

Spatio-temporal variability of larval abundance and settlement of *Perna perna*: differential delivery of mussels

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ABSTRACT: We examined larval availability and settlement of the intertidal mussel *Perna perna* simultaneously at different spatial and temporal scales using a nested design at 2 sites, 3 km apart on the south coast of South Africa. Each site had 3 locations (300 m apart) where 5 artificial settler collectors were placed about 20 cm apart. Collectors were replaced on temporal scales varying from fortnightly (for 16 mo) to daily (2 series of 15 to 20 d). Each intertidal location was paired with an inshore location (these too were 300 m apart) within 500 m of the shore, where larval availability was measured by 3 vertical plankton hauls collected on the same dates as for settler sampling. There was strong temporal variation in abundances of larvae and settlers, and no correlation (r always < 0.14) was found between the two. Larvae were abundant only at the start of sampling and rare for the rest of the study, while distinct peaks in settler numbers occurred later. No spatial effect was detected for larval availability, while there was strong spatial variation in settlement at the location level. These results indicate that, on scales of 100s of m to km, delivery of larvae from the nearshore water column onto the shore is strongly differential, with some locations consistently receiving more settlers than others. We conclude that, at these sites, the patchiness in settlement observed on scales of 100s of m depends on differential delivery, rather than differential offshore distribution of larvae. We suggest that differential delivery is due to the effect of nearshore bottom topography on local hydrodynamics.

KEY WORDS: *Perna perna* · Settlement · Larval availability · Differential delivery · Mussel · Spatial variability · Temporal variability

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INTRODUCTION

For the past 3 decades, assessing spatial and temporal variability has been one of the major goals of intertidal ecologists trying to understand community organization and population dynamics (Gaines et al. 1985, Menge & Farrell 1989). For sedentary marine species with planktonic larvae, settlement is a major process, regulating the dynamics and structure of benthic populations (Hunt & Scheibling 1998, Balch & Scheibling 2000) and acting as a bridge between dispersal and recruitment. In fact, intertidal community structure and population dynamics are strongly affected by the

arrival of larvae on the shore, as settlement rates may determine the final spatial variation of adults (Gaines & Roughgarden 1985). This is especially true if settlement rates are low (Menge & Farrell 1989, Menge 2000). Thus, abundances of pelagic larvae and/or settlers can influence the adult population size (Connell 1985, Pulfrich 1996, Norkko et al. 2001, Jeffery 2003) and an understanding of population regulation requires investigation of spawning, dispersal, larval availability and settlement (Gaines & Bertness 1992, Morgan 2001, Underwood & Keough 2001).

A problem in studying species with external fertilization and planktonic larvae is that the possible scales of

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dispersal of larvae may range considerably (Roughgarden et al. 1985, Norkko et al. 2001). Dispersal is highly complex and variable, as pelagic larvae exist in a highly patchy and extremely variable environment (McQuaid & Phillips 2000, Natunewicz & Epifanio 2001). While some bivalve larvae show active swimming behaviour, controlling their position in the water (Dobretsov & Miron 2001, Metaxas 2001 for review, Norkko et al. 2001), studies on mussels indicate that larvae are transported passively, without even the vertical migration shown by barnacle larvae (Scheltema 1995, McQuaid & Phillips 2000, Satumanatpan & Keough 2001).

Larval patchiness and dispersal are, however, affected by a wide range of factors including seasonality of reproduction (Hurlbut 1991), hydrodynamics, water temperature, geography and topography (Gaines et al. 1985, Pineda 1991, Bertness et al. 1996), which have effects at different spatial scales (Dobretsov & Miron 2001, Poulin et al. 2002). Temporal variation in larval availability and settlement is also important (Beukema et al. 2001); apart from long-term (e.g. seasonal) variation, abundance in the water column can change dramatically over a short time (e.g. day, hours, seconds; Pulfrich 1997, Poulin et al. 2002). In order to understand the scales of transport and delivery of larvae and to make possible predictions on the future of larvae and the structure of adult populations, dispersal and settlement should be assessed simultaneously (Seed 1976, Suchanek 1985, Underwood & Keough 2001). Despite this, few studies have inspected the processes, relationships or scales that regulate the passage of invertebrate larvae from nearshore waters to the intertidal, adult habitat (Gaines et al. 1985, Minchinton & Scheibling 1991, Miron et al. 1995, Pineda et al. 2002).

The present survey is one of the few examples of concurrent investigation of different scales of temporal (seasonal to daily) and spatial (100 m to km) variation in larval availability and settlement. This approach should provide a better understanding of the dynamics and scales that regulate the arrival of pelagic larvae onto the shore (Minchinton & Scheibling 1991, Jeffery & Underwood 2000, Jenkins et al. 2000).

