Short-term fate of seagrass and macroalgal detritus in *Arenicola marina* bioturbated sediments

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ABSTRACT: Seagrass meadows are globally important ecosystems for carbon (C) sequestration. However, bioturbation by benthic fauna can alter the distribution, degradation and overall preservation of C in the sediment. We performed a 4 wk laboratory experiment to investigate the short-term degradation and burial of 2 major C sources in bare sediments associated with seagrass ecosystems. Eelgrass *Zostera marina* and macroalgal (*Fucus vesiculosus*) detritus were amended in sediment with and without bioturbation by the common polychaete *Arenicola marina*. Bioturbation did not significantly affect the loss of eelgrass detritus (>0.5 mm), but caused a rapid burial of this material as a discrete layer (55% recovery) at sediment depths ranging from 8 to 14 cm. *A. marina* effects on macroalgal detritus were more pronounced, resulting, in total, in an 80% loss of macroalgal detritus by microbial degradation and worm ingestion. We conclude that *A. marina* bioturbation effectively buries eelgrass detritus into deep anoxic sediments, but we cannot confirm that this leads to enhanced C preservation in coastal ecosystems. In contrast, *A. marina* bioturbation significantly increases the degradation of macroalgal tissue, and it is unlikely that this detritus is a major source for permanent C burial.

KEY WORDS: Carbon degradation · Carbon burial · Lugworm · Bioturbation · *Zostera marina* · *Fucus vesiculosus*

1. INTRODUCTION

Coastal ecosystems including seagrass beds, saltmarshes, and mangrove forests are highly productive and sequester large amounts of organic carbon (Corganic) in the sediment, serving an important role in the global carbon (C) cycle and climate change mitigation (Duarte et al. 2005, Fourqurean et al. 2012). Degradation and burial of Corganic are key factors that control the storage of C in coastal ecosystems (Duarte et al. 2013). Aerobic microbial Corganic degradation, which is constrained to the upper few mm of oxic sediment, is fast and efficient as it involves oxidised radicals with the ability to depolymerize relatively refractory organic molecules, such as lignin (Benner et al. 1984, Kristensen et al. 1995). Below the oxic sediment surface, microbial degradation is dominated by slower and less efficient anaerobic processes (Canfield 1994, Kristensen & Holmer 2001). Therefore, redox conditions and sediment burial have decisive implications on the degradation of Corganic.

The Corganic that is deposited in coastal seagrass ecosystems originates from various sources, including seagrass leaves, phytoplankton, macroalgae and terrestrial vegetation (Kennedy et al. 2010, Trevathan-Tackett et al. 2015). Seagrass detritus and vascular terrestrial plants rich in structural polymers (i.e. lignocellulose) are resistant to microbial degradation under anoxic conditions, and may persist in sediments (Trevathan-Tackett et al. 2015). Macroalgae (i.e. Fucales, Laminariales and Plocamiales) are other important primary producers in coastal ecosystems as they
can contribute up to 50–75% of the stored C\textsubscript{org} in eutrophic areas (Gacia et al. 2002, Kennedy et al. 2010, Hill et al. 2015, Krause-Jensen & Duarte 2016). Since macroalgal tissues are dominated by labile polysaccharides and lack structural polymers (i.e. lignin), they degrade at much higher rates compared with seagrass detritus (Kristensen 1994, Kristensen & Holmer 2001). Accordingly, seagrass detritus is usually a better source for long-term C\textsubscript{org} burial than macroalgal detritus (Trevathan-Tackett et al. 2015).

Bioturbating benthic fauna are a natural component of coastal ecosystems. The role of bioturbation on sediment processes has long been recognised and is accepted as a key driving factor for biogeochemical processes in coastal and estuarine habitats (Wang et al. 2010, Kristensen et al. 2012). Large conveyor-belt feeding bioturbators, such as the polychaete Arenicola marina (hereafter referred to as ‘Arenicola’), are among the most common species occurring in sandy coastal temperate areas, such as intertidal flats, lagoons, semi-enclosed bays and fjords and often rapidly colonize empty spaces in the sediment inside or adjacent to eelgrass meadows (Philippart 1994, Valdemarsen et al. 2011, Govers et al. 2014). Arenicola is considered an ecosystem engineer due to its large capacity for both sediment particle-reworking and burrow ventilation (Kristensen et al. 2012). It significantly alters the sediment texture by selectively ingesting particles <1 mm and burying larger sizes down to ~10–40 cm depth, i.e. below the feeding funnel (Baumfalk 1979, Valdemarsen et al. 2011). Simultaneously, its intense burrow ventilation increases the oxygen availability in deep sediment layers, and inhibits anaerobic processes, such as sulphate reduction (Banta et al. 1999). However, few studies have reported in detail on how particle subduction and bioirrigation by Arenicola affect the fate of seagrass and macroalgal detritus recently buried at the sediment surface (Kristensen 2001).

In area-specific C budgets of coastal ecosystems, it is important that both bare and vegetated sediments are considered. Assessments for climate mitigation potential of marine environments are fundamentally based on the C\textsubscript{org} content in the upper metre of the sediment, which is often extrapolated from 10–25 cm deep sediment cores (Fourqurean et al. 2012). This is the depth range potentially affected by Arenicola bioturbation. Therefore, it is critical to understand how Arenicola affects degradation and subduction of various C\textsubscript{org} sources above and below its burial depth.

We investigated the effects of Arenicola bioturbation on the short-term fate of eelgrass Zostera marina and macroalgal Fucus vesiculosus detritus in sandy coastal sediments. This study had 2 goals: to examine the impact of Arenicola on (1) the degradation and (2) the burial of eelgrass and macroalgal detritus. We hypothesised that Arenicola bioturbation stimulates the degradation of macroalgal detritus to a higher extent than eelgrass detritus. Consequently, a higher fraction of eelgrass than macroalgal detritus would be buried deep into bioturbated sediments.

