

Effects of *Ulva prolifera* blooms on the carbonate system in the coastal waters of Qingdao

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ABSTRACT: This study investigated the effects of a bloom of the macroalga *Ulva prolifera* on the carbonate system in the Yellow Sea. Two cruises were carried out in the coastal waters of Qingdao, one during *U. prolifera* late bloom stages in 2015 and one after the bloom. In parallel, an 8 d incubation experiment was conducted to determine the variation in the seawater carbonate system during the late bloom period. During the late bloom period, pH was decreased by 0.06 units, while the concentrations of dissolved inorganic carbon (DIC), total alkalinity (TA) and the values of partial pressure of CO₂ (pCO₂) were increased by 27 μmol kg⁻¹, 57 μmol kg⁻¹ and 77 μatm in surface waters compared with those after the bloom, respectively. The mean (±SD) air–water CO₂ flux during the first cruise was estimated to be 0.82 ± 0.88 mmol m⁻² d⁻¹, while it declined to -1.31 ± 2.41 mmol m⁻² d⁻¹ during the second cruise. During both cruises, the pH decreased with depth, and this trend was reversed for DIC, whereas small variations in TA were detected throughout the water column. Comparing final data of the incubation experiment with initial values showed that pH was reduced by 0.23 ± 0.01 units, while DIC and TA increased on average by 71 ± 37 and 37 ± 18 μmol kg⁻¹, respectively. The release of CO₂ from *U. prolifera* occurred rapidly during the late bloom period, which resulted in conversion of Qingdao coastal waters from a sink to a weak CO₂ source.

KEY WORDS: *Ulva prolifera* · Carbonate system · Southern Yellow Sea · Late bloom period

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INTRODUCTION

Ulva prolifera is the main causative species of coastal green macroalgal blooms that result from the accumulation of free-floating macroalgae. This phenomenon has been a persistent feature in eutrophic coastal waters and estuaries worldwide (Lin et al. 2011, Guidone & Thornber 2013). Qingdao, a major city on the west coast of the Yellow Sea, has experienced a *U. prolifera* bloom every summer since 2007. The bloom not only substantially alters marine community structure and function, but also produces noxious odors and noxious-smelling deposits on beaches, causing economic losses for the tourism industry and environmental problems for local citi-

zens and governments (Ye et al. 2011, Hu et al. 2015, Y. Li et al. 2016).

U. prolifera results in an extremely proliferative, albeit ephemeral community. It typically originates from aquaculture of the red macroalga *Porphyra yezoensis* along the Jiangsu coastline during mid-April to early May, then drifts northward and rapidly expands to bloom in the southern Yellow Sea for 1.5 months, and finally reaches the coast near Qingdao in July or August (Liu et al. 2010, Keesing et al. 2011, H. Li et al. 2016). In recent years, numerous surveys have been conducted to study the origin of the *U. prolifera* bloom (Pang et al. 2010, Zhang et al. 2011, Xiao et al. 2013), associated carbon fixation mechanisms (Xu et al. 2012), and growth and nutri-

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ent uptake of *U. prolifera* in the southern Yellow Sea (Teichberg et al. 2010, Shi et al. 2015, H. Li et al. 2016, 2017). Additionally, laboratory experiments showed that it could survive in a wide range of temperature, salinity, pH and irradiance conditions (Wang et al. 2007, Lin et al. 2011).

The outbreaks of algal blooms cause ecosystem changes in the region, especially changes in the chemical environment. For example, a significant drawdown in the partial pressure of CO_2 ($p\text{CO}_2$) in the Pearl River estuary was attributed to a phytoplankton bloom (Dai et al. 2008). An increase in dissolved inorganic carbon (DIC) and a decrease in pH, as well as abnormal total alkalinity (TA) values were reported in the southern Yellow Sea during the *U. prolifera* post-bloom period (Hu et al. 2015). To the best of our knowledge, there have been no reports about the carbonate system during the *U. prolifera* bloom period, and only 2 cruises were carried out by Hu et al. (2015) during the post-bloom period in 2008. In the present study, 2 cruises were conducted in coastal waters of Qingdao during the late bloom period and after the bloom in summer 2015. The aim was to better understand the influence of the *U. prolifera* bloom on the carbonate system and evaluate the effects of *U. prolifera* green tides on the ecological environment of a small region that was close to the coast but not subjected to the influence of terrestrial inputs. A parallel incubation experiment was also performed onboard the research vessel, to determine variations in the characteristics of the carbonate system in response to the decay of *U. prolifera*. This study thus aimed to improve our understanding of the biogeochemical response to this bloom event,

with particular focus on the changes in the carbonate system in the context of ecosystem metabolism, and the response of the source or sink of CO_2 to the *U. prolifera* bloom during the late bloom period.

MATERIALS AND METHODS

Study area

Qingdao coastal waters, part of the southern Yellow Sea, are surrounded by the Shandong Peninsula and Jiaozhou Bay to the west (Fig. 1). Due to the shallow water depth and complex structure of the region, Qingdao coastal waters are dominated by semi-diurnal tidal currents (Zhao et al. 2011). Generally, *Ulva prolifera* initially appears in the Subei Shoal and then drifts to the southern Yellow Sea, where it grows rapidly into a green tide under favorable conditions. It is then transported to the Qingdao coast by surface ocean currents and the southeast monsoon in summer, and finally decays slowly and disappears in mid to late August (Hu et al. 2010, Keesing et al. 2011, Xu et al. 2016, Hu et al. 2017). The drift path and distribution of the bloom are provided by the North China Sea Branch of the State Oceanic Administration (SOA) (www.ncsb.gov.cn/n1/n127/n139/n39/index_5.html) (Fig. 2).

Two cruises were conducted on the RV 'Haidiao 235' during (1) the late bloom period, when thick accumulations of the green alga could still be observed at some stations, and (2) after the bloom period, when *U. prolifera* had disappeared from the coastal waters of Qingdao. The 2 cruises were carried

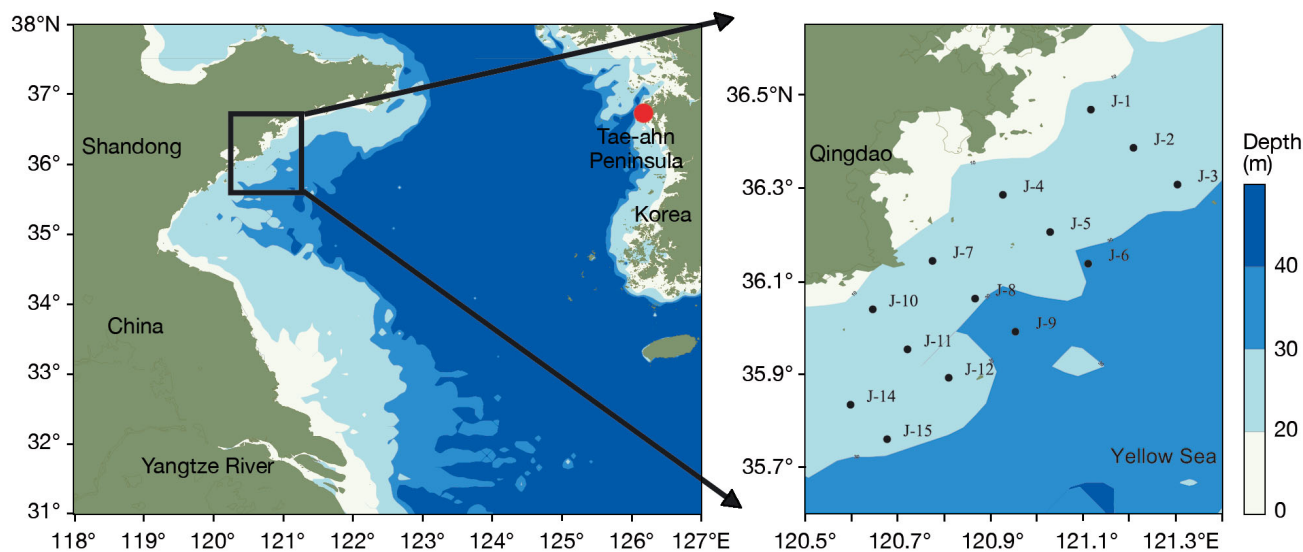


Fig. 1. Sampling stations (Stns J1–J15) in the coastal waters of Qingdao, southern Yellow Sea. Vertical color bar: water depth