Adult mussel densities differ enormously on within-shore scales (Lawrie & McQuaid 2001, Erlandsson & McQuaid 2004). This could be due to post-settlement processes, but there is evidence that settlement differs systematically among locations separated by 100s of m (S. Antrobus unpubl. data). Differential settlement could be due to larvae arriving from different pools of larvae, or to differential delivery from nearshore waters to intertidal localities. This study was designed to compare these 2 hypotheses on larval availability and settlement of *Perna perna*.

MATERIALS AND METHODS

Study sites. Larval availability and settlement of the mussel *Perna perna* were investigated (March 2000 to June 2001) on the south coast of South Africa, using a nested approach at 2 sites, Kenton (hereafter KE) and High Rocks (hereafter HR), approximately 3 km apart (33° 41' S, 26° 40' E). At each site, inshore larval abundance was quantified as close to the shore as possible, approximately within 500 m of the shore, at 3 locations 300 m apart (positions maintained using a GPS) opposite intertidal locations where settlement was measured at 3 low shore locations with similar mussel cover (approximately 80% adult cover), 300 m apart on Aeolian dune rock platforms. All intertidal locations were characterised by the presence of the foliose red alga *Gelidium pristoides*, with which mussel patches were interspersed.

Larval availability. Inshore larval abundance was quantified from plankton samples collected in the morning on receding spring tides, prior to settlement sampling on the same day, immediately behind the surf zone (within 500 m of the shore). The water column depth at all offshore locations ranged between 8 and 11 m. Due to possible larval stratification within the water column (McQuaid & Phillips 2000), sampling was by means of vertical hauls using a plankton net (80 µm mesh, mouth size 30 × 30 cm), raised from approximately 1 m above the seabed (7 to 10 m), in order to minimise sand collection, to the surface at 0.5 m s⁻¹. Sampling lasted approximately 1 h. At each location, 3 replicate samples were collected, filtering volumes of water that ranged between 0.5 and 0.8 m³ per replicate. Samples were preserved in 40% formalin: 60% seawater. In the laboratory, a few drops of 0.38% Rose Bengal (acid red, CI45440), in 95% alcohol, were added to stain the organic material (Garland & Zimmer 2002). Temporal variation in larval abundance was inspected at different scales from a coarse scale long-term study (fortnightly sampling at spring tides for 14 mo from March 2000 to April 2001 giving 23 sampling events), to fine scale daily studies in 2000 and 2001. Daily sampling was carried out, weather permitting, from 20 March to 10 April 2000 and 6 February to 11 March 2001, when larval abundance was expected to be highest (S. Lawrie unpubl. data).

Settlement. Settlement was investigated from March 2000 to June 2001 using plastic scouring pads as larval collectors (Menge 1992, Menge et al. 1994). Pads were soaked in seawater for 1 to 2 wk, before we deployed them among adult mussels. At each location, 5 pads (20 cm apart) were secured to metal screws using plastic cable ties. Pad position changed quite frequently, due to loss of screws. In the laboratory, collected pads

were agitated and rinsed in 75 μm filtered fresh water and the contents examined under a dissecting microscope. *Perna perna* settlers were identified by shell morphology, counted, measured (0.05 mm precision) and preserved in 70% alcohol. The shell morphology of *P. perna* larvae and settlers is similar to that of *Mytilus galloprovincialis*, which also occurs here. However, morphological differentiation between the two has been confirmed by genetic identification (S. Bownes unpubl. data). *P. perna* settlement consists mainly of primary settlement. However, larger individuals can appear on artificial collectors exposed for just 24 h (see 'Results'), so a degree of secondary settlement, or dislodgment and re-attachment, may occur.

Temporal variability in settlement was explored at 2 scales including a long-term study (fortnightly sampling at spring tides for 16 mo from March 2000 to June 2001 giving 27 sampling events) and 2 short-term studies with daily sampling in 2000 and 2001. For the long-term study pads were collected after 2 d on the shore, while for the daily studies pads were replaced after 24 h.

The analyses of larval availability and settlement used data sets with different numbers of dates because the settlement data needed to be balanced due to the loss of pads on some days.

Statistical analysis. Nested ANOVA was performed using STATISTICA 6. Location (3 levels) was nested within Site (2 levels) and orthogonal to Time (23, 17 and 22 levels respectively for the long-term, 2000 and 2001 daily larval studies; 27, 17 and 22 levels for the settlement studies). Spatial factors were treated as random, since the sites used were 'ecological replicates' of the nearshore area and location was nested within site (Zar 1996).