## 2. MATERIALS AND METHODS

### 2.1. Sediment and animal collection

Well-sorted sandy sediment with a median grain size of 210 µm and total 0.4 % organic matter (loss on ignition) (Papasyrou et al. 2007), and individuals of juvenile Arenicola were collected on separate days from Bregnør Bay in Odense Fjord, Denmark. The top-most sediment layer (~1 cm) was removed to ensure that highly labile benthic microalgae were excluded before sediment was sampled with a spade down to ~30 cm. Sediment was wet-sieved through a 0.5 mm mesh and homogenised on site. Arenicola juveniles were dug out of the sediment and recovered through careful sieving (1.0 mm mesh), and intact individuals were transferred to buckets for transport. In the laboratory, Arenicola individuals were acclimatized in Petri dishes containing seawater (salinity of 20) at 15°C for 36 h and then weighed. Only healthy juveniles of Arenicola (still active after the 36 h acclimation) were used in experiments.

### 2.2. Macrophyte collection

Eelgrass Zostera marina (hereafter referred to as ‘Zostera’), was collected at Enebærodde in the outer part of Odense Fjord, Denmark (Petersen et al. 2009). Green Zostera leaves were gathered from an eelgrass bed at approximately 1 m water depth and kept in seawater during transport back to the laboratory. Fronds of the brown alga Fucus vesiculosus (hereafter referred to as ‘Fucus’) were harvested in the inner part of Odense Fjord. Leaves (Zostera) and fronds (Fucus) were gently cleaned of epiphytes, washed with distilled water and cut into ca. 15 mm long and 5 mm wide pieces. Subsequently, the macrophyte pieces were arranged into portions of 5 g wet weight (WW), each representing an experimental core as described below. Portions of mixed macrophyte detritus consisting of 2.5 g WW of each type were also prepared. Samples of the fresh macrophyte
material (both Zostera and Fucus) were stored initially for total C and nitrogen (N) analysis.

2.3. Characterisation of macrophyte detritus used for sediment amendment

Initial samples of Zostera and Fucus detritus (n = 5) were analysed for total C and N content, and wet weight to dry weight relationships (WW:DW). Approximately 2 g WW of each macrophyte was thoroughly washed in distilled water, blotted on paper tissue to remove excess water and weighed. Inorganic C from fouling animals, such as barnacles, bivalves and bryozoans, was not observed on the macrophyte samples. Samples were then dried at 60°C for 48 h and re-weighed to establish WW:DW.

Dried and unacidified samples were ground using a clay mortar and pestle and analysed using a LECO TruSpec CN analyser.

2.4. Mesocosm experimental set-up

A 27 d laboratory experiment was set up to assess the effect of Arenicola bioturbation on the degradation and burial of Zostera and Fucus detritus in the sediment. A total of 8 experimental treatments with 3 replicates each were established, consisting of faunated (Arenicola) and defaunated (Control) sandy sediment cores, without enrichment (Non-amended) or enriched with eelgrass (+Zostera), macroalgae (+Fucus) or a mixture of eelgrass and macroalgal detritus (+Mix). Sediment cores were prepared by adding homogenised wet sediment into 24 acrylic (30 cm long, 8 cm inner diameter) core tubes, to a depth of 18 cm. Cores were left to compact overnight (approximately 16 h at 15°C). The following day, 60 g of the same homogenised wet sediment, corresponding to ~1 cm layer, was added to non-amended cores (n = 6). The three different batches of macrophyte detritus (5 g WW) were mixed with the wet homogenised sediment (60 g WW) and added to the remaining cores in a 1 cm thick layer to prepare the +Zostera, +Fucus and +Mix treatments. All cores were then left to settle for 4 h before an additional 3 cm layer of the wet homogenised sediment (180 g WW) was added on top, providing 22 cm of final sediment depth and the layer of macrophyte detritus located at 3–4 cm depth. This procedure prevented detritus loss by water currents and assured that all detritus in defaunated controls initially was buried in anoxic sediment. Each core was then topped up with seawater, and all 24 sediment cores were transferred to 4 seawater tanks (90 l capacity, salinity of 20) with water level 1 cm above the core rim. The cores were distributed in 4 tanks: 2 tanks with Arenicola cores and 2 tanks with defaunated cores. Each tank contained random, but not equal, core replicates from each of the 4 treatments. The cores with Arenicola were separated from defaunated cores to prevent migration of Arenicola to defaunated cores. Water was circulated below resuspension level using stirring magnets (1.5 cm long) fitted to the internal wall 2 cm below the upper rim of each core tube and driven by a central rotating external magnet (~60 rpm). Each tank held a total of 6 cores and was equipped with 2 air stones to aerate and mix the water. The set-up was kept at 15°C in the dark for the entire experimental period to restrict growth of benthic microalgae.

2.5. Flux measurements

Exchange of O₂, total CO₂ (TCO₂) and dissolved organic carbon (DOC) between the sediment and overlying water was determined by conducting flux incubations before (Day 0), and 1, 6, 13, 20, and 27 d after Arenicola was added to the sediment cores. Cores were left to compact for approximately 48 h before the first flux measurement was performed on Day 0. Right after the Day 0 flux measurements, 2 small-sized Arenicola individuals (0.28–0.56 g WW) were added to each faunated core (total of 0.6–0.7 g WW). The worms were allowed to construct burrows for 24 h before the second flux measurement was initiated on Day 1. During flux incubations, cores were sealed with gas-tight rubber bungs and water was mixed with continuous stirring as described in Section 2.4. Incubations were carried out for 2–3 h, and water O₂ concentration at the beginning and end of incubation was determined for individual cores using a fibreoptic O₂ dipping probe connected to a Pre-Sens Microx 4 transmitter. Water samples were taken from each core using a 60 ml syringe connected to a silicone tube at the start and end of incubation to determine rates of O₂ uptake, TCO₂ production and DOC exchange. Start water samples were taken while the unsealed cores were submerged in the tanks. End water samples were taken after removing the cores from the tank one by one, opening the lid for O₂ measurement and subsequent water sampling, before placing the cores back to the tank. Subsamples for TCO₂ analysis were transferred to 3 ml gas-tight glass vials and preserved with saturated HgCl₂ (30 µl).

