

Propagule pressure and functional diversity: interactive effects on a macroalgal invasion process

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ABSTRACT: Invasive species are considered an increasing global threat to marine ecosystems. Understanding which factors can accelerate or reduce invasion success is one of the main goals of invasion ecology. The present study investigated the interactive effects of propagule pressure and native functional diversity on the invasion success of the invasive alga *Sargassum muticum* (Yendo) Fensholt over a 1 yr invasion process (i.e. settlement, recruitment and colonisation). The invasibility of macroalgal assemblages with different functional diversity was assessed under high and low propagule pressure. Synthetic assemblages resembling natural marine macroalgal assemblages were built and used as model systems. Recruitment discs were used to record early settlement in the assemblages. Recruitment and colonisation success were monitored in the field 2 and 10 mo after exposure to propagule rain in the laboratory. The availability of resources (space and photosynthetically active radiation [PAR]) was also monitored for each assemblage and used as predictive variables in the analyses. The effects of propagule pressure and functional diversity varied among the invasion stages. Space availability did not play a significant role in the recruitment success of the invader. However, PAR was a key resource in its invasion success. Overall, the invasion success of *S. muticum* was influenced differently by different traits of the recipient assemblages. Hence, invasion success may be better explained by the interaction of environmental factors that co-vary with species diversity and identity at different invasion stages.

KEY WORDS: Invasion success · Propagule pressure · Resources · Functional diversity · *Sargassum muticum* · Macroalgae

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INTRODUCTION

Increasing intensity of anthropogenic activities is promoting diversity loss and enhancing the introduction and spread of exotic organisms to new areas outside of their native range (Vitousek et al. 1997, Chapin et al. 2000). Beyond the changes in species range, which occur constantly in nature (Vermeij 1991), it is the accelerating rate and the magnitude of those changes that pose a threat to global biodiversity (Lodge 1993). Linked to human-

induced modifications, invasions of non-indigenous species (NIS) all over the world are enhancing habitat homogenisation (Vitousek et al. 1997, Olden & Poff 2004), with unknown consequences for Earth's biota and ecosystem functioning (Olden et al. 2004).

The invasion process, however, is a very complex phenomenon and consists of several transitional stages (Williamson 2006), such as arrival, establishment and success in spreading. Survivorship through each stage of the invasion process is the

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ultimate goal of any NIS. Understanding which factors influence invasion success through the different stages, and to what extent, is extremely important in invasion ecology (Kolar & Lodge 2001, White & Shurin 2007). Overall, 3 factors are usually cited as determining the fate of invasions: the biology of the introduced species (species invasiveness), the number and frequency of introductions (propagule pressure [PP]) and the susceptibility of the native community to invasion (community invasibility) (Lonsdale 1999). Nonetheless, invasion studies are usually focused on only one of these factors (Lonsdale 1999).

It is widely accepted that highly diverse communities are more resistant to invasions because of a more complete utilisation of resources, the so-called biotic resistance theory (see Elton 1958, Levine 2000, Fridley et al. 2007 for a review of diversity-invasibility relationships). For benthic algal communities as sessile photosynthetic organisms, space and light are important limiting resources (Sousa 1979). For example, lack of space and low light availability have been shown to negatively affect the invasion success of the brown macroalga *Sargassum muticum* (Stæhr et al. 2000, Britton-Simmons 2006). However, there is some ongoing debate about which component of diversity is more relevant in preventing community invasions and the mechanisms involved (e.g. Davis et al. 2000, Arenas et al. 2006, Stachowicz et al. 2007). For example, as a component of biodiversity, functional-group diversity has been suggested to have a greater impact on ecosystem processes than the total number of species per se (Tilman et al. 1997). Moreover, several studies indicate that the composition, i.e. identity effect, of a community may determine its susceptibility to invasion (Symstad 2000, Arenas et al. 2006).

It has been suggested that a large input of propagules is needed for invasion to occur in systems where environmental resistance is strong (D'Antonio et al. 2001). Thus, besides the importance of biotic resistance on the success or failure of invasions, there is a general consensus on the importance of PP (also termed 'introduction effort') as a strong predictor of invasion success (Colautti et al. 2006, Eschtruth & Battles 2011). PP can be defined as the total number of individuals introduced at a given location (Williamson 1996). This introduction of propagules into a location may occur through several releases. Thus, calculations of PP should include estimates of the number of individuals released in an event, i.e. propagule size, and the

number of introduction events, i.e. propagule number (Lockwood et al. 2005). A positive relationship between PP and invasion success has been suggested (Lonsdale 1999, Knights et al. 2012), although the true shape of the relationship is unclear (Lockwood et al. 2005), and its interaction with other factors (e.g. disturbance or resource availability) is mostly unknown (Sakai et al. 2001, Britton-Simmons & Abbott 2008).