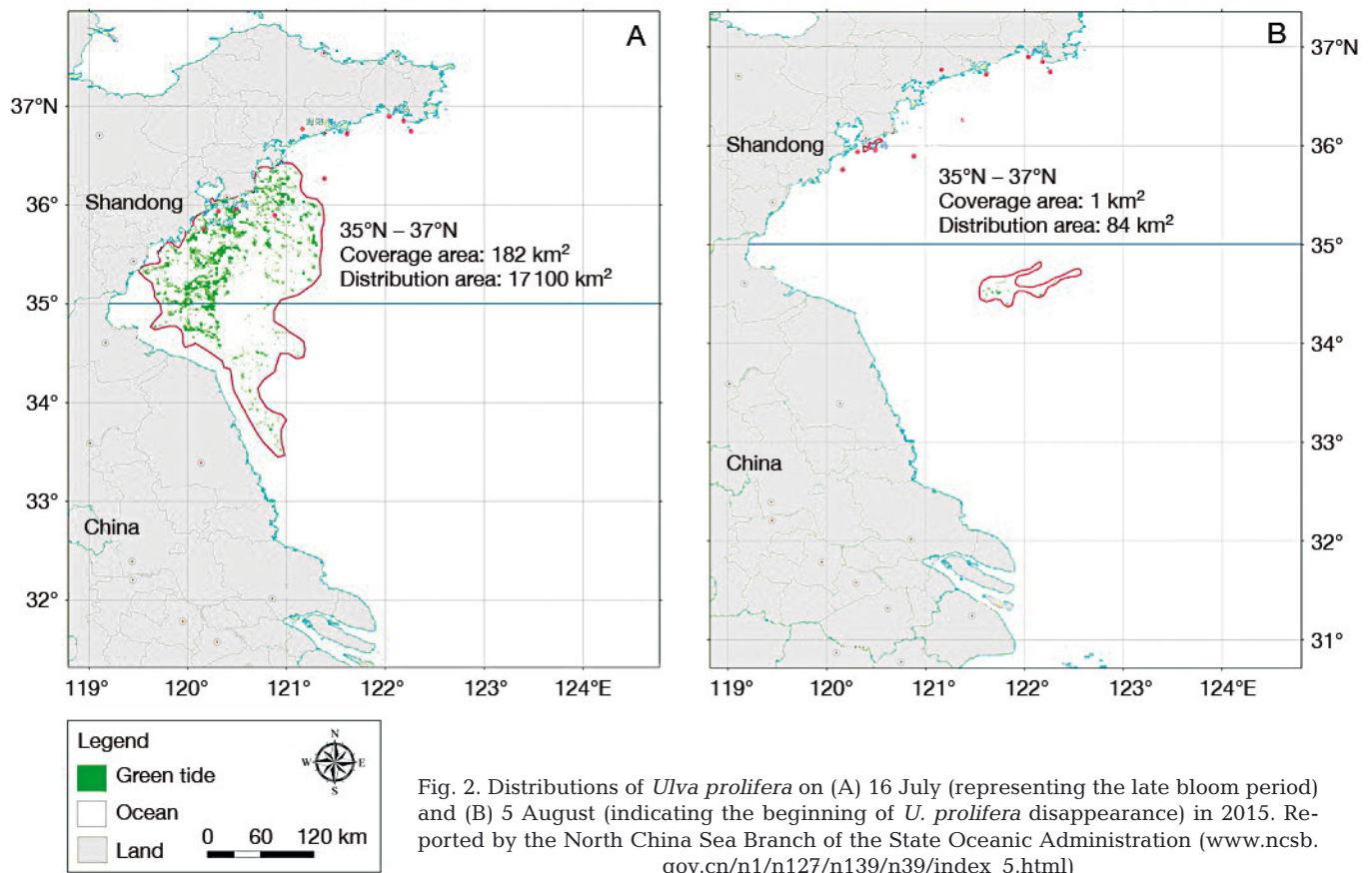


Fig. 2. Distributions of *Ulva prolifera* on (A) 16 July (representing the late bloom period) and (B) 5 August (indicating the beginning of *U. prolifera* disappearance) in 2015. Reported by the North China Sea Branch of the State Oceanic Administration (www.ncsb.gov.cn/n1/n127/n139/n39/index_5.html)

out on 16–17 July and 28–29 August 2015, respectively. The study area and observation stations are shown in Fig. 1. The distributions of *U. prolifera* on 16 July (representing the late bloom period) and 5 August (when the *U. prolifera* began to disappear) in 2015 are shown in Fig. 2. Thus, via these 2 cruises, the present study investigated the responses of the carbonate system to this *U. prolifera* bloom in the coastal waters of Qingdao.

Sampling

During the 2 cruises, temperature and salinity were recorded *in situ* with a conductivity, temperature, depth (CTD) sensor (Manta2; Eureka). When collecting samples for dissolved oxygen (DO), pH, DIC and TA, all bottles were allowed to overflow at least twice their volume to minimize contact with air. Samples for DO determination were collected using 125 ml glass bottles and immediately treated with the Winkler reagents (see below), and then sealed and immersed in ambient seawater. Discrete samples for DIC analysis were collected in 30 ml glass vials and TA samples (unfiltered and filtered; the TA values

reported in this article refer to the former) were collected using 100 ml rigid polyethylene bottles; DIC and TA samples were preserved with 100 μ l saturated HgCl₂ solution. An aliquot of 30 ml samples for dissolved organic carbon (DOC) was filtered through a 0.7 μ m-pore Whatman glass fiber filter (GF/F) and stored frozen in glass bottles at -20°C . Chlorophyll *a* (chl *a*) samples were collected by filtering 300 ml of seawater through 0.7 μ m-pore GF/F filters under low pressure (<15 kPa); filters were then frozen until analysis. In this study, the surface layer sampled corresponded to a depth of 0.5 m, the middle layer to a depth of 10 m and the bottom layer to a depth of 2 m from the bottom. All samples were taken onshore to the home laboratory and parameters were measured within several days of collection.

Incubation experiment

The effect of floating *U. prolifera* on the carbonate system of seawater was also examined via *in situ* incubation experiments in 5 l polyethylene bottles, filled with Qingdao coastal waters collected from 17 to 25 July 2015 and filtered onboard through a 0.2 μ m

membrane filter ($\phi = 47$ mm; Pall Corporation). The *U. prolifera* specimens were thoroughly washed and their epiphytes removed following macroalgal collection. According to the microscopic examination and identification by the Key Laboratory of Marine Genetics and Gene Resource Exploitation of the Ocean University of China, the growth of the algae was consistent with the standard characteristic of *U. prolifera*. Group 1 consisted of a blank with only filtered seawater, used as a control. Groups 2 and 3 were treated by adding 3 and 6 g wet weight of *U. prolifera* to 3 l of 0.2 μm -filtered seawater. All polyethylene bottles were maintained at *in situ* temperature by using a circulatory device whose water was provided continuously by shipboard ballast water. Cultivation of *U. prolifera* was established under the initial conditions of pH 8.03, DIC 2262 $\mu\text{mol kg}^{-1}$, TA 2342 $\mu\text{mol kg}^{-1}$, salinity 31.3, phosphate (PO_4^{3-}) 0.20 $\mu\text{mol l}^{-1}$, nitrate (NO_3^-) 3.16 $\mu\text{mol l}^{-1}$, nitrite (NO_2^-) 0.15 $\mu\text{mol l}^{-1}$, and ammonium (NH_4^+) 5.71 $\mu\text{mol l}^{-1}$. The ranges of temperature and light intensity were 15.46 to 25.92°C and 0.3 to 66300.0 lx during the course of incubation, respectively. The bottles were gently shaken 3 times d^{-1} to ensure well-mixed nutrients. All treatments were conducted in duplicate and the incubation lasted for 8 d. The carbonate system parameters in the culture medium, including pH, DIC and TA, were measured every 2 d during the incubation period. In addition, statistical analyses were done using SPSS v.19.0, and results were considered significant at $p \leq 0.05$.

