Environmental drivers of salp *Thalia democratica* population dynamics from *in situ* observations

Maria Pascual¹,* , Michael G. Neubert², José Luis Acuña³, Andrew R. Solow², Carlos Domínguez-Carrió¹, Joaquín Salvador¹, Alejandro Olariaga¹, Verónica Fuentes¹

¹Institut de Ciències del Mar, 08003 Barcelona, Spain
²Woods Hole Oceanographic Institution, Woods Hole, MA 02543, USA
³Universidad de Oviedo, 33003 Oviedo, Spain

**ABSTRACT:** *Thalia democratica* blooms are a recurrent phenomenon in many coastal areas of the Mediterranean Sea and have significant ecological effects. To better understand the environmental drivers of salp blooms, we conducted 8 surveys to sample *T. democratica* in contrasting seasonal, temperature and chlorophyll conditions. In each survey, short-term variations in the abundances of different salp stages were assessed by sampling the same population at 30 min intervals. Using these data, we estimated the parameters in a set of stage-classified matrix population models representing different assumptions about the influence of temperature and chlorophyll on each stage. In the model that best explains our observations, only females are affected by changes in water temperature. Whether this is a direct influence of temperature or an indirect effect reflecting low food availability, female reproduction cessation seems to slow population growth under unfavourable conditions. When conditions become favourable again, females liberate the embryo and change sex to male, allowing for mating under extremely low salp densities and triggering the bloom. In contrast to previous findings, our results suggest that females, rather than oozooids, are responsible for the sustainability of salp populations during latency periods.

**KEY WORDS:** Matrix population models · Pelagic tunicate ecology · Population latency · Gelatinous zooplankton blooms

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**INTRODUCTION**

Salps are filter-feeding pelagic tunicates present in many of the world’s seas. Salp blooms are common in most oceans and have important ecological consequences. They generate episodic but intense downward fluxes of faecal pellets and dead bodies (Lebrato et al. 2012, Henschke et al. 2013, Smith et al. 2014) and serve as prey and/or hosts for many pelagic and benthic organisms (Harbison 1998). Salp blooms can have important economic impacts as well, as high salp densities can (1) negatively impact fish farms (Giesecke et al. 2014), (2) clog the cooling systems of power plants and (3) potentially reduce tourism due to their jellyfish-like appearance (Boero et al. 2013). These adverse effects explain the growing interest in understanding the mechanisms that drive salp blooms.

Salp populations exhibit ‘boom and bust’ cycles due to the alternation between an asexual solitary stage (oozooid) and a sexual aggregated stage (blastozooid) during their life cycle (Fig. 1). Each blastozooid in a chain starts as a female and is impregnated almost immediately after its release. The female becomes male after giving birth to a single, free-swimming juvenile oozyoid. Once the oozyoid becomes sufficiently large, it becomes productive and sequentially releases up to 3 chains of blastozooids, closing the life cycle (Heron 1972a). The particularities of this cycle, combined with high individual growth rates (up to
Salp blooms are population-level processes resulting from individual physiological responses to triggering environmental factors. Laboratory experiments have provided a wealth of information on how the growth, reproduction and survival rates of individual salps respond to environmental factors, usually temperature and food availability (e.g. Heron 1972a, Deibel 1982, Braconnot et al. 1988, Madin & Purcell 1992). Those studies provide forcing functions which can be extrapolated to field conditions, allowing first-order approximations of population growth. However, laboratory measurements of physiological rates are highly variable and usually far from those derived from field studies (Heron & Benham 1984, Madin & Deibel 1998, Everett et al. 2011). In addition, the typical spatial and temporal scales of laboratory experiments preclude manipulation of whole salp populations. For these reasons, direct observation remains the best tool to study population-level phenomena, such as salp blooms (e.g. Heron, 1972a,b, Heron & Benham 1984, 1985, Le Borgne & Moll 1986, Tsuda & Nemoto 1992, Loeb & Santora 2012). In this regard, field observation has revealed associations between active salp population growth, low temperature and high food availability (Heron & Benham 1985, Andersen & Nival 1986, Lavaniegos & Ohman 2003, Licandro 2006, Deibel & Paffenhöfer 2009, Henschke et al. 2014). Comparatively less attention has been given to population persistence during adverse conditions, when temperature is relatively high and food availability is low. Based on field observations, Heron & Benham (1985) postulated that in latent, low-density populations of Thalia democratica, oozooids grow slowly, nursing chains of blastozooids that will release under favourable conditions.