In the present study, we explored how the traits of the recipient assemblage, i.e. functional diversity (FD) and identity, interact with PP to shape the success of an invasion. Our model systems were synthetic macroalgal assemblages with varying levels of FD and, as the invader, we used the Japanese seaweed *Sargassum muticum*. We tested the hypotheses that PP affects the invasion success of *S. muticum* at different invasion stages and that this effect is shaped by the FD of the recipient native assemblages. Invasibility of macroalgal assemblages is expected to decrease with FD (increased biotic resistance) and increase with greater PP (D'Antonio et al. 2001). In addition, we tested whether the availability of resources (space and photosynthetically active radiation [PAR]) would influence the invasion success.

MATERIALS AND METHODS

Study area

Field work was carried out in a large rock pool (11 × 2 m, average depth of 35 cm) in the mid-intertidal shore (~1.5 m above chart datum) at Viana do Castelo (41° 42' 25.72" N, 8° 51' 42.71" W, northwest coast of Portugal). The intertidal zone along the Portuguese coastline is influenced by a semidiurnal tidal regime. Native algal assemblages from low shore and rock pools are predominantly of 2 types: (1) understory algae dominated by turf-forming algae (e.g. *Corallina* spp., *Chondracanthus* spp., *Gelidium* spp. and *Ceramium* spp.) coexisting with encrusting coralline algae like *Lithophyllum incrustans* and (2) subcanopy species like *Bifurcaria bifurcata*, *Mastocarpus stellatus* and *Chondrus crispus* (Araújo et al. 2006). In the study area, *Sargassum muticum* is found in mid-intertidal rock pools, lower intertidal channels and subtidal zones. Herbivores include a variety of molluscs, such as snails (mostly *Gibbula* spp.), limpets (e.g. *Patella* spp.) and the sea hare *Aplysia punctata* Cuvier, 1803 (Monteiro et al. 2009a, 2012).

The invader, *Sargassum muticum*

Sargassum muticum is a marine macroalga native to SE Asia (Yendo 1907). Considered an invasive species all around the world (Norton 1976, Critchley et al. 1983), it is now one of the dominant seaweeds in many low intertidal and subtidal communities along the northern Portuguese coast (Monteiro et al. 2009a). This invasive species is monoecious and self-fertile (Norton 1976) and is considered a pseudo-perennial alga due to the presence of its lateral branches only for a portion of the year. Reproduction takes place in spring/summer. In the region of study, peak reproduction occurs between mid-April and late July, while senescence of lateral branches begins in mid-June (Arenas et al. 1995). Fertilised eggs are attached to the external surface of the receptacles (reproductive structure) and remain there until the development of tiny germlings with adhesive rhizoids (Norton 1981). After this, the embryo is liberated, sinks and reattaches to the substratum within a few meters of the parent plant (Deysner & Norton 1982). Adult individuals stay attached to the substratum by a holdfast.

Macroalgal assemblages and experimental design

Synthetic assemblages were created to manipulate functional group diversity. Based on previous knowledge of macroalgal communities inhabiting rock-pools in the northwest coast of Portugal (Araújo et al. 2006, Monteiro et al. 2009a), we selected 3 morpho-functional groups (modified from Steneck & Dethier 1994): (1) encrusting coralline species, such as *Lithophyllum incrustans*, (2) turf-forming species from the genera *Corallina* (usually 5 cm length) and (3) subcanopy species like *Chondrus crispus* or *Mastocarpus stellatus* (may reach 20 cm length). Small boulders bearing species of the selected morpho-functional groups were collected in the intertidal zone of Viana do Castelo and taken to the laboratory. Using a commercial tile cutter, rock pieces were cut into $2 \times 2 \times 2$ cm cubes and were attached to polyvinyl chloride (PVC) plates using fast-setting underwater cement and screws. Experimental assemblages consisted of $12 \times 17 \times 1$ cm PVC plates with 16 pieces of rock surrounded by 1 cm PVC pieces for support and protection (Fig. 1). Each rock piece represented 1 functional group characterised by a percentage of cover $>50\%$ or, in the case of subcanopy species, the presence of ≥ 1 adult individuals was considered a representation of that functional group.