Water
samples were stored in darkness at 4°C for <7 d and analysed for TCO\textsubscript{2} by flow injection (Hall & Aller 1992). Other subsamples for DOC analysis were transferred to 15 ml pre-combusted glass vials and stored at −20°C until analysis on a Shimadzu TOC-L total organic C analyser after acidification with 2 M HCl (pH <3) to remove TCO\textsubscript{2}. The exchange of solutes between sediment and water was calculated as the difference between end and start concentrations assuming constant rates throughout incubations.

2.6. Sediment core slicing, worm recovery and macrophyte detritus extraction

Each sediment core was sliced on Day 27 of the mesocosm experiment. Individual cores were retrieved from the experimental set-up and sliced within 2 h. A water sample was collected before core slicing to establish overlying TCO\textsubscript{2}, DOC and sulphate (SO\textsubscript{4}\textsuperscript{2−}) concentrations in each of the 4 tanks. The remaining overlying water in the cores was syphoned off ca. 2 h after gently flattening conical faecal casts to ensure a distinct sediment surface before cores containing Arenicola were sliced. This procedure was done with minimal sediment resuspension. The accumulated faecal casts in cores containing Arenicola elevated the sediment surface by about 0.7 to 2.0 cm compared with the original surface. All sediment cores were then sliced by extruding and cutting the sediment elevation resulting from the faecal casts and the sediment in 1 cm intervals to 6 cm depth from the original surface, and in 2 cm intervals to 22 cm depth. Recovery of intact Arenicola was attempted, but most individuals were damaged due to the slicing procedure. Each slice was then gently homogenised with a metal spoon, and after manual recovery of all visible large pieces of macrophyte detritus, subsamples of every section were taken for porewater extraction (25 ml), as well as for evaluation of sediment characteristics (4 ml).

Two fractions of macrophyte detritus were recovered from the sediment at the end of the experiment: (1) large particulate organic carbon (POC, >0.5 mm) and (2) excess fine POC (<0.5 mm). This division was adopted since the recovered Fucus was broken into various sizes ranging from <0.5 to 15 mm, while Zostera detritus did not visually change in size (i.e. 15 mm). The remaining sediment from each slice and the subsample left after porewater extraction were sieved through a 0.5 mm mesh to assure that all large macrophyte fragments were collected. Together with the detritus recovered visually from subsamples, the detritus retained by the sieve was then washed in distilled water and dried at 60°C for 48 h for later POC >0.5 mm analysis, as described in Section 2.3. The material that passed the sieve was discarded. By assuming that the recovered detritus was homogeneously distributed in each slice, the recovered POC >0.5 mm was adjusted to the volume of full sediment slices by correcting for the sediment used for other purposes (i.e. sediment characteristics and porewater extraction). Any fine particulate fractions derived from macrophyte material (POC <0.5 mm) that was not accounted for in the sieving procedure was assessed using the total sediment C\textsubscript{org} (= inherent C\textsubscript{org} + POC <0.5 mm), analysed in separate non-sieved sediment subsamples free of large particles as described in the section below. The excess fine C\textsubscript{org} representing POC <0.5 mm was calculated as the difference between the depth-integrated C\textsubscript{org} in amended (+Zostera, +Fucus and +Mix) and non-amended cores (inherent C\textsubscript{org}) in sediments with and without Arenicola.

2.7. Sediment characteristics

Sediment dry density, water content and porosity of each slice were determined by weighing 4 ml sediment free of visible detritus particles before and after drying at 60°C for 48 h. Dried subsamples (0.5 g each) were collected for C\textsubscript{org} determination, ground with a mortar and pestle, acidified by fumigation and re-dried according to Komada et al. (2008). The dried sediment samples may have contained a few large particles that were not removed manually. This might underestimate the excess fine C\textsubscript{org} (POC <0.5 mm) fraction in +Fucus, but should not affect the results in +Zostera treatments. The C\textsubscript{org} content of each sediment sample was measured using a CN analyser as previously described, and converted to µmol cm\textsuperscript{−3} using the C molar mass and the sediment dry density (Lavery et al. 2013). Selected sediment intervals from non-amended and control treatments (0–2, 3–6, 8–10 and 18–20 cm; including sediment accumulated on top of the original surface −2 to 0 cm from cores containing Arenicola) were analysed for median grain size using a Malvern Mastersizer 3000 particle size analyser.

2.8. Porewater analysis

Porewater was extracted by centrifuging ca. 25 ml wet sediment subsamples in double centrifuge tubes
containing GF/C filters for 10 min at 500 × g. Sub-samples of the extracted porewater were fixed with saturated HgCl₂ (1:100 v/v) and stored for TCO₂ analysis as described in Section 2.5. The remaining porewater was stored in glass vials at −20°C for DOC and SO₄²⁻ analyses. Porewater DOC concentrations were measured on a Shimadzu TOC-5000 analyser. SO₄²⁻ concentrations were analysed using ion chromography and standardised against chloride concentrations as described by Martin & Banta (1992).