Henschke et al. (2015) recently built a population dynamic model for T. democratica in which the vital rates were considered constant and based on published laboratory or field estimates. Since vital rates may vary with oceanographic conditions, the stage-classified model should include environment-depen-
dent transition rates. In this regard, salp population models built by means of differential equations lend support to a role of chlorophyll and/or temperature levels as drivers of salp population dynamics (Andersen & Nival 1986, Henschke et al. 2015).

The goal of the present study was to evaluate how *Thalia democratica* population dynamics change under different environmental conditions. For this purpose, we conducted short time-series surveys around different *T. democratica* blooms under contrasting environmental conditions and used these observations to fit temperature- and chlorophyll-dependent stage-classified matrix models. Our main findings are in contrast to previous hypotheses (Heron & Benham 1985) that pointed to the asexual stage as being responsible for population latency: we propose that females limit their reproduction when conditions turn unfavourable, thus lowering the population growth rate. When favourable conditions resume, females liberate their embryo and become males, allowing for mating at low salp densities and triggering a bloom.

**MATERIALS AND METHODS**

**Sampling area**

All sampling stations were located in the Catalan Sea, NW Mediterranean. The most relevant hydrographical structure in this area is the density front produced by differences in salinity between coastal and oceanic waters. Associated with this shelf-slope front, a current flowing southward is formed (Font et al. 1988). Catalan coastal waters can be considered oligotrophic, although substantial nutrient sources, mostly coming from river runoff, turn them into a more productive area than the open sea (Arin et al. 2013, Saiz et al. 2014). Primary production in the NW Mediterranean exhibits a strong seasonality, mainly forced by changes in surface temperature (Duarte et al. 1999). During autumn and winter, stronger and colder winds cool down the superficial waters, inducing vertical mixing of the water column, which in turn enriches the euphotic layer with nutrients. In late winter and early spring, water starts to warm and the stability of the water column increases rapidly, triggering a rise in primary production. The pronounced warming and stratification in summer causes a depletion of nutrients in the euphotic layer, which decreases phytoplankton growth. Finally, the thermocline starts to break again in autumn and a second, but less intense, phytoplankton bloom occurs (Ménard et al. 1994, Estrada 1996).

Two locations (both approximately 1.6 km from shore) were selected for the different surveys. The first sampling site, Cadaqués (42°18’34.5” N, 3°19’19.3” E; bottom depth = 60 m), is located in front of the Cap de Creus area, an area where the continental shelf is particularly narrow and oceanic waters exert a strong influence. In contrast, the second sampling site, Barcelona (41°20’50.6” N, 2°17’53.3” E; bottom depth = 40 m), is located in an area with a wider continental shelf, where the shelf-slope front is relatively far from shore (Fig. 2). Two main rivers have an influence on the nutrient concentrations: the Rhône River in the case of Cadaqués, with maximum flows in autumn and spring (Ulses et al. 2008), and the Besós River in the case of Barcelona, with discharge peaks in autumn and winter (Arin et al. 2013). Sampling positions and depths may have varied slightly since a drifting buoy was followed during sampling.
Sample collection

Monthly time series

To determine the pattern of temporal variation of *Thalia democratica* populations, we conducted time-series surveys in both locations where monthly samples were collected during 2013 and 2014. Temperature and salinity profiles were obtained by deploying a CTD down to 20 m depth. Chlorophyll *a* (chl *a*) concentrations were measured in water samples collected 1 m below the sea surface using a 5 l Niskin bottle. Zooplankton samples were collected using a bongo net (40 cm diameter, 300 µm mesh size), obliquely towed for 10 min from 10 m depth to the surface. Sampling deeper layers or at certain times of the day was unnecessary, as *T. democratica* is considered a non-migrant species and mainly occurs in surface waters (Heron 1972b, Tsuda & Nemoto 1992, Paffenhöfer et al. 1995). Samples were preserved in 5% formalin immediately after being collected.

