COMMENT

Appropriate scales and data to manage seabird–fishery interactions: Comment on Torres et al. (2013)

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ABSTRACT: Torres et al. (2013; Mar Ecol Prog Ser 473:275–289) use fine-scale data on tracked locations of one species of albatross and fishing vessels to make a variety of assertions on interactions between them, especially in relation to risk of bycatch. Some of these assertions are incorrect, and we seek to clarify the issues and perspectives involved. We argue that while a fine-scale approach can provide interesting insights into foraging behaviour, large-scale risk analysis is needed because management measures (e.g. for bycatch mitigation) need to operate across all vessels in a fishery and across all seabird species at risk from the fishing operation. In addition, an estimate of 10% time spent in close proximity to vessels cannot be used to infer low bycatch risk alone: such an inference would need comparison to bycatch rate data. The analysis also does not take into account a number of factors known to affect the nature and duration of the association of albatrosses with fishing vessels, which limits the conclusions that can be drawn. However, fine-scale studies can provide important insight into factors affecting individual bycatch events, and studies at fine and broad scales will be complementary.

KEY WORDS: Bycatch · Satellite tracking · Albatrosses

Introduction

Torres et al. (2013) use fine-scale tracking data of an albatross species and periodic locations of fishing vessels, in waters off southern New Zealand, to infer interactions between the birds and the vessels. They assume that such interactions would be a good proxy for bycatch risk. On this basis they conclude that assessment of risk of interaction between seabirds and fishing vessels is better undertaken by inferring interactions at the level of individual seabirds and vessels than by assessment at the scale of fisheries and seabird populations. This represents a misunderstanding of the properties and utilities of fine-scale and larger scale analysis in relation to pragmatic management of fisheries at appropriate scales.

In addition, Torres et al. (2013) assert that conservation strategies, such as the reduction of non-target bycatch in commercial fisheries, should 'minimize
burden on the fishing industry’ (p. 276). We view the reduction of bycatch of non-target species as a requirement of the fishing industry in view of its obligation to undertake sustainable and responsible fishing (FAO 1995).

**Analysis of seabird–vessel interaction**

The analysis by Torres et al. (2013) of interaction between Buller’s albatross and vessels, stratified by fishery type, sex and year, does not take into account many factors known to be of high importance in influencing the nature and duration of the association of albatrosses with fishing vessels. These include: (1) vessel practice with respect to discard and offal management (the nature and frequency of these having a strong influence on the duration of vessel attraction for seabirds; Abraham et al. 2009); (2) use by vessels of measures to deter or prevent albatrosses from accessing fishing gear; (3) time of day of active fishing operations (albatrosses are largely inactive at night, e.g. Jimenez et al. 2009); (4) weather (low wind speed reduces transit and foraging rates in albatrosses; Weimerskirch et al. 2000); (5) the number and species of other seabirds in the vicinity which may compete for space behind a vessel (especially important for small albatrosses like Buller’s, readily out-competed by the co-occurring congener white-capped albatross *Thalassarche steadi*; Pierre et al. 2012); and (6) the availability of alternative food sources for foraging seabirds in the area. Lack of consideration of these variables severely limits the conclusions drawn by Torres et al. (2013), although they concluded that (2) and (5) were possible explanations for the birds foraging independently of vessels. In addition, the attractiveness of fisheries to seabirds is strongly influenced by the type of fishery. Thus, the low mean inferred overlap rate (10%) calculated by Torres et al. (2013) will partly reflect that only 2% of the vessel events related to longliners (of intense interest to seabirds; Anderson et al. 2011), whereas 90% related to trawling operations, only parts of which (late stages of hauling and when factory processing is active) are of high interest to seabirds (Sullivan et al. 2006, Watkins et al. 2008).

Second, while the authors are correct that congruent simultaneous high densities of seabirds and fishing vessels do not necessarily indicate interaction, they neglect to mention any of the numerous studies, using direct observation, which have shown that high densities of seabirds susceptible to bycatch in the vicinity of trawl vessels do lead to high levels of interaction (Sullivan et al. 2006, Watkins et al. 2008), including examples from New Zealand trawl fisheries (Abraham & Thompson 2009, Abraham et al. 2009, Pierre et al. 2012).

**Scales appropriate for interaction assessment for fisheries management**

The authors conclude that their results ‘demonstrate the increased precision and value of quantifying seabird–fishery overlap using a fine-scale approach, rather than relying on conventional large-scale analysis’ (Torres et al. 2013, p. 276). However, while a fine-scale approach can provide interesting insights into foraging behaviour, large-scale analysis is needed for effective risk assessment. This is because, inter alia: (1) management measures (e.g. for bycatch mitigation) need to operate across all vessels in a fishery and often across multiple fisheries operating at different spatial and temporal scales; and (2) measures need to apply to all seabird species at risk from the fishing operation (rarely, if ever, a single species of
albatross) and to individuals of the full range of age, stage and status.

Torres et al. (2013) criticize the large-scale overlap approach for sampling bias and providing only qualitative evaluation. However, the references given to support this statement are all from the pioneering (and preliminary) stages of this work (BirdLife International 2004, Phillips et al. 2006, Copello & Quintana 2009). Since then, very substantial advances have been made in quantitative analytical methods and their application (BirdLife International 2009, 2010, 2011), including overlap analysis, under the auspices of major regional fisheries management organizations (Waugh et al. 2012, ACAP unpubl. data) and in relation to modelling relationships between fishing effort and albatross population dynamics (Tuck et al. 2001, Thomson et al. 2009). Furthermore, the results of this analysis of Buller’s albatross during chick-rearing should not be generalised to other fisheries, situations and species (as occurs in the paper with the substitution of ‘seabird’ for ‘Buller’s albatross’). However, the paper does provide interesting insights into aspects of the interactions of adult, chick-rearing Buller’s albatrosses with vessels in the subantarctic waters of southern New Zealand.

### Conclusion

We show here that the premises and assertions in Torres et al. (2013) lack a sound basis, leading to misinterpretation of the data. In addition, the paper does not mention the possibility that studies at different scales may be complementary and not interchangeable, that most effective management of fisheries is perforce at relatively large scales, and that pragmatic fishery management at all scales has a greater need of direct observation of interactions than of inferences based on proximities from remote-recording data.

We agree that studying the fine-scale behaviour of seabirds in relation to relatively fine-scale data on the location of fishing vessels provides interesting insights. However, management of seabird bycatch is best addressed by combining quantification of overlap at medium (and large, where appropriate) scales with data on rates of seabird–vessel interaction and bycatch.

### LITERATURE CITED


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