2.9. Determination of Arenicola metabolism

The contribution of Arenicola respiration to TCO₂ and O₂ flux was determined in a separate experiment. One Arenicola individual weighing 0.2–0.4 g WW was transferred to each of 5 replicate 50 ml Winkler bottles filled with O₂ saturated seawater. Water samples for O₂ and TCO₂ were taken (as described in Section 2.5) at the start and end of 2 h incubation. Incubation bottles were held in darkness at a constant temperature (15°C) and salinity (20) throughout the experiment. Samples were analysed for O₂ and TCO₂ as described in Section 2.5.

2.10. Statistical analysis

The impact of Arenicola on the degradation of Zostera and Fucus was tested with separate 2-way ANOVAs on total detritus recovery (depth integrated) and flux data (time integrated). The addition of Arenicola (control and cores containing Arenicola), and macrophyte detritus (+Zostera, +Fucus, +Mix and Non-amended) were used as fixed independent factors. Levene’s test for homogeneity of variance was performed on data prior to analysis. When appropriate, Tukey’s post hoc test was used to establish variables with a significantly different interaction. All tests were performed at a significance level of α = 0.05, using IBM SPSS Statistics (Ver. 22).

3. RESULTS

3.1. Visual observations

Upward movement of sediment as a result of Arenicola subsurface feeding and surface defecation added an average of 1.2 ± 0.5 cm (mean ± SD) sediment on the surface of faunated cores. There was a clear appearance of light-coloured oxidised burrow walls in all Arenicola cores, and the faecal material on the sediment surface was oxidised. In non-amended cores, the oxidised sediment layer extended approximately 2–3 mm from the surface and was met by a dark-coloured sediment band of iron sulphides, indicating that sulphide produced by sulphate reduction in sediments below diffused upwards, reacting with iron (Preisler et al. 2007). Control cores and especially cores containing macrophyte detritus developed small white patches on the surface within the first week, indicative of sulphide-oxidising bacteria. These were largely absent in cores containing Arenicola. A 100% recovery of the added Arenicola individuals was attained by the end of the 27 d mesocosm experiment, although most of them were too damaged by the slicing procedure to provide reliable final biomass values.

3.2. C and N content of macrophyte detritus

Sediments amended with eelgrass (+Zostera) received 5.4 mol C m⁻² of Zostera C (33.4 ± 1.0% C of DW leaf material) while those amended with macroalgae (+Fucus) initially received 5.3 mol C m⁻² of Fucus C (41.4 ± 3.0% C of DW frond material), and those with mixed macrophyte material (+Mix) received 2.7 mol C m⁻² of Zostera C and 2.6 mol C m⁻² of Fucus C. Before the experiment, C:N ratios of the 2 macrophyte tissues were 30.6 ± 1.3 for eelgrass and 21.6 ± 1.5 for macroalgae.

3.3. Recovery of large macrophyte detritus (POC >0.5 mm)

All macrophyte detritus in defaunated sediments was recovered from 0–5 cm sediment depth (Fig. 1), which was around the depth where it was originally placed (3–4 cm), with only slight vertical displacement in either direction. Sediment reworking by Arenicola radically altered the distribution of detritus in all faunated cores. Most Zostera detritus was recovered 4–10 cm deeper than the deposition depth, which was 8–14 cm below the original surface, while Fucus detritus was evenly distributed throughout the bioturbated sediment from the faecal casts at the top and down to 14 cm. The same pattern was observed when the 2 types of macrophyte detritus were added as a mixture to sediment containing Arenicola. Zostera detritus appeared visually almost unaltered compared to the initial state. Larger pieces of Fucus (>5 mm) were only found in defaunated cores, while
Fucus detritus was predominantly recovered as small particles (<5 mm) in cores containing Arenicola, suggesting maceration of Fucus detritus by the activities of Arenicola. Overall, more Zostera detritus was recovered from the sediment than Fucus (p < 0.001; Fig. 1). Recovery of Zostera material was not significantly affected by the presence of either Arenicola (p = 0.380) or buried Fucus (p = 0.672). Recovery of Fucus was, however, significantly reduced by both the presence of buried Zostera (p = 0.002) and Arenicola (p < 0.001). In the presence of Arenicola, more than half (55%) of initially added Zostera detritus was found buried 8–14 cm into the sediment. Conversely, loss of Fucus (80%) was evident in sediment containing Arenicola. The total recovery varied from 47 to 32% in +Mix treatments without and with Arenicola, respectively. Of this, more Zostera (62–67%) and less Fucus (17–27%) detritus was recovered than in sediment containing only 1 type of detritus.

3.4. Sediment characteristics, C_\text{org} and excess fine C_\text{org} (POC <0.5 mm)

Dry sediment density varied from 1.36 to 1.54 g cm\(^{-3}\) and porosity ranged from 0.26 to 0.39 in all treatments. Control treatments had relatively homogeneous grain size distribution (median grain size 0.23–0.25 mm, Fig. 2) and silt+clay content (4.7–5.5%) at all depths. Treatments with Arenicola showed slightly lower grain size within the reworked sediment layer (0.20 mm from −2 to 10 cm depth) and 2-fold higher grain size (0.43 mm) in the 18–20 cm sediment depth zone. Accordingly, the silt+clay content in these treatments was highest in the top 0–2 cm layer (7.2%) followed by a gradual decrease below (from 6.2 to 1.5%) reaching the lowest level at 18–20 cm depth.