Analytical methods

The DIC concentrations of water samples were determined using a DIC analyzer (AS-C2; Apollo Sci-Tech), with a precision of 0.1 to 0.2% (Cai & Wang 1998, Cai et al. 2004, Zhai et al. 2005). A 0.75 ml aliquot of the sample was pumped into the reactor and acidified by addition of 10% H_3PO_4 , and the concentration of the extracted CO_2 gas was subsequently quantified with a Li-Cor CO_2 infrared detector (Wang & Cai 2004). The concentration of TA was determined using a 20 ml sample by Gran titration operated by a computer-controlled Kloehe digital pump; titration precision was 0.1% (Cai et al. 2004, Dai et al. 2008). Both the DIC and TA analyzers were calibrated using certified reference materials from A. G. Dickson's laboratory (Scripps Institution of Oceanography). For pH measurements, water samples were collected in narrow-mouth glass bottles and kept in a thermal bath ($25 \pm 0.1^\circ\text{C}$) for about 30 to 60 min. The pH was then measured onboard using a Ross Orion combination

electrode (Ross-8102) and a Fisher pH meter (Star A211) on a National Bureau of Standards (NBS) scale (precision: ± 0.005) and finally the values were corrected to the *in situ* temperature. The $p\text{CO}_2$ was calculated from the pH and DIC using CO_2SYS software (Pierrot et al. 2006) with the equilibrium constants determined based on Mehrbach (1973), refitted by Dickson & Millero (1987), and potassium bisulfate (KHSO_4) as described by Dickson (1990).

Air–water CO_2 fluxes (F , $\text{mmol m}^{-2} \text{d}^{-1}$) were estimated according to the following equation:

$$F = k \times \alpha \times \Delta p\text{CO}_2 \quad (1)$$

where k (cm h^{-1}) is the gas transfer velocity of CO_2 , α is the solubility coefficient of CO_2 at the *in situ* temperature and salinity, and $\Delta p\text{CO}_2$ is the difference between the water and air $p\text{CO}_2$. The atmospheric $p\text{CO}_2$ was acquired from the values measured at the Tae-ahn Peninsula (36.731°N , 126.131°E ; red dot in Fig. 1), which was adjacent to the southern Yellow Sea. Positive flux values indicate the net CO_2 exchange from the water body to the atmosphere, while negative values represent the transfer of CO_2 from the atmosphere to the water. The calculations of k were based on the Wanninkhof (1992) empirical function of wind speed (since the *in situ* k value was not available), as follows:

$$k = 0.39 \times U \times (Sc / 660)^{-0.5} \quad (2)$$

where U (m s^{-1}) is the wind speed at 10 m height, here represented by long-term wind speed; Sc is the Schmidt number of CO_2 in seawater, and 660 is the Sc value in seawater at 20°C .

The DOC concentration was determined by catalytic high temperature oxidation using a total organic carbon analyzer (Shimadzu TOC-V_{CPH}; Shimadzu) with platinum catalyst at 680°C . In order to eliminate the inorganic carbon component, the samples were acidified using 2 M HCl, and injected into the furnace by the autosampler; the combusted products (CO_2) were detected by the non-dispersive infrared gas analyzer. The precision was $< 2.0\%$ (Yang et al. 2010).

Dissolved oxygen was measured using the Winkler method (Dickson 1995); MnCl_2 and alkaline KI solutions were added to DO bottles immediately after sampling with care taken not to introduce bubbles. Samples were maintained in the dark for at least 1 h before titration with the $\text{Na}_2\text{S}_2\text{O}_3$ solution under H_2SO_4 conditions. All samples were analyzed within 24 h of collection.

Chl *a* on the membrane was extracted with 90% acetone for 12 h in the dark at 4°C , centrifuged for 10 min, and then measured with a fluorescence

spectrophotometer (F-4500; Hitachi) according to Parsons et al. (1984). Because samples were filtered through Whatman GF/F membranes after the removal of *U. prolifera*, the concentrations of chl *a* in this study represent the biomass of micro-phytoplankton.

RESULTS

Temperature and salinity in coastal waters of Qingdao

The ranges and mean values of temperature and salinity are listed in Table 1. The vertical temperature profile decreased significantly with depth, while salinity showed a very slight increasing trend from the surface to the bottom layer. However, the water temperature and salinity did not show a detectable difference between the late bloom period and after the bloom. In the horizontal distribution, the temperature generally decreased from the coast to offshore waters, and salinity showed an opposite trend during both cruises (Figs. 3 & 4). Furthermore, the temperature and salinity were within suitable ranges for the growth of *Ulva prolifera* (temperature: 15 to 30°C, salinity: 16 to 40) during the 2 cruises (Gao et al. 2014).

Distributions of DO, chl *a* and DOC in coastal waters of Qingdao

The oxygen content of the water column indicated oversaturated conditions during both cruises. The highest DO value was obtained in the middle layer during the first cruise, while it exhibited a minimum value in the middle layer during the second cruise. It is notable that DO concentrations throughout the water column were lower in July than those in August, as shown in Table 1.

Concentrations of chl *a* were in the range of 0.01 to 0.21 $\mu\text{g l}^{-1}$ in the surface seawater, with an average ($\pm\text{SD}$) value of $0.10 \pm 0.06 \mu\text{g l}^{-1}$ during the first cruise. However, chl *a* concentrations increased to between 0.35 and 0.64 $\mu\text{g l}^{-1}$ in the surface layer when *U. prolifera* was no longer present during the second cruise. Compared with the study period without *U. prolifera*, the average values of chl *a* were lower in all layers during the late bloom period (Table 1), which could be attributed to both the competition of *U. prolifera* with microalgae for nutrients and the inhibition of microalgal photosynthesis by the *U. prolifera* cover, and thus light attenuation in surface water.

Horizontal distributions of DOC are shown in Fig. 5. During the late bloom period, concentrations

Table 1. Ranges and means (in brackets) of hydrographic and carbonate system data during the *Ulva prolifera* late bloom and after the bloom in the coastal waters of Qingdao. DO: dissolved oxygen; DOC: dissolved organic carbon; DIC: dissolved inorganic carbon; TA: total alkalinity; $p\text{CO}_2$: partial pressure of CO_2

Parameter	Cruise	Surface layer	Middle layer	Bottom layer
Temperature ($^{\circ}\text{C}$)	July	23.74–25.49 (24.39)	21.13–24.32 (22.86)	15.34–22.80 (19.93)
	August	22.67–26.38 (24.37)	18.85–24.53 (21.53)	15.34–21.54 (19.28)
Salinity	July	30.74–31.51 (31.14)	31.02–31.51 (31.26)	31.11–31.57 (31.34)
	August	30.82–31.57 (31.25)	30.98–31.46 (31.22)	31.02–31.38 (31.26)
DO (mg l^{-1})	July	6.44–6.88 (6.69)	6.26–7.13 (6.74)	5.33–7.13 (6.44)
	August	6.55–7.57 (7.16)	6.23–8.24 (7.08)	6.42–11.82 (8.97)
DOC ($\mu\text{mol kg}^{-1}$)	July	61–183 (131)	61–240 (138)	80–190 (118)
	August	67–154 (130)	72–157 (127)	92–156 (129)
Chl <i>a</i> ($\mu\text{g l}^{-1}$)	July	0.01–0.21 (0.10)	0.06–0.55 (0.25)	0.05–0.32 (0.18)
	August	0.35–0.64 (0.45)	0.23–0.35 (0.28)	0.23–0.42 (0.34)
DIC ($\mu\text{mol kg}^{-1}$)	July	2008–2152 (2087)	2064–2174 (2097)	2048–2148 (2113)
	August	2032–2096 (2060)	2036–2194 (2071)	2074–2228 (2160)
TA ($\mu\text{mol kg}^{-1}$)	July	2345–2414 (2385)	2348–2396 (2373)	2264–2455 (2388)
	August	2296–2349 (2328)	2297–2366 (2330)	2283–2356 (2333)
TA _(filtered) ($\mu\text{mol kg}^{-1}$)	July	2329–2392 (2365)	2319–2387 (2353)	2335–2356 (2317)
	August	2271–2335 (2311)	2271–2335 (2311)	2269–2343 (2319)
pH	July	7.99–8.11 (8.04)	7.90–8.06 (7.99)	7.84–8.02 (7.90)
	August	7.99–8.24 (8.10)	7.98–8.18 (8.09)	7.78–8.09 (7.88)
$p\text{CO}_2$ (μatm)	July	380.5–466.5 (427.0)	355.0–555.8 (450.2)	431.9–579.7 (506.6)
	August	245.5–471.8 (349.7)	224.1–441.9 (332.6)	316.5–680.5 (528.7)

