

A new model to assess the probability of occurrence of a species, based on presence-only data

G. Beaugrand^{1,2,*}, S. Lenoir¹, F. Ibañez³, C. Manté⁴

¹Centre National de la Recherche Scientifique, Laboratoire d'Océanologie et de Géosciences UMR LOG CNRS 8187, Station Marine, Université des Sciences et Technologies de Lille – Lille 1, BP 80, 62930 Wimereux, France

²Sir Alister Hardy Foundation for Ocean Science, The Laboratory, Citadel Hill, Plymouth PL1 2PB, UK

³Laboratoire d'Océanographie de Villefranche (LOV) BP 28, 06234 Villefranche-sur-Mer CEDEX, France

⁴Centre d'Océanologie de Marseille, UMR CNRS 6117 LMGEM, Campus de Luminy, Case 901, 13288 Marseille CEDEX 09, France

ABSTRACT: This study aims to describe a new nonparametric ecological niche model for the analysis of presence-only data, which we use to map the spatial distribution of Atlantic cod and to project the potential impact of climate change on this species. The new model, called the Non-Parametric Probabilistic Ecological Niche (NPPEN) model, is derived from a test recently applied to compare the ecological niche of 2 different species. The analysis is based on a simplification of the Multiple Response Permutation Procedures (MRPP) using the Generalised Mahalanobis distance. For the first time, we propose to test the generalized Mahalanobis distance by a non-parametric procedure, thus avoiding the arbitrary selection of quantile classes to allow the direct estimation of the probability of occurrence of a species. The model NPPEN was applied to model the ecological niche (*sensu* Hutchinson) of Atlantic cod and therefore its spatial distribution. The modelled niche exhibited high probabilities of occurrence at bathymetry ranging from 0 to 500 m (mode from 100 to 300 m), at annual sea surface temperature of from -1 to 14°C (mode from 4 to 8°C) and at annual sea surface salinity ranging from 0 to 36 (mode from 25 to 34). This made the species a good indicator of the sub-arctic province. Current climate change is having a strong effect on North Sea cod and may have also reinforced the negative impact of fishing on stocks located offshore of North America. The model shows a pronounced effect of present-day climate change on the spatial distribution of Atlantic cod. Projections for the coming decades suggest that cod may eventually disappear as a commercial species from regions where a sustained decrease or collapse has already been documented. In contrast, the abundance of cod is likely to increase in the Barents Sea.

KEY WORDS: Ecological niche models · Multiple response permutation procedure · Generalised Mahalanobis distance · Atlantic cod

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INTRODUCTION

The effects of climate change on living systems in both the terrestrial and the marine realms are now well documented (Parmesan & Matthews 2006, IPCCWG1 2007a). In the marine biosphere, current climate change is affecting the abundance, spatial distribution and phenology of species, as well as altering prey-predator interactions (Beaugrand et al. 2002, Beau-

grand et al. 2003, Edwards & Richardson 2004). The effect of climate change is seen from phytoplankton (Reid et al. 1998) to zooplankton (Beaugrand et al. 2007) and fish (Brander et al. 2003, Perry et al. 2005), and it translates from the physiological to the ecosystem level (Pörtner & Farrell 2008), affecting coupling between systems (i.e. benthic-pelagic coupling; Reid & Edwards 2001, Kirby et al. 2008). Pronounced climate change may become a confounding factor of fishing,

*Email: Gregory.Beaugrand@univ-lille1.fr

and both climate change and fishing may act in synergy to precipitate the collapse of fish stocks around the world (Beaugrand & Kirby 2010b,a). To better evaluate the effect of climate on a species, it is essential to know its spatial distribution; this information is often lacking in the marine realm.

One way to evaluate the spatial distribution of a species is to use Ecological Niche Models (ENMs). ENMs, also known as bioclimatic envelopes, are being more frequently used in the context of global change and are often based on the concept of the ecological realised niche described by Hutchinson (1957). The realised niche is the environmental envelope in which a species can be found when the effect of dispersal and interspecific relationships are considered. ENMs have been used in conservation to manage endangered species (Sanchez-Cordero et al. 2005), to predict the responses of species to climate change (Berry et al. 2002), to forecast past distribution (Bigg et al. 2008) and to estimate the potential invasion of a non-native species (Peterson & Vieglais 2001). When quantitative data are available, regression techniques such as Generalised Linear Models (GLMs; McCullagh & Nelder 1983) or Generalised Additive Models (GAMs; Hastie & Tibshirani 1990), ordination or neural networks have been frequently used (Guisan & Zimmermann 2000, Guisan & Thuiller 2005). When only binary (presence-absence) data are available there are far fewer techniques that can be applied, although regression techniques such as GAMs can still be utilised. Traditional models such as BIOCLIM (based on a multilevel rectilinear envelope) and DOMAIN (based on a point-to-point similarity metric; Carpenter et al. 1993) tend to be relatively simple, although more sophisticated models have been developed recently, such as Ecological Niche Factor Analysis (ENFA, (Hirzel et al. 2002) and MAXENT (Phillips et al. 2006), which are based upon Principal Component Analysis (PCA) and the principle of maximum entropy, respectively.

The objective of this study is to describe a new nonparametric ENM adapted to presence-only data. This technique is based on a modified version of the Multiple Response Permutation Procedure (MRPP; Mielke et al. 1981), using the Generalised Mahalanobis distance (Ibañez 1981). (1) We present a rationale and describe the technique. (2) We present a simple example of a calculation based on simulated data in order to illustrate the technique and to justify the use of the generalised Mahalanobis distance. (3) We use the technique to model the ecological niche of the Atlantic cod and map its probability of occurrence. Then we project the probability of cod occurrence for the middle and the end of the 21st century based upon IPCC scenarios for changes in sea surface temperature.

MATERIALS AND METHODS

Physical data

Bathymetry data were obtained from a global ocean bathymetry chart (1° longitude × 1° latitude; Smith & Sandwell 1997). This dataset is among the most complete, high-resolution image of sea floor topography currently available. The chart was constructed from data obtained from ships with detailed gravity-anomaly information provided by the satellites GEOSAT and ERS-1 (Smith & Sandwell 1997). Bathymetry data were considered because the spatial distribution of Atlantic cod, which occurs mainly over continental shelves (Sundby 2000), is explained partially by this parameter.

Salinity has a strong impact upon the distribution of most fishes. Annual Sea Surface Salinity (SSS) data for a depth ranging from 0 to 10 m, were obtained from the Levitus climatology (Levitus 1982). International Council for the Exploration of the Sea (ICES) data were used to complete the Levitus dataset in coastal regions (e.g. some regions of the eastern English Channel) where there was no assessment of annual SSS. ICES data were downloaded from www.ices.dk. We did not include temporal changes in salinity because the parameter is not presently well assessed in the Atmosphere-Ocean General Circulation (AO-GCM) models (M. Visbeck, pers. comm.). Furthermore, the spatial variance in the salinity is much more pronounced than the temporal variance.

The spatial distribution of cod is affected by temperature (Brander 2000, ICES 2007). Sea Surface Temperature (SST) data from 1960 to 2005 were taken from the International Comprehensive Ocean-Atmosphere Data Set (ICOADS) at <http://icoads.noaa.gov>. Longitudes had a spatial resolution of 1° longitude × 1° latitude (Woodruff et al. 1987). An annual mean was calculated for the period from 1960 to 2005. SST data were considered, as this parameter has a strong effect on the spatial distribution of cod (Brander 2000, ICES 2007). The use of SST to assess the niche of adult cod assumes that climate exerts its major influence on cod through the effects of temperature on larval development and plankton food availability, since the pelagic larval stage is a critical life cycle phase affecting recruitment (Beaugrand & Kirby 2010b,a).

To assess the potential effect of changes in SST, data were obtained for the period 1990 to 2100 from the AO-GCM ECHAM 4 (EC = European Centre and HAM = Hamburg; Roeckner et al. 1996). These data are projections of monthly skin-temperature equivalent above the sea to SST (<http://ipcc-ddc.cru.uea.ac.uk>). Data used here are modelled based on scenario A2 of the IPCC Special Report on Emissions Scenarios (SRES), which scenario has the atmospheric concentra-

tion of CO₂ reaching 856 ppm by 2100, and on SRES scenario B2, where atmospheric CO₂ reaches 621 ppm by 2100 (IPCCWG1 2007b). Scenario A2 supposes a rate of increase of CO₂ similar to that currently observed. SRES scenarios A2 and B2 reflect world populations of 15.1 and 10.4 billion people in 2100, respectively (IPCCWG1 2007b).