Short time-series observations

When salps were easily recognizable in the first zooplankton sample of each monthly visit, we followed a short time-series sampling methodology. From that point until the end of the bloom, we increased the sampling frequency, provided that weather conditions allowed. Following Heron (1972b), each survey comprised 6 consecutive hauls, performed at 30 min intervals, tracking a World Ocean Circulation Experiment (WOCE) drifting buoy (Hansen & Poulain 1996). Each haul started near the buoy that followed surface waters, thus tracking the salp population. If salp density was high enough to collapse the bongo net cod ends, towing time was reduced to 5 min. CTD and Niskin bottles were deployed near the drifting buoy after plankton samples were collected, as was also the case for the monthly surveys (i.e. 3 to 4 h after sampling had begun). Both samplings were conducted during daylight, avoiding dusk and dawn since these periods can affect salp distributions due to mating aggregations (Gibbons 1997). We assumed no net-avoidance by the salps given their slow swimming response (Ohman & Lavaniego 2002).

Sample analyses

Salps efficiently retain particles above 2 to 3 µm in size (Kremer & Madin 1992, but see Sutherland et al. 2010). To obtain an estimation of the concentration of phytoplankton available to salps, water samples were size-fractionated through 3 µm Whatman polycarbonate filters and glass fibre GF/F filters (0.45 µm). Chl *a* was extracted from those filters for 24 h at 4°C using 90% acetone (Venrick & Hayward 1984). Fluorescence of the acetone solution was measured using a fluorometer (Turner Designs). We only considered total chl *a* concentrations, since total chl *a* rendered a better model fit and both size fractions were highly correlated (Spearman’s coefficients >0.7). A better fit of population parameters with the smallest fraction agrees with the capacity of salps to retain submicrometre particles (Sutherland et al. 2010).

*T. democratica* individuals were identified, separated from the original sample and then scanned using a Zooscan (Grosjean et al. 2004). Individuals were measured from the posterior ridge of the gut to the opening (Foxton 1966) using Image J software (Abramoff et al. 2004), measuring up to a maximum of 1000 individuals each of blastozooids and oozooids. The number of buds chain⁻¹ was also counted (up to a maximum of 50 oozooids survey⁻¹) only when the oldest chain was clearly distinguishable from the stolon (Henschke et al. 2014). These values were then used to determine the maximum and minimum number of buds chain⁻¹ in the models.

Individuals were classified in 4 stages depending on the blastozooid/oozooid form and size as described by Henschke et al. (2015), after applying a 10% shrinkage correction due to the effects of formalin (Heron et al. 1988). Samples from all surveys were measured during the first month after collection. Size ranges were adjusted to our observations, adapting previous partitions to actual data from the population sampled. For example, the maximum length measured for a blastozooid carrying an embryo was 7 mm, and oozooids containing a developed chain reached lengths of 8 mm. These measures are similar to those found in other studies on *T. democratica* in the Mediterranean Sea (e.g. Braconnot & Jegu 1981). Therefore, we classified individuals as female (F, blastozooid, size range: 1–7 mm), male (M, blastozooid, size range: >7 mm), juvenile oozooid (J, oozooid, size range: 3–8 mm) and productive oozooid (PO, oozooid, size range >8 mm). The density of individuals (ind. 100 m⁻³) in each stage was calculated by dividing the number of individuals by the total volume of water filtered. When blastozooids or oozooids exceeded 1000 individuals, we multiplied their relative size frequencies by the total density of individuals to obtain the stage density. One individual count was added to all samples prior to calculation of densities.
to avoid logarithms of 0 in the parameter estimation procedure.

Model construction and parameter estimation

We modelled the population dynamics of *Thalia democratica* using a stage-classified matrix model based on the life cycle as represented in Fig. 1. First, we classified all individuals in the population at time *t* as either female (*F*), male (*M*), juvenile oozooid (*J*), or reproductive oozooid (*P*); and, gathered the counts of the classified individuals into a vector representing the population stage distribution at time *t*. The model projects the stage distribution from time *t* to time *t* + 1 via matrix multiplication:

\[
\begin{bmatrix}
F_{t+1} \\
M_{t+1} \\
J_{t+1} \\
PO_{t+1}
\end{bmatrix} =
\begin{bmatrix}
F_t & 0 & R_F & 0 \\
G_F & M_t & 0 & 0 \\
R_J & 0 & P_J & 0 \\
0 & 0 & G_J & P_{PO}
\end{bmatrix}
\begin{bmatrix}
F_t \\
M_t \\
J_t \\
PO_t
\end{bmatrix}
\]  

(1)

The entries in this matrix correspond to the rates of the transitions illustrated in Fig. 1. *P*; *P* and *P*; are the survival rates for males and productive oozooids. Females at time *t* either survive and remain female (with probability *P*;), or grow and become male (with probability *G*;), or die. Since a female sheds a juvenile oozooid and becomes a male simultaneously, the probability to release the embryo is *R* = *G*; *P*; and *G*; are the probability that an oozooid remains a juvenile and the probability that it grows to become a productive oozooid, respectively. *R*; and *R*; give the rate of production of new females by juvenile and productive oozooids. We assumed that reproduction is a birth-flow process sensu Caswell (2001).

