

Krill excretion and its effect on primary production

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ABSTRACT: During the austral summer, zooplankton excretion along the western Antarctic Peninsula was studied in a contrasting hydrographic regime including coastal and oceanic waters. In coastal waters, ammonium supply by mesozooplankton indicated a low contribution to fuel primary production. In oceanic waters, however, Antarctic krill *Euphausia superba* contributed a significant percentage to the nitrogen requirements of primary producers. Thus, the ontogenetic migration of adult krill during austral summer should be a key factor regulating the regenerated ammonium for primary production. A significant coupling of ammonium concentration in the water column and *in situ* krill biomass supported the significant role of krill excretion in the epipelagic realm. Results from short-term experiments with *E. superba* indicated that ammonium excretion rates were much higher than previously found. Because the use of experimental metabolic rates that are close to field rates would be more appropriate, we suggest to re-assess the ammonium supplied by the epipelagic marine biota. Moreover, the outcomes of experimental krill excretion rates, *in situ* measurements of ammonium and a review of data on primary production suggest that Antarctic krill sustain a high proportion of the daily phytoplankton production.

KEY WORDS: Krill · Ammonium · Primary production · *Euphausia superba* · Southern Ocean · Austral summer · Western Antarctic Peninsula

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INTRODUCTION

The Southern Ocean generally shows low primary production. Nevertheless, a high spatial variability in the phytoplankton biomass and production is found, especially along the Antarctic Peninsula where some regions associated with stationary phytoplankton blooms exhibit high productivity (Park et al. 1999, Varela et al. 2002, Garibotti et al. 2003). Light limitation (Dortch 1990), the low affinity of phytoplankton for nitrate owing to the low temperature (Reay et al. 1999) and the higher energetic cost of nitrate uptake compared with ammonium uptake (Dugdale 1976) or iron limitation (Martin & Fitzwater 1988) have been proposed to explain the low productivity of the

Southern Ocean. Low iron availability can be drawn upon as a limiting factor for primary production (Timmermans et al. 1994, De Baar et al. 1995, Boyd et al. 2000, Holm-Hansen et al. 2005). However, as coastal waters, particularly along the Antarctic Peninsula, are probably iron-replete (Martin et al. 1990, Sullivan et al. 1993, Moore & Abbott 2000, Varela et al. 2002), critical concentrations of ammonium may be a potential limiting factor as phytoplankton production is mainly sustained by this nutrient (Priddle et al. 1997, Whitehouse et al. 1999, Atkinson et al. 2001, Reay et al. 2001, Bode et al. 2002). Indeed, mesocosm experiments in Antarctic waters showed that ammonium additions greatly stimulated phytoplankton growth and biomass (Agustí et al. 2009).

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Nitrate and ammonium are important for primary production but ammonium can be a preferred source of nitrogen (Dugdale & Goering 1967, McCarthy et al. 1977). Bacterial recycling and zooplankton excretion are critical sources of ammonium for phytoplankton growth. While nitrate commonly dominates nitrogen concentrations in Antarctic waters (Priddle et al. 1997, Bode et al. 2002), ammonium excreted by zooplankton is thought to be important for primary production (Priddle et al. 1997, Alcaraz et al. 1998, Hernández-León et al. 2008), and in fact, at least 50 to 93% of the nitrogen is assimilated as ammonium in continental shelf waters (Koike et al. 1986).

Ammonium is the principal form of dissolved nitrogen excreted by planktonic crustaceans (Båmstedt 1985, Miller & Glibert 1998, Conover & Gustavson 1999). Therefore, zooplankton should control primary production through both grazing and excretion. Swarms of Antarctic krill *Euphausia superba* represent important sources of ammonium (Atkinson & Whitehouse 2000) since large nitrogen concentrations were found within the aggregation (Johnson et al. 1984, as cited in Priddle et al. 1997). Swarm-forming krill should cause a temporary depletion of both copepods and phytoplankton through grazing and predation but at the same time should induce optimal phytoplankton bloom-forming conditions through the release of ammonium (Tovar-Sánchez et al. 2007). Resolving krill excretion rates is, therefore, critical to estimate the contribution of ammonium to the nitrogen demand for Antarctic phytoplankton growth. Here, we try to quantify the contribution of krill ammonium excretion in the different hydrographical zones of the western Antarctic Peninsula in order to assess the role of krill in ammonium regeneration in these waters.