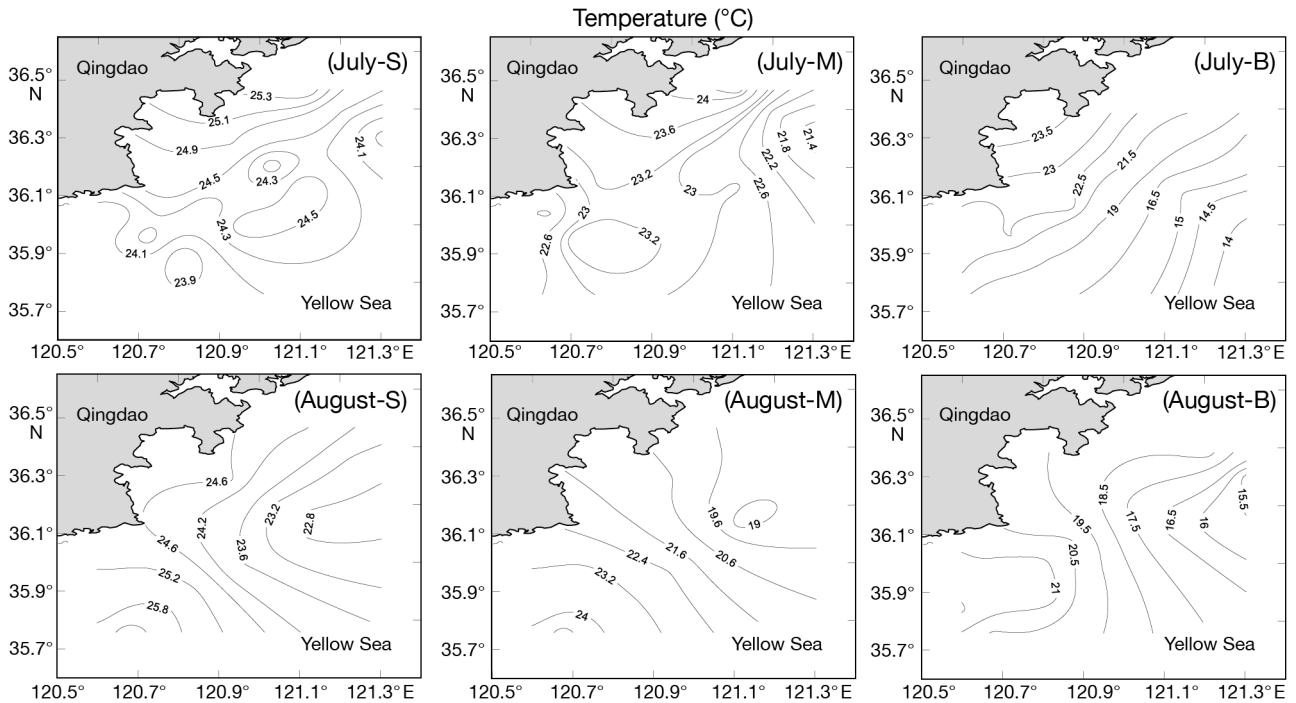


Fig. 3. Horizontal distributions of temperature ($^{\circ}\text{C}$) in the surface (S), middle (M), and bottom (B) layers of the coastal waters of Qingdao during July and August

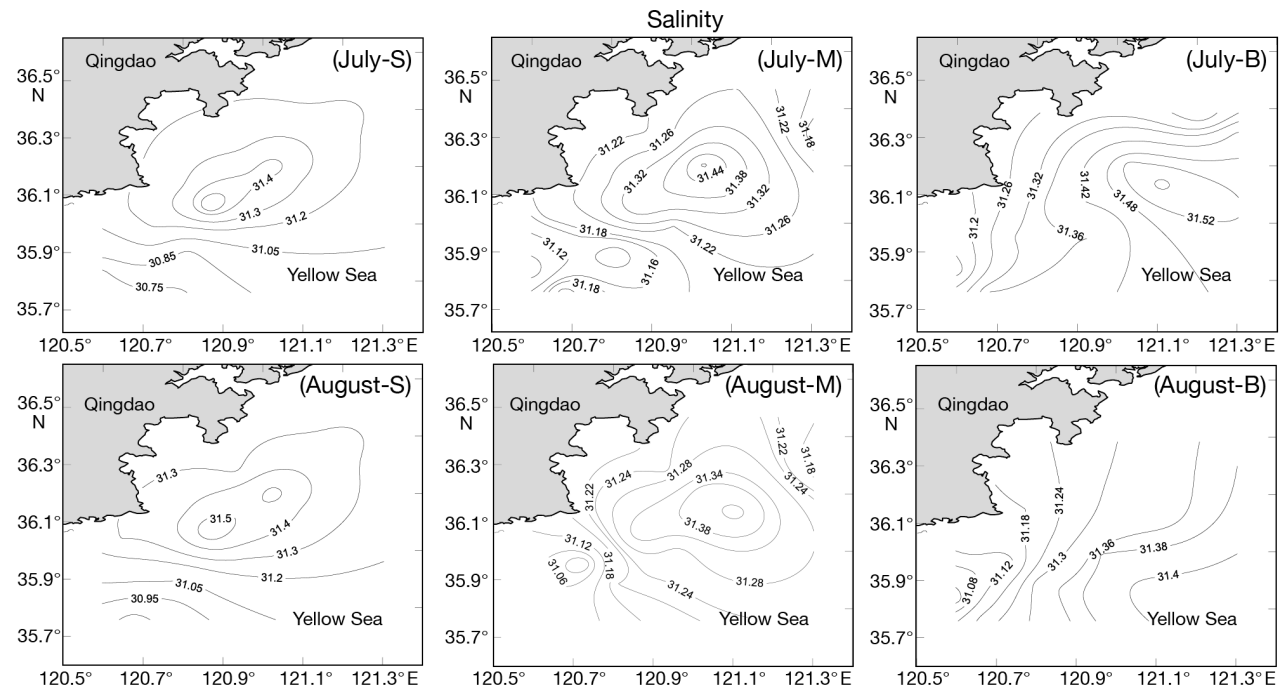


Fig. 4. Horizontal distributions of salinity in the surface (S), middle (M), and bottom (B) layers of the coastal waters of Qingdao during July and August

of DOC in surface waters generally decreased from inshore to offshore. The concentrations of DOC in the middle layer decreased at first and then increased from nearshore to offshore; the minimum value was

obtained near Stn J5 (Fig. 1). Relatively high DOC concentrations were observed at inshore stations, and the concentration gradient was perpendicular to the shoreline in the bottom layer. After the bloom

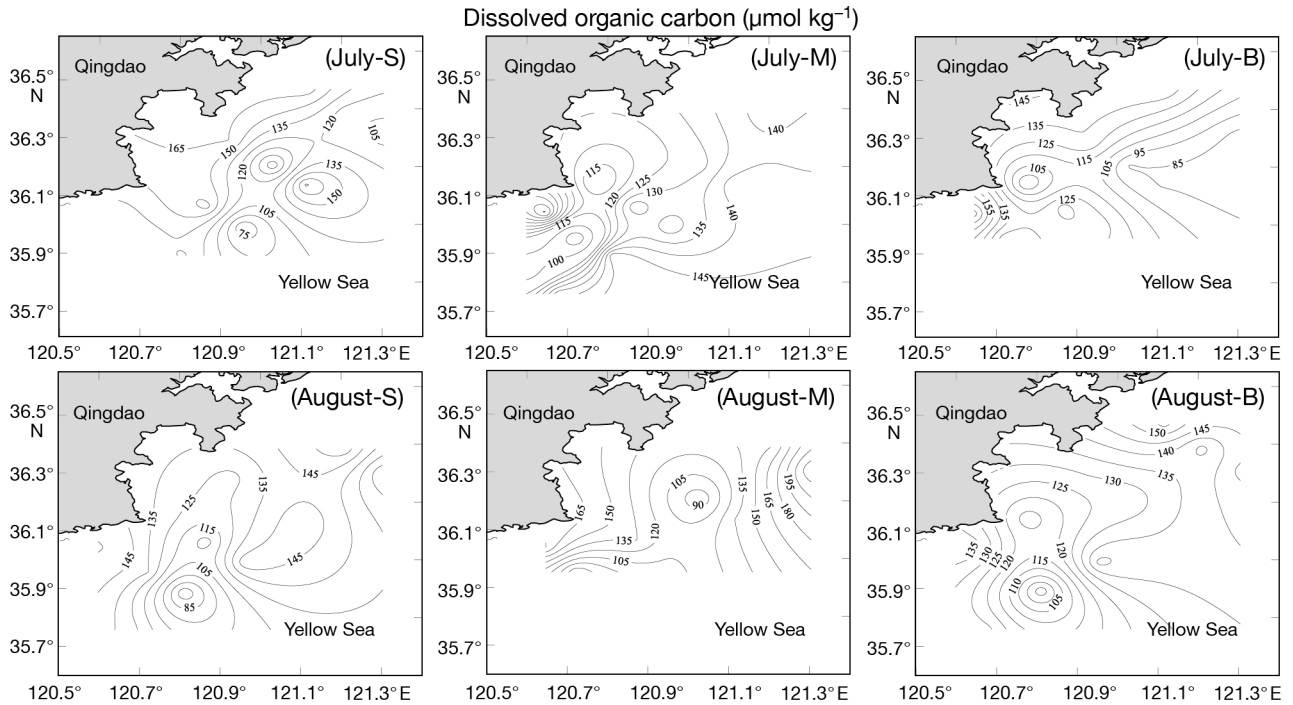


Fig. 5. Horizontal distributions of dissolved organic carbon (DOC; $\mu\text{mol kg}^{-1}$) in the surface (S), middle (M), and bottom (B) layers in the coastal waters of Qingdao during July and August

period, concentrations of DOC generally increased from south to north in the surface layer, and the location of the DOC minimum was in the vicinity of Stn J11 in the middle and bottom layers. Additionally, DOC concentrations in July were lower than those in August in the surface and bottom layers, while the concentrations of DOC in July were higher than those in August in the middle layer (Table 1).