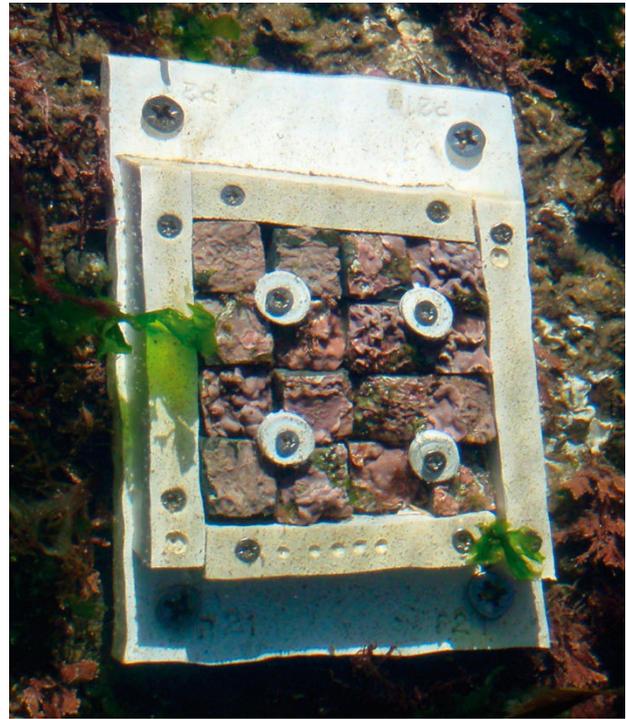


Fig. 1. A macroalgal assemblage plate used in the present study, in this case a monospecific assemblage of encrusting species

Although synthetic and relatively small assemblages were used, these assemblages resembled quite realistically the patchy pattern of intertidal algal assemblages, in which extremely small spatial scales of variation (<20 cm) seem to be among the most important sources of heterogeneity (Archambault & Bourget 1996).

Additionally, 4 recruitment discs (1.13 cm^2 each) were placed in each assemblage plate for assessment of *Sargassum muticum* early settlement. A total of 60 plates ($n = 4$) were built: 12 plates of only bare rock, 36 plates with only 1 functional group and 12 plates with 3 functional groups. The spatial distribution of the 3 functional groups within plates was random.

A 2-way factorial design was used to examine the interactive effects of FD and PP on different invasion stages (i.e. settlement, recruitment and colonisation). Both FD and PP were orthogonal fixed factors ($n = 4$). FD incorporated 4 levels: (1) monospecific assemblages of encrusting species, (2) monospecific assemblages of turf-forming species, (3) monospecific assemblages of subcanopy species and (4) 3 functional groups, i.e. high FD. In addition, we also prepared bare rock plates as controls. PP had 3 levels: (1) control/none, (2) low density and (3) high density.

Invasion procedure

PP level, here regarded as the density of propagule rain from the introduced species used in a single invasion event, was manipulated by placing different amounts of fertile individuals of *Sargassum muticum* (high density ≈ 25 kg; low density ≈ 13 kg; control = none) floating over the macroalgal assemblages. To do so, fertile individuals of *S. muticum* were collected from the field and transported to the laboratory, where they were rinsed with fresh water to eliminate grazers. To assure effectiveness of the PP treatment, only individuals with receptacles bearing exuded propagules were used. Experimental invasion was programmed to be around the new moon (22 July 2009) due to the semilunar periodicity of egg expulsion in *S. muticum* around new or full moons (Norton 1981, Monteiro et al. 2009b). Three tanks of ~ 300 l of seawater (i.e. 1 tank per treatment) were used for the experiment. A total of 20 macroalgal assemblages, 4 of each assemblage type plus bare rock plates, were randomly assigned to control, low and high PP treatments. Assemblage plates remained in experimental tanks for 1 wk. During the invasion procedure, *S. muticum* individuals remained at the low PP treatment for 3 d and at the high density treatment for 5 d. Additionally, 10 recruitment discs (5.11 cm^2) placed between assemblages in the tanks were used to assess the rain of propagules in each PP treatment. Low propagule rain consisted of an average ($n = 10$) of 195 ± 7.72 (\pm standard error [SE]) germlings cm^{-2} , while the high density level was $\sim 786 \pm 30.14$ (\pm SE) germlings cm^{-2} .