MATERIALS AND METHODS

Sampling and data processing

A survey along the western Antarctic Peninsula (WAP) region was conducted during the ICEPOS 2005 cruise (January to February) on board RV 'Hespérides'. Three regions with distinct hydrographic regimes were sampled along the WAP: Bellingshausen Sea (Stns Be2, Be3, Be4, Be5), Bransfield Strait (Br12, Br14, Br18) and the Antarctic Sound (As11), see (Fig. 1). Sampling was conducted during daytime for all stations (Table 1). Zooplankton was collected with a BIONESS net (Sameoto et al. 1980) equipped with 6 nets of 200 μm mesh size, and the

samples were used for biomass estimation. Oblique hauls started at 400 m depth, or within 50 m from the bottom for shallower stations, to the surface (10 m depth). The water volume filtered by the nets was determined with a calibrated flowmeter (General Oceanics) and varied between 601 and 1524 m^3 .

Zooplankton was preserved in buffered 4% formaldehyde solution just after collection, size-fractionated, counted by using a standard digital camera with a charge-coupled device (CCD) sensor (see Lehette & Hernández-León 2009 for details) and sorted into 5 taxonomic categories. Adopted resolutions (14.7 μm pixel size for copepods and larval krill, and 33.3 μm pixel size for salps and small and large euphausiids) were suitable for morphometric measurements and for semiautomated plankton recognition and classification. Image processing and machine-learning methods followed those described by Grosjean et al. (2004).

Ammonium experiments

Ammonium experiments were conducted on fresh specimens of Antarctic krill captured in oblique hauls with an Isaac-Kidd mid-water trawl (IKMT) net equipped with a mesh size of 1 cm. The net was towed from 100 m depth to the surface at a speed of 2 to 3 knots during the night. All individuals captured

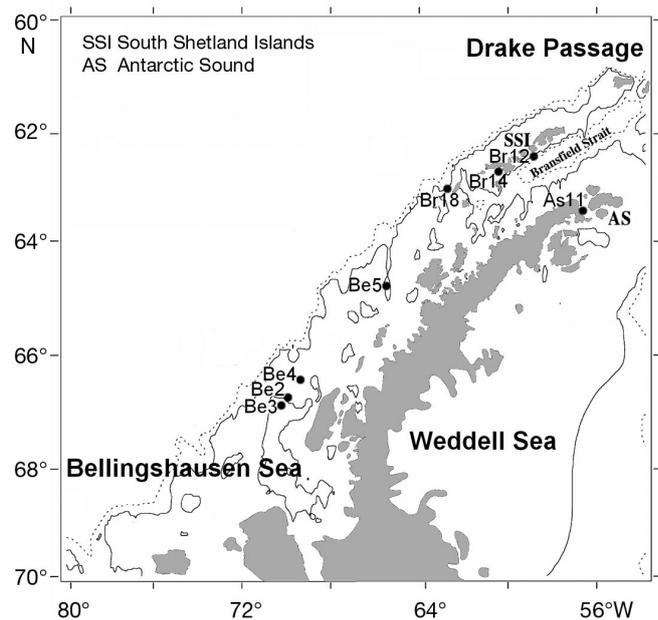


Fig. 1. Stations along the western Antarctic Peninsula which were sampled during the ICEPOS 2005 cruise. Depth contours: solid = 500 m, dashed = 1000 m

Table 1. BIONESS sampling dates, local time, geographic position, depth and category along the Antarctic Peninsula during austral summer 2005. See Fig. 1 for location of stations

Station	Sampling date (dd/mm/yy)	Local time (GMT – 3 h)	Latitude (S)	Longitude (W)	Bottom depth (m)	Category
Be2	03/02/2005	14:41 h	66° 34.14	69° 59.27	436	Midshelf
Be3	04/02/2005	14:48 h	66° 37.28	70° 09.31	464	Midshelf
Be4	05/02/2005	14:22 h	66° 10.65	69° 17.64	340	Midshelf
Be5	06/02/2005	13:06 h	64° 45.39	65° 42.27	387	Midshelf
As11	11/02/2005	18:17 h	63° 23.73	56° 44.33	289	Coastal
Br12	12/02/2005	15:38 h	62° 24.87	58° 48.56	1222	Coastal
Br14	14/02/2005	12:47 h	62° 44.35	60° 32.27	298	Coastal
Br18	22/02/2005	16:23 h	63° 00.73	62° 55.91	629	Coastal

were immediately transferred into on-deck aquaria with constantly renewed surface seawater. All the experiments were derived from the ‘water-bottle’ method (Omori & Ikeda 1984) by using 1 or 2 individuals per flask.