Distributions of carbonate system parameters in coastal waters of Qingdao

The ranges and mean values of carbonate parameters (DIC, TA, pH and $p\text{CO}_2$) showed clear spatial and temporal differences. These are presented in Table 1 and Figs. 6–9.

Concentrations of DIC and pH in surface waters varied from 2008 to $2152 \mu\text{mol kg}^{-1}$, and 7.99 to 8.11, respectively, with average values of $2087 \pm 44 \mu\text{mol kg}^{-1}$ and 8.04 ± 0.03 , respectively, during the first cruise. The vertical distribution of DIC increased gradually with depth, while pH exhibited an opposite trend throughout the water column. During the period of no detectable *U. prolifera*, DIC and pH in the surface layer were in the range of 2032 to $2096 \mu\text{mol kg}^{-1}$, and 7.99 to 8.24, respectively, averaging $2060 \pm 19 \mu\text{mol kg}^{-1}$ and 8.10 ± 0.08 , respec-

tively. The vertical variation in DIC and pH was generally similar to that observed during the late bloom period. In addition, TA values varied within a relatively narrow range, from 2345 to 2414 and 2296 to $2349 \mu\text{mol kg}^{-1}$, respectively, during these 2 cruises.

During the first cruise, high DIC concentrations prevailed in the surface layer of inshore waters, and the concentration gradient was perpendicular to the shoreline. However, DIC showed an opposite trend in the bottom layer, and low DIC concentrations primarily occurred in Stn J6 and in the middle layer of the southwest portion of the study area (Fig. 6). As shown in Fig. 7, the horizontal distribution of TA in the surface layer showed an increasing trend from inshore to offshore waters, whereas TA concentrations in the middle layer showed only minor variation (2348 to $2396 \mu\text{mol kg}^{-1}$) among the stations. The minimum value was obtained near Stn J10 in the bottom layer. During the second cruise, DIC showed distribution patterns similar to those observed during the first cruise (Fig. 6). The concentrations of TA increased from the coast to offshore waters in the surface and bottom layers, but not in the middle layer. The maximum concentration of TA was obtained near Stn J12 in the middle layer (Fig. 7).

The pH during the late bloom period increased in a northeast–southwest direction in surface water,

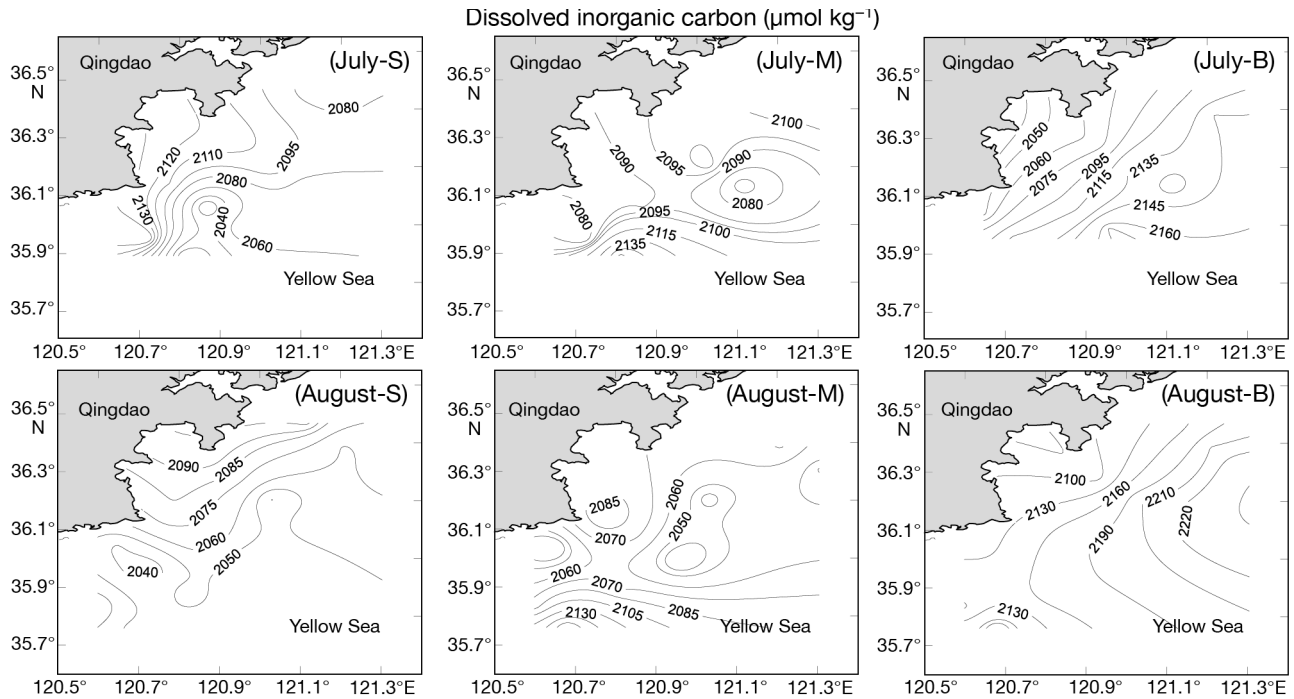


Fig. 6. Horizontal distributions of dissolved inorganic carbon (DIC; $\mu\text{mol kg}^{-1}$) in the surface (S), middle (M), and bottom (B) layers of the coastal waters of Qingdao during July and August

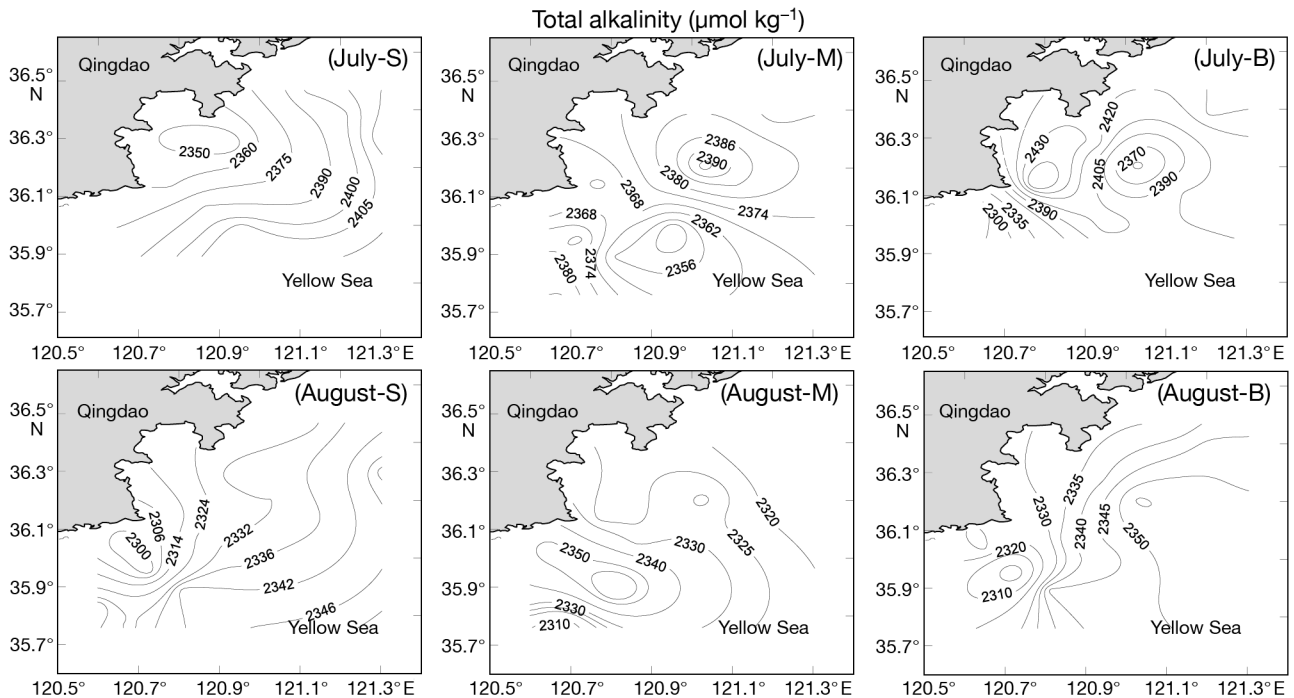


Fig. 7. Horizontal distributions of total alkalinity (TA; $\mu\text{mol kg}^{-1}$) in the surface (S), middle (M), and bottom (B) layers of the coastal waters of Qingdao during July and August