We also used data from the Hadley Centre Coupled Model, version 3 (HadCM3; Gordon et al. 2000). Data were used from both SRES scenarios A1B and B1. Scenario A1B reflects a world of rapid economic growth, low population growth and rapid introduction of new and more efficient technology, whereas scenario B1 reflects a world with rapid introduction of resource-efficient technologies (IPCCWG1 2007b). Two non-SRES scenarios were also utilised: PICNTRL (i.e. experiments run with constant pre-industrial levels of greenhouse gasses) and COMMIT (i.e. idealised scenario in which the atmospheric burdens of long-lived greenhouse gasses are held fixed at the 2000 level). The Hadley Centre Global Environmental Model, version 1 (HadGEM1), was also used with the non-SRES scenario 1PTO4x (1% to quadruple) in which greenhouse gasses increase from pre-industrial levels at a rate of 1% per year until the concentration has quadrupled and becomes constant thereafter (Johns et al. 2006). This is the most pessimistic of all scenarios we considered. In sum, a total of 7 scenarios (A2, B2, A1B, B1, PICNTRL, COMMIT and 1PTO4x) was used from 3 different AO-GCMs (ECHAM4, HadCM3, HadGEM1). The best estimate of global temperature increase is 0.6°C for COMMIT, 1.8°C for Scenario B1, 2.4°C for Scenario B2, 2.8°C for Scenario A1B and 3.4°C for Scenario A2 (IPCCWG1 2007b). These forecasted SST datasets were used (1) to examine how the probability of occurrence of cod varied as a function of the intensity of warming and AO-GCMs and (2) to identify regions most susceptible to the influence of intense warming.

All physical data were interpolated bilinearly on a spatial grid of 0.1° longitude × 0.1° latitude in a spatial domain ranging from 80.50°W to 70.50°E and from 35.50°N to 70.50°N (Fig. 1). We used a high spatial resolution to decrease potential bias that could arise from the averaging of bathymetry data in a large geographical cell. The high resolution also enabled us to better assess the probability of cod occurrence along coastline. While only one map

of bathymetry and SSS (annual climatology) was generated, a grid for each year of the period from 1960 to 2006 was built for SST. Therefore, the spatial distribution of cod varies according to the 3 abiotic parameters, while year-to-year changes were a function of SST only.

Fish data

Data of cod occurrence were taken from Fishbase (www.fishbase.org; Froese & Pauly 2009). This represented a total of 52630 data points. Unfortunately, while high densities of data are present in the dataset on the western side of the North Atlantic this is not the case on the eastern side. Of the 52630 data points taken from Fishbase, only 9638 were located to the east of 30°W. Therefore, we completed the dataset from our knowl-

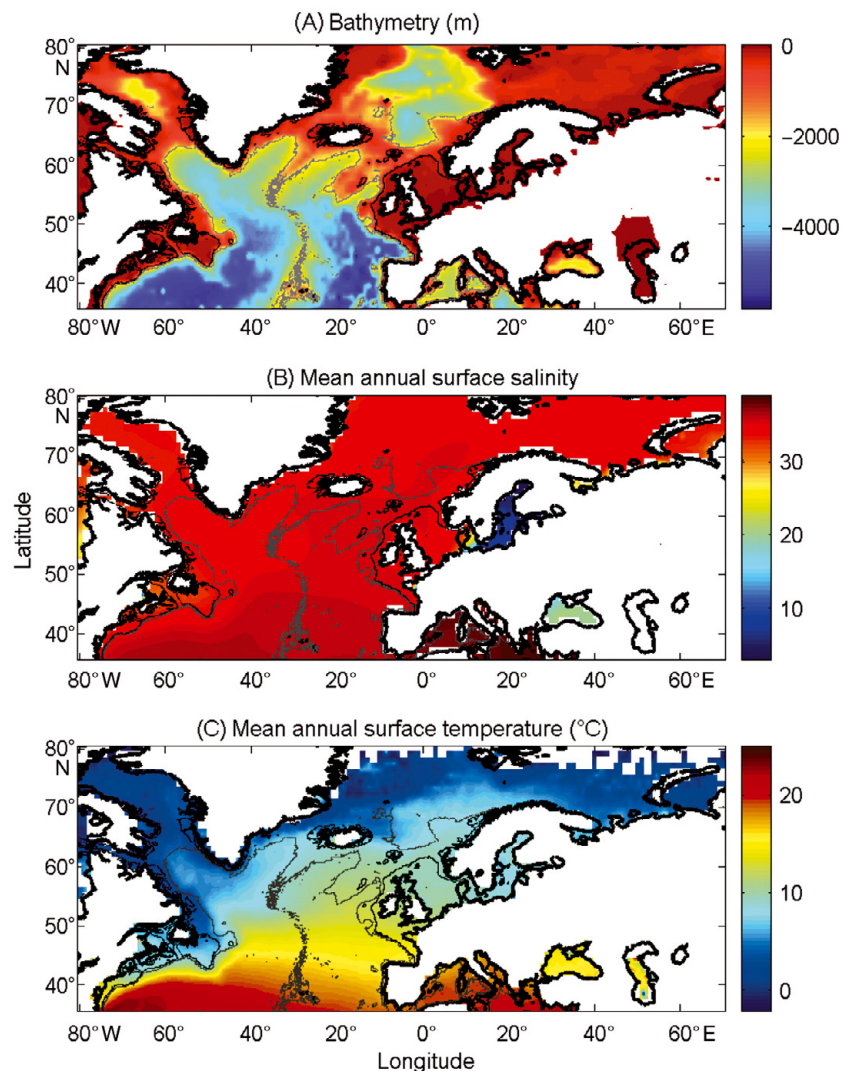


Fig. 1. Spatial distribution of (A) bathymetry, (B) mean annual sea surface salinity and (C) mean annual sea surface temperature. Isobaths: grey lines (200 and 2000 m)

edge of the spatial distribution of the species (ICES 2005, Brander et al. 2006, Heath & Lough 2007, ICES 2007) (Fig. 2). The data largely reflect the occurrence of cod over 1 yr (www.fishbase.org), although no distinction was made on age. Data therefore originated from scientific cruises, agencies, museums, university, nongovernmental organizations, commercial catches, occasional fishermen and expert knowledge. The total number of data points equalled 140026 observations. For each observation of cod occurrence, information on SST, bathymetry and SSS were added to each data point by interpolation of each environmental data point from the datasets described above (see section on physical data).

To compare results of our ecological niche model, we used probability data of cod occurrence obtained from the numerical procedure Aquamaps (www.fishbase.org). This derives from the Relative Environmental Suitability (RES) model that was initially developed for mapping mammalian species distribution (Kaschner et al. 2006) and that has been adapted subsequently to map the probability of occurrence of all marine organisms. A total of 62 160 data points were used to produce the probability map. Although no distinction was made on age, the data reflect mainly the occurrence of cod ≥ 1 yr old (www.fishbase.org).

Description of the Non Parametric Probabilistic Ecological Niche (NPPEN) model

The model is derived from a test recently applied to compare the ecological niche of 2 species (Beaugrand

& Helaouët 2008). The analysis is based on Multiple Response Permutation Procedures (MRPP), a test first proposed by Mielke et al. (1981). MRPP has been applied in conjunction with Split Moving Window Boundary analysis to detect discontinuities in time series (Cornelius & Reynolds 1991). This method has also been used to identify abrupt ecosystem shifts (Beaugrand 2004, Beaugrand & Ibañez 2004). Mathematically, MRPP tests whether 2 groups of observations in a multivariate space are significantly separated. Mielke et al. (1981) gave a full description of the test and Beaugrand & Helaouët (2008) have recently illustrated an adaptation of the test to compare 2 ecological niches.

The model we propose for assessing the probability of cod occurrence (and its changes in space and time) is in fact a simplification of MRPP. Instead of comparing 2 groups of observations, our new analysis tests whether one observation belongs to a group of reference observations we call here the reference matrix. The reference matrix is represented by a matrix $\mathbf{X}_{n,p}$ with n the number of reference observations and p the number of variables. Each row of the matrix represents the environmental conditions where a species is detected. It is crucial that the reference matrix cover the entire niche of a species to give a reliable probability (Thuiller 2004); we check this point, which is often forgotten in this kind of exercise, by compiling histograms. The predictive matrix $\mathbf{Y}_{m,p}$ encompasses m observations of the environment using p predictors. Each observation of \mathbf{Y} (environmental conditions) is then tested against \mathbf{X} (range of conditions where the species was detected). The model is applied in 4 main steps:

Step 1: Homogenization of the reference matrix.