Two series of experiments were carried out in filtered seawater and designed to have a significant response between the control and the experiment. For each experiment, 3 controls were done with 1 or 2 individuals per flask depending on the incubation time, i.e. 2 specimens per flask for the short incubations and 1 for the longer ones. The first series of experiments was carried out with surface seawater pumped through acid-cleaned Teflon tubing coupled to C-flex tubing (for the Cole-Parmer peristaltic pump head), filtered through an acid-cleaned polypropylene cartridge filter (0.22 μm , Calyx, Micron Separations) and collected in 2 l low density polyethylene (LDPE) bottles for experiments. Experimental and control bottles were incubated in the dark in an incubation chamber set at surface water ambient temperature ($\pm 1^\circ\text{C}$). In a class-100 high efficiency particulate air (HEPA) hood, water samples from the experimental bottles were collected at 2 h intervals, from 1 h up to 11 h from the onset of each experiment. At the end of the experiments, specimens were dried to constant weight in a drying oven set at 60°C and weighed to the nearest mg.

The second series was performed in filtered seawater using GF/F filters (Whatman) in 2 l bell-jar incubation bottles held in a thermostatic bath at $1 \pm 0.1^\circ\text{C}$ in dim light for 1 to 7 d. At the end of the incubation, excretion rates were calculated from differences in ammonium concentration between the control and experimental bottles. The health condition of all individuals was checked at the end of the experiment and krill individuals were then image-processed for dry weight (DW) determination (Lehette & Hernández-León 2009).

Determination of ammonium by fluorometry

Ammonium determination based on the reaction with orthophtaldialdehyde (OPA) and sulphite followed the procedure and recommendations described by Kerouel & Aminot (1997). Incubations of 50 min were made in PVC flasks in a thermostatic bath fixed at $37 \pm 0.1^\circ\text{C}$. The solution was then exposed to an excitation wavelength of 365 nm, which produced a fluorescent isoindole fluorophore detected at a specific wavelength of 425 nm by the spectrofluorometer (Shimadzu RF-5301 PC). After calibration, final concentration is calculated from the difference between the sample and the blank readings (milli-Q water). This automated analysis method is easy to use and highly accurate (< 0.5 nanomolar) for ammonium determination of discrete marine samples (Kerouel & Aminot 1997).

RESULTS

Zooplankton biomass showed sharp differences in its vertical distribution between the different regions of the WAP (Fig. 2). In the Bellingshausen Sea, the high zooplankton biomass was mainly related to large specimens of krill in the surface layer (Fig. 2A). Salps were abundant in the Bransfield Strait (Fig. 2B) with a biomass peak in the 70 to 100 m depth layer. In the Antarctic Sound (Fig. 2C), copepods and larval krill dominated the upper 100 m depth layer.

A coupling between the overall biomass of zooplankton in the water column and the *in situ* ammonium concentration was observed ($r^2 = 0.72$, Fig. 3A) with a slightly better coefficient of determination for the relationship of ammonium concentration to krill biomass ($r^2 = 0.78$, Fig. 3B).

As expected, the compilation of our experimental krill ammonium excretion rates with the results of