whereas it increased from nearshore to offshore waters in the middle layer of the study area, and exhibited a reverse trend in the bottom layer (Fig. 8). Conversely, $p\text{CO}_2$ increased in a stepwise

manner from the south to the north in the surface layer of the study area, ranging from 381 to 466 μatm . It decreased from the coast to offshore waters, ranging between 355 and 556 μatm in the

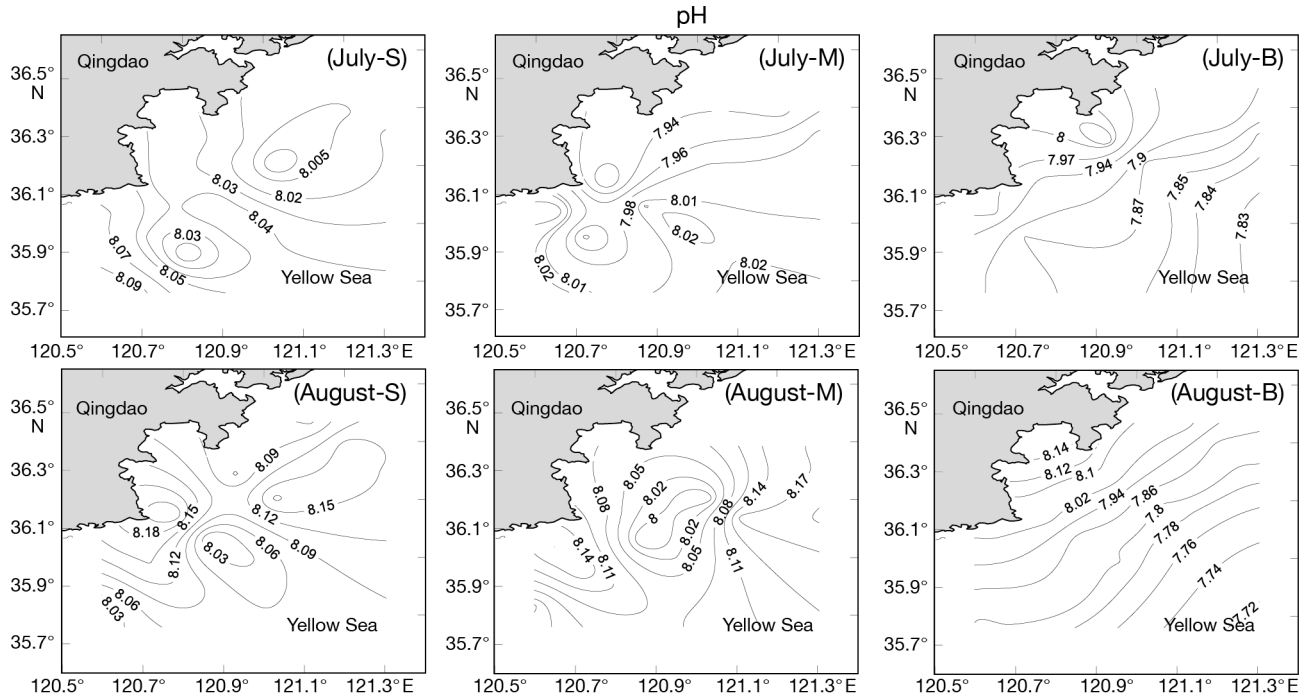


Fig. 8. Horizontal distributions of pH in the surface (S), middle (M), and bottom (B) layers of the coastal waters of Qingdao during July and August

middle layer. It initially increased and then decreased along an inshore to offshore direction in the bottom layer, and high $p\text{CO}_2$ values (580 and 591 μatm) were found at Stns J5 and J11, respectively (Fig. 9). Under conditions without *U. prolifera*, the pH in the surface layer generally decreased from inshore to offshore waters and relatively higher values (~ 8.20) were found along the coast-line. A low pH value of 7.98 was recorded in the vicinity of Stn J8 in the middle layer, and the distribution pattern of pH in the bottom layer was generally consistent with that of the late bloom period. The $p\text{CO}_2$ in the surface layer increased from inshore to offshore waters and from the northern to the southern region with a range of 245 to 472 μatm . This parameter again increased from the north to the south in the middle layer, and gradually increased from inshore to offshore waters in the bottom layer. In general, the location of the pH minimum coincided with that of $p\text{CO}_2$ maximum.

Variations of the carbonate system parameters during the incubation experiment

Fig. 10 shows the incubation results from 17 to 25 July 2015. The carbonate system parameters of the incubation medium with *U. prolifera* were signifi-

cantly higher than those of the control (1-way ANOVA, $p < 0.05$). There was no significant difference between incubations when 3 and 6 g wet weight of *U. prolifera* were added (1-way ANOVA, $p > 0.05$). To be specific, the pH of culture medium in Groups 2 and 3 increased gradually at first, attaining peak values of 8.35 ± 0.05 and 8.38 ± 0.02 on Day 4, and then decreased to 7.79 ± 0.08 and 7.81 ± 0.08 on Day 8, respectively. The pH values in Group 3 were slightly higher than those in Group 2 during the early stage of incubation, whereas the opposite trend was found during the later incubation period.

The DIC concentration in the culture medium of Groups 2 and 3 decreased by ca. 117 and 151 $\mu\text{mol kg}^{-1}$, respectively, within the first 2 d. Subsequently, the release of DIC from *U. prolifera* occurred rapidly, the concentration of DIC in the culture medium of Groups 2 and 3 increased by ca. 159 and 245 $\mu\text{mol kg}^{-1}$, respectively, from Day 2 to the end of the experiment. However, TA in the culture medium of Groups 2 and 3 increased to 2404 ± 4 and 2435 ± 7 $\mu\text{mol kg}^{-1}$ initially, then decreased to 2326 ± 9 and 2342 ± 6 $\mu\text{mol kg}^{-1}$, and finally increased to 2366 ± 10 and 2392 ± 6 $\mu\text{mol kg}^{-1}$, respectively. The values of TA in Groups 2 and 3 on Day 8 were higher than their corresponding initial values. Variation patterns of DIC and TA in Group 2 were highly consistent with those in Group 3 (Fig. 10).