The density of cod occurrence reported in some databases (e.g. Fishbase) depends on fishing activities, and it is clear that the density of data points is higher in fishing areas. Although this suggests that the resource is more abundant in those regions, this is not completely true. This phenomenon can potentially influence the outcome of any ecological niche model. Another phenomenon that can influence probability of occurrence is the inaccurate reporting of occurrence. In an attempt to overcome these drawbacks, we created a virtual cube (i.e. 3 controlling factors) with intervals of SST of 1°C between -2°C and 18°C, intervals of bathymetry of 20 m between 0 and 800 m and intervals of annual SSS of 2 between 0 and 40. We retained one data occurrence when more than one observation was made in the crossed intervals of annual SST, bathymetry and annual SSS. This threshold was fixed to eliminate the influence of one single misreporting. The resolution might, at first sight, appear to be coarse. However, the practice of the

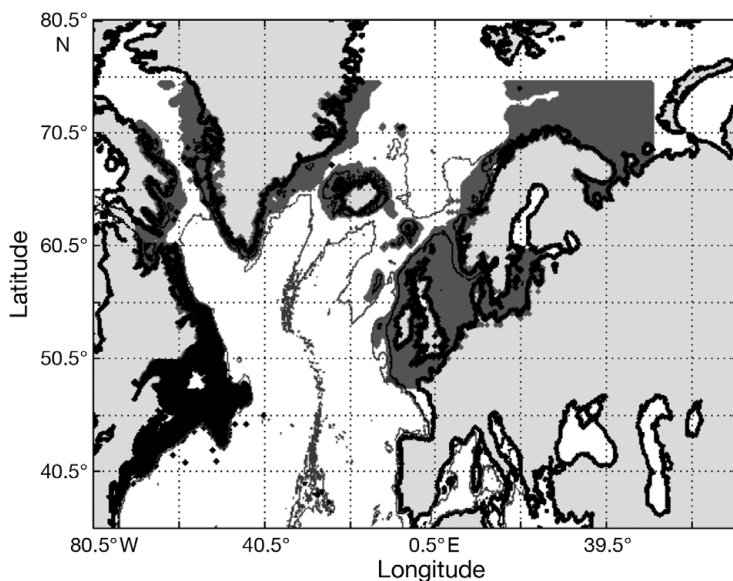


Fig. 2. *Gadus morhua*. Spatial distribution of occurrence data points: observed (Fishbase data, 52 630 datapoints, black zones along the coastlines) and inferred (104 642 datapoints, dark grey zones). Isobaths: 200 m (dark grey line) and 2000 m (light grey line)

ENM indicates that the probability remains similar, even at a lower resolution (see Fig. 4), which depends on the size of the reference matrix. Here, the size was equal to 140 026 observations. This amount divided by 16 000 (20 SST intervals \times 40 bathymetric intervals \times 20 SSS intervals) = 8.75 observations per geographical cell. Such a calculation obviously assumes that observations are equally distributed, which was clearly not the case. A resolution of 0.1°C for SST, 1 m for bathymetry and 0.1 unit of SSS would appear to be too sensitive, as only 0.02 observations per crossed interval are expected if observations were randomly distributed. The determination of the threshold also depends on the uncertainties of the physical variables.

Step 2: Preparation of data. A matrix called $\mathbf{Z}_{n+1,p}$ is created for each observation of \mathbf{Y} to be tested against \mathbf{X} . For the first observation, the following matrix is constructed:

$$\mathbf{Z}_{n+1,p} \begin{bmatrix} Y_{1,1} & Y_{1,2} & \cdots & Y_{1,p} \\ X_{1,1} & X_{1,2} & \cdots & X_{1,p} \\ \cdot & \cdot & \cdot & \cdot \\ \cdot & \cdot & \cdot & \cdot \\ X_{n,1} & X_{n,2} & \cdots & X_{n,p} \end{bmatrix} \quad (1)$$

where $x_{i,j}$ is the observation in matrix \mathbf{X} and $y_{i,j}$ is an observation of matrix \mathbf{Y} . The building of matrix \mathbf{Z} is repeated m times, corresponding to the m observations of \mathbf{Y} .

Step 3: Calculation of the mean multivariate distance between the observation to be tested and the reference matrix. MRPP was first proposed for application with an Euclidean distance, a squared Euclidean distance or a chord distance (Mielke et al. 1981). To first illustrate the technique, we use a Euclidean distance. Obviously, if variables do not have the same unit or dimension, such a distance should be avoided. The Euclidean distance is calculated as follows:

$$d(z_i, z_j) = \sqrt{\sum_{j=1}^p (z_{1,j} - z_{i,j})^2} \quad (2)$$

With $z_{1,j}$ the first observations for the j th variable, originally the observation of the variable of matrix \mathbf{X} , $1 \leq j \leq p$; $z_{i,j}$, the observation i of the variable j in matrix \mathbf{Z} with $2 \leq i \leq n+1$ and $1 \leq j \leq p$. Then, the average observed distance ε_o is calculated as follows:

$$\varepsilon_o = \frac{\sum_{i=1}^n d_i}{n} \quad (3)$$

where n , the total number of Euclidean distances, is equal to the number of observations in the training set \mathbf{X} .

Step 4: Calculation of the probability that the observation belongs to the reference matrix. The mean Euclidean distance is tested by replacing each obser-

vation of \mathbf{X} by y in \mathbf{Z} from row 2 to $n+1$. The number of maximum permutations is equal to n . After each permutation, the mean Euclidean distance ε_s is recalculated, with $1 \leq s \leq n$. A probability v can be assessed by looking at the number of times a simulated mean Euclidean distance is found to be greater than or equal to the observed mean Euclidean distance between the observation and the reference matrix \mathbf{X} .

$$v = \frac{q_{\varepsilon_s \geq \varepsilon_o}}{n} \quad (4)$$

where the probability v is the number of times the simulated mean Euclidean distance was found greater or equal to the observed mean distance. When $v = 1$, the observation has environmental conditions that represent the centre of the species niche. When $v = 0$, the observation has environmental conditions outside the species niche. It is essential to emphasise that the niche and its borders have to be correctly assessed. Applying the procedure to each observation of $\mathbf{Y}_{m,p}$ leads to a matrix $\mathbf{V}_{m,1}$ of probability. It is important to have a large reference matrix so that the resolution of the probability is as high as possible. The resolution R of the probability is:

$$R = \frac{1}{n} \quad (5)$$

where n is the number of reference observations in \mathbf{X} . Ideally, R should be < 0.05 .

Simple example of application of the model using the Euclidean distance. To illustrate the principle of the technique, we present a hypothetical case where the reference matrix \mathbf{X} has $n = 3$ observations and $v = 2$ controlling factors while the predictive matrix \mathbf{Y} has $m = 1$ observation (Fig. 3). Calculations of the 3 Euclidean distances between y and the reference observations x give $d_{(y-x_1)} = 2.236$, $d_{(y-x_2)} = 2.236$ and $d_{(y-x_3)} = 1.803$ (Fig. 3a). The average observed distance ε_o is = 2.092. The simulated distances are $\varepsilon_{s1} = 1.589$ (Fig. 3b), $\varepsilon_{s2} = 1.383$ (Fig. 3c) and $\varepsilon_{s3} = 1.140$ (Fig. 3d). The probability is therefore equal to 0. Observation y has environmental conditions not compatible with the species ecological niche inferred here from 2 variables.

Selection of a better coefficient of distance for Step 2. Mielke et al. (1981) used mainly the Euclidean, squared Euclidean and chord distances. However, in the context of habitat modelling, the use of the Euclidean (squared or not) distance in step 2 is inappropriate in most (if not all) cases and the chord distance is often a better approach (Beaugrand & Helaouët 2008). The computation of the chord distance is achieved by normalizing each vector of \mathbf{Z} to 1 prior to the calculation of the Euclidean distances; this is a special kind of scaling (Legendre & Legendre 1998). Each element of the vector is divided by its length, using the Pythagorean formula to ensure that

each variable carries the same weight in the analysis. In our study, the normalization of elements of $Z_{n+1,p}$ per Eq. (1) would be:

$$Z_{n+1,1}^* = \frac{1}{\sqrt{Y_{1,1}^2 + x_{1,1}^2 + \dots + x_{n,1}^2}} \begin{bmatrix} Y_{1,1} \\ x_{1,1} \\ \vdots \\ x_{n,1} \end{bmatrix} \quad (6)$$

where x and y are as in Eq. (1). Here however, we prefer the use of the Mahalanobis generalised distance, which is independent of the scales of the descriptors (as is the chord distance) but which also takes into consideration the covariance (or the correlation) among descriptors (Ibañez 1981). The Mahalanobis generalised distance has been frequently used recently in this context (e.g. Nogués-Bravo et al. 2008). Prior to the calculation of the distance, standardisation of Z is accomplished by the following transformation:

$$z_{ij}^* = \frac{z_{ij} - \bar{z}_j}{S_j} \quad (7)$$

where z_{ij} are observation i of the j th variables in Z , \bar{z}_j the average value of variable j and S_j the standard deviation of variable j in Z . To calculate the Mahalanobis generalised distance between each observation of the environment y_i ($1 \leq i \leq m$) and all observations of the training set x_j ($1 \leq j \leq n$), we used a particular form of the generalized distance, giving the distance between any observation and the centroid of a unique group (Ibañez 1981):

$$D_{z_1, z_{n,p}}^2 = k'R^{-1}k \quad (8)$$

where $R_{p,p}$ is the correlation matrix of the standardized table Z^* (mean 0 and variance 1), $k_{1,p}$ is the vector of the differences between values of the p variables at z_1^* of standardized matrix Z^* and the mean $\bar{z}_{i \neq 1, j}^*$ of the p variables in the standardized matrix Z^* . Therefore in Step 3, the Euclidean distance was replaced by the use of the Mahalanobis generalised distance.

Analyses

Analysis 1. A comparison of the model based on a chord distance and the Mahalanobis generalised distance was performed using an example of 2 variables and in 2 cases: no correlation between the 2 variables ($r = 0$, a training set of 25 observations), and a strong correlation between the 2 variables ($r = 0.82$, a training set of 13 observations) (Fig. 4).