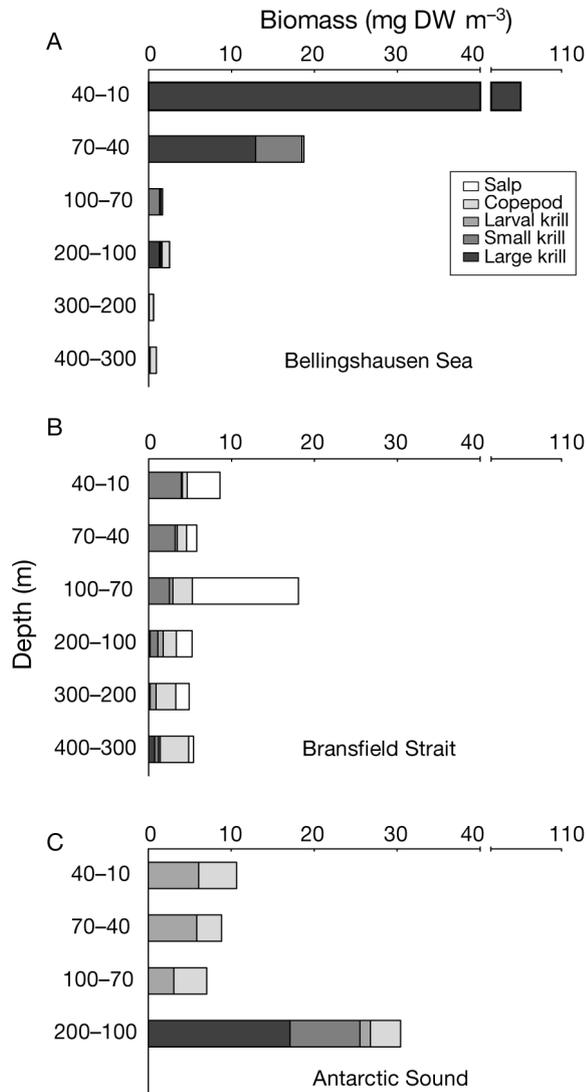


Fig. 2. Vertical distribution of the 5 dominant groups of zooplankton in terms of biomass (mg DW m^{-2}) in the (A) Bellingshausen Sea, (B) Bransfield Strait and (C) Antarctic Sound. DW = dry weight

Huntley & Nordhausen (1995) and those of Atkinson & Whitehouse (2000) indicated a decreasing krill excretion rate with the length of incubation (Fig. 4). Maximum rates ($30.0 \pm 6.3 \text{ nmol NH}_4^+ \text{ mg}^{-1} \text{ DW h}^{-1}$, mean \pm SD) were measured during the first hour of incubation and were ~ 12 -fold higher than standard rates measured after 24 h ($2.45 \pm 1.1 \text{ nmol NH}_4^+ \text{ mg}^{-1} \text{ DW h}^{-1}$). The relationship between incubation time (h) and excretion rates was given by the expression:

$$\text{NH}_4 \text{ excretion (nmol NH}_4^+ \text{ mg}^{-1} \text{ h}^{-1}) = 17.04h^{-0.49}, \\ r^2 = 0.76$$

Combining the biomass of the main zooplankton groups with the excretion rates for each sampling

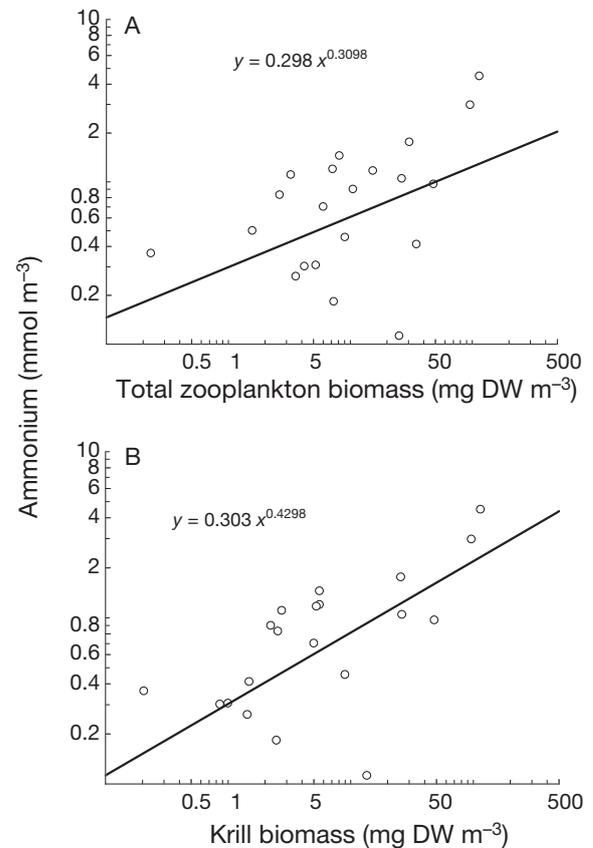


Fig. 3. Relationship between ammonium concentrations (mmol m^{-3}) and (A) total zooplankton biomass (mg DW m^{-3}) ($r^2 = 0.72$, $n = 24$) and (B) krill biomass (mg DW m^{-3}) ($r^2 = 0.78$, $n = 24$). Values from field measurements are plotted on a log-log scale. DW = dry weight