For all studies, the temporal factor, Time, was treated as random. ANOVA tables for random models were made manually following Zar (1996) in order to calculate the degrees of freedom, mean squares and *F*-ratios. The values obtained through the calculation of synthetic mean squares were very similar to the normal construction of mean squares and, therefore, the results of the synthetic means are not presented here. Normality and homogeneity of variances were determined using the Kolmogorov-Smirnov and Cochran's tests respectively. When assumptions of normality and homogeneity could not be satisfied, the dependent variable was logarithmically transformed (Zar 1996, Underwood 1997).

Significant results were tested by post-hoc comparisons using the Student-Newman-Keul's test.

When necessary, data on settlement were balanced using 1 of 2 options. When 1 or 2 of the 5 pads were lost, the mean from the remaining pads was substituted for the missing values. Dates when 3 to 5 pads

were lost at 1 or more locations were eliminated from the analysis.

Coupling between the abundance of *Perna perna* larvae in the water column and settlers on scouring pads was investigated. For each temporal survey, the mean values of larvae collected from the water column on a sampling date were correlated with those for settlers on the shore at the same location and date and the Pearson correlation coefficients (*r*) calculated. Subsequently, cross-correlation analysis was run for each temporal survey to explore correlations using all possible spatial and temporal lags.

RESULTS

Larval availability

Mussels in the plankton ranged in size from 200 μm to 1.3 mm, but most (99.87%) were between 200 and 400 μm and only these individuals were used in the analysis. This range almost overlapped that of animals on settlement pads, representing the size of competent larvae.

Long-term study

Long-term collections showed peaks in larval distribution at the start of the sampling period, but only on the first 2 dates of collection (20 March and 3 April 2000). For the rest of the study, larvae were extremely rare at both sites (Fig. 1A,B).

Neither site nor location had a significant influence on larval availability ($p > 0.05$ for both; Table 1). Nevertheless, the abundance of larvae in the water changed dramatically both at sites and locations during specific collections. There were significant differences in larval variability between sites and among locations on specific sampling dates ($p < 0.0001$ for the interactions between Date and Location and Date and Site, Table 1). Total values at HR were >3 times greater than at KE (Fig. 1A,B), yet Site alone did not have a significant effect, possibly because the effect of Site was masked by the low numbers of larvae during most of the study and the high temporal variability (effect of Date, $p < 0.0001$, Table 1).

Daily study (2000)

The numbers of larvae found in the water during the 2000 daily sampling period varied considerably from one day to the next (Fig. 2A,B). Although larval abundance differed among locations, it was not affected by