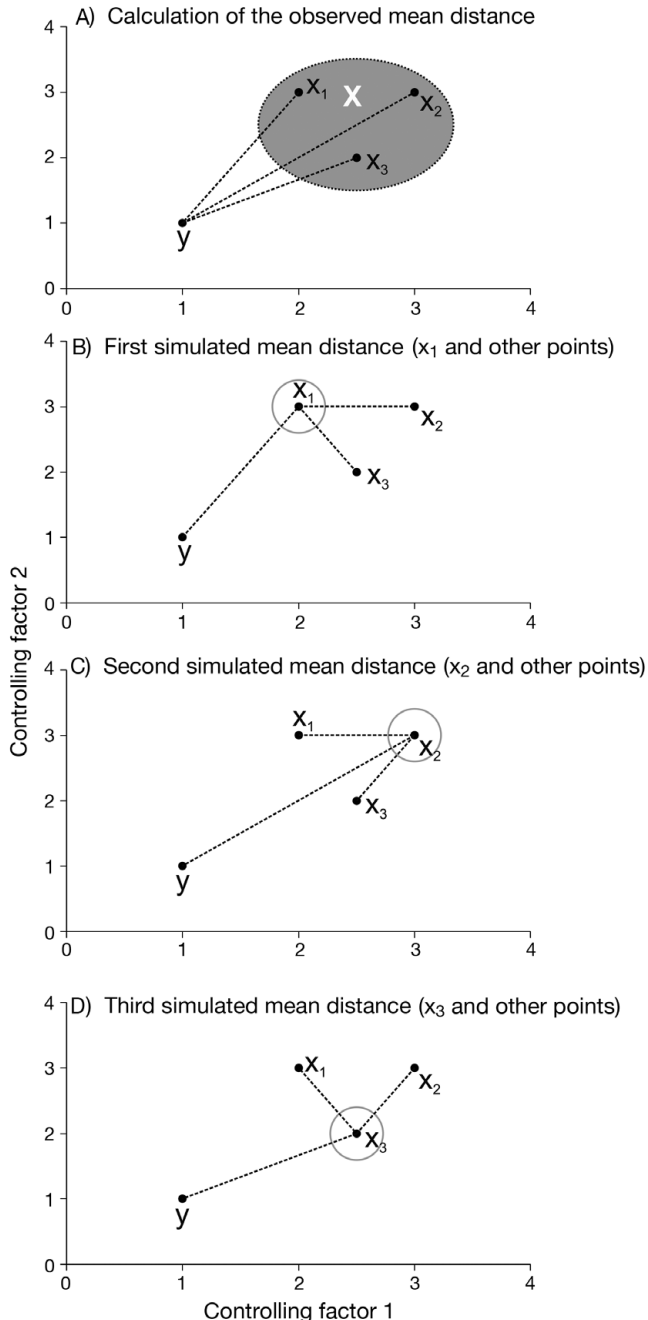


Fig. 3. Principles of the calculation of the niche model that lead to probability of occurrence of a species. (A) Hypothetical observations to be tested against a training set (X) composed of 3 observations in the space of 2 controlling factors. Three Euclidean distances are first calculated and then the average observed distance between the observation to be tested and the ones of the training set is assessed. (B) Recalculation of the mean distance after permutation of the first observation (x_1) of the training set by the observation to be tested. (C) Recalculation of the mean distance after permutation of the second (x_2) observation of the training set by the observation to be tested. (D) Recalculation of the mean distance after permutation of the last observation (x_3) of the training set X by the observation to be tested. All calculated Euclidean distances are indicated by a dashed line. The number of times the simulated mean distance is found inferior to the observed mean distance defines the probability of finding the species in a region

Analysis 2. The procedure of homogenization was illustrated by compiling histograms of each predictive variable (annual SST, annual SSS, bathymetry) for all geographical cells of the spatial domain covered by this study, and for both the original and corrected training (or reference) set (Fig. 5).

Analysis 3. The model was then applied to project the spatial distribution of occurrence of Atlantic cod in order to enable the characterization of its ecological niche (realized niche) as a function of annual SST, annual SSS and bathymetry (Fig. 6). The ecological niche was then projected in the spatial domain as a combined function of the 3 environmental parameters (observed annual SST, annual SSS and bathymetry) for the 1960s and the period from 2000 to 2005 (Fig. 7). Some projections of changes in the spatial distribution of cod occurrence, based on modelled SST (scenarios A2 and B2; annual SSS and bathymetry), were provided for the 1990s (for comparison purposes), the 2050s and the 2090s (Figs. 8 & 9).

RESULTS

Analysis 1

A simple example of a training set composed of 2 variables (correlated or not) illustrated well the difference between the chord and the Mahalanobis generalised distances (Fig. 4). When the correlation between 2 parameters was not different from 0, the ecological niche model based on the chord distance and the Mahalanobis generalised distance gave similar results (Fig. 4A,B). However, when the variables of the training set were highly correlated, the model gave improved results when it was based on the Mahalanobis generalised distance (Fig. 4C,D). This fictional example shows that the chord distance should not be used when ecogeographical variables are correlated. Therefore, we did not calculate the probability of cod occurrence based on this distance.

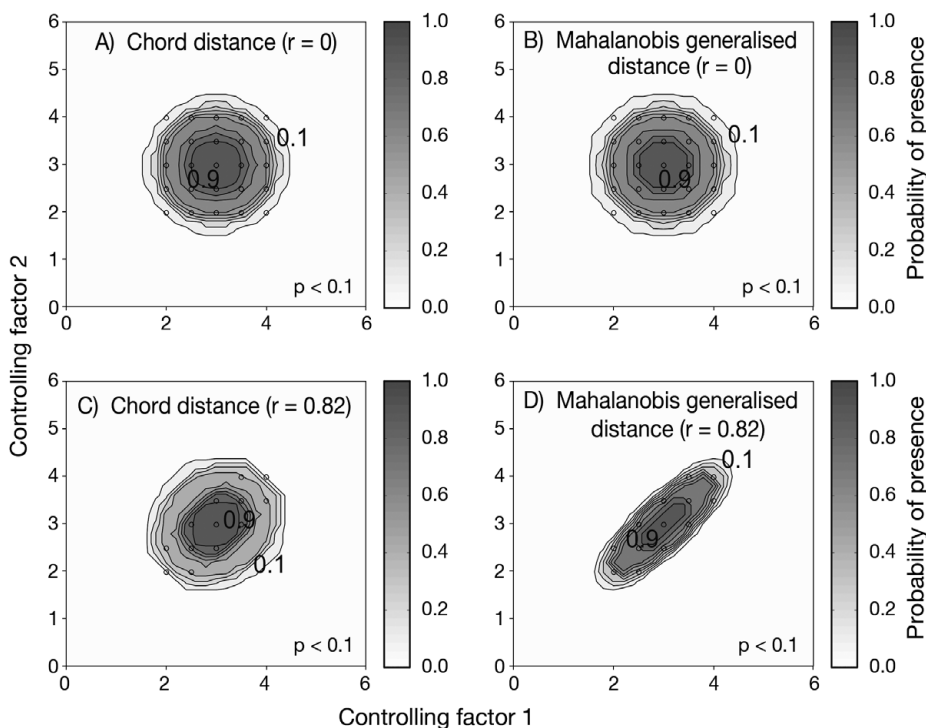


Fig. 4. Fictive examples that show the better performance of the Mahalanobis generalised distance in comparison to the chord distance, justifying the choice of the distance coefficient in the ecological niche model NPPEN. First, the reference matrix is composed of 25 observations with 2 controlling factors. The correlation between the 2 controlling factors is null (A and B). (A) Probabilities based on the chord distance ($r = 0$). (B) Probabilities based on the Mahalanobis generalised distance ($r = 0$). Second, the reference matrix is composed of 13 observations with 2 parameters. The correlation between the 2 controlling factors is high ($r = 0.82$; C and D). (C) Probabilities based on the chord distance ($r = 0.82$). (D) Probabilities based on the Mahalanobis generalised distance ($r = 0.82$). \circ = reference observations (reference matrix). High probabilities are located at the centre of the reference matrix, denoting the centre of the ecological niche (*sensu* Hutchinson) and probabilities < 0.1 are situated outside (white)

Analysis 2

Cod individuals were mainly reported over neritic regions (Fig. 5; see also Fig. 2). Most of the reported cod occurrences (79.87% in the original reference matrix) were in regions shallower than 200 m, and the frequency of cod occurrence increased when the region became shallower. The homogenisation procedure did not radically alter the shape of the distribution (Fig. 5). Only a fraction of the total records (6.41% of reported cod occurrence) were deeper than 800 m, the threshold of bathymetry selected in this study. The sharpness of the continental slope may explain this small percentage since it is likely that the species can make short incursions over the shelf-edge (small mistakes on the spatial coordinates could increase this percentage slightly). Cod are rarely seen in oceanic regions and some authors have proposed a limit of 600 m, which corresponds to 12.48% of the records (Sundby 2000). To be more conservative, the bathymetric threshold was increased to 800 m.