The sediment in all treatments had total C_\text{org} content varying from 0.1 to 0.3% of DW after removal of large macrophyte material. The treatment control +Fucus had higher levels of C_\text{org} (230–250 µmol C cm\(^{-3}\)) in the top 4 cm of the sediment, compared to all...
layers below this depth (150–190 µmol C cm⁻³ Corg; Fig. 3). Corg within all other control treatments (non-amended and +Zostera) was relatively constant (150–180 µmol C cm⁻³) at all sediment depths. Concentrations of Corg in all cores containing Arenicola were relatively high and variable (140–247 µmol C cm⁻³) in the top 15 cm, but lower and less variable (80–130 µmol C cm⁻³) in the bottom quarter (15–20 cm) of the sediment. Feeding activities by Arenicola therefore seemed to have transferred Corg in all cores from the bottom to the surface. In defaunated controls, the depth-integrated excess fine Corg (POC <0.5 mm) varied from 235 mmol m⁻² in +Zostera to 2448 mmol m⁻² in +Fucus treatments. In sediments with Arenicola, excess fine Corg was only found for +Zostera treatments (467 mmol m⁻²).

### 3.5. Solute exchange

O₂ was consumed and TCO₂ released by all sediment cores (Fig. 4). O₂ uptake was almost identical initially (Day 0) at about 5–6 mmol m⁻² d⁻¹ in the non-amended and +Zostera treatments, while the +Fucus treatment showed 2- to 3-fold higher initial rates. Arenicola addition instantly increased the O₂ uptake by 4- to 10-fold in all treatments and more or less maintained those rates throughout the experiment. Therefore, the time-integrated O₂ uptake was significantly higher in all Arenicola treatments compared with defaunated controls (p < 0.001). In terms of macrophyte amendment, the O₂ uptake in Arenicola treatments was highest in +Fucus (54–68 mmol m⁻² d⁻¹) followed by +Mix (44–59 mmol m⁻² d⁻¹), non-amended (33–57 mmol m⁻² d⁻¹) and +Zostera (33–44 mmol m⁻² d⁻¹; p < 0.001).

Defaunated +Fucus treatments produced more TCO₂ initially (71 mmol m⁻² d⁻¹) than all other sediments (1–12 mmol m⁻² d⁻¹; Fig. 4). Arenicola bioturbation resulted in 5- to 30-fold increased TCO₂ release in all treatments immediately after the worms were introduced (Day 1). Subsequently, the TCO₂ release in treatments containing Arenicola decreased markedly throughout the experimental period, but remained significantly higher than in defaunated treatments (p < 0.001), except for amended treatments on the final date. TCO₂ release from sediment containing Arenicola was always higher when amended with Fucus (87–244 mmol m⁻² d⁻¹) compared with +Mix (67–164 mmol m⁻² d⁻¹), +Zostera (52–153 mmol m⁻² d⁻¹) and non-amended treatments (51–239 mmol m⁻² d⁻¹; p < 0.001).

DOC flux was variable throughout the experimental period, fluctuating in all treatments between release and uptake across the sediment–water interface (Fig. 4). DOC flux varied in all treatments between −17 and 60 mmol m⁻² d⁻¹, with no clear trends over time.
Vertical distribution of porewater solutes was strongly influenced by the organic amendment and presence of *Arenicola* (Fig. 5). Defaunated sediment amended with macrophyte detritus displayed steep solute concentration gradients within the upper 10 cm of the sediment. TCO$_2$ in amended sediments increased from 3–9 mM near the sediment surface to 30–40 mM at 4–5 cm depth (highest in the +Fucus treatment) followed by constant or slightly decreasing levels below. DOC in amended sediments increased from 1–2 mM near the surface to 2–10 mM at 3–4 cm depth, followed by a rapid decrease to ~2 mM below. Highest DOC levels were recorded in the +Fucus treatment and lowest in the +Zostera treatment. SO$_4^{2−}$ decreased from 10–15 mM near the sediment surface to 1–9 mM at 4 cm depth (lowest in the +Mix treatment) and returned almost to the surface level below 10 cm depth. Only TCO$_2$ changed markedly in defaunated sediment without macrophyte amendment, showing an increase with depth to ~20 mM. The flushing of porewater solutes by *Arenicola* ventilation straightened the profiles of any solute in all faunated treatments to such an extent that concentrations at any depth were not significantly different from that of the overlying water (Fig. 5).

### 3.7. *Arenicola* metabolism

Respiration by *Arenicola* resulted in a consumption of $60.6 \pm 7.7 \mu$mol O$_2$ g$^{-1}$ d$^{-1}$ and production of $90.0 \pm 11.9 \mu$mol TCO$_2$ g$^{-1}$ d$^{-1}$. When extrapolated to the initial biomass (Table 1), *Arenicola* contributed with $7.8 \pm 0.9$ mmol m$^{-2}$ d$^{-1}$ to O$_2$ consumption and $11.5 \pm 1.4$ mmol m$^{-2}$ d$^{-1}$ to TCO$_2$ release. This represented 12–29% of the initial O$_2$ consumption and 5–8% of the initial TCO$_2$ release measured in faunated treat-
For each replicate, the final contribution of Arenicola metabolism on Day 27 was calculated to be 8.1 ± 1.0 mmol O$_2$ m$^{-2}$ d$^{-1}$ (14–30% of final O$_2$ consumption) and 12.0 ± 1.4 mmol TCO$_2$ m$^{-2}$ d$^{-1}$ (21–37% of final TCO$_2$ release).