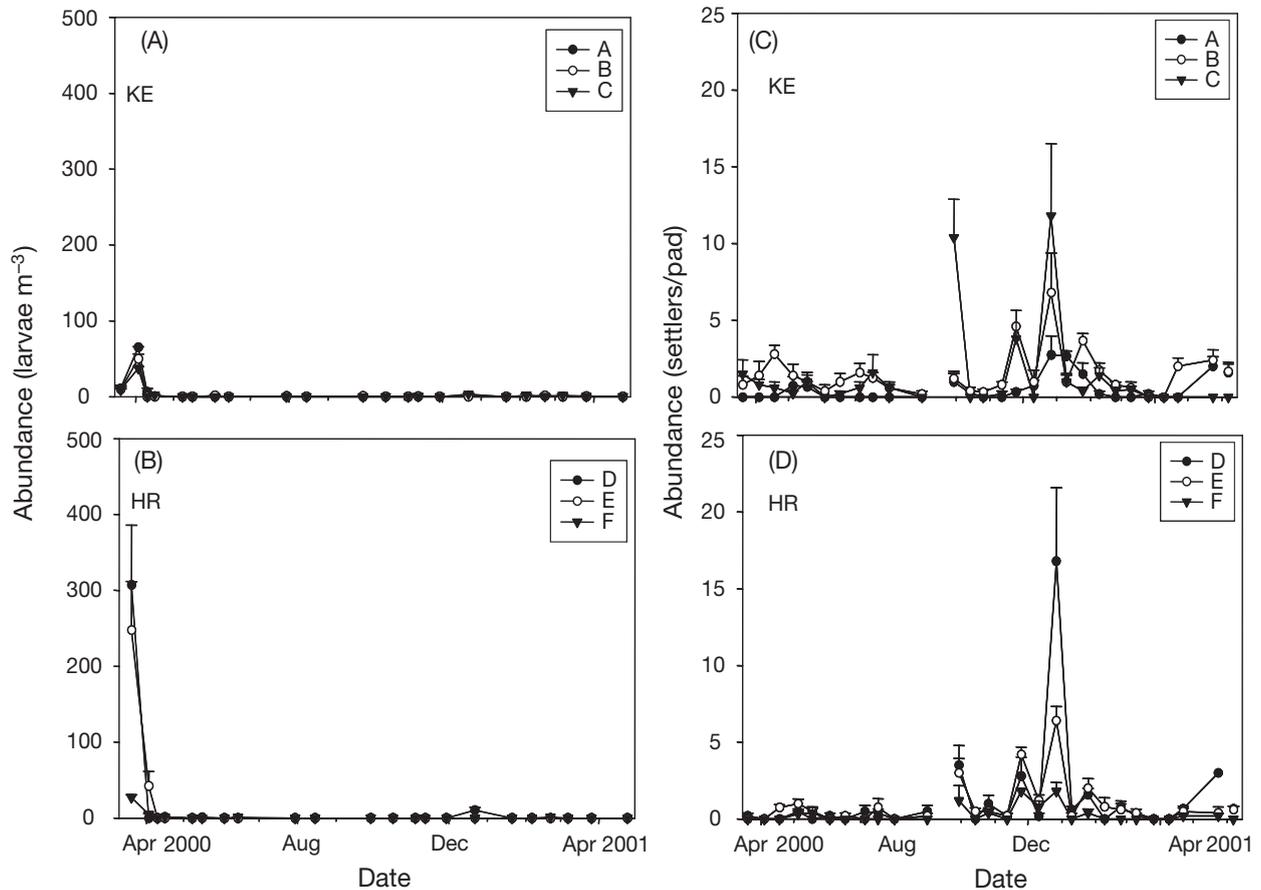


Fig. 1. *Perna perna*. Long-term larval (A,B) availability and (C,D) settlement at Kenton (KE) and High Rocks (HR). Letters in legend boxes refer to locations in Kenton (A, B, C) and High Rocks (D, E, F). Points indicate mean values and error bars show standard errors (SE)

either Location or Site ($p > 0.05$ for both, Table 1), despite drastic differences between sites on the same day and among locations within sites ($p < 0.0001$ for the interactions between Date and Location and Date and Site, Table 1). When larval numbers were high, values peaked at all locations simultaneously, with the single exception of 24 March 2000 at KE. Thus, the interaction between Day and Location reflects the weak spike in larval numbers at KE on 24 March and the fact that, while some locations showed strong peaks, others showed weaker peaks on the same day.

Daily study (2001)

Generally, only small numbers of larvae were collected at both sites and all locations, with extreme variability on a daily scale (Fig. 2C,D). In contrast to 2000, there were no significant effects involving spatial factors and the only significant effects were temporal ($p < 0.0001$; Table 1).

Settlement

Settler sizes were more variable than for larvae, with occasional individuals (23%) of ≥ 2 mm shell length. However, most individuals (60%) were identified morphologically as primary settlers between 200 and 400 μm , and this was the size class used in the analysis.

Long-term study

Settler abundance was generally extremely low throughout the study. Although there were clear differences in abundances at different locations, settlers collected at the 2 sites generally showed similar patterns, with a small peak in abundance on 29 September and a bigger one on 27 December 2000 (Fig. 1C,D).

For the settlement study, Time and the smallest spatial scale investigated (Location) affected the variability of settler abundance ($p < 0.0001$ for both; Table 2). On the whole, similar numbers of settlers were found

at the 2 sites, but the numbers varied significantly among locations ($p < 0.0001$, Table 2).

Daily study (2000)

Again, numbers were very low at both sites, with mean maximum values of about 2 larvae per pad d^{-1} (Fig. 3A,B). Even with the low numbers of settlers collected at both sites, there was variability among days and among locations (Fig. 3A,B).

In the present study only Location affected settlement variability significantly ($p < 0.0001$, Table 2). Location KE B experienced higher values than all other locations (Fig. 3A,B, Student-Newman-Keul's post-hoc comparison).

Daily study (2001)

Again, numbers of settlers were very low with high variability on a daily basis and at the location level (Fig. 3C,D).

The results of the 2001 daily settlement study confirmed the strong spatial variability at location level.