In general, each of these transitions, and so each of the positive elements of the transition matrix, may depend on temperature (*T*) or chlorophyll concentration (*chl*), or both. We incorporate these potential environmental effects through the parametric models listed in Table 1. Each parametric model is a logistic or multinomial logistic function, since each element in the transition matrix is the probability of an event to happen among 2 or more alternatives (i.e. dichotomous or polytomous data). For example, *P*; represents the male survival probability to the following time step. Thus, it consists of 2 potential alternatives (to survive or not) and is accordingly modelled by a binomial logistic function. In contrast, *P*; represents the probability of an event — remaining female — among a set of 3 different possibilities: to die, to remain in the same state or to grow and become a male. Accordingly, a multinomial logistic function was used. Note that *G*; represents the probability of a different event — to grow to male — among exactly the same set of possibilities as in *P*;. Blastozooid production (*R*; and *R*;)) includes a mean oozooid fecundity term (*f*) and a mean newborn blastozooid survival term (*S*) within the same time interval, since oozooids produce chains continuously. To comport with our observations, the minimum (*f*) and maximum (*f* max) number of blastooids per oozooid were constrained between 14 (1 chain with only 14 blastozooids) and 258 (3 chains with up to 68 blastozooids each).

The general model (Eq. 1) involves a total of 25 parameters. Different restrictions on these parameters correspond to different hypotheses about the role of temperature and chlorophyll in the population dynamics of different life stages of *Thalia democratica* (see life cycle illustrated in Fig. 1). To select among these hypotheses, we fit a total of 48 models. These models correspond to combinations of temperature-dependence only, chlorophyll-dependence only and

<table>
<thead>
<tr>
<th>Transition</th>
<th>Function</th>
</tr>
</thead>
<tbody>
<tr>
<td><em>P</em>;</td>
<td>[ \frac{1}{1 + e^{\beta - \alpha T - \delta chl}} ]</td>
</tr>
<tr>
<td><em>G</em>;</td>
<td>[ \frac{1}{1 + e^{\beta - \alpha T - \delta chl}} ]</td>
</tr>
<tr>
<td><em>R</em>;</td>
<td>[ \frac{1}{1 + e^{\beta - \alpha T - \delta chl}} ]</td>
</tr>
<tr>
<td><em>R</em>;</td>
<td>[ \frac{1}{1 + e^{\beta - \alpha T - \delta chl}} ]</td>
</tr>
</tbody>
</table>

Table 1. Dependence of matrix elements (cf. Eq. 1) on temperature (*T*) and chlorophyll concentration (*chl*) by means of binomial or multinomial functions. *f*; fecundity or number of blastozooids of salp *Thalia democratica* produced by an oozooid; *S*; is the survival of the newborn blastozooids until the start of the next time interval. Note that *α*, *β*, *γ*, *δ*, *ε*, *f* min, and *f* max are model parameters. Transition matrix elements are explained in Fig. 1.
Table 2. Summary of the model selection results. Digits in the first 4 columns indicate whether transitions out of the female (F), male (M), juvenile (J) and productive oozooid (PO) stages of salp Thalia democratica have (1) or have not (0) been modelled with chlorophyll and/or temperature effects. The following 9 columns report the residuals of the cross-validation method (Cv), of the non-linear least squares optimization (Opt) and the total number of parameters to optimize (Θ) for models including temperature, chlorophyll and both. The selected model is highlighted in bold.