The frequency of occurrence showed a mode over 34 for annual SSS, a mode that corresponds to the one identified when all regions of the Atlantic were taken together. Another smaller mode, more visible after homogenisation, appeared around 8 (Fig. 5). This mode corresponds to the salinity observed in the Baltic Sea. While annual SSTs in the regions of the North Atlantic vary from -2°C to 22°C , with a mode around 2°C , the range of temperature in which cod occurrence was more frequently reported were $>8^{\circ}\text{C}$ but ranged from 2°C to 14°C in the homogenised reference matrix. The procedure of homogenisation (Step 1) made it clear that the thermal optimum of the species is about 8°C .

Analysis 3

The model NPPEN was first applied to reproduce the ecological niche of Atlantic cod as a function of annual

SST (observed data), annual SSS and bathymetry (Fig. 6). The niche exhibited high probabilities of occurrence at a bathymetry that ranged between 0 and 500 m (mode between 100 and 300 m), an annual SST of from -1°C to 14°C (mode between 4 and 8°C) and an annual SSS ranging between 0 and 36 (mode between 25 and 34) (Fig. 6).

When projected on the geographical space (Fig. 7), the modelled spatial distribution of the probability of cod occurrence was congruent with the main location of the cod stocks in the North Atlantic sector (Sundby 2000, Bigg et al. 2008). However, there was a notable exception in the Baltic Sea, where low probabilities of cod occurrence were detected (Fig. 7A). A significant decrease in the probability of cod occurrence was evident between the 1960s and the period from 2000 to 2005 in the North Sea, while no such change was noted on the western side of the Atlantic at the southern range of the species' spatial distribution. At the north-

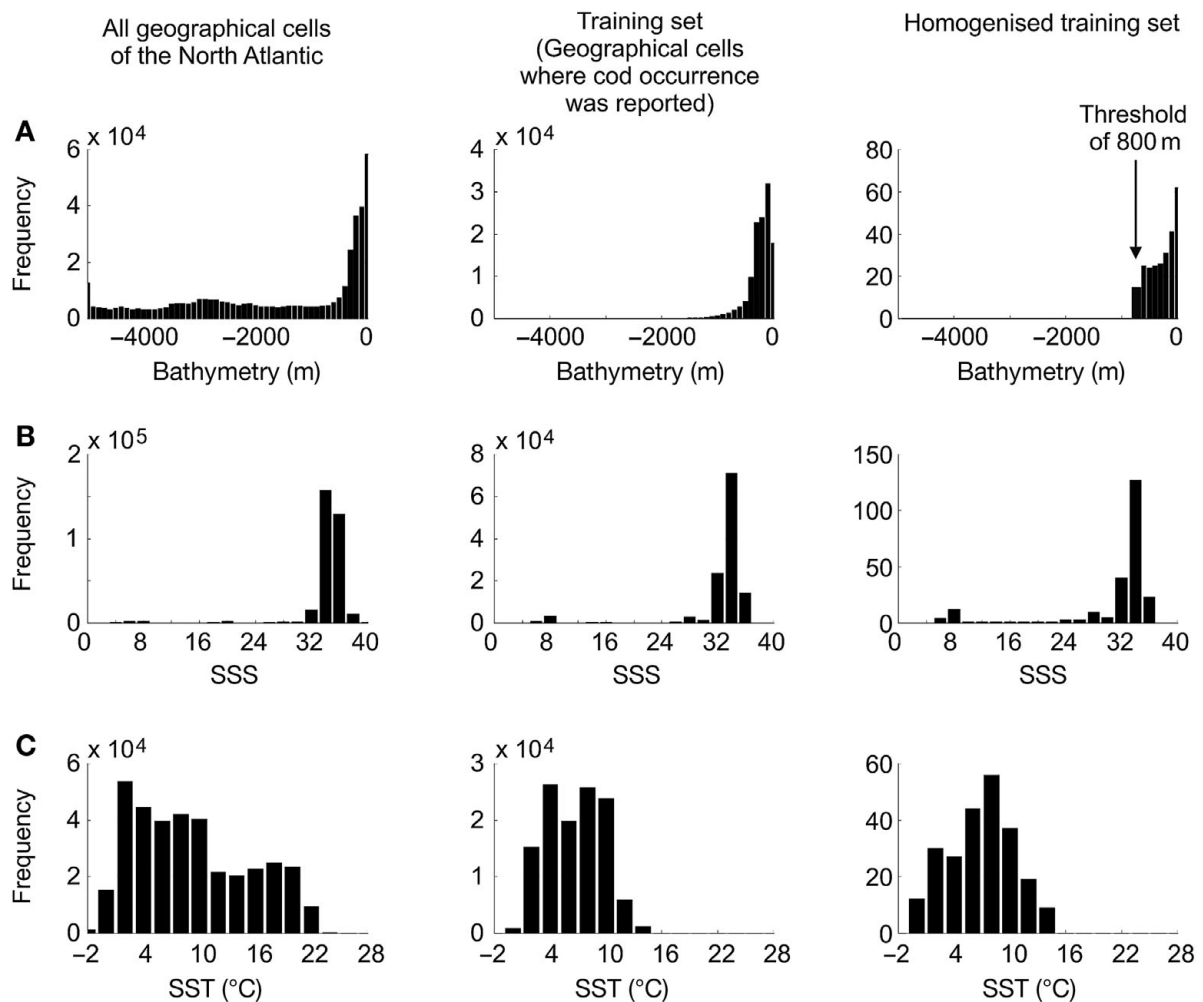


Fig. 5. Frequency distribution of (A) bathymetry, (B) sea surface salinity (SSS) and (C) sea surface temperature (SST) in the North Atlantic (from 80.50°W to 70.50°E and from 35.50°N to 70.50°N ; left panels), in the original reference matrix (i.e. geographical pixels where cod occurrence was detected) and in the homogenised reference matrix (see 'Materials and methods')

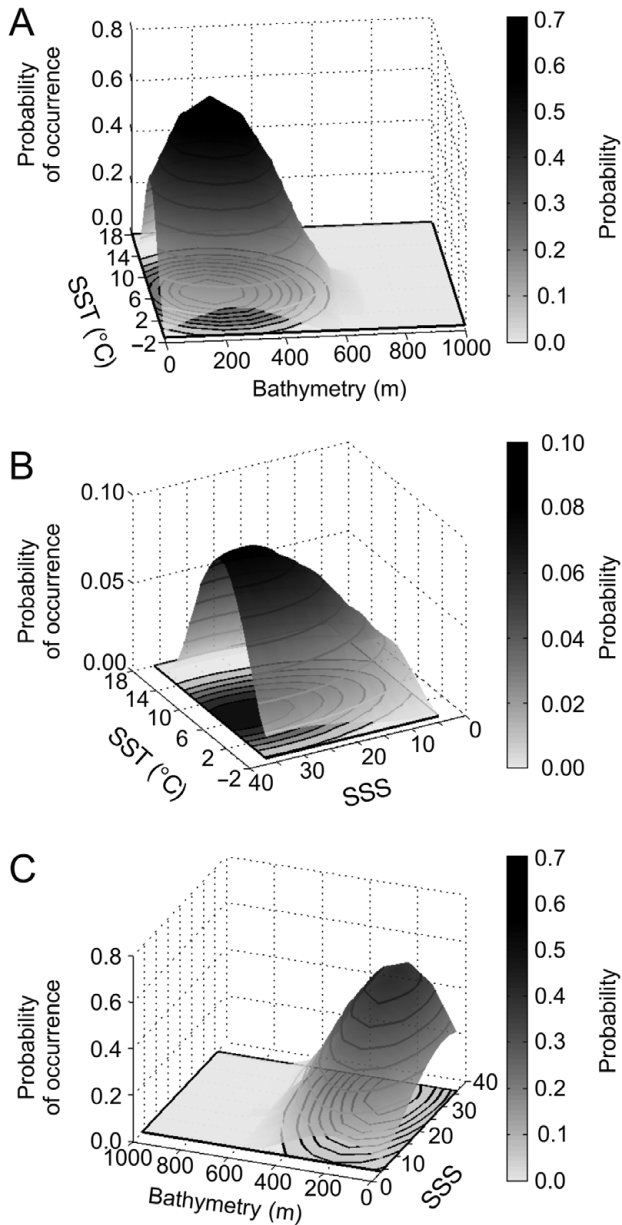


Fig. 6. Realised niche (*sensu* Hutchinson) of the Atlantic cod. Probability of cod occurrence as a function of (A) bathymetry and mean annual sea surface temperature (SST), (B) mean annual sea surface salinity (SSS) and mean annual SST, or (C) mean annual SSS and bathymetry

ern edge of the species' spatial distribution, probability of cod occurrence increased along the Greenland coast—especially in the western regions of this country—and in the Barents Sea. Probabilities around the Faeroes and Iceland remained stable.

Probabilities of cod occurrence based on modelled annual SST (ECHAM 4, scenario B2) and observed annual SST were similar ($r = 0.95$, $p < 0.01$). However, some discrepancies were noted on the western side of the Atlantic (Fig. 8D; Beaugrand et al. 2008). Based on

modelled data, the probabilities of cod occurrence appear much lower than observed data would suggest during the 1990s on the Georges Bank, the Eastern Scotian Shelf and the Grand Bank (Fig. 8D).