Table 1. ANOVA for larval availability in *Perna perna*. D: Date, S: Site, Loc: Location, d: Day, E: Error, ns: not significant

	df	MS	F	p	Error terms
(a) Long-term study					
D	22	9.071	5.537	<0.0001	D × S
S	1	0.493	5.386	ns	Loc(S)
Loc(S)	4	0.091	0.214	ns	D × Loc(S)
D × S	22	1.638	3.827	<0.0001	D × Loc(S)
D × Loc(S)	88	0.428	2.352	<0.0001	E
E	138	0.182			
(b) Daily study (2000)					
d	16	15.770	4.779	<0.0001	d × S
S	1	2.334	1.786	ns	Loc(S)
Loc(S)	4	1.307	1.363	ns	d × Loc(S)
d × S	16	3.300	3.441	<0.0001	d × Loc(S)
d × Loc(S)	64	0.959	2.411	<0.0001	E
E	102	0.398			
(c) Daily study (2001)					
d	21	13.330	4.298	<0.0001	d × S
S	1	0.363	0.080	ns	Loc(S)
Loc(S)	4	4.545	1.314	ns	d × Loc(S)
d × S	21	3.101	0.896	ns	d × Loc(S)
d × Loc(S)	84	3.459	1.285	ns	E
E	132	2.691			

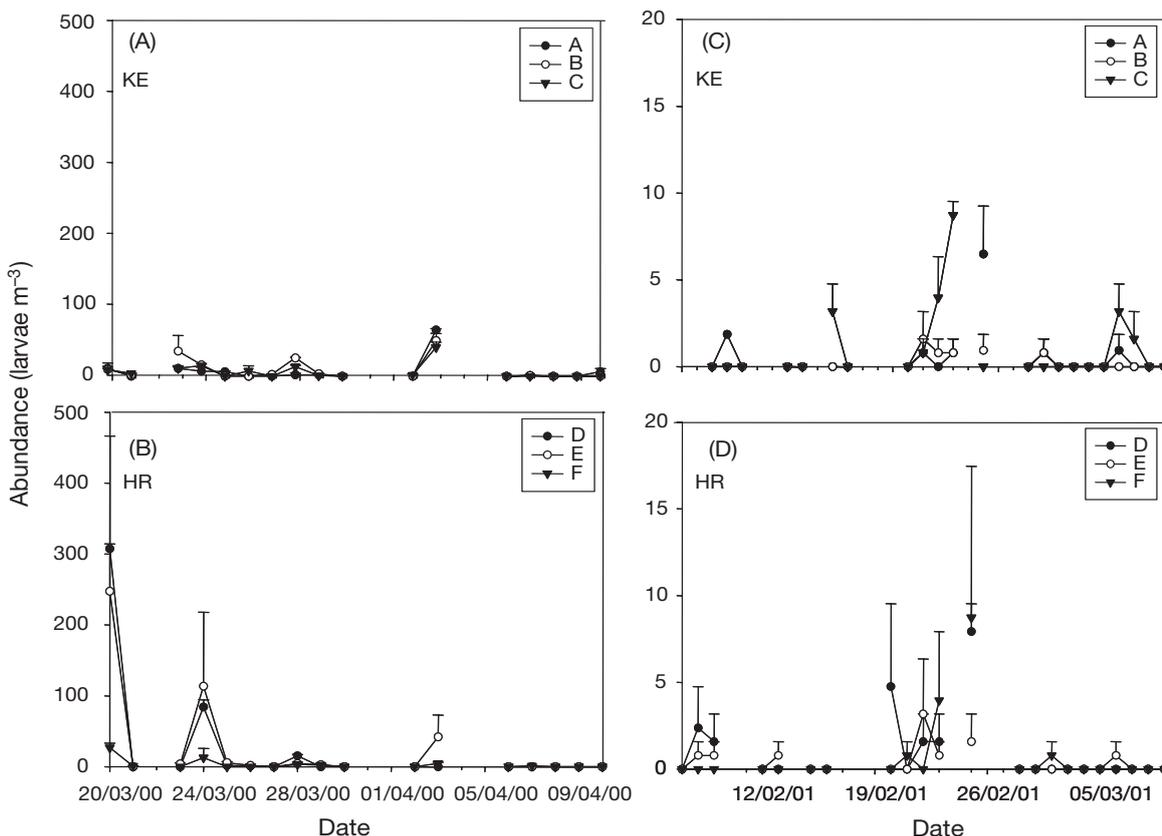


Fig. 2. *Perna perna*. Daily (A,B) 2000 and (C,D) 2001 larval availability at Kenton (KE) and High Rocks (HR). Letters in legend boxes refer to locations in Kenton (A, B, C) and High Rocks (D, E, F). Points indicate mean values and error bars show standard errors (SE). Dates given as d/mo/yr