<table>
<thead>
<tr>
<th>Model</th>
<th>Temperature</th>
<th>Chlorophyll a</th>
<th>Temperature and chlorophyll a</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td>Cv</td>
<td>Opt Θ</td>
<td>Cv</td>
</tr>
<tr>
<td>♂</td>
<td>0 0 0 0</td>
<td>172.178</td>
<td>146.768</td>
</tr>
<tr>
<td></td>
<td>0 0 0 1</td>
<td>184.466</td>
<td>146.653</td>
</tr>
<tr>
<td></td>
<td>0 0 1 0</td>
<td>176.232</td>
<td>144.965</td>
</tr>
<tr>
<td></td>
<td>0 1 0 0</td>
<td>163.726</td>
<td>142.754</td>
</tr>
<tr>
<td>♂</td>
<td>0 0 0 0</td>
<td>160.141</td>
<td>137.4457</td>
</tr>
<tr>
<td>♂</td>
<td>0 1 0 1</td>
<td>165.647</td>
<td>145.760</td>
</tr>
<tr>
<td>♂</td>
<td>0 1 1 0</td>
<td>168.627</td>
<td>149.9376</td>
</tr>
<tr>
<td>♂</td>
<td>1 1 0 0</td>
<td>165.896</td>
<td>135.748</td>
</tr>
<tr>
<td>♂</td>
<td>0 0 1 1</td>
<td>176.591</td>
<td>143.684</td>
</tr>
<tr>
<td>♂</td>
<td>1 0 0 1</td>
<td>162.665</td>
<td>135.411</td>
</tr>
<tr>
<td>♂</td>
<td>1 1 0 1</td>
<td>161.869</td>
<td>129.340</td>
</tr>
<tr>
<td>♂</td>
<td>0 1 1 0</td>
<td>168.727</td>
<td>139.624</td>
</tr>
<tr>
<td>♂</td>
<td>1 1 1 0</td>
<td>170.798</td>
<td>125.549</td>
</tr>
</tbody>
</table>

Both temperature- and chlorophyll-dependence operating on all 16 possible combinations of life stages (Table 2).

To estimate the parameters in this model, we must incorporate stochasticity to account for variability in the data. Dennis et al. (1995) described the statistical advantage of adding noise on the log scale. Doing so transforms Eq. (1) to:

\[
\begin{bmatrix}
F_{t+1} \\
M_{t+1} \\
J_{t+1} \\
PO_{t+1}
\end{bmatrix} =
\begin{bmatrix}
e^{E_F} & 0 & 0 & 0 & P_F & 0 & R_J & R_{PO} & F_t \\
0 & e^{E_M} & 0 & 0 & G_F & P_M & 0 & 0 & M_t \\
0 & 0 & e^{E_J} & 0 & R_F & 0 & P_J & 0 & J_t \\
0 & 0 & 0 & e^{E_{PO}} & 0 & 0 & G_J & P_{PO} & PO_t
\end{bmatrix}
\]

(2)

where the vector \([E_F \ E_M \ E_J \ E_{PO}]^T\) has a multivariate normal distribution with 0 mean and variance-covariance matrix \(\Sigma\). On the log scale we then have

\[
\ln(F_{t+1}) = \ln(P_F \ F_t + R_J \ J_t + R_{PO} \ PO_t) + E_F
\]

(3)

\[
\ln(M_{t+1}) = \ln(G_F \ F_t + P_M \ M_t) + E_M
\]

(4)

\[
\ln(J_{t+1}) = \ln(R_F \ F_t + P_J \ J_t) + E_J
\]

(5)

\[
\ln(PO_{t+1}) = \ln(G_J \ J_t + P_{PO} \ PO_t) + E_{PO}
\]

(6)

where \(E\) is the stochastic error for each stage. We fit the models described by Eqs. (3–6) and Table 2 by nonlinear least squares. We compared models using the cross-validated 1-step prediction errors of the log-transformed population counts. Cross-validation is commonly used in comparing models, like those in Table 2, that differ in the number of fitted parameters (Claeskens & Hjort 2008). Otherwise, using non-cross-validated sum of squared prediction errors would tend to favour models with more parameters. We then used the selected model to estimate the transition matrix, the population growth rate (\(\lambda\)) and the elasticity matrix for selected environmental conditions. We constructed approximate 0.95 confidence intervals for the fitted values of the stage counts conditional on the initial counts by the percentile bootstrap method based on 200 bootstrap samples of the residuals from the fitted model (Efron & Tibshirani 1993). \(\lambda\) was calculated as the first eigenvalue of the transition matrix, and reflects the long-term behaviour of the population under constant conditions. The population decreases exponentially when \(0 < \lambda < 1\) and increases when \(\lambda > 1\). The elasticity of \(\lambda\) with respect to one of the matrix elements indicates the proportional contribution of this matrix element to \(\lambda\), and was computed following standard methods (Caswell 2001).