Assemblages from the control treatment were used to check for invasion by *Sargassum muticum* individuals in the field. Synthetic assemblages were transported to Viana do Castelo after laboratory experimental manipulation (control plates at Day 6 and plates from PP treatments at Day 8). Assemblage plates were randomly placed and screwed to the bottom of a large intertidal rock pool, separated by ~ 50 cm, where they remained for 12 mo. The selected rock pool was *S. muticum* free, as all of the individuals had been ripped off by their holdfasts days before the start of the experiment. During the course of the experiment, the rock pool was continuously monitored, and new individuals of *S. muticum* were removed with a scraper.

Measurements of invasibility and sampling procedure

Invasion of *Sargassum muticum* was monitored over 3 stages of invasion: settlement, recruitment and

colonisation. Settlement of *S. muticum* germlings was assessed 2 to 4 d after laboratory exposure to propagule rain, prior to transportation of assemblage plates to the field. Recruitment discs from each assemblage plate were used to record germling densities with the use of a stereomicroscope with additional dimmed light. Plates were then transported to the field. To assess the effect of transport, 26 additional recruitment discs (13 per PP treatment) with known densities of *S. muticum* germlings were taken to the field and then returned to the laboratory, where germlings were recounted. Mortality rate (\pm SE) due to transportation ($4.3 \pm 0.91\%$) was insignificant when compared to the 2 mo field natural mortality rate of $\sim 99\%$ (± 0.34) ($F_{1,50} = 2673.16$, $p < 0.0001$). Thus, mortality due to transportation was not taken into account when interpreting the results.

Two months after the laboratory exposure to propagule rain (September 2009), recruits were visible to the naked eye, and recruitment was assessed by counting *Sargassum muticum* recruits in the assemblage plates at the substratum level. Colonisation success was assessed by quantifying adult *S. muticum* (> 20 cm length) survivorship in assemblage plates from September 2009 to April 2010.

In addition, we estimated the percentage cover of algae (primary and secondary cover) and bare rock surface. Algal cover was estimated by visually dividing each rock cube in 4 and attributing a score from 0 to 4 for the functional groups present and adding up the 16 estimates (Dethier et al. 1993). The above canopy area was measured as secondary cover when subcanopy species were present.

Available resources (space and PAR) were also monitored for each macroalgal assemblage prior to the invasion procedure. Space availability was assessed during algal cover monitoring by quantifying the percentage of bare rock at the substratum level. PAR penetration (measured as the ratio of PAR under the algal canopy to that above it, or in the case of the encrusting species functional group, at the substratum level to the level 10 cm above it) was registered with a spherical quantum scalar irradiance sensor (1.9 cm diameter sensor, QSL-2100 Biospherical Instruments). PAR availability at substratum level was related to the irradiance intercepted per functional group and represented a measure of how much PAR reached the substratum. Experimental plates were submerged in a 5 l tank during PAR availability measurements. For each plate, the sensor was moved throughout the substrata (below assemblage canopy) and 10 cm above (or over the assemblage canopy) for 30 s each.

Statistical analyses

Differences in resource availability among FD treatments were assessed with a 1-way analysis of variance (ANOVA). FD was a fixed factor with 4 levels (i.e. encrusting, turf, subcanopy and high FD). Both space and PAR data had homogeneous variances (Cochran's test, $C = 0.3518$ and $C = 0.3511$, respectively; $p > 0.05$).

No *Sargassum muticum* recruitment was recorded in plates from the control PP treatment over the 12 mo experimental period. All new *S. muticum* recruits counted were assumed to come from the laboratory artificial invasion, and thus analyses were only performed for low and high PP treatments. Generalised linear models (GLM) were fitted for settlement and recruitment data to investigate the influence of the explanatory variables on the variation in the number of *S. muticum* germ-lings and recruits, respectively. For both sets of count data, we assumed a negative binomial distribution (an extension of GLM) to account for high overdispersion using the `glm.nb` function from the *MASS* package (Venables & Ripley 2002). Colonisation success was assessed with presence-absence data, in which 1 represented survivorship of *S. muticum*, and 0 represented no *S. muticum* survivorship to adult stage. For presence-absence data, we used a binomial GLM and a logit link function. Initial fitted models were built using PP (2 levels, high vs. low), FD (5 levels: bare rock, monospecific assemblages of encrusting species, monospecific assemblages of turf-forming species, monospecific assemblages of subcanopy species and high FD assemblages) and 2 continuous variables, PAR and space, as predictors. The continuous variable space availability was not included in the predictive model performed for settlement due to the use of recruitment discs in this stage. We also included biologically reasonable 2-way interactions in the initial model.

