8). The closest available published sequences were from *Synechococcus* Strains WH5701 (Fuller et al. 2003), P211 (Vincent et al. 2000) and BO0014 (Ernst et al. 2003), which were 95 to 96% similar. Similarities to other sequences in the larger *Synechococcus* clade ranged from 93 to 95%.
The Antarctic Lake strains had DNA base compositions of 56 to 57 mol%. The strains also represented a single genospecies as they shared high levels of DNA hybridisation. The Ace Lake and Pendant Lake strains hybridised at 88 ± 10%, the Ace Lake and the Lake Abraxas strain hybridised at 77 ± 8%, and the Pendant Lake and Lake Abraxas strains hybridised at a level of 74 ± 5%.

Fig. 6. *Synechococcus*. HPLC lipid-soluble pigment profile of (A) Ace Lake particulates from depth of 10 m; (B) Ace Lake strain under strong light (100 µmol photons m⁻² s⁻¹); (C) Ace Lake strain under low light (5 µmol photons m⁻² s⁻¹). Peak identification = a: me-chlorophyllide; b: unknown carotenoid; c: unknown carotenoid; d: zeaxanthin; e: chlorophyll a; and f: β,β-carotene. Peaks identified by comparison with authentic standards and cultures

Fig. 7. *Synechococcus*. Theoretical growth curve, obtained from 4 parameter, square-root growth kinetic model of Ratkowsky et al. (1983), showing actual data points obtained during temperature gradient experiments. Irradiance was 20 µmol photons m⁻² s⁻¹. Upper and lower root mean-square error (RMSE) limits for model based on the fitted data are shown

DISCUSSION

Population dynamics

Although meromictic Ace Lake has been much studied, reports of picocyanobacteria in this lake and the other maritime lakes of Vestfold Hills have been sparse and of a preliminary nature (Rankin et al. 1997, Vincent et al. 2000). Using flow cytometry, distinct phycocerythrin-containing picocyanobacterial blooms could be readily visualized as distinctly stratified populations (Fig. 1C) in certain saline, meromictic Vestfold Hills lakes (Table 1, Fig. 4). Stratified distributions of phycocerythrin-rich picoplankton have also been observed in various tropical and temperate meromictic lakes (Maeda et al. 1992, Venkateswaran et al. 1993), and are maintained by active cell division overcoming net losses from grazing and sinking. The picocyanobacteria here were identified as members of form genus *Synechococcus* (Herdman et al. 2001) on the basis of pigment content, cell morphology and binary cell division, which occurred only across a single plane. Blooms of *Synechococcus* in Ace Lake, occurred within the stable zone below the pycnocline (Fig. 1B,C) and corresponded in response to increases in PAR (Fig. 1A) and water temperature (Fig. 1B) during spring, a phenomenon that has also been observed in freshwater polar and subpolar lakes (Vezina & Vincent 1997). Populations of marine picocyanobacteria typically reach maxima of 10⁶ to 10⁷ cells ml⁻¹ in open ocean regions and 10⁴ to 10⁶ cells ml⁻¹ in temperate and trop-
Synechococcus spp. and related strains. Antarctic lake strain sequences compared were 1389 nucleotides in length, and distances were generated using maximum-likelihood algorithm and clustered using neighbour-joining method. Clusters are designations for potential new genera making up form genus Synechococcus (Herdman et al. 2001). GenBank accession numbers are given after each strain. Bootstrap values (>70%) are shown at branch nodes and were calculated from 500 replicates. Distance bar = 0.1 nucleotide substitutions per sequence position.

Experiments were performed to observe the interactions of light, salinity, nutrients and temperature (which effectively incorporated grazing loss) on Synechococcus cells in Ace Lake. Populations exhibited very low growth or net losses at 6 m depth whether the light intensity was low or high (Fig. 3). Since culture studies indicate that growth at 3°C is about 0.06 d⁻¹, grazing could account for low levels of biomass production. A combination of decreased light and increased temperature, which occurred at a depth of 10 m in Ace Lake, resulted in increased growth rates. The increase in growth rate compensated for grazing losses and resulted in net biomass production and the highest net population growth rates (Fig. 3). Preliminary data from the seasonal survey of the Synechococcus cell populations showed that the mixed-zone populations increased by only about 10-fold from September to October. This small population increase may have been affected by grazing; however, the effect of temperature may have been a critical factor resulting in a smaller net gain over the spring bloom period. At 10 to 11 m, at which the greatest cell numbers occurred, the specific growth rate during the bloom

Fig. 8. 16S rRNA gene-based phylogram comparing various Synechococcus spp. and related strains. Antarctic lake strain sequences compared were 1389 nucleotides in length, and distances were generated using maximum-likelihood algorithm and clustered using neighbour-joining method. Clusters are designations for potential new genera making up form genus Synechococcus (Herdman et al. 2001). GenBank accession numbers are given after each strain. Bootstrap values (>70%) are shown at branch nodes and were calculated from 500 replicates. Distance bar = 0.1 nucleotide substitutions per sequence position.