\[\text{RESULTS}\]

The temperature range in Barcelona was wider than in Cadaqués (12.2−24.9°C vs. 11.6−23.7°C), while total chl \(a\) concentrations were slightly lower in Barcelona (0.07–1.03 vs. 0.07–1.25 µg l\(^{-1}\); Fig. 3). Since environmental conditions can be considered
relatively similar, both locations were treated as survey replicates. *Thalia democratica* appeared from May to June/July, and between September and October in 2013 and 2014 (Fig. 3). Abundances in Cadaqués during spring time were higher than those recorded in Barcelona or during autumn. *T. democra-
tica* was the only salp forming blooms in all sampling events except for spring 2013, when *Salpa fusiformis* was also present at both locations (Fig. 3). When salp abundance was low (1 ind. 100 m$^{-3}$ or less), populations were mostly composed of blastozooids.

Salp densities and meteorological conditions favoured the development of 8 short time-series observations (marked with asterisks in Fig. 3; see size distributions in Fig. S1 in the Supplement, available at www.int-res.com/articles/suppl/m561p189_supp.pdf), each covering 5 time transitions (0–0.5, 0.5–1.0, 1.0–1.5, 1.5–2.0 and 2.0–2.5 h). The total number of time transitions was 40 (i.e. 5 time transitions in each of the 8 surveys), all of which were used to fit the models. The model that best explained the variability observed in our data included only females affected by temperature (Table 2). This effect can be estimated using the model equations from Table 3. Accordingly, the probability of females growing to males ($G_F$) decreased with increasing temperature (Fig. 4a). In contrast, the probability of remaining a female ($P_F$) was higher in warmer waters. Within the temperature range of this study (14–22°C), $G_F$ varied between 0.1 and 7.6% for a 30 min time step. The number of buds per oozooid per 30 min time step was inversely correlated to temperature and ranged from 31 to 52. Population growth rate ($\lambda$) smoothly decreased with temperature from 0.946 to 0.784. Elasticities (proportional sensitivity of $\lambda$ to variation in a particular matrix element) also varied with temperature. Values were more balanced at low temperatures, with slightly higher elasticities for $P_{PO}$ (Fig. 4b). In contrast, elasticity values for $P_{F}$ were clearly higher at high temperatures.

For the sake of comparison with existing literature, it is possible to derive the stage-dependent survival rates from our model probabilities to remain in the same stage ($P$) and to grow until the next stage ($G$), following the methods given by Caswell (2001). An individual will remain in the same stage ($P$) if it survives

<table>
<thead>
<tr>
<th>Matrix element</th>
<th>Function or fixed value</th>
</tr>
</thead>
<tbody>
<tr>
<td>$P_F$</td>
<td>$e^{-2.809 + 0.193 \times T} \times f \times {1 + e^{-2.803 + 0.193 \times T + e^{3.945 - 0.415 \times T}}}$</td>
</tr>
<tr>
<td>$G_P/R_F$</td>
<td>$e^{3.945 - 0.415 \times T} \times {1 + e^{-2.803 + 0.193 \times T + e^{3.945 - 0.415 \times T}}}$</td>
</tr>
<tr>
<td>$P_{M}$</td>
<td>0.739</td>
</tr>
<tr>
<td>$R_J$</td>
<td>$0.011 \times f$</td>
</tr>
<tr>
<td>$G_J$</td>
<td>0.204</td>
</tr>
<tr>
<td>$P_J$</td>
<td>0.615</td>
</tr>
<tr>
<td>$R_{PO}$</td>
<td>$0.009 \times f$</td>
</tr>
<tr>
<td>$P_{PO}$</td>
<td>0.639</td>
</tr>
<tr>
<td>$f$</td>
<td>$14.001 + 222.26 \times f$</td>
</tr>
<tr>
<td>$S_0$</td>
<td>$1 / (1 + e^{0.112 \times T}) = 0.011$</td>
</tr>
</tbody>
</table>
and does not grow \((1 - g)\), thus \(P = s \times (1 - g)\). An individual will be transferred into the next stage \((G)\) if it survives and grows \(g\), that is, \(G = s \times g\). By using \(P\) and \(G\) estimates from the best model, we could calculate survival rates \(s\) by solving a 2-equation system. Survival values of each stage are shown in Table 4. Note that \(g\) is the probability of growing to the next stage and not an individual increase in length; thus, these values are not comparable with growth rates reported in previous studies.