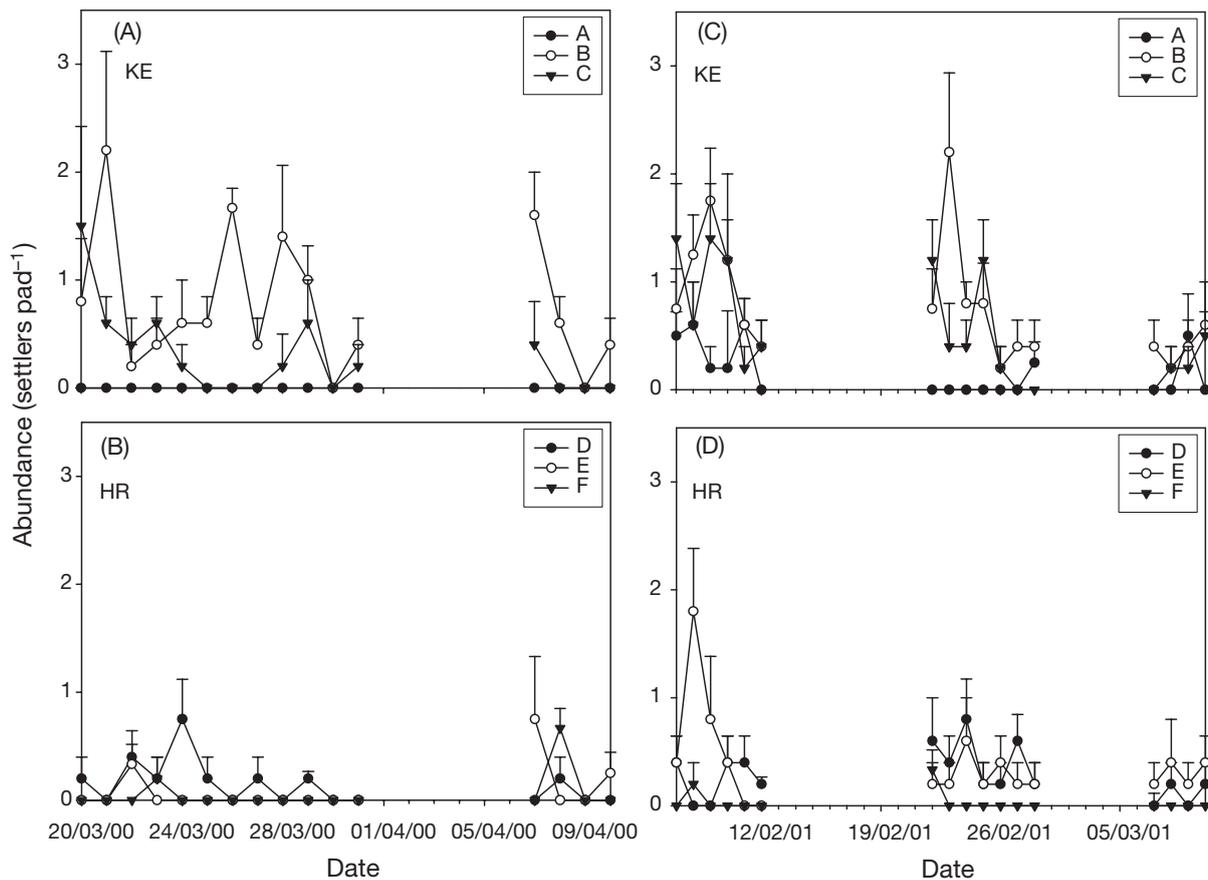


Fig. 3. Daily (A,B) 2000 and (C,D) 2001 settlement at Kenton (KE) and High Rocks (HR). Letters in legend boxes refer to locations in Kenton (A, B, C) and High Rocks (D, E, F). Points indicate mean values and error bars show standard errors (SE). Dates given as d/mo/yr

Table 2. ANOVA for settlement in *Perna perna*. D: Date, S: Site, Loc: Location, d: Day, E: Error, ns: not significant

	df	MS	F	p	Error terms
(a) Long-term study					
D	26	3.848	12.751	<0.0001	D × S
S	1	4.110	1.207	ns	Loc(S)
Loc(S)	4	3.404	8.223	<0.0001	D × Loc(S)
D × S	26	0.301	0.729	ns	D × Loc(S)
D × Loc(S)	104	0.414	2.388	<0.0001	E
E	648	0.173			
(b) Daily study (2000)					
d	15	0.632	0.909	ns	d × S
S	1	8.030	1.322	ns	Loc(S)
Loc(S)	4	6.075	11.714	<0.0001	d × Loc(S)
d × S	15	0.695	1.340	ns	d × Loc(S)
d × Loc(S)	60	0.518	1.850	ns	E
E	384	0.280			
(c) Daily study (2001)					
d	17	1.112	1.772	ns	d × S
S	1	10.027	1.727	ns	Loc(S)
Loc(S)	4	5.805	9.991	<0.0001	d × Loc(S)
d × S	17	0.628	1.080	ns	d × Loc(S)
d × Loc(S)	68	0.581	1.616	<0.001	E
E	432	0.359			