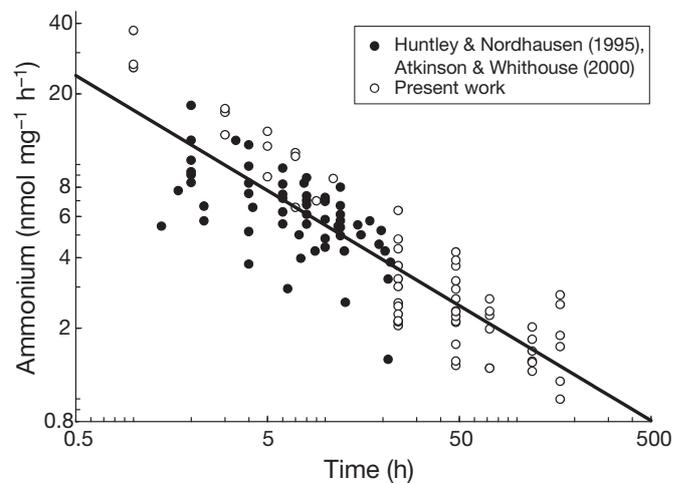


Fig. 4. *Euphausia superba*. Experimental krill ammonium excretion rates including data of Atkinson & Whitehouse (2000) and Huntley & Nordhausen (1995). Data are plotted on a log-log scale. The relationship between excretion rates and time of incubation ($r^2 = 0.76$, $n = 117$) is given by the equation: $\text{nmol NH}_4^+ \text{ mg}^{-1} \text{ h}^{-1} = 17.042h^{-0.4906}$

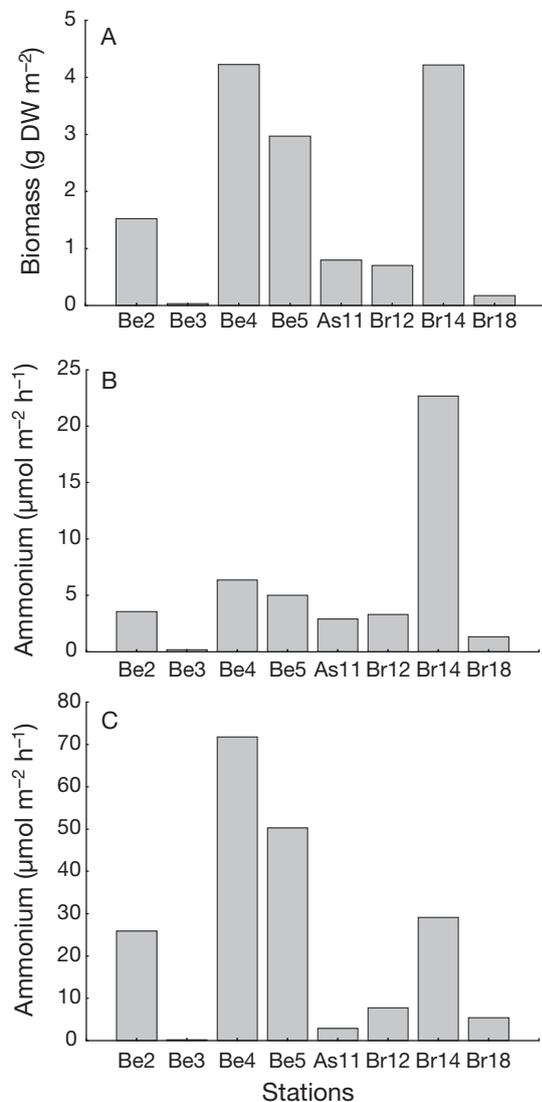


Fig. 5. (A) Total zooplanktonic biomass (g DW m⁻²) in the upper 100 m layer. Total ammonia excretion rates (μmol m⁻² h⁻¹) by zooplankton in the upper 100 m layer (B) determined from literature data and (C) using experimental rates by small and large krill and data from literature for other taxa (salps and copepods). Note the difference in the excretion rates for Bellingshausen sector. For details on stations see Table 1. DW = dry weight

station (Fig. 5A) (Table 2) indicated that the Bellingshausen stations exhibited rather low ammonia production rates (Fig. 5B). In contrast, salps in the Bransfield Strait provided high ammonia excretion rates in the top 100 m depth layer. Applying the experimental excretion rates of krill during the first hour of incubation (17.04 nmol NH₄⁺ mg⁻¹ DW h⁻¹), krill in the Bellingshausen Sea (Sites Be2, Be4 and Be5, see Fig. 1) exhibited much higher ammonia excretion rates than those in the other locations sampled (Fig. 5C).