The examination of long-term decadal changes in the probability of cod occurrence projected for this century suggests a clear northward movement of this species (Fig. 9). Interestingly, the probability of cod occurrence decreased substantially at the southern edge of the range of the species: the North Sea, Georges Bank, Eastern Scotian Shelf, Grand Bank and Newfoundland. The probabilities increased in the Barents Sea and in areas close to Greenland, while no major changes were detected for Iceland. The model also suggested a substantial decrease in the probability of cod occurrence around the Faeroes at the end of this century if long-term changes in SST follow Scenario B2 (Fig. 9D).

The probability of cod occurrence was sensitive to the intensity of warming (Fig. 10). This sensitivity was not constant in space. Major changes occurred in areas located at the southern or northern limit of the spatial distribution of the species. When the level of anthropogenic warming remained unchanged (Fig. 10A,B), the probability of cod occurrence remained high in regions located at the southern limit (and low at the northern limit) of the spatial distribution of the species. However, even with a moderate scenario (Scenario B1, Fig. 10C), the probability of cod occurrence decreased in key regions such as the North Sea and the Georges Bank. When the level of anthropogenic warming increased, this had a pronounced influence in regions such as the North Sea and the Eastern Scotian Shelf (Fig. 10D–E). The coefficient of variation calculated per geographical cell using all 7 simulations issued from the 3 models (ECHAM4, HadCM3, HadGEM1) showed the areas where the changes are expected to be the most prominent (Fig. 10F). This analysis showed that modifications are expected to be pronounced at the periphery of the current spatial distribution of species. High values were observed in the Barents Sea, where a strong increase in the probability of cod occurrence is expected in the case of an intense warming, and, to a lesser extent, in areas such as the North Sea and the Georges Bank.

DISCUSSION

To predict how the range of marine species may change with climate, it is essential to understand the factors that limit their spatial distributions. One way to achieve this is to use ecological niche models. However, only a limited number of models can deal with presence-only data. The model BIOCLIM use a recti-

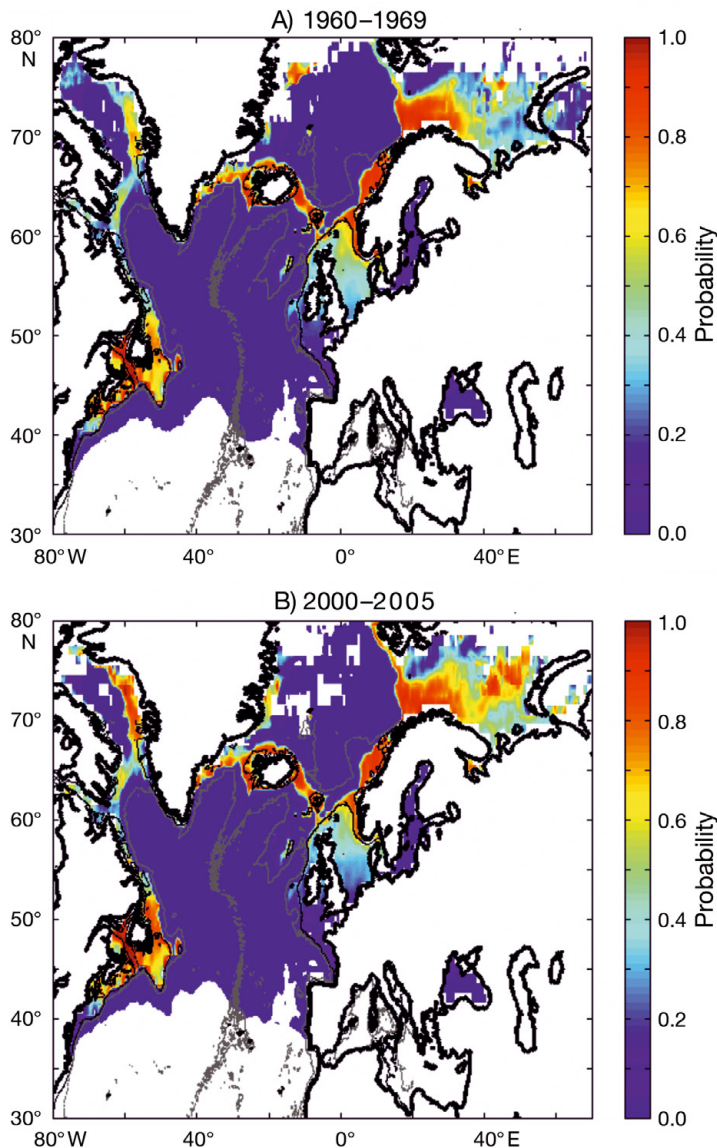


Fig. 7. Modelled spatial distribution in the probability of cod occurrence for the periods (A) 1960–1969 and (B) 2000–2005

linear volume (Carpenter et al. 1993). The main drawback of this simple model is the imposed shapes, which can be the cause of a non-justified exclusion or inclusion of a geographical point from the predicted distribution (Carpenter et al. 1993). The model DOMAIN can solve this problem by use of a point-to-point similarity metric. However, the metrics used (e.g. the Gower metric; Carpenter et al. 1993, Legendre & Legendre 1998) do not take into account the correlation between descriptors (Legendre & Legendre 1998). Furthermore, a threshold is used to map the modelled distribution of the species. The RES model (Kaschner et al. 2006) uses a trapezoid shape that constitutes a good compromise between species with a shorter or a unimodal ecological niche and migratory species with a larger or a

bimodal niche (Kaschner et al. 2006). However, the bounds of this trapezoid need to be precisely defined, implying often arbitrary choices, and thus it requires a good knowledge of the species. While the Ecological Niche Factor Analysis (ENFA; Hirzel et al. 2002) could be adapted for our study, this technique requires the multinormality of the ecogeographical variables to extract the eigenvectors to calculate marginality and specialization factors; for this reason a transformation of the ecogeographical variables (e.g. Box-Cox) is often needed prior to analysis. Finally, the statistical technique MAXENT (Phillips et al. 2006), which is based on the maximum-entropy principle, also requires accurate threshold definition and shows some application restrictions.

NPPEN offers a number of advantages over the above-mentioned methods. Firstly, unlike models such as RES (Kaschner et al. 2006) or the mixed model of Cheung and colleagues (Cheung et al. 2008), our simple model does not need an *a priori* knowledge of the species biology. Our technique is also based on a non-parametric test that does not require the multinormality of ecogeographical variables. Although the use of the Generalised Mahalanobis distance is not new in this kind of model (Farber & Kadmon 2003, Cayuela 2004, Etherington et al. 2009), this is the first time that this distance metric has been embedded into a non-parametric test. For example, Cayuela (2004) rescaled the Mahalanobis distance into quantiles to produce a map of probability, and Nogués-Bravo et al. (2008) converted the distance into quartiles. Ibañez (1981) and Farber & Kadmon (2003) tested this distance by approximating this measure by a χ^2 distribution with $n-1$ degrees of freedom. Legendre & Legendre (1998) also described the conversion of the D2 by the Hotelling T2 (Hotelling 1931) statistic and its test by the F statistic. These tests require the distribution to be multinormal. Although these authors stated that the test can tolerate some degrees of deviation from this assumption, it can be seen from histograms that the bathymetry data (Fig. 5) were very far from the normality. Finally, the procedure does not need the selection of arbitrary thresholds and is fully statistical. The technique simply tests whether an observation belongs to a group of observations, called here a training set, or a reference matrix. NPPEN therefore can be used very quickly as an exploratory analysis to give a first approximation of the spatial distribution of a species. The test is also appropriate for many species for which no information on its physiology exists. The only caveat is that our