The temporal scale investigated here, Day, did not significantly affect settlement, although numbers varied amongst locations from day to day (Table 2). Again, KE B showed highest settler numbers (post-hoc $p < 0.05$); values at all other locations were generally very low (Fig. 3C,D).

The results emphasise the strong effect of location on settlement at all temporal scales. The larger spatial variable, Site, never influenced settler abundance, though a significant interaction between Time and Site revealed the importance of Site to larval availability during specific sampling events. Neither of the spatial variables had a significant effect as a main factor in any analysis of larval availability. Time, as expected, had an important effect on both larval availability and settlement, with the single exception of the daily settlement studies.

Coupling

There was no correlation ($r < 0.14$ for all studies) between larvae in the water and settlers arriving

ashore. The raw data suggested that the lack of correlation was not due to a lag effect, and this was confirmed by cross-correlation analysis, from which no clear patterns emerged. The long-term data did not show any significant correlation with any lag period. The daily data showed few significant correlations. In 2 cases there were significant correlations between larvae and settlers at the same location with a lag of 2 d. Likewise, analyses comparing settler numbers with larval abundance at other locations occasionally (6 out of 78 spatial comparisons) gave significant results, with lag periods of 1 or 2 d.

DISCUSSION

Menge & Farrell (1989) demonstrated a strong biogeographic component to the intensity of recruitment in marine invertebrates, with much less intense recruitment at lower latitudes, though recent studies on spatial variation of barnacle recruitment over a European scale show that this is not always the case (e.g. O'Riordan et al. 2004). Menge & Farrell (1989) also showed that low settlement rates generally support low adult abundances, although post-recruitment processes (e.g. predation) can also contribute towards the final community structure (Menge & Farrell 1989, Menge 2000). This highlights the importance of larval dispersal, delivery and settlement in determining whether populations are strongly structured by post-settlement factors such as intraspecific competition for space, or whether they are recruitment limited (Connell 1985). Larvae potentially disperse over large distances (Widdows 1991), but hydrology is important and there is evidence that dispersal may be quite limited (Bhaud et al. 1995). For example, mussel larvae on the south coast of South Africa appear relatively poorly dispersed, with most larvae dispersing over scales of less than 10 km (McQuaid & Phillips 2000). In addition, there is evidence of significant differences in recruitment rates on spatial scales of 100s of m within a shore (Lawrie & McQuaid 2001). This could be explained by differences among locations in post-settlement mortality, but the results of the present study also indicate strong differences in daily settlement rates on such scales. Differences in settlement rates over scales of 100s of m could be explained by settlers coming from different concentrations of larvae offshore, or by differential delivery of larvae to the shore. Our study was designed to compare these 2 hypotheses.

Although the results of McQuaid & Phillips (2000) (based on few sampling events) showed drifting of larvae parallel to the shoreline, the present study did not show any correlation between wind and larval abundance or settlement (T. Phillips unpubl. data). We

therefore cannot assume an influence of wind on larval transport. The intertidal stations were physically opposite offshore stations. No correlation was found between numbers of larvae and settlers at paired locations. This is unsurprising, as this type of study requires different sampling techniques for the different life stages examined. Our measurements of settlement were cumulative samples, while larval samples were instantaneous spot samples. So, for example, there were occasions when relatively high numbers of settlers were collected, even though the plankton samples contained few or no larvae. Nevertheless, the lack of correlation suggests the importance of the scales of larval transport and delivery as regulated by physical factors, especially local hydrodynamics. Small-scale vertical migrations can be used to influence dispersal in highly stratified environments such as estuaries (DiBacco et al. 2001), but the nearshore water column on this coast is highly mixed due to strong wind-forcing (Goschen & Schumann 1994) and there is no indication that mussel larvae exhibit diel or other vertical migrations (McQuaid & Phillips 2000). It thus seems most likely that they are dispersed as passive particles (Bourget 1988). Furthermore, Caceres et al. (1994) found that mussel settlers do not initially show active substratum selection, but tend to stay on the substratum they first encounter through random collisions, and that wave action can control settlement through subsequent mechanical dislodgment. Similarly, laboratory studies by Pernet et al. (2003) showed that mussels can select substrata at low flow rates, but that at high flow rates, as in nature and especially on this coastline, they behave as passive particles.