DISCUSSION

Krill excretion rates measured shortly after capture were relatively high compared with published excretion rates (Hirche 1983, Huntley & Nordhausen 1995, Hernández-León et al. 1999, Atkinson & Whitehouse 2000). In the absence of food, *Euphausia superba* has proven its resilience to starvation for >200 d (Ikeda & Dixon 1984) by not only adopting strategies of moulting, cannibalism and shrinking in size (Ikeda & Mitchell 1982), but also by reducing its excretion rates to extremely low levels (Hirche 1983, George & Fields 1984, Atkinson & Whitehouse 2000). The metabolism of freshly caught krill may be 2 to 4 times higher than that of starved and inactive animals (Chekunova & Rynkova 1974, Sameoto 1976, Ikeda & Dixon 1984). As the effect of starvation during incubations is rather fast (Le Borgne 1979, Head et al. 1988, Lehette & Hernández-León 2010), experiments performed for long periods (>6 h) underestimate metabolic activity (Head et al. 1988), which accounts for the low excretion rates previously observed when long incubation times were used. The exponential decline in excretion rates with incubation time is clearly observed in our experiment (see Fig. 4) and is consistent with that reported in the literature (see Atkinson & Whitehouse 2001, Tovar-Sanchez et al. 2007, Lehette & Hernández-León 2010). Lehette & Hernández-León (2010) observed a factorial scope (ratio between active metabolism and standard metabolism) of 11 for subtropical copepods, which is similar to the short-term excretion rates of ~12 in krill obtained after capture and observed in the present work. Short-term incubation periods yielding higher metabolic rates are appropriate given the continuous feeding habit of krill and would be most suitable to represent *in situ* metabolic rates of zooplankton (Satomi & Pomeroy 1965, Webb & Johannes 1967, Biggs 1977, Le Borgne 1979, Omori & Ikeda 1984, Båmstedt 1985). However, high metabolic rates in experiments without acclimation periods could be due to an effect of stress after capture. In that regard, Ikeda & Skjoldal (1980) found that this decline in metabolic rates is due to the effect of food shortage during incubation rather than to the effect of stress after capture. Thus, immediate measurement after the catch is thought to be a more realistic approach to assess field rates (Ikeda et al. 2000), and should also minimize the environmental condition, nutritional status and starvation effect on metabolic rates. Therefore, using this short-term assessment of excretion, rates were expected to be in the upper range of published krill excretion rates in the WAP region during

Table 2. Ammonia excretion rates of the dominant groups of zooplankton as compiled from the literature and experimental results of this study

Group	Dry weight range	Excretion rates (nmol NH ₄ ⁺ ind. ⁻¹ h ⁻¹)	Source
Salps	10.6–14.4 mg	59.78–80.23	Alcaraz et al. (1998)
Copepods	1.9–138.1 µg	0.33–0.74	Ikeda et al. (2001) ^a
Larval krill	233.3–544.6 µg	1.94–4.54	Meyer et al. (2002) ^b
Small krill	3.6–27.9 mg	11.65–59.85	Atkinson & Whitehouse (2000)
Large krill	104.7–186.3 mg	172.16–272.83	Atkinson & Whitehouse (2000)
Small and large krill	3.6–186.3 mg	61.34–475.42	Present study ^c

^aAssuming equal proportions of the following copepod species: *Calanus acutus*, *C. propinquus*, *Metridia gerlachei* and *Rhincalanus gigas*; ^bcalculated using the average value of larval krill excretion rate; ^cdetermined using experimental close-to-field excretion rates of about 17 nmol NH₄⁺ mg⁻¹ DW h⁻¹

winter (>15 nmol NH₄⁺ mg⁻¹ DW h⁻¹; George & Fields 1984, Huntley & Nordhausen 1995). In this sense, metabolic rates by polar invertebrates in summer are supposed to reflect maximal annual metabolic rates (Clarke & Prothero-Thomas 1997).

The tight coupling between *in situ* ammonium measurements and krill biomass is consistent with our independent calculations pointing to a prevalence of krill contributions compared with those of other groups (Fig. 3B). For instance, Atkinson et al. (2001) suggested that copepods and krill excretion can supply up to 80% of the ammonium required for primary production and that this was related to the abundance of krill. Priddle et al. (1997) drew similar conclusions but also pointed out the importance of krill at low densities that could sustain elevated ammonium concentrations in near-surface waters. Recently, Whitehouse et al. (2011) also confirmed similar evidence of enhanced *in situ* ammonium concentration at higher krill densities. The uncertainties of their krill acoustic detection and their underestimation of excretion rates (Whitehouse et al. 2011) should explain the slight discrepancies with our data. Accurate estimates of zooplankton abundance and biomass still remain a challenge since a large variability occurs between net and acoustic methods or even within net sampling (Burd & Thomson 2012). Thus, optical devices could be integrated with net samplers to increase both accuracy and efficiency of plankton surveys (Broughton & Lough 2006).