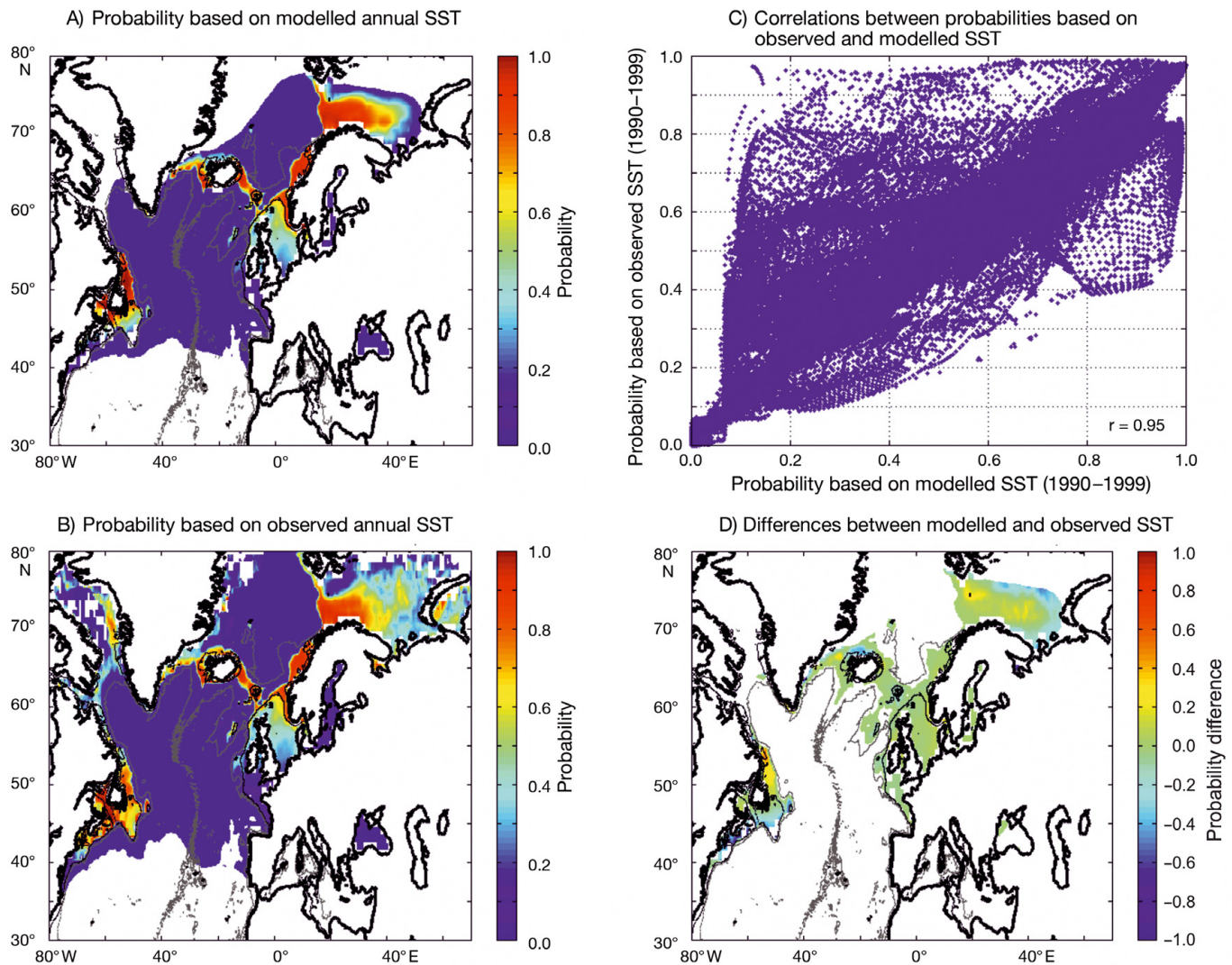


Fig. 8. Modelled spatial distribution in the probability of cod occurrence for the period 1990–1999. (A) From modelled annual SST (ECHAM 4, scenario B2). (B) From observed annual SST. (C) Relationships between probabilities based on observed and modelled (ECHAM 4, scenario B2) annual SST. The Pearson coefficient of correlation is indicated. Probabilities assessed from Scenario A2 gave similar conclusions. (D) Difference between probability of cod occurrence based on modelled SST (A) and probability of occurrence based on observed SST (B)

model, like others, does not fully resolve the problem of autocorrelation (SAR). The spatial autocorrelation can inflate significantly the probabilities inferred from ENMs (Bahn & McGill 2007). We think that the presence-only technique of ENMs is much less subject to this problem than other types of ENMs (e.g. GLMs). The problem is that only a few studies have considered local functions of autocorrelation (Beaugrand & Ibañez 2002, Dormann et al. 2007). Most corrections applied are based on the global function of autocorrelation with an underlying assumption of isotropy (e.g. Moran's Index, global semi-variograms), which is rarely the case in biogeography (Beaugrand & Ibañez 2002).

Our technique is currently restricted to presence-only data. Although some adjustments could be made

in the application of the method (e.g. calculating the test for a different category of abundance), it is probably preferable in such a case to use other techniques such as GLMs (McCullagh & Nelder 1983) or GAMs (Hastie & Tibshirani 1990). Guisan & Zimmermann (2000) provided an extensive review of different techniques used to assess the spatial distribution of a species. Another limitation of this technique may lie in the fact that it should only be employed with a limited number of ecogeographical variables. If a high number of variables are used it would be preferable to use a principal component analysis prior to the application of the test, or use the Mahalanobis distance factor analysis (MADIFA; Calenge et al. 2008) to better understand the contribution of the ecogeographical variables. This

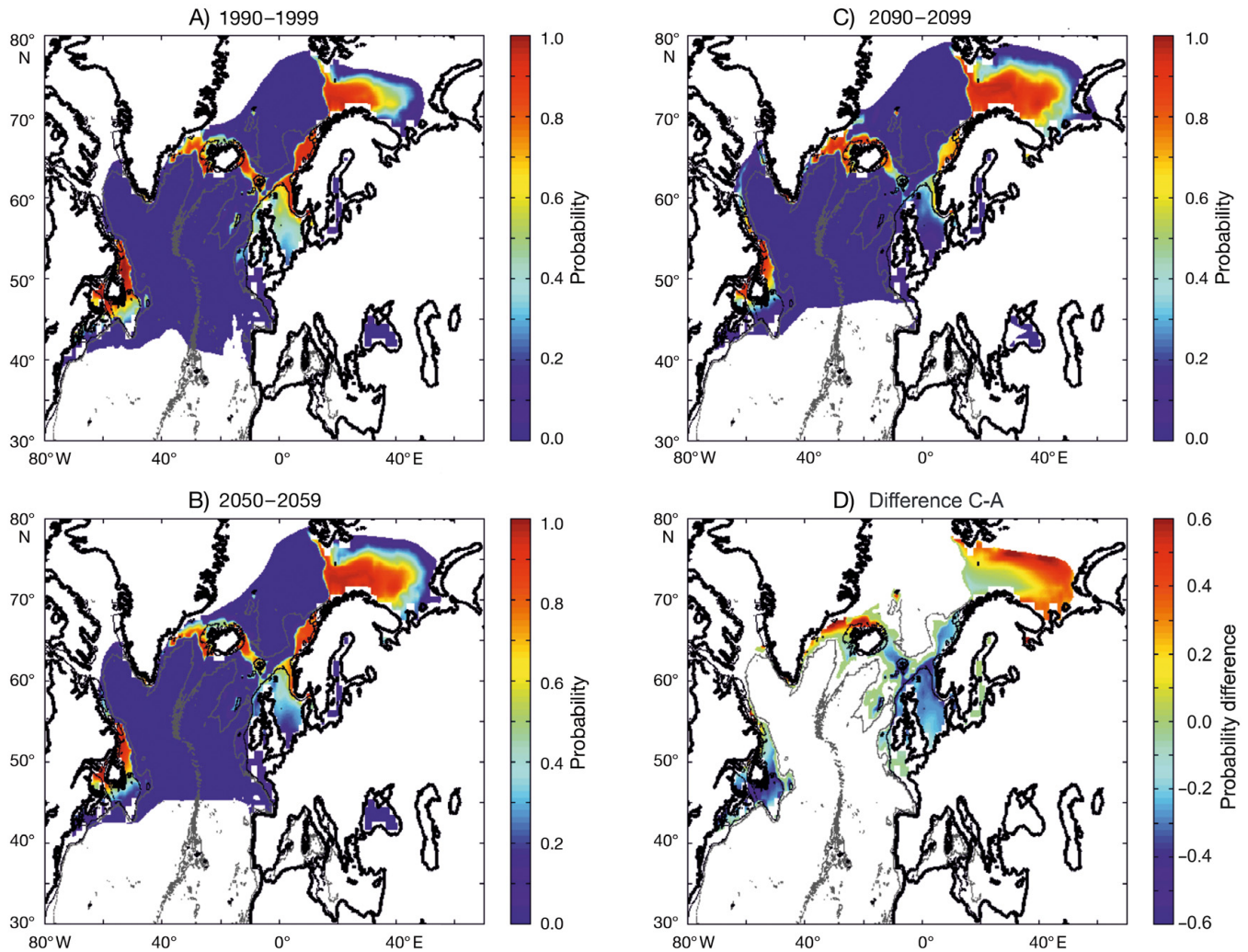


Fig. 9. Projected long-term decadal changes in the probability of cod occurrence from modelled annual SST (Scenario B2). (A) Period 1990–1999. (B) Period 2050–2059. (C) Period 2090–2099. Probabilities assessed from Scenario A2 gave similar conclusions. (D) Difference of modelled probability of occurrence between periods 2090–2099 and 1990–1999

can also be done *a posteriori* by calculating the correlation—here the rank correlation coefficient of Spearman or Kendall (Legendre & Legendre 1998) between the modeled probability and each environmental factor. NPPEN might also be subjected to what might be described as a 'border effect'. Indeed, the modelling of the niche of Atlantic cod (see Fig. 6) showed a reduction in the probability of cod occurrence towards shallow regions, which is unexpected based upon our knowledge of the species (Sundby 2000). Indeed, the technique works in such a way that maximum probability is concentrated towards the middle of the niche. Therefore some borders of the multidimensional niche might be underestimated. Although this problem is difficult to circumvent, it could be overcome partially by modelling the absence of the species (i.e. by estimating the probability of the absence of the species). Probabilities issuing from such a modelling approach would be

less sensitive to the border effect discussed above and would be complementary, by assessing the fundamental niche, whereas the ENMs applied on presence data estimate the realized niche (Pulliam 2000, Helaouët & Beaugrand 2009).