Fig. 5. Vertical porewater profiles of total carbon dioxide (TCO$_2$), dissolved organic carbon (DOC) and SO$_4^{2-}$ in sediment cores without enrichment (non-amended), and with eelgrass (+Zostera), macroalgae (+Fucus) and mixed macrophyte material (+Mix). The 2 curves in each graph show profiles in sediment cores with fauna (+Arenicola; solid symbols) and without (Control; open symbols). The grey bar at 3–4 cm depth interval represents the initial burial depth of macrophyte material. The straight reference line (0 cm depth) represents the original sediment surface and values above it represent overlying water concentrations. Error bars: SE (n = 3).
Table 1. Initial and estimated final biomass of *Arenicola* in faunated treatments without (Non-amended) and with eelgrass (+*Zostera*), macroalgae (+*Fucus*) and mixed (+Mix) macrophyte detritus, and their calculated metabolic contributions to total carbon dioxide (TCO2) and O2 fluxes. Values represent mean ± SD (n = 5)

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<th>Contribution to O2 (%)</th>
<th>Contribution to TCO2 (%)</th>
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<tr>
<td></td>
<td>Initial     Final</td>
<td>Initial      Final</td>
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<tr>
<td>Non-amended</td>
<td>132 ± 11    137 ± 12</td>
<td>29 ± 8      16 ± 1</td>
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<tr>
<td>+<em>Zostera</em></td>
<td>124 ± 1     129 ± 1</td>
<td>18 ± 3      30 ± 15</td>
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<tr>
<td>+<em>Fucus</em></td>
<td>131 ± 17    136 ± 17</td>
<td>12 ± 0      14 ± 1</td>
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<td>+Mix</td>
<td>127 ± 2     132 ± 2</td>
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4. DISCUSSION

Seagrasses and macroalgae are recognised as important coastal C sinks contributing greatly to climate change mitigation (McLeod et al. 2011, Fourqurean et al. 2012, Krause-Jensen & Duarte 2016). However, only a few studies have investigated the effects of bioturbation on the fate of detrital C in bare coastal sediments around seagrass ecosystems and macroalgal beds. In this study, we found that bioturbation by *Arenicola marina* affects the degradation and burial of 2 types of detritus differently. A large part of added *Zostera marina* detritus (55%) was buried almost intact in a distinct layer below 8 cm sediment depth, while within 4 wk, *Fucus vesiculosus* detritus was rapidly degraded by microbes or consumed and assimilated by the worms (80%). Both organic C sources typically constitute a large proportion of the C storage observed in sediments within and surrounding seagrass meadows (Gacia et al. 2002, Kennedy et al. 2010, Greiner et al. 2016). Therefore, the strong impact of bioturbators such as *Arenicola* must be considered when C burial and sequestration potential are investigated in coastal ecosystems.

4.1. *Arenicola* and the fate of *Zostera* and *Fucus* detritus

*Arenicola* is an upward conveyor feeder that is known to completely rework the top 6–40 cm of the sandy sediment it inhabits (Riisgaard & Banta 1998, Kristensen 2001, Valdemarsen et al. 2010). Accordingly, experimental sediment cores containing *Arenicola* are rapidly restructured due to the substantial sediment reworking. The strong grain size selectivity of *Arenicola* creates vertical stratification within the sediment (Baumfalk 1979, Wendelboe et al. 2013). It typically only ingests small particles <0.5 mm (Andresen & Kristensen 2002) due to anatomical constraints of the mouth and pharynx. Larger particles such as coarse sand grains, gravel, shells and large organic debris are instead buried below the feeding depth, forming a graded bedding (Valdemarsen et al. 2011). The sorting capacity of *Arenicola* was also noticed in the present study, with 2-fold higher median grain size (0.43 mm) at 18–20 cm depth than in the upper 10 cm mixed layer (0.20 mm). This pattern was associated with a vertical stratification of silt+clay and Corg, exhibiting higher levels above than below 15 cm depth.

The physical reworking and active grain selection combined with intense bioirrigation by *Arenicola* affect the burial and subsequent degradation of *Zostera* and *Fucus* detritus differently. *Arenicola* is capable of efficient burial of *Zostera* detritus, as more than half of the deposited material at the end of our experiment was recovered in a discrete layer below 8–14 cm depth in the sediment. The mechanism of particle selection noticed for grain size and Corg can also explain the rapid burial of *Zostera* detritus, which apparently was not consumed by the worms but rather treated as large particles. This resulted in deposition of *Zostera* detritus at the base of the feeding funnel and thus burial below the bioturbation zone, as observed previously with plastic particles (Valdemarsen et al. 2011, Gebhardt & Forster 2018). In fact, the almost identical recovery in control cores suggests that *Zostera* detritus was buried fairly intact into the deep sediment, despite its exposure to the effects of bioirrigation in the experimental set-up. The burial of the lignocellulose fraction of seagrass tissues can therefore be markedly stimulated by *Arenicola* bioirrigation (Godshalk & Wetzel 1978, Kristensen & Holmer 2001). However, we cannot confirm enhanced preservation of *Zostera* detritus by *Arenicola* in the deeper sediment zones observed here (8–14 cm) due to experimental constraints required for extended anoxic conditions such as time and sediment height.

The lower recovery of *Fucus* detritus and its homogeneous distribution within the bioturbated zone, on the other hand, was a consequence of microbial degradation, reworking and ingestion by *Arenicola*. It seems that tissue softening induced by degradation and combined with the extensive movement of sand grains by *Arenicola* resulted in maceration of the fragile *Fucus* detritus into smaller particles (Kristensen 1994, Lopes et al. 2011). The presence of *Fucus*
detritus in faecal casts in the ~2 to 0 cm depth layer indicates that macroalgal detritus was ingested by the worms. The maceration, mixing and ingestion of labile Fucus detritus stimulated its degradation by increasing the relative surface area for microbial colonization and its exposure to oxic conditions (Aller 1994, Sun et al. 2002). This can explain the 30% higher degradation of Fucus detritus in bioturbated compared to defaunated treatments, which is comparable with previous studies reporting twice as fast degradation of F. serratus under oxic than anoxic sediment conditions (Kristensen & Mikkelsen 2003).