### DISCUSSION

We have explored the population dynamics of the salp *Thalia democratica* by combining a stage-specific matrix model with empirical modelling of the transition matrix elements. Our study highlights the key role of the female stage during periods of population latency, a conclusion that departs from previous hypotheses (Heron & Benham 1985). Our approach required samples of relatively dense salp populations (at least 60 ind. 100 m\(^{-3}\)) to achieve meaningful parameter estimates. This method, however, does not need well-defined size distributions since it is not based on identified cohort peak displacements, which would require greater abundances (Heron 1972a). *T. democratica* populations in the NW Mediterranean reach abundances up to 11000 ind. 100 m\(^{-3}\) (Licandro 2006), much denser than those found in the present study (Fig. 3). There is the possibility that the highest salp densities may have been missed by the coarse, monthly sampling frequency of our study. However, maximal annual abundances reported by Licandro (2006) from 1974 to 1999 varied among years, mostly from 10 to 1000 ind. 100 m\(^{-3}\), and densities greater than 1000 ind. 100 m\(^{-3}\) were only recorded in a few occasions. We sampled only blooms, because salp abundance during non-bloom periods ranges from 0 to <10 ind. 100 m\(^{-3}\) (Sardou et al. 1996), which is our approximate operational threshold. It can be assumed that population growth rates <1 suggest that our observed, dense populations might have been entering senescence. However, if population growth rates were >1, we would still expect a similar effect of temperature on population growth. Bearing in mind these experimental limitations, our 1-time-step expected densities fitted reasonably well with our observations (Fig. 5), and the estimated stage-dependent survival rates were in consonance with those obtained from previous model-derived rates (e.g. Henschke et al. 2015; Table 4).

### Effect of temperature on salp dynamics

The range of temperatures recorded was sufficient to evaluate its effects on salp dynamics. The model
Fig. 5. Observed densities of salp *Thalia democratica* at time $t+1$ (black dashed line) and their 1-time-step expected densities given the observed densities at previous time $t$ corresponding to the selected model (blue triangles). Longest segments of each triangle define maximum and minimum confidence intervals for each 1-time-step expected densities. Columns indicate the different life stages (F: female, M: male, J: juvenile, PO: productive oozooid); rows correspond to each sampling day; dates are given as dd/mm/yyyy.
that best explained the variability observed in the population dynamics of *T. democratica* pointed to a direct, negative effect of temperature on females. Low temperatures were associated with favourable conditions for salps, while high temperatures corresponded to periods of salp population arrest (Fig. 4). Although the effects of temperature seemed clear, these results can be attributed to either direct influence on salp physiology or an indirect, seasonal correlation with primary production. Temperatures above 20°C negatively affect feeding rates in *Salpa fusiformis* (Andersen 1986), and greater excretion rates in warmer waters could have been detrimental for the salp population (Ménard et al. 1994). On the other hand, temperature has traditionally been inversely related to primary production in the Mediterranean Sea (Bosc et al. 2004, Saiz et al. 2014). Both spring and autumn phytoplankton peaks normally preceded a salp bloom (Ménard et al. 1994), but salps in autumn (high temperature) would have less food available than in spring (low temperature). We have not seen salp blooms in winter, coinciding with the lowest water temperature registered, probably because thermocline formation and the rise of primary production have not yet started (Fig. 3).

Our results agree with other studies that found a negative correlation between temperature and salp abundance (Ménard et al. 1994, Licandro 2006). In addition to high temperatures, water column stability and irradiance may contribute to the termination of salp blooms (Ménard et al. 1994). Nevertheless, all of these factors seem to be associated with the annual cycle of the water column in the NW Mediterranean. Our study contrasts with others in which chl *a*, a measure of food availability, was correlated with salp population dynamics (Heron & Benham 1984, Andersen & Nival 1986, Deibel & Paffenbøker 2009, Henschke et al. 2014). The chl *a* range used in our study may have been too narrow, or alternatively there may be a lagged response of salp populations to food availability (Licandro 2006). It is possible that primary production would describe food availability better than chl *a*, but the methodology needed to obtain these values was logistically impossible in our study.