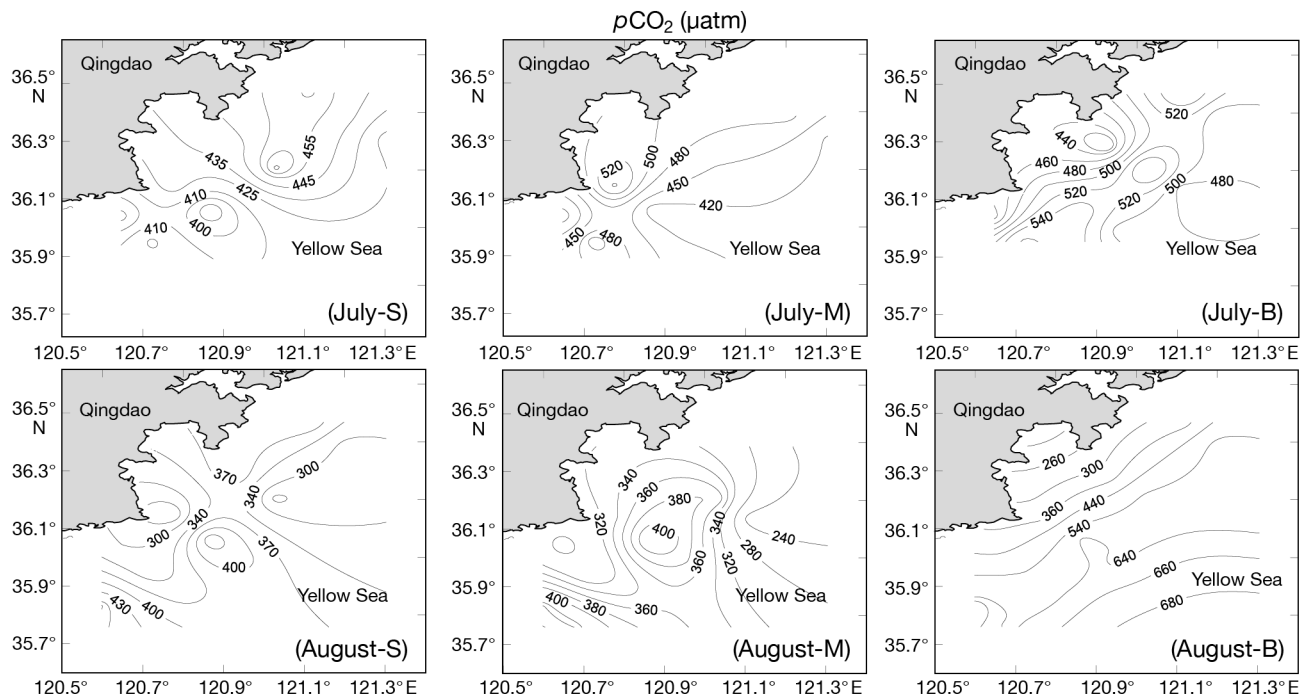


Fig. 9. Horizontal distributions of partial pressure of carbon dioxide ($p\text{CO}_2$; μatm) in the surface (S), middle (M), and bottom (B) layers of the coastal waters of Qingdao during July and August

DISCUSSION

Influence of the *Ulva prolifera* bloom on DO and pH

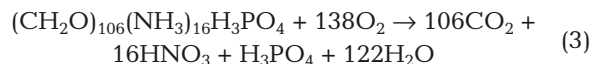
The content of DO in seawater is an important indicator of the biological growth and pollution status of a body of water. The main source of DO in seawater is the atmosphere, although DO can also be produced by planktonic and benthic algae via photosynthesis in the sea (Xia et al. 2009). As shown in Table 1, our entire study area was found to be rich in oxygen due to the fact that coastal waters were well-mixed by tidal currents (Qu et al. 2015). Nevertheless, the concentrations of oxygen throughout the water column in July were lower than those in August. This was most likely related to the fact that *Ulva prolifera* cover hindered the air–water exchange to some extent, which led to the limited decrease in DO in the late bloom regions (Gao et al. 2014).

As noted in Table 1, pH values in July were lower than those in August except for those in the bottom layer. This discrepancy could be attributed to floating of *U. prolifera* in the surface layer, which released a large amount of CO_2 via respiration and decomposition processes. Concurrently, the pH value on Day 8 of the incubation experiment was reduced by ~ 0.23 units compared with the initial value. This result was consistent with that of field observations obtained in

the present study in the coastal waters of Qingdao, and it also agrees well with the estimate by Hu et al. (2015).

Response of the carbonate system to *U. prolifera* late bloom conditions

A plot of TA versus DIC was generated to account for the influence of water mixing. The TA–DIC relationship should be a straight line during simple water mixing (Dai et al. 2008). Fig. 11 portrays the variation of TA–DIC in the surface layer during the 2 cruises; however, there was no detectable correlation between the 2 parameters. According to the equation for biological oxygen consumption and respiration (Wang et al. 2003):



every unit of Redfield-like organic matter produced or respired is associated with 106 units of DIC change, but only a small TA change (i.e. $\Delta\text{DIC}/\Delta\text{TA} = -106/17$) when significant photosynthesis or respiration occurs. Therefore, major departures from the straight line of mixing towards higher DIC values caused by biological release were observed during the late bloom period.

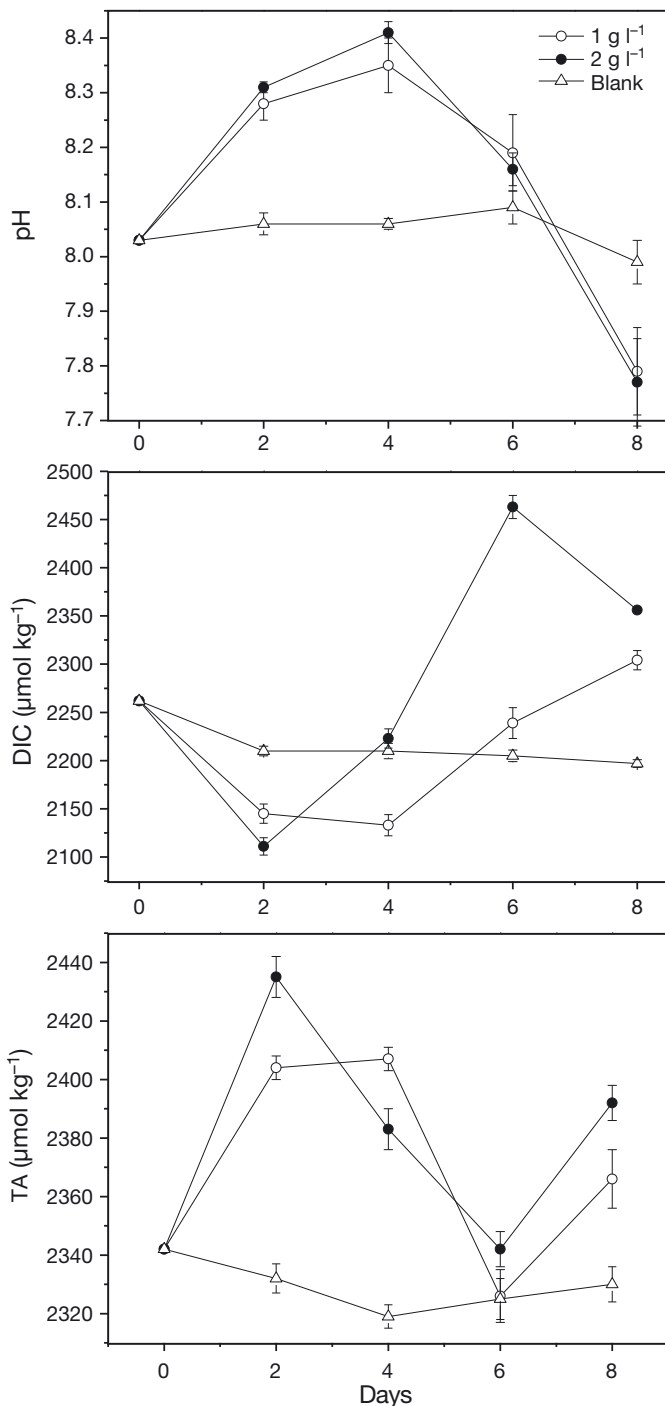


Fig. 10. Mean (\pm SD) variation in pH, dissolved inorganic carbon (DIC) and total alkalinity (TA) in the incubation experiments ($n = 2$)

On the other hand, it is well known that most phytoplankton surfaces exhibit a net negative charge at neutral or high pH, because there are more abundant negatively charged carboxyl and phosphate groups than positively charged amino groups in sea-

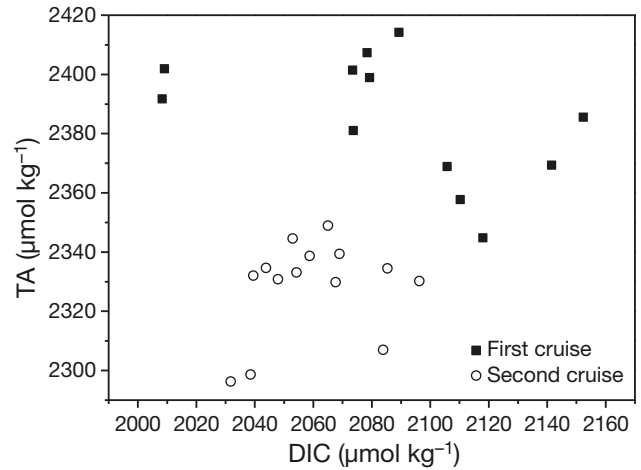


Fig. 11. Relationships between total alkalinity (TA) and dissolved inorganic carbon (DIC) in the surface layer of the coastal waters of Qingdao during the 2 cruises