A methodological problem with the present study is that the time scales of sampling could have partially hidden correlations between larval and settler abundances of *Perna perna*. An integrated sampling design recording larval abundance in the inshore waters over a period of time is highly desirable (Yund et al. 1991, Setran 1992, Whitlatch & Osman 1998, Todd 2003) and the use of intertidal larval traps (Castilla & Varas 1998) was attempted. However, this approach was not possible, as traps became completely clogged with sand within a matter of hours. Similarly, moored plankton nets did not work because of strong wave action.

Our results confirmed high variability in time and space for both larval and settler abundances. Temporal variability in both was great on coarse (seasonal) and fine (daily) scales, with no obvious structure to it. Consequently, reliable or precise predictions of the presence or arrival of larvae cannot be made on this coast. The best we can do is to identify seasonal periods when settlement rates *may* be high. Throughout the whole study, larvae were abundant only for the period March to April 2000. Otherwise, larvae were rare in

the water column, in contrast to other studies on the same species in the same area (F. Porri unpubl. data) and 300 km farther east (Lasiak 1986). Even within months, day-to-day variability in larval abundance was very high in both years. There were also often significant interactions between Time and the smallest spatial scale investigated, Location. Larvae were often abundant at a particular location on a particular day, but not the day after or before.

In contrast to larval data, there were distinct peaks in settler abundance throughout the 16 mo study, with highest abundances in summer. However, even in summer, hydrodynamic conditions in the southern hemisphere are probably less predictable than in the northern hemisphere (Dame 1996), and there was considerable variation in settlement intensity during this season. Coarse temporal effects due to changes in currents from one month to another and among seasons are not unexpected, but fine scale temporal variation in the arrival of larvae on the shore is also influenced by local hydrodynamics (Gaines et al. 1985). Localities may experience settlement at certain times, but not hours, minutes or seconds later, when the water may be calmer or the state of the tide may have changed (Pineda 2000).

Many studies suggest that large spatial scale variation controls the distribution and abundance of larvae offshore, while larval settlement is more influenced by fine scale variability (e.g. Connell 1985, Gaines & Roughgarden 1985, Pineda 2000). There was little pattern in the spatial variability of larval abundance and, although larval densities varied between sites and locations, neither the site (km) nor the location (100s of m) scales had significant effects. In contrast, location consistently showed a significant influence on settler abundances. The effect of location on settler abundance indicates differential delivery of larvae to this shore at scales of hundreds of meters, i.e. larvae are preferentially transported to some locations rather than others, in this case KE B. Possible explanations for this are that the arrival of larvae at the shore could be locally controlled by coastal morphology, local topography, wave action on the shore (Connell 1985, Jeffery & Underwood 2000) and by finer scale local micro hydrodynamics, influenced by surface roughness, rock cracks, presence of sessile individuals, pools, channels, protected and exposed shelves (Pineda 2000).

There are other factors that could have influenced differential arrival of larvae among locations. In fact, although we used a standard substratum for settlement, our results could have been influenced by the intensification effect shown in barnacles (Pineda & Caswell 1997), i.e. settlement patterns influenced by differential availability of suitable substrata. However, there were no obvious differences in mussel cover

among locations and differential delivery is a simpler explanation.

Moreover, active choice of substratum or aggregation behaviour, as shown for adult mussels (de Vooy 2003), has not been demonstrated for mussel larvae. Searching for settlement spots or differential availability of substratum can dramatically influence spatial patterns of settlement and survival. For example, on very small spatial scales (cm to μm), substratum topography or quality (e.g. presence of algal films) can strongly affect survival after settlement of barnacles, hydrozoans and bryozoans (Walters & Wetthey 1996). However, as we used standard artificial settlement substrata (plastic scouring pads) such effects should not influence our results.

Larval availability and settlement are among those parameters with crucial biological importance for the regulation of the spatial and temporal distribution and abundance of marine mussel populations (Morgan 2001, Underwood & Keough 2001). The results of this study indicate that the physical scales of variability in larval and settler abundances differ, suggesting that, on scales of 100s of m, differential delivery, rather than differential supply of larvae in the water, determines the final distribution of settlers on the shore. We suggest that differential delivery, caused by the effects of small-scale topography on local hydrodynamics, is the best explanation for the transport of larvae to certain locations rather than others.

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