The rather robust potential relationship between krill biomass and *in situ* ammonium concentration could confirm the importance of krill excretion at higher densities. Under conditions of high food availability, krill aggregation is expected to be higher (Daly & Macaulay 1991) and should directly affect local nutrient concentration in the surrounding waters. The importance of the swarming behaviour of Antarctic krill to fuel primary production should

also be considered. Indeed, Antarctic krill are generally found in aggregations at depth during daytime mostly to avoid predation (Ritz et al. 2003), dispersing in the upper layer to feed during the night period (Kalinowski 1978, Tomo 1983, as cited in Zhou & Dorland 2004). Nevertheless, this pattern may change owing to physical and biological constraints (Zhou & Huntley 1996, Zhou & Dorland 2004). It seems that krill living solitarily are likely to be stressed, and they will respond to neighbouring individuals by decreasing their metabolic rate and saving energy (Ritz 2000, Ritz et al. 2001). Aggregation behavior is believed to confer lower predation risk (O'Brien 1987, Ritz 2000), to reduce energy expenditure of hydrodynamic processes (Ritz 2000) but also to lower food intake owing to intraspecific competition (Morris et al. 1983, Ritz 2000). Thus, the behaviour of swarm-forming krill should probably affect their metabolic activity by lowering it. Some observations showed that krill swarms can alter the structure of zooplankton communities, decreasing copepod abundance, but also other zooplankton taxa (Nordhausen 1994, Atkinson et al. 1999, Priddle et al. 2003). Hence, the high *in situ* ammonium concentration at high krill biomass (see Fig. 3B) is in agreement with the carnivorous feeding of krill on N-rich copepods (Granéli et al. 1993).

Thus, in order to determine *in situ* ammonia excretion rates of krill and because zooplankton metabolism in the field is never maximal (Hernández-León & Ikeda 2005), we suggest that field excretion rates be represented by those calculated after 1 h of incubation (i.e. 10 to 30 nmol NH₄⁺ mg⁻¹ DW h⁻¹, Fig. 4). By applying those rates to observed krill biomass, oceanic waters off the Bellingshausen Sea were observed to have supported much higher ammonium production rates (Fig. 5B,C) than did waters of the Bransfield Strait region. During the transition from bloom to post-bloom conditions, small krill inhabit

coastal waters (Siegel 2005) as observed in the Bransfield Strait and Antarctic Sound, whereas larger specimens inhabit the open waters of the Bellingshausen Sea (Fig. 2A). The high ammonium concentration found in 10 to 40 m depth of the Bellingshausen Sea (up to 4.5 mmol m^{-3} , data not shown) was closely linked to the high krill biomass (Fig. 2A). Thus, large krill performing their ontogenetic migration offshore (Siegel 1988) and preying upon large copepods (Hernández-León et al. 2001, 2008) probably have an important role in nutrient regeneration (Atkinson et al. 2001).

We also proposed to estimate the contribution of ammonium produced by the main taxonomic groups to primary production in the top 100 m depth layer. Primary production data (Alcaraz et al. 1998, Varela et al. 2002) for the different regions of the WAP were extracted from the literature (Table 3), and carbon to nitrogen units were transformed using a Redfield C:N ratio of 100:16 (Goldman & McCarthy 1978). In general, this supply was low but in the range of previous studies (Atkinson & Whitehouse 2001). Indeed, using excretion rates from the literature, 3.3% of the ammonium required to support primary production was regenerated by Antarctic krill in the Bellingshausen Sea. The Antarctic Sound presented unexpected low values owing to the location of krill below the top 100 m layer (Fig. 2C) and an assumption of no migration to the upper layers. By contrast, copepod ammonium excretion showed, in this area, relatively high values compared with other taxa (Table 3). Unlike other areas, salps were present only in the Bransfield Strait and contributed close to 75% of the total zooplankton excretion, showing a rather low contribution to primary production (Table 3). Hence, the

contribution of the zooplankton community to phytoplankton nitrogen demand depends on the spatial and vertical distribution of the main taxonomic groups, and this contribution generally represents a minor role in the regeneration of reduced nitrogen (Alcaraz et al. 1998).