To obtain the minimum adequate model, we performed stepwise variable selection minimising Akaike's information criterion (AIC) and then hypothesis testing (Zuur et al. 2009). *A posteriori* contrasts were performed to assess the statistical significance of differences between the various levels of the categorical factor FD (Crawley 2009). To assess collinearity between variables, we calculated the variance inflation factor (VIF) with the `vif` function in the *car* package (Fox & Weisberg 2011). All VIF values were below 8, indicating low collinearity, so we kept all variables in the analyses. Residuals from the selected

model were plotted against the fitted values for model validation.

Rates of instantaneous mortality of the invader were used to test whether or not the mortality rates were constant over the different invasion stages and were calculated following Viejo et al. (1999):

$$\ln[(N_t/N_{t_0}) + 0.0001]/t \quad (1)$$

where N_t/N_{t_0} is the proportion surviving per macroalgal assemblage, and t is the time after invasion (0 to 2 mo: settlement period; 2 to 6 mo: recruitment period; 6 to 12 mo: colonisation period). At the end of the experiment, 2 of the 60 experimental plates were lost due to rough weather, and thus the colonisation data were analysed with 3 replicates instead of the initial 4 in 2 treatments. The design was balanced by adding the averaged values of the other 3 replicates to the model and recalculating the F -values (see Underwood 1997). Rates of instantaneous mortality were analysed with a 3-way ANOVA with PP (2 levels), time (3 levels) and FD (5 levels) as orthogonal fixed factors. Variance was homogeneous (Cochran's test, $C = 0.1109$; $p > 0.05$).

GLMs were carried out using R 2.12.1 (R Development Core Team 2010). ANOVA for resource availability and instantaneous mortality was performed with WinGMAV 5 (Underwood & Chapman 1998).

RESULTS

Resource availability

PAR availability at the substratum level varied significantly among FD treatments (ANOVA, $F_{3,28} = 35.20$, $p < 0.001$). Monospecific assemblages of encrusting species showed the greatest values of PAR reaching the substratum level (Fig. 2a). PAR availability was particularly low for assemblages composed of subcanopy species, i.e. monospecific assemblages of subcanopy species and assemblages with high FD, whereas assemblages of turf-forming species showed intermediate levels (Fig. 2a).

Space availability varied significantly among FD treatments (ANOVA, $F_{3,28} = 29.52$, $p < 0.0001$). Monospecific assemblages of subcanopy species and turf-forming species presented around 40 and 30%, respectively, of bare rock surface, while monospecific assemblages of encrusting species exhibit the lowest values (~9%). Plates assembled with the 3 functional groups presented intermediate levels of availability of primary space (Fig. 2b).

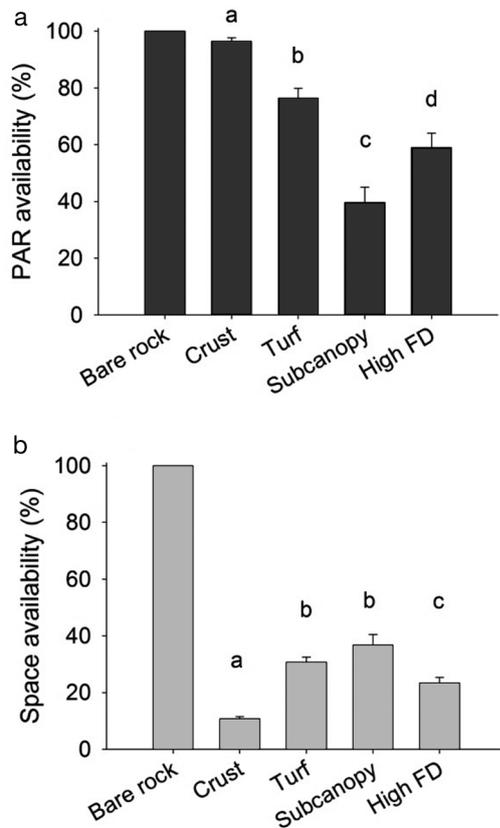


Fig. 2. Availability of resources (mean + SE, $n = 8$) at the substratum level over the different functional diversity treatments, including those made only of bare rock. FD: functional diversity. (a) Photosynthetically active radiation (PAR) availability; (b) space availability as percentage cover of bare rock. Different letters indicate significant differences at $p < 0.05$

Invasion success at different invasion stages

Invasion success of *Sargassum muticum* was partially explained by some of the predictive variables included in the models, namely PP, FD and PAR availability. However, the variables included in the minimum adequate models differed among the 3 stages of invasion, i.e. settlement, recruitment and colonisation. A different order of inclusion of the variables did not affect the results.