Arenicola individuals used in our experimental setup were small-sized juveniles (0.2–0.4 g WW ind.−1) and added at a density of 400 ind. m−2 to simulate the typical juvenile abundance found in nature (Flach & Beukema 1994). Adult individuals with much higher biomass (4–10 g WW ind.−1) are only found in abundances up to 80 ind. m−2 in coastal areas (Beukema & De Vlas 1979, Valdemarsen et al. 2011, Delefosse et al. 2012). Based on the initial and final burial depth of Zostera detritus, we roughly estimated that reworking rates in our cores were 1.4 cm3 g−1 d−1, which is lower than found previously for adult individuals (3–4 cm3 g−1 d−1, Valdemarsen et al. 2011). Reworking rates of Arenicola vary as a function of abundance and size of individuals, but environmental factors including temperature and food availability may also play a role. However, a model considering sediment reworking in relation to abundance and biomass indicates that a population consisting of many small individuals may have the same population burial effect as few larger individuals (Valdemarsen et al. 2011).

The losses of Zostera (42%) and Fucus (56%) detritus in defaunated treatments are probably caused by DOC leaching and by microbial excretion of exoenzymes and subsequent degradation and hydrolysis of labile components in the sediment (Kristensen 1994, Lopes et al. 2011). In fact, DOC production of the added detritus was clearly evident at the end of the experiment as elevated porewater DOC levels exactly where the macrophytes had been deposited. DOC was then probably consumed rapidly by the aerobic degradation processes near the sediment surface (Kristensen 2000). Otherwise, dominating anaerobic microbial processes were responsible for the degradation of both types of detritus in the absence of bioirrigation (Kristensen & Mikkelsen 2003). The significantly higher loss of Fucus than Zostera in the +Mix treatment suggests that addition of a reactive, labile source (Fucus) diverts microbial metabolism away from sources more resistant to degradation, such as refractory Zostera detritus (Canfield 1994, Zonneveld et al. 2010). This response is apparently the opposite of the microbial priming effect that has been observed in other studies (Kuzyskov et al. 2000, Thomson et al. 2019).

4.2. Arenicola and the transport and exchange fluxes of solutes

The drastically changed distribution of porewater solutes (TCO2, SO42− and DOC) in treatments with Arenicola indicates strong bioirrigation. The effect was distinct even in sediment immediately surrounding the Zostera and Fucus amended layer (0–5 cm), where accumulation (i.e. of TCO2 and DOC) and depletion (i.e. of SO42−) of solutes was substantial in defaunated sediment. Similar intense porewater flushing effect has been observed previously in sediments inhabited by Arenicola (Banta et al. 1999, Papaspyrou et al. 2007). The porewater transport caused by Arenicola ventilation was so powerful that solute concentrations in sediments containing Arenicola were completely uniform and comparable with overlying water levels. Thus, mineralisation products (TCO2) generated within the sediment are quickly flushed to the overlying water by Arenicola, at the same time as electron acceptors, e.g. SO42−, are replenished (Quintana et al. 2013).

The intense detritus burial and porewater transport in sediment influenced by Arenicola can explain the pattern of sediment–water TCO2-efluxes. Arenicola induced high TCO2 release in +Fucus treatments, which was initially more than 300% higher than the corresponding defaunated treatment, and decreased rapidly by about 50% (corrected for worm contribution) on Day 7 before reaching zero difference by the end of the experiment. The initial strong enhancement of microbial degradation was a consequence of high Fucus detritus lability (Kristensen 1994), combined with the intense particle reworking and bioirrigation. Thus, the Fucus detritus was exhausted much faster in the presence than absence of Arenicola. Although the rate of TCO2 release observed initially in +Zostera and +Mix treatments with Arenicola was lower than in the +Fucus treatment, it was still more than 400% higher than the corresponding defaunated treatments. The worm stimulation in these treatments reached 100–200% on Day 14 before being reduced to zero at the end. Mechanisms other than changes in microbial stimulation contributed to the decline in this case. The rapid burial of Zostera detritus to deep sediment layers may have prevented aerobic microbial degradation, since TCO2 release was reduced to levels of
defaunated treatments. In the non-amended treatment, the presence of *Arenicola* caused continuous high levels of inherent detritus degradation throughout the experiment as indicated by a steady 200–630 % enhanced TCO₂ release compared to defaunated controls. A similar effect has been documented previously for *Arenicola* (Banta et al. 1999). The stimulatory effect of *Arenicola* is 2- to 4-fold higher than for other polychaetes, such as *Marenzelleria viridis* and *Nereis (Hediste) diversicolor*, in estuarine sandy sediments (Kristensen et al. 2014, Quintana et al. 2018). Therefore, the present results support the contention that *Arenicola* is an important ecosystem engineer when considering the processing of POC either by microbial degradation or burial in coastal sediments.

### 4.3. C budget

The loss of added particulate detritus (POC >0.5 mm) over the experimental period was considerable in all treatments, with lowest values for the +Zostera and highest for +Fucus treatments, and with the +Mix as intermediate (Table 2), which is comparable with the losses found in previous reports (Kristensen 1994). The presence of *Arenicola* enhanced the POC loss of Fucus detritus by 30 %, while no difference in POC loss was evident between Zostera treatments and defaunated controls. The lost POC during degradation must be converted to DOC and TCO₂ (Kristensen 2000, Kristensen & Holmer 2001). The excess production of total dissolved C (TDC = DOC + TCO₂) in detritus-amended compared with non-amended treatments should therefore be consistent with POC loss and balance the C budget (Table 2). TDC production over the entire experiment was calculated as the sum of the TDC flux and TDC accumulated in the porewater (Table 2). The excess TDC production due to detritus addition can then be estimated by subtracting TDC generation in the non-amended treatments to deduct the degradation of inherent Corg (excess TDC = TDC in amended sediment − TDC in non-amended sediment). Unfortunately, excess TDC showed no clear pattern among treatments and with a deficit of C compared with the POC loss. The deficit (= POC loss [%] − excess TDC [%]) was most pronounced in *Arenicola* treatments, ranging from 30–48 % of the initially added POC, while only 5–28 % was missing in the defaunated treatments. The excess fine Corg (POC <0.5 mm, 235–2448 mmol m⁻²) corresponding to the fine fraction of added detritus not captured by the sieving may partly explain this deficit. Other possible causes for the deficit include: (1) initial rapid pulses of TDC generated by *Arenicola* between incubations may not have been recorded within our time series; (2) assimilation and incorporation of detrital C into worm biomass was not measured due to worm damage during slicing; and (3) loss of fine particles due to water currents in the experimental set-up. Nevertheless, we can rely on the POC loss results and conclude that Zostera detritus is buried in anoxic sediment to a larger extent than Fucus detritus. The rapid burial of the Zostera detritus promoted by *Arenicola* to deeper sediments (i.e. 8–14 cm) may prevent resuspension and fast oxic degradation. Conversely, Fucus detritus is degraded rapidly in the presence of *Areni-