Modelling the absence of a species has never been done, as far as we know. It could, however, be as informative as modelling the presence of a species, especially in the case of an exploited species such as Atlantic cod. Indeed, modelling the probability of absence is very informative for policymakers and fisheries scientists. It should not, however, be assessed from a map of the probability of presence, but instead should be based on physiological evidence (Bigg et al. 2008, Helaouët & Beaugrand 2009). When presence-only data are available to model the spatial distribution of cod, several known scientific facts, including physiological data, are available for use in NPPEN. First, the

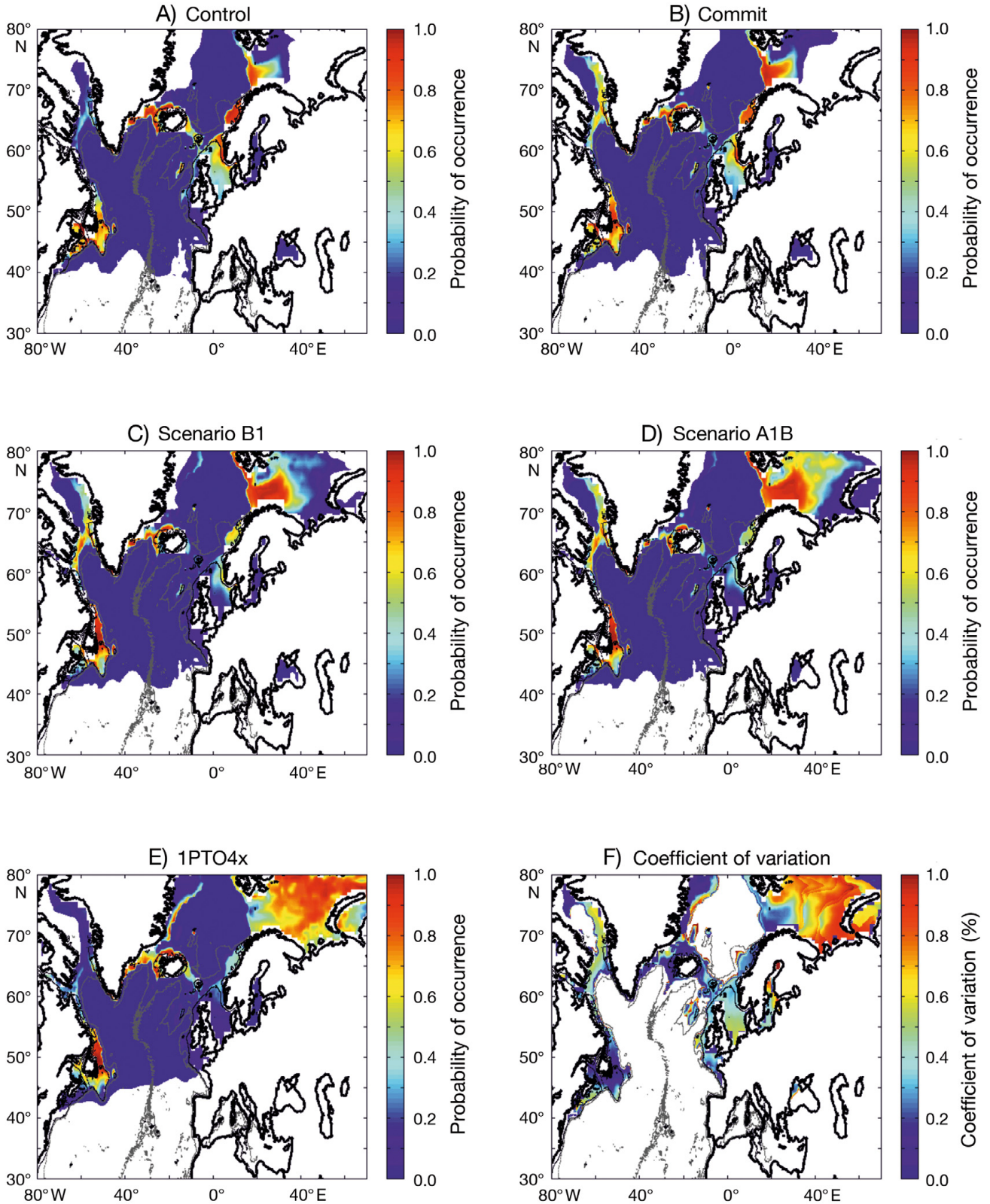


Fig. 10. Projected long-term decadal changes in the probability of cod occurrence for the period 2090–2099. (A) scenario ‘Control’. (B) Scenario COMMIT. (C) SRES scenario B1. (D) SRES scenario A1B. (E) 1PTO4x. (F) Coefficient of variation based on the 7 scenarios and 3 atmosphere-ocean general circulation models described under ‘Materials and methods’. The intensity of warming increases from A to E

species is generally found where the bathymetry is shallower than 800 m (see Fig. 5). While some authors have found a sharp decrease in the frequency of occurrence of this species at 400 m and an absence below 600 m (Bigg et al. 2008), this is not a serious constraint and we can therefore be conservative. From field and experimental studies, we know that cod are unable to reproduce at salinities below 11 because their sperm become immobile and their eggs sink (KM Brander, pers. comm.). We can therefore predict confidently that cod will cease to reproduce in areas where salinity falls below these levels. With respect to temperature, different thresholds could be used. Beaugrand et al. (2008) found a pronounced increase in the variance of Atlantic cod when the thermal regime was between 9 and 12°C with maximum variance between 9 and 10°C. Brander (2005), in a synthesis report on this species, found maximum spawning temperatures of 12.7°C in Georges Bank. Pepin et al. (1997) found a sustained decrease in the percentage of egg survival in the laboratory between 10 and 12°C. Here also, it would be logical to select the threshold of 12.7°C (as monthly SST) in order to remain conservative.

Our model explains in part the pronounced decrease observed in the abundance of cod in the North Sea by Brander et al. (2006) although the decline modelled from our study seems less pronounced. Two main features could explain this result. (1) Our model does not incorporate information on plankton. Recent studies have shown that incorporating plankton amplifies the effect of temperature increase (Beaugrand & Kirby 2010b, 2010a). If the incorporation of plankton exacerbates the effect of temperature increase, our model might be too conservative. (2) Overfishing has exerted a sustained pressure on the stock, which has probably increased its sensitivity to climate change (Hsieh et al. 2006, Perry et al. 2010). It is also expected that these 2 types of forcing act in synergy to reduce the size of the stock (Kirby & Beaugrand 2009). Our model only explains in part the collapse of the cod fishery observed in the eastern part of North America (e.g. Georges Bank, the Eastern Scotian Shelf and Newfoundland). This is mainly observed when our model is based on modelled SST data (Fig. 8). Recently, Beaugrand & Kirby (2010a) also showed that some plankton indicators decreased at the same time as the observed collapse of cod stocks in these regions. However, the region is complex and plankton variables were insufficient to explain completely this collapse. Here also, overfishing has had a well documented effect (Myers et al. 1996, Hsieh et al. 2006, Perry et al. 2010). The stock may resist sustained pressure up to a point when environmental conditions become less favorable and trigger the collapse of the stock. While it is impossible to compensate directly for either direct or indirect

effects of global warming on the ocean, the consideration of change in the carrying capacity of the ecosystem in the management of the species should be made more explicit in ecosystem fisheries-based management (Pikitch et al. 2004).

Projections of change in fish distribution suggest major modifications in the spatial distribution of fish that could exceed 10° of latitude by the end of the century (Lenoir et al. 2010). Our scenarios of change in species distribution are highly sensitive to the intensity of warming (Fig. 10). If temperatures remain unchanged, our model suggests that cod is likely to stay commercially exploitable in the North Sea. If warming is moderate (Scenario B1), cod is likely to remain exploitable in the northern part of the North Sea. In case of strong warming, cod will inevitably disappear from the North Sea as a commercial species (see Fig. 10). While the current edges of the spatial distribution are highly responsive to the intensity of warming, the centre of the spatial distribution is expected to be much less sensitive. This result was emphasized recently in Beaugrand & Kirby (2010a,b).

Our study revealed that climate change may strongly alter cod stocks; other studies have suggested similar effects of climate on the same species (Clarck et al. 2003, Drinkwater 2005, Cheung et al. 2008). Each species has a unique spatial distribution, which is determined by its bioclimatic envelope (Beaugrand & Kirby 2010b), and therefore, climate controls the spatial distribution of fish. IPCC projections indicate a likely warming of the earth of more than 2°C by the end of this century, and possibly even 5°C, which would represent the temperature difference between the last glacial maximum and the current interglacial period (Cronin 1999). Even warming to half this difference will have a major impact on the spatial distribution of many species on the planet. In the case of cod, we doubt that this species will be able to adapt to such level of warming on a short time scale.

Our projections indicate that in the case of moderate to strong warming cod might be strongly reduced in both abundance and distribution, and reach the level of commercial extinction in the North Sea. These results tend to suggest that the rebuilding of cod stocks in the North Sea might be difficult. Instead, our effort should perhaps focus on what resource is likely to become available over the next decades to enable fishermen to anticipate changes in these resources, should the earth continue to warm.

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