Abundances of *Sargassum muticum* individuals in the different experimental treatments for the different stages of invasion are presented in Fig. 3. Settlement of *S. muticum* germlings was interactively influenced by PP and functional group diversity (Table 1a), with much higher settlement in high PP treatments (Fig. 3a). Overall, the explanatory variables from the model explained ~92% of the variability in the number of *S. muticum* settled germlings

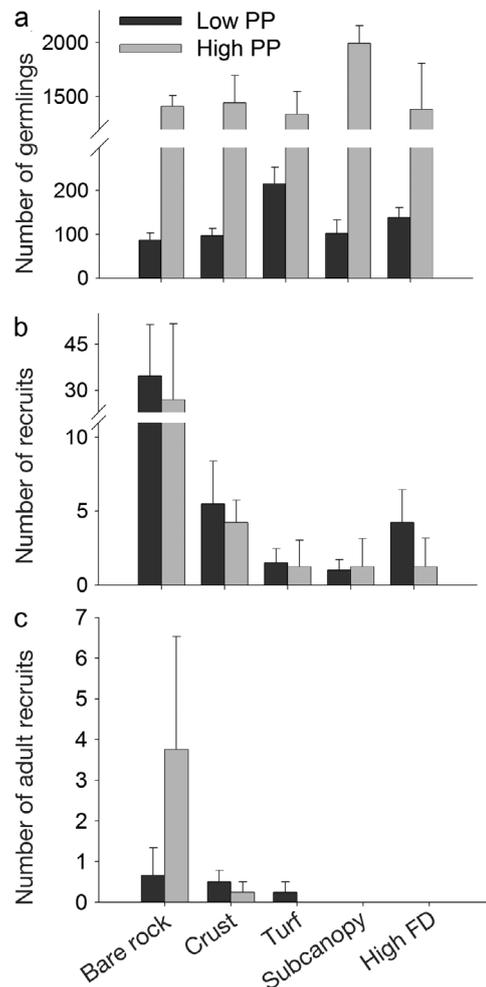


Fig. 3. *Sargassum muticum*. (a) Settlement (per 2.26 cm²), (b) recruitment (per 64 cm²) and (c) colonisation (per 64 cm²). Mean (+SE, $n = 4$ per functional group and bare rock) number of individuals on assemblages of different functional diversity, including those made only of bare rock. FD: functional diversity; PP: propagule pressure

(Table 1a). The interactive effects of PP and FD resulted in significantly contrasting numbers of settled germlings between macroalgal assemblages at low and high PP. In the low PP treatment, monospecific assemblages of turf-forming species and high FD assemblages registered significantly more settled germlings compared to the other assemblages (Fig. 3a). However, in the high PP treatment, the opposite pattern was observed (Fig. 3a).

Unlike the results observed for the settlement stage, PP was not a significant predictor of *Sargassum muticum* recruitment success. Thus, the fitted model for recruitment data, i.e. *S. muticum* juveniles recorded on assemblages at substratum level, inclu-

Table 1. Analysis of deviance revealed by generalised linear models fitted to the explanatory variables propagule pressure (PP), functional diversity (FD), photosynthetically active radiation (PAR) and space availability. Variable parameters are described for the minimum adequate model for the response variables (a) number of *Sargassum muticum* germlings, (b) number of *S. muticum* recruits and (c) presence-absence data for survivorship of *S. muticum* to adult stage. AIC: Akaike's information criterion

Variables	df	Deviance	Residual df	Residual deviance	p	AIC
a) Settlement						
Model			39	498.68		
PP	1	434.47	38	64.21	<0.0001	
FD	4	9.12	34	55.09	0.058	
PP:FD	4	14.28	30	40.81	0.006	
Minimum adequate model						530.77
Full model:						531.43
PP × FD + PAR + PP:PAR + FD:PAR, family = negative binomial						
b) Recruitment						
Model			39	88.372		
FD	4	46.302	35	42.070	<0.0001	
Minimum adequate model						209.7
Full model:						228.57
PP × FD + PAR + Space + PP:PAR + PP:Space + FD:PAR + FD:Space, family = negative binomial						
c) Colonisation						
Model			39	42.653		
PAR	1	15.425	38	27.228	<0.0001	
Minimum adequate model						31.29
Full model:						53.44
PP + FD + PAR + Space + PP:FD + FD:PAR, family = binomial						

ded only FD as a significant predictor and explained 52.4% of the variability in the number of *S. muticum* recruits (Table 1b). In particular, bare rock assemblages had significantly higher number of recruits than the other assemblage types (Fig. 3b).