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period was estimated at 0.07 d\(^{-1}\). This growth rate was quite similar to the growth rate in laboratory cultures of the Antarctic lake strains at 7 to 8°C, and similar to the growth rates of Synechococcus spp. in polar oceans (Neuer 1992). In comparison, growth rates for Synechococcus spp. in temperate and tropical locations can be >1 d\(^{-1}\) (Waterbury 1992). Growth rates estimated during the \textit{in situ} lake experiments often exceeded those determined from laboratory culture studies conducted at the same temperature (e.g. at 7°C in Ace Lake at 10 m growth was up to 0.34 d\(^{-1}\), while the culture studies estimate only 0.07 d\(^{-1}\) for an equivalent temperature). This large difference could have been due to experimental artefacts introduced into either the field or laboratory experiments such as the effects of the containment of the populations in the sample bags, or self-shading.

**Effect of irradiance, nutrients and salinity**

Primary production in Ace Lake is relatively low compared to that in Antarctic coastal waters, being estimated at 0.5 to 0.7 µg C l\(^{-1}\) h\(^{-1}\) (Laybourn-Parry & Perriss 1995). \textit{Mesodinium rubrum} is the most dominant phytoplankter in the Ace Lake mixed zone (Fig. 2). \textit{Synechococcus} cells made the greatest contribution to Ace Lake primary production, below the pycnocline (Fig. 2). \textit{Synechococcus} cells in the Vestfold Hills lakes appeared to prefer overall lower light intensities. In spring, \textit{Synechococcus} cells growth was initiated at between 1 and 5 µmol photons m\(^{-2}\) s\(^{-1}\), with greatest cell concentrations occurring in zones with light intensities of 5 to 20 photons µmol photons m\(^{-2}\) s\(^{-1}\), corresponding to about 1% of the incident light at the time. Culture studies indicated that growth yield was slightly lower under dim light (5 µm photons m\(^{-2}\) s\(^{-1}\)), but growth rates and cell volumes were unaffected. Flow cytometry of Ace Lake samples also showed that the \textit{Synechococcus} cells were phototactic and have high explosive growth with increasing PAR. Although light itself alone was not a significant factor affecting growth (p > 0.35) in the \textit{in situ} experiments, when combined with temperature it was (p = 0.008). Overall, the light intensities encountered by the \textit{Synechococcus} cells at 11 m in Ace Lake were essentially optimal for cell yield; in comparison, the effect on growth rates were not as significant. The enhancement in the levels of 2 structurally unidentified zeaxanthin-like carotenoids in the \textit{Synechococcus} cell cultures may be linked to protection against photo-oxidative stress (Baroli et al. 2003), since the proportionate levels of these carotenoids were markedly elevated at 100 µm photons m\(^{-2}\) s\(^{-1}\) (Fig. 6B).

Within Ace Lake, the \textit{Synechococcus} cell bloom was associated with oxygen supersaturation, probably due to active oxygenic photosynthesis. \textit{Synechococcus} spp. can potentially exhibit light-dependent sensitivity to oxygen, preferring lower oxygen when PAR is high (Glover & Morris 1981). In this study, the \textit{Synechococcus} cell numbers were low in those parts of the lake subjected to high irradiance levels, and populations were already in decline at the time of the annual light maximum in January. The growth of populations did not seem to be affected by the high oxygen level, perhaps because the PAR was relatively low and any oxidative damage to the photosynthetic apparatus was efficiently repaired (He & Hader 2002). The high dissolved inorganic carbon concentrations in Ace Lake (Gibson 1999), which include bicarbonate ion concentrations higher than those found in seawater (typically 2.4 mM; ZoBell 1946), suggest that photosynthesis would not be limited on the basis of CO\(_2\) availability. Ammonia, derived from the deeper anoxic zone, has also been found to be elevated in the middle layer of Ace Lake (5 to 25 µM; H. R. Burton pers. comm.), while SRP was quite high below the pycnocline. N and P thus seem to be more than sufficient to support higher \textit{Synechococcus} cell populations in Ace Lake. This was suggested by the \textit{in situ} experiment, which indicated that nutrient addition did not significantly stimulate growth. It is possible that a lack of response could have been due to slow acclimation to the extra nutrients provided. Alternatively, the relatively slow growth rate of the \textit{Synechococcus} cells also suggests that nutrient depletion over the 48 h period of the experiment would not be significant.