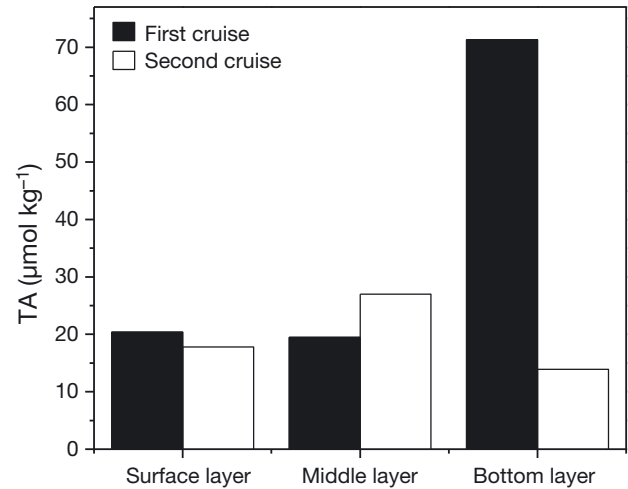


Fig. 12. Average difference in total alkalinity (ΔTA : $\text{TA}_{\text{unfiltered}} - \text{TA}_{\text{filtered}}$) in different layers of the coastal waters of Qingdao during the 2 cruises

water (Kleijn & Leeuwen 2000). The former reacts with protons during the titration of unfiltered seawater with hydrochloric acid, causing an increase in the measured alkalinity, especially during the late bloom period. In the present study, ΔTA ($\text{TA}_{\text{unfiltered}} - \text{TA}_{\text{filtered}}$) was maintained within a narrow range, except for the abrupt increase to ca. 70 $\mu\text{mol kg}^{-1}$ in bottom water during the first cruise (Fig. 12). This suggests that biogenic CaCO_3 and organic particles, suspended in deep waters, also contributed to the increase in the magnitude of ΔTA during the late bloom period (Kim et al. 2006).

Results from the incubation experiment provide clear evidence indicating that there was rapid

absorption of DIC by *U. prolifera* during the first 2 d, and rapid release of DIC during the next 6 d in the culture medium. For example, it can be estimated that the average absorption rates of DIC from *U. prolifera* at different densities were $59 \mu\text{mol l}^{-1} \text{d}^{-1}$ (at an algal density of 1 g l^{-1}) and $76 \mu\text{mol l}^{-1} \text{d}^{-1}$ (at an algal density of 2 g l^{-1}), during the first 2 d of the incubation experiment, while in the control group they decreased by $26 \mu\text{mol l}^{-1} \text{d}^{-1}$ during the first 2 d and subsequently remained relatively constant in the culture medium. In turn, the average release rates of DIC by *U. prolifera* were $27 \mu\text{mol l}^{-1} \text{d}^{-1}$ (at an algal density of 1 g l^{-1}) and $41 \mu\text{mol l}^{-1} \text{d}^{-1}$ (at an algal density of 2 g l^{-1}), respectively, during the next 6 d in the culture medium. This indicated that DIC absorption or the release rate by *U. prolifera* was positively correlated with the biomass of *U. prolifera* (over the relatively narrow biomass range tested) during the late bloom period. However, DIC absorption and release rates by *U. prolifera* were underestimated, since the atmosphere became a CO_2 source and atmospheric CO_2 entered the water body when DIC was absorbed by *U. prolifera*, while the atmosphere became a CO_2 sink and CO_2 in water entered atmosphere when DIC was released by *U. prolifera*. Moreover, the absorption of DIC by *U. prolifera* during the incubation experiment, when *U. prolifera* was decaying, was consistent with the results of the field study during the late bloom period.

Implications for changes of $p\text{CO}_2$ and air–water CO_2 fluxes during the *U. prolifera* late bloom

The average $p\text{CO}_2$ ($\pm \text{SD}$) in surface water was ca. $350 \pm 75 \mu\text{atm}$ during the second survey, but it rose to a maximum of ca. $427 \pm 32 \mu\text{atm}$ in the surface layer during the late bloom period. According to the atmospheric CO_2 data (402 ppm in July 2015) from the flask measurements at the Tae-ahn Peninsula (36.731°N , 126.131°E) adjacent to the southern Yellow Sea, a sink of CO_2 in this region (Qu et al. 2015) was converted into a weak CO_2 source during the late bloom period. Concomitantly with the increase in $p\text{CO}_2$, a slight reduction in DO was also recorded, demonstrating that enhanced respiration by *U. prolifera* occurred during late bloom conditions. A lack of correlation between $p\text{CO}_2$ and temperature was observed during both cruises (Fig. 13). There was no statistically significant difference between $p\text{CO}_2$ and temperature during the first ($p = 0.086$) or second ($p = 0.226$) cruise, although temperature is the primary controlling factor for $p\text{CO}_2$ in seawater (Takahashi et

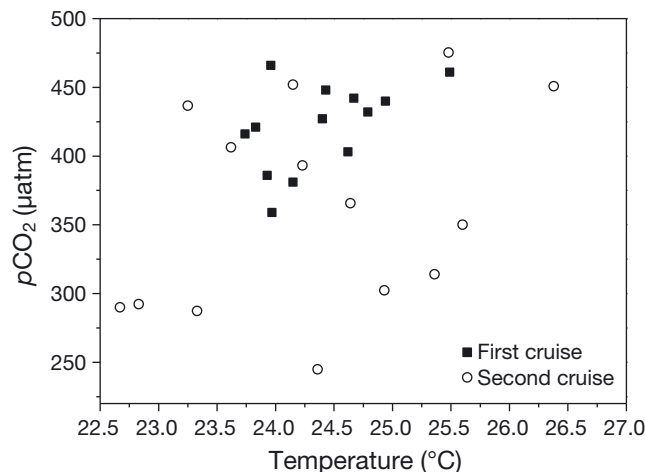


Fig. 13. Relationship between partial pressure of CO_2 ($p\text{CO}_2$) and temperature during the 2 cruises in the coastal waters of Qingdao

al. 1993). In addition, the finding that water temperature did not show a detectable difference between the 2 cruises indicates that the significant difference in $p\text{CO}_2$ between the 2 cruises was not caused by temperature. In this study, biological activity (the process of decline of *U. prolifera*) was considered the main cause for this finding. It was attributed to the fact that the *U. prolifera* respiration exceeded photosynthesis and more CO_2 was released rather than absorbed during the late bloom period.

The average wind speeds (i.e. 3.73 and 3.65 m s^{-1}) in Qingdao coastal waters in July and August, respectively (Saha et al. 2014), were used to calculate the wind speed-dependent CO_2 gas transfer coefficient of Wanninkhof (1992). The estimated air–water CO_2 fluxes of the coastal waters off Qingdao ranged from -0.72 to $2.17 \text{ mmol m}^{-2} \text{d}^{-1}$ and from -5.06 to $2.36 \text{ mmol m}^{-2} \text{d}^{-1}$, with mean values of 0.82 ± 0.88 and $-1.31 \pm 2.41 \text{ mmol m}^{-2} \text{d}^{-1}$ during the late bloom and after bloom periods, respectively. In general, the $p\text{CO}_2$ in seawater exceeded the $p\text{CO}_2$ in the atmosphere, which led to outgassing of CO_2 from seawater under the late bloom conditions. Based on reports that the areal cover of the green tide in coastal waters off Qingdao attained up to 182 km^2 on 16 July 2015 (North China Sea Branch of the SOA 2015) (Fig. 2), the carbon fluxes from the coastal waters off Qingdao to the atmosphere were calculated at $\sim 1.79 \text{ t C d}^{-1}$ during the late bloom period. Thus, the rapid release of CO_2 from the relatively short, late bloom *U. prolifera* period could have a significant effect on the ecosystem in coastal waters and directly affected the biogeochemical cycle of carbon in seawater.

CONCLUSIONS

Compared with the conditions recorded after the bloom, DIC, TA and $p\text{CO}_2$ were increased while pH was decreased during the *Ulva prolifera* late bloom period. Likewise, the observed variations in the carbonate system during the incubation experiment matched well with results obtained during the field surveys. The sink of atmospheric CO_2 in coastal waters off Qingdao was converted to a weak source of CO_2 as the *U. prolifera* decayed during the late bloom period. Furthermore, quantitative assessment of CO_2 fluxes during such intense bloom events, especially in coastal waters, should be taken into consideration in the assessment of air–water CO_2 fluxes. To enhance the understanding of the effects of such *U. prolifera* blooms on the carbonate system, continuous *in situ* monitoring on a larger spatial and temporal scale should be conducted in the future.

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