At the colonisation stage, the survivorship of *Sargassum muticum* was mainly influenced by PAR availability at the substratum level (Table 1c). Survivorship of *S. muticum* to the adult stage was only observed at high PAR availability (Fig. 4).

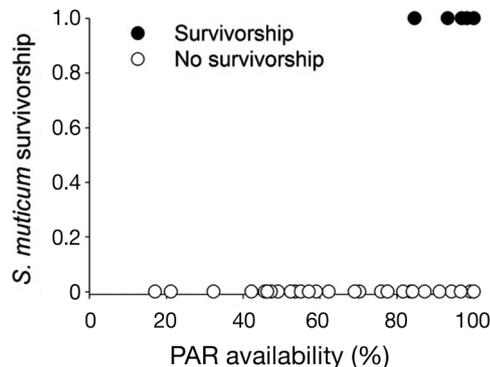


Fig. 4. *Sargassum muticum*. Colonisation success: variation in the survivorship of *S. muticum* to adult stage (from 2 to 9 mo) with photosynthetically active radiation (PAR) availability at substratum level on macroalgal assemblages

Instantaneous mortality rates

The instantaneous mortality rate varied over time independently from the FD of assemblages or PP (ANOVA, $p < 0.0001$; Table 2). Post hoc multiple comparisons showed that instantaneous mortality rates were significantly greater during the first 2 mo (mortality of settlers) than over the recruitment period (2 to 12 mo after invasion). In addition, instantaneous mortality rates were also greater in the first 4 mo (2 to 6 mo after invasion) compared to the last 6 mo of the experiment (Fig. 5).

Table 2. Results of the 3-way ANOVA testing for the effects of PP, time (T) and FD on rates of instantaneous mortality of *Sargassum muticum*. *Two missing replicates were replaced by the average value of the 3 replicates remaining in the treatment, and 2 df were subtracted from residuals

Source	df	MS	F	p
PP	1	0.082	0.11	0.741
T	2	136.747	177.22	<0.0001
FD	4	0.112	0.14	0.967
PP × T	2	0.209	0.27	0.764
PP × FD	4	0.076	0.10	0.982
T × FD	8	0.712	0.92	0.504
PP × T × FD	8	1.433	1.86	0.076
Residuals	88*	0.772		

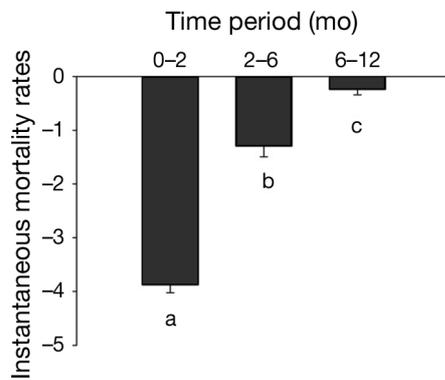


Fig. 5. *Sargassum muticum*. Instantaneous mortality rates (mean + SE, $n = 40$) during the settlement (0 to 2 mo), recruitment (2 to 6 mo) and colonisation period (6 to 12 mo after invasion). Different letters indicate significant differences at $p < 0.05$

DISCUSSION

The main and interactive effects of PP and the functional structure of assemblages varied over the different stages of invasion, i.e. settlement, recruitment and colonisation. Moreover, PAR availability at the substratum level was an important variable explaining the invasion success of *Sargassum muticum*. In contrast to previous studies of *S. muticum* (e.g. Stæhr et al. 2000, Britton-Simmons 2006), space availability did not play a significant role in explaining the recruitment success of the invader.

The role of PP in the invasion process was only relevant over the settlement stage (Clark & Johnston 2009). Although a high percentage cover of canopy, i.e. monospecific assemblages of subcanopy species, may act as increased biotic resistance under low PP, high PP seemed to overwhelm biotic resistance to invasion. The swaying of the algal fronds induced by water movement could free the substratum they covered, allowing propagules to reach the substratum. Here, because of the use of discs for settlement assessment, it was impossible to separate and model the effect of space availability on the number of settled germlings. Over the settlement stage, the observed larger number of *Sargassum muticum* germlings in algal turfs than in areas covered by encrusting species has also been registered for *Cystoseira* spp. (Fucales, Phaeophyta) (Benedetti-Cecchi & Cinelli 1992). Explanations, such as the idea that algal turf might provide a refuge from water displacement (Brawley & Johnson 1991) or from herbivores (Benedetti-Cecchi & Cinelli 1992), are not applicable within laboratory-based artificial invasion.