A survey of various Vestfold Hills water bodies and coastal areas emphasized the preference of the Antarctic lake strains for a relatively narrow salinity range, corresponding to levels approaching that of seawater (Table 1). The absolute requirement for seawater is a property of form genus \textit{Synechococcus} Clusters 4 and 5.1. Clusters 1 and 2 require freshwater and have low salt tolerance, while Cluster 3 is halotolerant but does not require Na\(^+\) ions for growth. Sea-water requirement, however, varies between strains of Cluster 5.2 (Herdman et al. 2001); for example, Strains BO0014 and P211, phylogenetic neighbours of the Antarctic strains were isolated from freshwater habitats. This suggests that salt requirement is a phylogenetically shallow trait amongst strains related to Clusters 5.1 and 5.2, and that salinity requirements are species-specific.

The incidence of \textit{Synechococcus} cell populations was also dependent on PAR, with populations exhibiting highest cell yields under the same light regimes (Fig. 4) as those observed in Ace Lake (Fig. 1B). Like Ace Lake, \textit{Synechococcus} cell populations were strati-
cies in Lake Abraxas below a pycnocline and could access nutrients diffusing up from an anoxic lower zone. In comparison, Pendant Lake exhibited high populations of *Synechococcus* throughout much of its water column, possibly because of the deep mixing in this lake, with populations slightly higher below the pycnocline at 11 to 12 m.

**Community interactions**

The *Synechococcus* cell bloom may have had the consequence of increasing bacterial heterotrophic populations in Ace Lake at 10 m (Fig. 1A), possibly driven by availability of DOC from photosynthate, lysing cells and from the ‘sloppy’ feeding of grazers. The subsequent growth of heterotrophs also reduced DOC levels by about 25% by the end of the summer season. An antibiotic selection method used in this study has been employed to determine grazing rates on *Synechococcus* cells in non-polar regions (Xiuren & Vaulot 1992). However, data obtained here showed no significant effects between treatments. This suggests that grazing potentially had a negligible effect on populations in Ace Lake at 10 m (Fig. 1A), possibly due to increasing bacterial heterotrophic growth, possibly driven by the availability of DOC from photosynthate, lysing populations in Ace Lake at 10 m. The antibiotic method used, however, makes several major assumptions: firstly that organisms in situ grow exponentially; secondly, that they are not lysed significantly by the antibiotics; and finally that the antibiotics selectively inhibit the predator but not prey populations. Antibiotic concentrations used in the study had been pre-tested thoroughly in the laboratory on the Ace Lake strain and a cultured *Cryotomonas* sp. common in Ace Lake. It is quite possible that under field conditions, manipulations used and logistical constraints (samples had to be fixed and cryopreserved before analysis) led to population changes which obscured grazing effects. *Synechococcus* cell grazing rates have been estimated at 0.23 to 0.44 d\(^{-1}\) in various temperate and tropical coastal and oceanic areas (Xiuren & Vaulot 1992, Liu et al. 1995), and often equal or exceed the specific growth rate. It is possible that the low growth rates in Ace Lake also added to the difficulty in estimating grazing rates. Ace Lake grazer populations have not been studied in great detail, but are known to include the ciliate mixotroph *Mesodinium rubrum*, copepods such as *Paralabidocera antarctica*, and heterotrophic microflagellates (Laybourn-Parry & Perriss 1995, Swadling & Gibson 2000). Peak grazer populations usually occurred at 11 m, where the maximum *Synechococcus* cell density is found. This could mean that total *Synechococcus* cell growth rates could be substantially greater than was measured directly in natural populations and equate better with the *in situ* growth experiments (e.g. 0.07 d\(^{-1}\) compared to 0.34 d\(^{-1}\)). If this were the case, there is a contradiction between these results and the culture growth rate measurements, which indicated maximal growth rates at only 0.13 d\(^{-1}\) at a *T*\(_{opt}\) of 19.7°C. Nevertheless, the high populations achieved by the *Synechococcus* cells suggest that they are an integral part of the microbial loop in Ace Lake.