The length of incubation is, therefore, an important factor for the estimation of metabolic rates. Standard rates normally used to estimate the contribution of zooplankton to primary production gives rise to rather low values. The excretion rates obtained by our experimental procedure for krill, which were close to field rates, provided a large difference in the Bellingshausen Sea when rates of fed animals were used (see Table 3). This suggests that Antarctic krill play an important role in regenerating ammonium since they may provide up to 33% of ammonium required by primary producers, at least in the open waters of the Bellingshausen Sea. Moreover, krill biomass was particularly low (1.7 to $13.4 \text{ mg DW m}^{-3}$), and it is known from acoustic surveys, that small nets such as the one used in this study underestimate the true biomass (Watkins et al. 2000). A much higher biomass is normally found in the region (Siegel & Loeb 1995, Siegel 2000, Shreeve et al. 2002, Quetin & Ross 2003, Hewitt et al. 2004, Siegel et al. 2004) and should raise the true ammonium concentration to much higher levels. As stated by Huntley & Lopez (1992), accurate measurements of biomass seem to be a major outstanding issue when estimating the production of zooplankton communities. Krill excretion, particularly along the western Antarctic Peninsula, should be evaluated by taking into account its variability in biomass, which should be much higher than our excretion rate measurements. Also, krill

Table 3. Primary production, ammonium regenerated by dominant groups of zooplankton and contribution of krill to ammonium demand in the different areas along the western Antarctic Peninsula. Biomass values are mean \pm SD where appropriate

Variable	Bellingshausen Sea	Bransfield Strait	Antarctic Sound
Zooplankton biomass in the top 100 m layer (g DW m^{-2})	2.19 ± 1.81	1.70 ± 2.20	0.80
Primary production ($\text{mmol N m}^{-2} \text{ d}^{-1}$)	2.67–6.67 ^a	6.67–13.33 ^a	8.95 ^b
Ammonium regenerated by salps in the top 100 m layer ($\text{mmol m}^{-2} \text{ d}^{-1}$) ^c	0	0.16	0
Ammonium regenerated by copepods in the top 100 m layer ($\text{mmol m}^{-2} \text{ d}^{-1}$) ^d	0.002	0.028	0.058
Ammonium regenerated by krill in the top 100 m layer ($\text{mmol m}^{-2} \text{ d}^{-1}$) ^e	0.089	0.025	0.012
Percent contribution of krill to ammonium demand ^e	1.13–3.32	0.19–0.38	0.13
Percent contribution of copepods to ammonium demand	0.03–0.08	0.21–0.43	0.64
Percent contribution of salps to ammonium demand	0	1.24–2.47	0
Ammonium regenerated by krill in the top 100 m layer ($\text{mmol m}^{-2} \text{ d}^{-1}$) ^f	0.89	0.15	0.012
Percent contribution of krill to ammonium demand ^f	13.30–33.25	1.09–2.18	0.13

^aFrom Varela et al. (2002), ^bfrom Alcaraz et al. (1998), ^cdetermined using salp excretion rate given by Alcaraz et al. (1998), ^dfrom Ikeda et al. (2001), ^edetermined using excretion rate of freshly caught krill (Atkinson & Whitehouse 2000), ^fdetermined using experimental close-to-field krill excretion rate (large and small specimens) of $\sim 17 \text{ nmol NH}_4^+ \text{ mg}^{-1} \text{ DW h}^{-1}$

swarms of >1000 ind. m⁻³, extending to >1 km² and swimming over distances of 100 km (Kils 1979) in a relatively short period (Kanda et al. 1982), were previously observed close to the Antarctic Peninsula (Nicol 2003). Despite the interannual biomass variability, krill swarms may represent an overwhelming source of regenerated nitrogen through grazing and excretion (Hirche 1983, Huntley & Nordhausen 1995) since high nitrogen concentrations are found within krill aggregations (Johnson et al. 1984, as cited in Priddle et al. 1997). Thus, the high quantities of iron (Tovar-Sánchez et al. 2007) and ammonium released by the krill community as observed in the present work provide optimal conditions for phytoplankton growth (Timmermans et al. 1994, Priddle et al. 1997, Whitehouse et al. 1999). The observation of higher excretion rates than those previously estimated indicate that krill play a much greater role in ammonium production in the Antarctic Peninsula than hitherto considered. These results suggest that high krill biomass fuels regenerated primary production in Antarctic waters.

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