During the process of settlement, propagules have first to locate the substratum and then to establish surface contact (Fletcher & Callow 1992). Hence, a possible explanation is that the highly compacted algal turfs entrapped the germlings (Fletcher & Callow 1992), allowing a higher probability of surface contact. In contrast, greater biomass of subcanopy species could either prevent *S. muticum* germlings from reaching the substratum (Deyscher & Norton 1982) or remove germlings due to scour (Vadas et al. 1992). Nevertheless, the results suggest that the positive effect of turf in highly diverse assemblages could exceed the negative effect of subcanopy fronds. It is possible that *S. muticum* could settle over the fronds (Deyscher & Norton 1982), but that information was not substantiated in the present study. However, no recruits were observed over fronds later on.

The invader, *Sargassum muticum*, presents a highly successful reproductive strategy by releasing large amounts of germlings intermittently at intervals of ~15 d throughout its fertile period (Fletcher 1980, Norton 1981). This strategy significantly increases the chances of at least one release coinciding with favourable conditions for settlement, as suggested by studies manipulating disturbance (Britton-Simmons & Abbott 2008, Clark & Johnston 2009). Several studies have reported that space and light availability are important resources limiting *S. muticum* recruitment (Andrew & Viejo 1998, Britton-Simmons 2006, Stæhr et al. 2000). The results from the present study, however, suggest that the number of *S. muticum* at recruitment stage is not significantly affected by either resource. Nonetheless, the fact that bare rock plates showed the greatest number of *S. muticum* recruits suggests that space availability is a determinant for germling settlement. Encrusting species have been described as suitable settlement substrate in some studies (Ang 1985, Arenas et al. 2006), while in other studies they have been suggested to act as an effective barrier to colonisation (Deyscher & Norton 1982). In the present study, monospecific assemblages of encrusting species had the second greatest number of *S. muticum* recruits, highlighting the suitability of encrusting species as a substratum for *S. muticum*. In fact, these results corroborate our field observations that indicate the presence of *S. muticum* adults over a crustose substratum. The structure of assemblages explained ~52% of *S. muticum* recruitment; the other 50% may be related to other factors, such as nutrient availability, herbivory or wave action.

The colonisation success of *Sargassum muticum* was positively affected by the percentage of PAR

reaching the substratum. This result, together with the fact that monospecific assemblages of subcanopy species significantly decreases the percentage of PAR reaching the substratum, suggests that high values of canopy biomass may have a negative effect on the invader survivorship. Kendrick (1994) indeed showed that canopy significantly affects the survival of recruits >6 mo of age, possibly explaining our non-significant results for 2 mo old recruits regarding PAR availability.

Although highly diverse communities are expected to achieve a more complete utilisation of resources (the complementary resource-use hypothesis), our experiment based on functional attributes of species provided little support for that. Nonetheless, colonisation of *Sargassum muticum* was more successful in less complex assemblages, i.e. monospecific assemblages of encrusting species, which presented higher levels of both of the studied resources, PAR and space availability. Additionally, the rate of instantaneous mortality of *S. muticum* was significantly greater during the first 2 mo of age, indicating that early post-settlement processes are decisive in terms of overall invasion success. This fact suggests that post-settlement mortality that is not related to our predictors (e.g. density dependent mortality) eliminates the effect of PP in the final success of the invasion.

Overall, our results did not support the theory that more diverse assemblages are less susceptible to invasion by non-native species (Elton 1958, Lodge 1993, Naeem et al. 2000). However, the inherent traits of each functional group have strong effects on specific resources (Symstad 2000, Britton-Simmons 2006). This fact and our results suggest that the additive effects of each functional role in highly diverse assemblages are important in modifying resource availability and may therefore interfere with community invasibility (Arenas et al. 2006). Our findings show that the invasion success of *Sargassum muticum* over the different invasion stages is influenced differently by different recipient assemblage traits (Davis et al. 2000, D'Antonio et al. 2001). Throughout the invasion process, resource quality may be just as important as resource quantity, as previously implied by Clark & Johnston (2009). Hence, the interaction of environmental factors that co-vary with species diversity and identity may better explain invasion success (Eschtruth & Battles 2009).

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