**Systematic biology of *Synechococcus* from Vestfold Hills lakes**

Phylogenetically, the Antarctic lakes strains were most closely related to strains of Cluster 5.2 (Marine Cluster B), including *Synechococcus* Strain WH5701 (Herdman et al. 2001) (Fig. 8). Although, morphologically they were indistinguishable and had similar fatty acid profiles (Merritt et al. 1991), other data, including DNA-base composition, pigment profiles, growth rates and light preference indicated that the Lake strains were quite different (Herdman et al. 2001). Strain WH5701 (Fuller et al. 2003) is the reference culture for *Synechococcus* Group 5.2 (Herdman et al. 2001), and is a halotolerant, *C*-phycoerythrin-lacking strain originally isolated from Long Island Sound, New York. Closely related strain *Synechococcus* sp. BO0014 (Ernst et al. 2003) is by comparison *phycoerythrin*-rich and isolated from freshwater (Lake Constance, Germany). Strain P211 is another freshwater strain isolated from a pond on Bylot Island in the Canadian high Arctic (Vincent et al. 2000). This strain lacks phycoerythrin, has a specific growth of 0.41 d\(^{-1}\), and prefers relatively high light intensities (60 µmol photons m\(^{-2}\) s\(^{-1}\)). The Antarctic lake strains had a DNA-base composition substantially lower than the 66 mol% value of Strain WH5701 (Herdman et al. 2001), further emphasizing the genetic difference between the strains.

Assuming that the nomenclature members of form genus *Synechococcus* will be revised, the lake strains would constitute at least a distinct species as they represent a distinct genospecies because of a DNA hybridisation level of >70% (Wayne et al. 1987). This is also evident from the physiological, genotypic and phylogenetic distinctiveness of the Antarctic strains. Data for the Ace lake strain (Fig. 7) also indicated that the lake strains were cold-adapted, with cardinal temperatures well below those of temperate and tropical marine *Synechococcus* (Moore et al. 1995). The Antarctic lake strains were relatively slow-growing for *Synechococcus* spp., which typically grow at doubling times of 0.4 to 1.4 d\(^{-1}\). However, the rates are similar to those of *Prochlorococcus* spp. (Moore et al. 1995) and polar picocyanobacteria (Tang et al. 1997). The acqui-
sition of psychrophily appears to result in physiological trade-offs, resulting in slower growth rates. This is possibly due to the energy requirements for maintenance of globular protein structural integrity (Rat-kowsky et al. 2004). The Vestfold Hills saline lakes formed subsequent to deglaciation and marine uplift only about 8000 yr ago (Fulford-Smith & Sikes 1996). It is thus possible that the *Synechococcus* strains studied here were trapped in the marine-derived ecosystems and subsequently thrived in locations that had environmental conditions favourable for their growth. Similar hypotheses have been posited for other bacteria isolated from saline Vestfold Hills lakes (Franzmann 1996).

**CONCLUSIONS**

It is possible that the *Synechococcus* sp. studied here is adapted to life at the bottom of the oceanic euphotic zone because of its preference for lower light conditions. However, low temperature would probably restrict it to negligible populations in Antarctic Zone waters (Marchant et al. 1987). Ace Lake, Pendant Lake and Lake Abraxas are all in close proximity to Long Fjord, and were isolated from the sea at about the same time only 8000 yr ago. It would be of interest to sample the coastal and fjord zones at depth with various PCR-based molecular methods (DGGE, real-time PCR etc.) to determine if the picocyanobacteria there correspond to the lake strains. Ultimately, *Synechococcus* populations in polar marine waters are highly dependent on a balance of rates of gain and loss (Vincent 2000). In the surface of Ace Lake, like that of the Southern Ocean, the rates of loss (grazing, turbulence of water) are high and the rates of gain (temperature) are low, resulting in low populations. Below the pycnocline of Ace Lake and Lake Abraxas, the loss rates are markedly offset by the stable water column, elevated temperatures and with optimal light levels leading to high cell yields during summer. In Pendant Lake the dynamics differ, with high *Synechococcus* sp. populations occurring at low temperatures; this suggests that the Pendant strain could be more cold-adapted or that the grazing pressure here is less than in the other lakes.

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