<table>
<thead>
<tr>
<th>Added POC (&gt;0.5 mm)</th>
<th>Recovered POC (&gt;0.5 mm)</th>
<th>POC loss</th>
<th>TCO₂ efflux</th>
<th>TCO₂ Δ PW</th>
<th>DOC efflux*</th>
<th>DOC Δ PW</th>
<th>TDC generation</th>
<th>POC loss (% of initial)</th>
<th>Excess TDC (% of initial)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Control</td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
<tr>
<td>Non-amended</td>
<td>na</td>
<td>na</td>
<td>408 ± 60</td>
<td>663 ± 18</td>
<td>−81 ± 120</td>
<td>37 ± 7</td>
<td>1027 ± 142</td>
<td>na</td>
<td>na</td>
</tr>
<tr>
<td>+Zostera</td>
<td>5411 ± 8</td>
<td>3162 ± 33</td>
<td>2249 ± 38</td>
<td>1142 ± 98</td>
<td>1337 ± 36</td>
<td>−51 ± 92</td>
<td>36 ± 4</td>
<td>2466 ± 198</td>
<td>42 ± 1</td>
</tr>
<tr>
<td>+Fucus</td>
<td>5293 ± 20</td>
<td>2076 ± 242</td>
<td>3217 ± 258</td>
<td>1823 ± 133</td>
<td>1687 ± 100</td>
<td>304 ± 239</td>
<td>107 ± 31</td>
<td>4001 ± 442</td>
<td>61 ± 8</td>
</tr>
<tr>
<td>+Mix</td>
<td>5326 ± 7</td>
<td>2507 ± 171</td>
<td>2819 ± 167</td>
<td>1147 ± 73</td>
<td>1164 ± 68</td>
<td>−17 ± 197</td>
<td>69 ± 6</td>
<td>2363 ± 165</td>
<td>53 ± 5</td>
</tr>
</tbody>
</table>

| Arenicola           |                          |          |             |           |             |           |                |                        |                         |
| Non-amended         | na                       | na       | 1992 ± 162  | −127 ± 2  | −175 ± 184 | −38 ± 4  | 1652 ± 62     | na                     | na                      |
| +Zostera            | 5426 ± 7                 | 2974 ± 624| 2452 ± 631  | 2347 ± 280| −131 ± 2    | 252 ± 89  | −15 ± 14       | 2453 ± 310             | 45 ± 20                  | 15 ± 6                  |
| +Fucus              | 5293 ± 20                | 1122 ± 82| 4711 ± 78   | 3210 ± 307| −138 ± 3    | 287 ± 52  | −39 ± 8        | 3320 ± 358             | 79 ± 3                   | 32 ± 7                  |
| +Mix                | 5338 ± 3                | 1726 ± 335| 3612 ± 334  | 2931 ± 173| −132 ± 6    | 151 ± 83  | −38 ± 4        | 2912 ± 201             | 68 ± 11                  | 24 ± 4                  |
coli because of fragmentation within the bioturbation zone and lack of net burial.

4.4. Ecological implications

The results of this study have significant implications for the understanding of C cycling in bare and vegetated coastal ecosystems. It is demonstrated that bioturbating animals such as Arenicola may play a crucial role as they bury a large fraction of eelgrass detritus deep into the sediment, where it will likely be stored for long time. In this study, eelgrass detritus was buried to 8–14 cm depth by Arenicola juveniles, but it may travel even deeper down in areas with adult individuals having burrow depths up to 40 cm (Baumfalk 1979). Thus, the graded bedding effect of Arenicola feeding and reworking may contribute to C preservation by burial of Zostera detritus below the zone of porewater flushing. Macroalgal (i.e. Fucus) detritus, on the other hand, is rapidly lost though degradation, which is enhanced considerably by maceration through Arenicola activities. The fate of C$_{org}$ in sediments is therefore highly source dependent and sensitive to the processing mechanisms induced by Arenicola, which confirms our hypothesis.

The well-studied estuary, Odense Fjord (Denmark), can provide evidence of the short-term impact of Arenicola bioturbation when extrapolated to the ecosystem scale (Petersen et al. 2009). The average Zostera biomass in the 62 km$^2$ fjord estimated based on 3% coverage is about 400 kg C yr$^{-1}$ (Fyns Amt 2001, Carstensen et al. 2016). By assuming that 25% of eelgrass primary production is exported to bare sediments (Kennedy et al. 2010, Duarte & Krause-Jensen 2017), the results of our experiment indicate that Arenicola can bury 50 kg C yr$^{-1}$ eelgrass detritus in the fjord. This burial rate is 25-fold higher than the C sink capacity of the entire coastal ecosystem. This burial rate is 25-fold higher than the C sink capacity of the entire coastal ecosystem. Unvegetated sediments to achieve correct assessments should therefore account for the burial in bioturbated unvegetated sediments to achieve correct assessments of the C sink capacity of the entire coastal ecosystem.

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