**Stages contributing to population growth under contrasting conditions**

Elasticity analysis has shown that the survival of oozooids stimulated population growth during cold, favourable periods (Fig. 4b). These results are in agreement with other studies pointing to oozooid survival as the most sensitive vital parameter of the population dynamics (Henschke et al. 2015), as well as asexual reproduction, a key parameter for exponential growth (Alldredge & Madin 1982, Andersen & Nival 1986). Under unfavourable conditions (high temperature and low food availability), females invest their low energy input into slow growth while arresting reproduction, leading to slow population growth (Fig. 4b). In this sense, we observed that low-density populations (≤1 ind. 100−3) were exclusively dominated by blastozooids, indicating the potential role of females as a switch for population growth. The decrease in female reproduction also fits our size distributions, where surveys with high temperatures (>19°C) had no males, indicating that females were not reproducing (see Fig. S1g–j in the Supplement). These findings are in agreement with other studies that found populations mostly dominated by females in autumn and winter samples (Heron & Benham 1985, Tew & Lo 2005). Interestingly Heron & Benham (1985) also observed that blastozooids were not reproducing, although they concluded that, in the typical situation for population latency (what they called ‘overwintering’), oozooids arrested their growth while retaining their chains until the onset of favourable conditions.

In our study, the probability of remaining in the same stage (P) exhibited higher elasticities than reproductive rates (R) (Fig. 4b). In other words, shortening or enlarging the residence time in a given stage has more influence on population dynamics than producing more or less offspring (Heron 1972b). *T. democratica* could control its population increase rates (r) through time, rather than clutch manipulation (sensu Aksnes & Giske 1990), as other pelagic tunicates like the appendicularian *Oikopleura dioica* would do (Subramaniam et al. 2014; but see Troedsson et al. 2002).

**Importance of females in salp dynamics**

In contrast to hypotheses centred on the role of the oozooid stages (Heron & Benham 1985), our results suggest that *T. democratica* females unfold a clockwise sequence of processes that initiates the bloom (Fig. 1). Once conditions improve, the female liberates an oozooid when still physically close to its daughters and becomes a male. This situation would favour sex encounters in conditions of extreme population dilution. In this sense, we assume that (1) the embryo is nearly mature when released (i.e. carrying
a primordial chain of daughter females) (Alldredge & Madin 1982, Miller & Cosson 1997); (2) testes develop when blastozooids are still nursing the embryo and become mature after giving birth (Madin & Purcell 1992); and (3) there is a partial overlap between the time in which oozooids start liberating chains, males release the sperm and females liberate the embryos (Miller & Cosson 1997). In an isolated group of closely related individuals, a sex ratio extremely skewed towards females (i.e. 1 male, many daughters) should maximize fitness (Hamilton 1967). If true, this mechanism would imply inbreeding during the early stages of a bloom, a process that could be detected using genetic markers. It should also increase fertilization success among males and daughters. These testable predictions may set the basis for future studies on salp dynamics.

CONCLUSIONS

We used an inverse method combining a stage-classified matrix population model, combined with empirical in situ observations, to understand how the vital rates of Thalia democratica vary with changing environmental conditions. Our results point to females, not the asexual oozooid as previously hypothesized, as the key stage to sustaining the population under the latency period. Productive oozooid survival accounted for high population growth under favourable conditions, while growth arrest in females lowered population growth under unfavourable conditions. In both scenarios, salps control population growth by time rather than by clutch manipulation. After a latency period, females tend to release the oozooids, which are already generating chains of females, and then become male. Therefore, we postulate that females may be the triggering mechanism to end the latency periods and initiate the salp bloom. This is the first study that has attempted to analyse how the processes inside the population vary in a changing environment. Understanding the latency period is essential, since it sustains the population under unfavourable environments, allowing the occurrence of new blooms.

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LITERATURE CITED

Bosc E, Bricaud A, Antoine D (2004) Seasonal and interannual variability in algal biomass and primary production in the Mediterranean Sea, as derived from 4 years of SeaWiFS observations. Global Biogeochem Cycles 18:GB1005
57. Springer Science+Business Media, Dordrecht


Everett JD, Baird ME, Suthers IM (2011) Three-dimensional structure of a swarm of the salp *Thalia democratica* within a cold-core eddy off southeast Australia. J Geophys Res 116:C12046


Troedsson C, Bouquet JM, Aksnes DL, Thompson EM (2002) Resource allocation between somatic growth and reproductive output in the pelagic chordate *Oikopleura dioica*


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